# 10. Assessment of the Alaska Plaice stock in the Bering Sea and Aleutian Islands 

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## Summary of Changes in Assessment Inputs

## Changes in the input data

1) The catch data have been fully updated through October 17, 2021. For purposes of modeling and projection, the full-year 2021 catch was estimated by projecting the remaining annual catch based on the average weekly catch during September \& October.
2) The 2021 eastern Bering Sea (EBS) shelf bottom trawl survey biomass estimates, uncertainty, and length composition were included in the assessment. There was no survey in 2020 due to the coronavirus pandemic.
3) The 2019 survey ages were read and were added to the assessment; no otoliths were collected in 2020.
4) The 2019 and 2020 fishery length compositions were also added.

## Changes in the assessment methodology

No modifications were made for this assessment.

## Summary of Results

1) The survey biomass estimate for 2021 ( $333,830 \mathrm{t}$ ) was $9 \%$ lower than the 2019 estimate and is the lowest value in the survey time series. Similarly, model estimates of female spawning biomass ( $158,090 \mathrm{t}$ in 2021) continued their decline since 2013.
2) In contrast, model estimates of total biomass ( $455,187 \mathrm{t}$ in 2021) show an increasing trend since 2019.
3) These results are likely due to estimates of relatively strong recruitment since 2017, a pattern which began to emerge in the 2019 assessment. There is substantial uncertainty surrounding these recruitment estimates, reflected in the large confidence intervals and the reduced model fit to some recent age and length compositions.
4) The 2021 projection model indicates slightly higher reference fishing mortality rates; combined with higher total biomass estimates they result in slightly increased OFL and ABC recommendations relative to 2019 despite the decline in the survey biomass estimates.
5) Alaska plaice continue to be found in high abundance in the northern Bering Sea (NBS); the 2021 NBS estimate of $344,578 \mathrm{t}$ exceeded the EBS estimate for the first time.

| Quantity | As estimated or specified last year for: 2021 2022 |  | As estimated or recommended this year for: 20222023 |  |
| :---: | :---: | :---: | :---: | :---: |
| $M$ (natural mortality rate) | 0.13 | 0.13 | 0.13 | 0.13 |
| Tier | 3 a | 3a | 3a | 3 a |
| Projected total (3+) biomass (t) | 427,587 | 430,164 | 442,946 | 454,030 |
| Female spawning biomass (t) | 166,528 | 160,150 | 141,838 | 144,767 |
| $B_{100 \%}$ | 335,172 | 335,172 | 286,587 | 286,587 |
| $B_{40 \%}$ | 134,069 | 134,069 | 114,635 | 114,635 |
| B35\% | 117,310 | 117,310 | 100,306 | 100,306 |
| $F_{\text {OFL }}$ | 0.160 | 0.160 | 0.170 | 0.170 |
| $\operatorname{maxF}_{A B C}$ | 0.132 | 0.132 | 0.140 | 0.140 |
| $F_{A B C}$ | 0.132 | 0.132 | 0.140 | 0.140 |
| OFL (t) | 37,924 | 36,928 | 39,305 | 39,685 |
| $\operatorname{maxABC}(\mathrm{t})$ | 31,657 | 30,815 | 32,697 | 32,998 |
| ABC (t) | 31,657 | 30,815 | 32,697 | 32,998 |
|  | As determ | $t$ year for: | As determined | s year for: |
| Status | 2019 | 2020 | 2020 | 2021 |
| Overfishing | no | $\mathrm{n} / \mathrm{a}$ | No | $\mathrm{n} / \mathrm{a}$ |
| Overfished | n/a | no | $\mathrm{n} / \mathrm{a}$ | no |
| Approaching overfished | n/a | no | $\mathrm{n} / \mathrm{a}$ | no |

## Responses to SSC and Plan Team Comments on Assessments in General

There were no relevant general comments.

## Responses to SSC and Plan Team Comments Specific to this Assessment

## From the December 2019 SSC minutes:

The SSC ... recommends continuing to track survey biomass trends in the NBS. The assessment indicates that sampling in the NBS in 2017 by a NPRB project showed differential age-at-maturity and size-at-age compared to the EBS. For the next full assessment, the SSC requests that the authors investigate differences in length composition and sex ratios between the NBS and EBS surveys. In addition, the SSC recommends analysis of genetic information to inform whether there is evidence of stock structure between the survey regions.

Response: The assessment includes updated information on and discussion of Alaska plaice in the NBS. The assessment now also includes a comparison of EBS and NBS size compositions, which reveals the persistent presence of small Alaska plaice that are not present in the EBS. No progress has been made on developing genetic analysis of population structure.

## Introduction

Alaska plaice (Pleuronectes quadrituberculatus) are primarily distributed on the Eastern Bering Sea continental shelf, with only small amounts found in the Aleutian Islands region. In particular, the summer distribution of Alaska plaice is generally confined to depths < 110 m , with larger fish predominately in deep waters and smaller juveniles ( $<20 \mathrm{~cm}$ ) in shallow coastal waters (Zhang et al., 1998). The Alaska plaice distribution overlaps with northern rock sole (Lepidopsetta polyxystra) and yellowfin sole (Limanda aspera), but the center of the distribution is north of the center of the other two species and seems to be positioned further north in warm years and more southern in cold years. Substantial amounts of Alaska plaice were also found between St. Matthew and St. Lawrence Islands in the 2010-2021 northern expansions of the annual Bering Sea shelf trawl surveys.

Prior to 2002, Alaska plaice were managed as part of the "other flatfish" complex. Since then an agestructured model has been used for the stock assessment allowing Alaska plaice to be managed separately from the "other flatfish" complex as a Tier 3 single species. There has been no research on stock structure for this species.

## Fishery

Since implementation of the Fishery Conservation and Management Act (FCMA) in 1977, Alaska plaice have been lightly harvested in most years as no major commercial target fishery exists for them. Catches of Alaska plaice increased from approximately $1,000 \mathrm{t}$ in 1971 to a peak of $62,000 \mathrm{t}$ in 1988, the first year of joint venture processing (JVP) (Table 10.1; Figure 10.1). Part of this apparent increase was due to increased species identification and reporting of catches in the 1970s. Because of the overlap of the Alaska plaice distribution with that of yellowfin sole, much of the Alaska plaice catch during the 1960s was likely caught as bycatch in the yellowfin sole fishery (Zhang et al. 1998). With the cessation of joint venture fishing operations in 1991, Alaska plaice have been harvested exclusively by domestic vessels. Catch data from 1980-89 by its component fisheries (JVP, non-U.S., and domestic) are available in Wilderbuer and Walters (1990).

Relative to their stock condition, Alaska plaice are lightly exploited averaging only $3 \%$ from 1975-2021. In $2020,83 \%$ and $12 \%$ of the Alaska plaice catch occurred in the yellowfin sole and northern rock sole fisheries, respectively. In 2021, most of the annual TAC for Alaska plaice was harvested during February and April as bycatch in the yellowfin sole fishery (at levels well below ABC). Catch rates were much lower over the rest of the year with a small increase during September. The total 2021 is predicted to be $16,760 \mathrm{t}$ (based on a catch of $15,411 \mathrm{t}$ as of October 17, 2021 and an additional catch of 123 t /week through the end of the year). This is well below the 2021 TAC of $24,500 \mathrm{t}$ and ABC of $31,657 \mathrm{t}$ (Table 10.1).

For monitoring of Pacific halibut bycatch, Alaska plaice are grouped with the rock sole, flathead sole, and other flatfish fisheries under a common prohibited species catch (PSC) limit, with seasonal and total annual allowances of prohibited species bycatch by these flatfish fisheries applied to the fisheries within the group. Before 2008, these fisheries were closed prior to attainment of the TAC due to the bycatch of halibut, and typically were also closed during the first quarter due to a seasonal bycatch cap. Since the implementation of Amendment 80 in 2008 where catch and bycatch shares were assigned to groups of fishing vessels (cooperatives), these fisheries have not been subjected to time and area closures (with the exception of a halibut closure in 2010).

Substantial amounts of Alaska plaice were discarded in various eastern Bering Sea target fisheries in past years due to low market interest. Retained and discarded catches were reported for Alaska plaice for the
first time in 2002, and indicated that of the $12,176 \mathrm{t}$ caught only 370 t were retained, resulting in a retention rate of $3.0 \%$ (Table 10.2). Similar patterns were observed for 2003-2005 (4\%,5\% and 6\%, respectively). The discard patterns have now changed, with increased retention each year. Since 2014 retention has exceeded $80 \%$ every year and has exceeded $90 \%$ every year. Most of the discards occur in the yellowfin sole fishery.

## Data

In summary, the data available for Alaska plaice are:

| Source | Data | Years |
| :---: | :---: | :---: |
| NMFS Eastern Bering Sea shelf survey | Survey biomass and standard error | 1982-2021; no survey was conducted in 2020 due to the coronavirus pandemic |
|  | Age Composition (by sex) | $\begin{aligned} & \text { 1982, 1988, 1992-1995, 1998, 2000-2002, 2005- } \\ & 2014,2016-2019 \end{aligned}$ |
|  | Length Composition (by sex) | $\begin{aligned} & \text { 1983-1987, 1989-1991, 1996-1997, 1999, 2003, } \\ & 2004,2015,2021 \end{aligned}$ |
| Fisheries | Catch | 1975-2021 |
|  | Age Composition (by sex) | 2000, 2002, 2003 |
|  | Length Composition (by sex) | 1978-89, 1995, 2001, 2008-2020 |

## Fishery:

This assessment uses fishery catches from 1975 through 2021 (Table 10.1). Fishery length compositions from 1978-89, 1995, 2001 and 2008-2020 for each sex were also used, as well as sex-specific age compositions from 2000, 2002 and 2003. The number of ages and lengths sampled from the fishery are shown in Table 10.3.

The non-commercial catch of Alaska plaice (i.e. catches in scientific surveys, subsistence fishing, recreational fishing, and fisheries managed under other FMPs) from $1977-2021$ is shown in Table 10.4.

From September 25- October 17, 2021 the Alaska plaice catch averaged 123 t per week. Alaska plaice are usually caught as bycatch in the yellowfin sole fishery. Yellowfin sole catch is still well below the TAC and fishing is ongoing. Since catches of Alaska plaice continued to accumulate into October, it seemed reasonable to assume that Alaska plaice would continue to be caught at a similar rate to the previous 3 weeks through the end of December. The catch at October 17 was 15,411; the average catch during the 3 weeks prior to October $17=123 \mathrm{t} /$ week. It was therefore projected that the Alaska plaice catch would reach $16,760 t$ by the end of 2021 .

## Survey:

Because Alaska plaice are usually taken incidentally in target fisheries for other species, CPUE from commercial fisheries is considered unreliable information for determining trends in abundance for these species. It is therefore necessary to use research vessel survey data to assess the condition of these stocks.

Large-scale bottom trawl surveys of the Eastern Bering Sea continental shelf have been conducted in 1975 and 1979-2019 by NMFS. Survey estimates of total biomass and numbers at age are shown in Tables 10.5 and 10.6, respectively. It should be recognized that the resultant biomass estimates are point
estimates from an "area-swept" survey. As a result, they carry the uncertainty inherent in the technique. It is assumed that the sampling plan covers the distribution of the fish and that all fish in the path of the trawl are captured. That is, there are no losses due to escapement or gains due to gear herding effects in the survey abundance calculations and catchability is therefore assumed to have a value of 1.0 for the design-based survey biomass estimates (catchability used in the model is discussed below).

The trawl gear was changed in 1982 from the 400 mesh eastern trawl to the 83-112 trawl, as the latter trawl has better bottom contact. This may contribute to the increase in Alaska plaice seen from 1981 to 1982, as increases between these years were noticed in other flatfish as well. Due to the differences in catchability between these two survey trawls, this assessment only uses the survey estimates from 19822019.

Survey estimates exhibit a relatively stable trend from 1982 to 2012 but have been in a declining trend since 2012. The last three surveys have estimated a decreasing population with the 2021 estimate $9 \%$ less than 2019; the 2021 estimate is the lowest in the time series (Table 10.5 and Figure 10.2).

Assessments for other BSAI flatfish have suggested a relationship between bottom temperature and survey catchability (Wilderbuer et al. 2002), where bottom temperatures are hypothesized to affect survey catchability by affecting either stock distributions and/or the activity level of flatfish relative to the capture process. Temperature was not expected to affect Alaska plaice catchability since they are a "cold loving" species with an anti-freeze protein that inhibits ice formation in their blood (Knight et al. 1991). This relationship was investigated for Alaska plaice by using the annual temperature anomalies from surveys conducted from 1982 to 2017. Examination of the residuals from the model fit to the bottom trawl survey relative to the annual bottom temperature anomalies did not indicate a positive correlation between the two data series (correlation $=-0.26$ ). This was also the result from a past assessment (Spencer et al. 2004) where a fit with a LOWESS smoother indicated that little correspondence exists between the two time series, and the cross-correlation coefficient ( -0.18 ) was not significant at the 0.05 level. Thus, the relationship between bottom temperature and survey catchability was not pursued further.

In 2010, 2017, 2019, and 2021 the Alaska Fisheries Science Center extended the annual bottom trawl survey to the northern Bering Sea (NBS) past St. Lawrence Island by the additional sampling of 142 stations. Substantial amounts of Alaska plaice were encountered in the northern area with a total biomass estimate of $302,976 \mathrm{t}$ in 2010, $330,728 \mathrm{t}$ in 2017, 321,571 t in 2019, and 344,578 t in 2021 (Figures 10.2 and 10.3). The proportion of Alaska plaice in the NBS relative to the EBS has been increasing since the 2010 survey, and in 2021 the NBS slightly exceeded the EBS estimate for the first time. Because the NBS estimate has been relatively stable during this time, it is likely that this change in distribution is due to declines in the EBS portion of the population rather than migration to the north. The population is also more concentrated in the NBS (Figure 10.3).

The size composition data from the 4 surveys (Figure 10.4) indicates that while the dominant size mode is similar between the EBS and NBS, many more small Alaska plaice occur in the NBS. In addition, there are slight differences in growth and maturation rates between areas. The implications of these differences for Alaska plaice population dynamics have yet to be explored. Since the northern Bering Sea has only been surveyed four times in the past ten years and because the area is closed to fishing, biomass estimates from only the standard eastern Bering Sea survey area are used in this assessment (Table 10.5).

In this assessment, the estimated population numbers at length from the trawl survey were multiplied by the age-length key in order to produce a matrix of estimated population numbers by age and length, from which an unbiased average length for each age can be determined. These population estimates by length
and sex were used to fit the model for years when age composition data were not available. The numbers of age and length samples obtained from the surveys are shown in Table 10.7.

## Analytic Approach

## General Model Structure

This catch at age model was developed with the software program Automatic Differentiation Model Builder (ADMB; Fournier et al. 2012). The age-structured assessment model is configured to accommodate the sex-specific aspects of the population dynamics of Alaska plaice, because the sexspecific weight-at-age diverges after the age of maturity (about age 10 for $50 \%$ of the stock) with females growing larger than males (Table 10.9). The model is coded to allow for the input of sex-specific estimates of fishery and survey age composition and weight-at-age and provides sex-specific estimates of population numbers, fishing mortality, selectivity, fishery and survey age composition and allows for the estimation of sex-specific natural mortality and catchability. The catch-at-age population dynamics model was used to estimate several population variables of the Alaska plaice stock, including recruitment, population size, and catch. Population size in numbers at age $a$ in year $t$ was modeled as

$$
N_{t, a}=N_{t-1, a-1} e^{-Z_{t-1, a-1}} \quad 3 \leq a<A, \quad 3 \leq t \leq T
$$

where $Z$ is the sum of the instantaneous fishing mortality rate $\left(F_{t, a}\right)$ and the natural mortality rate $(M), A$ is the maximum modeled age in the population, and $T$ is the terminal year of the analysis. Ages 3 through 25 were included in the Model. The numbers at age $A$ are a "pooled" group consisting of fish of age $A$ and older, and are estimated as

$$
N_{t, A}=N_{t-1, A-1} e^{-Z_{t-1, A-1}}+N_{t-1, A} e^{-Z_{t-1, A}}
$$

Recruitment was modeled as the number of age 3 fish. The efficacy of estimating productivity directly from the stock-recruitment data (as opposed to using an SPR proxy) was examined in a past assessment (Wilderbuer et al. 2008) by comparing results from fitting either the Ricker or Beverton-Holt forms within the model and choosing different time-periods of stock-recruitment productivity. This analysis is described in more detail in the 2008 assessment.

The numbers at age in the first year are modeled with a lognormal distribution

$$
N_{1, a}=e^{\left(\text {meaninit }-M(a-1)+\gamma_{a}\right)}
$$

where meaninit is the mean of the recruitments that made up the initial age comp and $\gamma$ is an age-variant deviation.

The mean numbers at age within each year were computed as

$$
\bar{N}_{t, a}=N_{t, a} *\left(1-e^{-Z_{t, a}}\right) / Z_{t, a}
$$

Catch in numbers at age in year $t\left(C_{t, a}\right)$ and total biomass of catch each year $\left(Y_{t}\right)$ were modeled as

$$
\begin{aligned}
& C_{t, a}=F_{t, a} \bar{N}_{t, a} \\
& Y_{t}=\sum_{a=1}^{A} C_{t, a} w_{a}
\end{aligned}
$$

where $w_{a}$ is the mean weight at age for Alaska plaice.
A conversion matrix was derived from the von Bertalanffy growth relationship, and used to convert the modeled numbers at age into modeled numbers at length. There are 51 length bins ranging from 10 to 60 cm , and 23 age groups ranging from 3 to $25+$. For each modeled age, the conversion matrix (TR) consists
of a probability distribution of numbers at length, with the expected value equal to the predicted length-atage from the von Bertalanffy relationship. The variation around this expected value was derived from a linear regression of coefficient of variation (CV) in length-at-age against age, where the CV were obtained from the sampled specimens over all survey years. The estimated linear relationship predicts a CV of 0.14 at age 3 and a CV of 0.10 at age 25 . The conversion matrix $\left(\mathrm{TR}^{\mathrm{T}}\right)$, vector of mean numbers at age, and survey selectivity by age were used to compute the estimated survey length composition $\left(\mathrm{NL}_{\mathrm{t}}\right)$, by year, as

$$
\overline{\mathbf{N L}}_{t}=\left(\text { srvsel } * \overline{\mathbf{N A}}_{t}\right) * \mathbf{T R}^{\mathbf{T}}
$$

where srvsel is a vector of survey selectivity by age.
Estimating certain parameters in different stages enhances the estimation of large numbers of parameters in nonlinear models. For example, the fishing mortality rate for a specific age and time $\left(F_{t, a}\right)$ is modeled as the product of an age-specific selectivity function (fishsel $l_{a}$ ) and a year-specific fully-selected fishing mortality rate. The fully selected mortality rate is modeled as the product of a mean $(\mu)$ and a yearspecific deviation $\left(\varepsilon_{t}\right)$, thus $F_{t, a}$ is

$$
F_{t, a}=\text { fishsel }_{a} * e^{\left(\mu+\varepsilon_{t}\right)}
$$

In the early stages of parameter estimation, the selectivity coefficients are not estimated. As the solution is being approached, selectivity was modeled with the logistic function:

$$
\text { fishsel }_{a}=\frac{1}{1+e^{(-\operatorname{slope}(a-f i f t y)}}
$$

where the parameter slope affects the steepness of the curve and the parameter fifty is the age at which $\operatorname{sel}_{a}$ equals 0.5 . The selectivity for the survey is modeled in a similar manner.

## Parameters Estimated Outside the Assessment Model

The parameters estimated independently include the natural mortality $(M)$ and survey catchability ( $q_{-} s r v$ ). Fish from both sexes have frequently been aged as high as 25 years from samples collected during the annual trawl surveys. Zhang (1987) determined that the natural mortality rate for Alaska plaice is variable by sex and may range from 0.195 for males to 0.27 for females. In past assessments natural mortality was fixed at 0.25 based on an earlier analysis of natural mortality (Wilderbuer and Walters 1997, Table 8.1).

In the 2010 assessment, the natural mortality rate of Alaska plaice was re-estimated using 3 methods from the literature based on the life history characteristics of maximum life span (Hoenig 1983), average age (Chapman and Robson 1960) and the relationship between growth and maximum length (Gislason et al. 2008). The results are summarized below and suggest a range of natural mortality values from 0.08 to 0.13 for males and 0.08 to 0.29 for females.

| Method | Males | Females |
| :--- | :--- | :--- |
| Hoenig (1983) | 0.11 | 0.11 |
| Chapman and Robson (1960) | 0.08 | 0.08 |
| Gislason et al. 2008 | 0.12 | 0.29 |
| Model profiling | 0.13 | 0.13 |

In the 2016 assessment, the model was again run for different combinations of male and female M to discern what value provides the best fit to the data components in terms of $-\log$ (likelihood). The best fit to the observable population characteristics occurred at $\mathrm{M}=0.13$ for both sexes (Fig. 10.5). This value of natural mortality is close to those estimated from the other three methods and also is consistent with the natural mortality used in other assessments of Bering Sea shelf flatfish which have similar life histories, growth and maximum ages. Therefore a value of $\mathrm{M}=0.13$ was used to model natural mortality for both males and females in this assessment.

Herding experiments in the eastern Bering Sea have demonstrated that many of the flatfish encountered in the area between the outer end of the footrope and where the bridles contact the sea floor (outside the trawl path) are herded into the path of the bottom trawl in varying degrees (Somerton and Munro 2001). Although Alaska plaice were not among the seven species that were explicitly studied, it is assumed that their behavior is similar to the other studied species which all exhibited herding behavior. The mean herding effect from all seven species combined resulted in a bridle efficiency of 0.234 . This assessment incorporates a herding effect into the stock assessment model by fixing survey catchability (q) at 1.2, close to the mean value from the combined flatfish species in the herding experiment.

Alaska plaice exhibit sex-specific dimorphic growth after the age of sexual maturity with females attaining a larger size than males. The von Bertalanffy parameters fit to the population length at age and the length-weight relationship of the form $W=a L^{b}$ were estimated as:

|  | Length at age fit |  |  | Length-weight fit |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathrm{L}_{\text {inf }}(\mathrm{cm})$ | k | $\mathrm{t}_{0}$ | a | b | n |
| males | 49.9 | 0.06 | -4.02 | 0.1249 | 2.98 | 866 |
| females | 50.1 | 0.127 | 0.35 | 0.0055 | 3.23 | 1,381 |

The combination of the length-weight relationship and the von Bertalanffy growth curve produces an estimated weight-at-age relationship that is similar to that used in previous Alaska plaice assessments. Minor changes in weight-at-age were made in this assessment relative to the 2016 assessment to exactly match the von Bertalanffy parameters. The sex-specific weight-at-age relationship calculated from the average population mean length at age and the length-weight relationship, by sex, are shown in Figures 10.6 and 10.7.

A maturity schedule is available for this assessment from samples obtained in 2012 (Table 10.8). These histologically determined estimates of proportion mature at age (TenBrink and Wilderbuer 2015) replace the previously used anatomically-derived estimates (Zhang 1987). Both studies estimated similar results differing in estimated 2013 female spawning biomass by only $4 \%$.

## Parameters Estimated Inside the Assessment Model

Parameter estimation is facilitated by comparing the model output to several observed quantities, such as the age compositions of the fishery and survey catches, the survey biomass, and the fishery catches. The general approach is to assume that deviations between model estimates and observed quantities are attributable to observation error and can be described with statistical distributions. Each data component provides a contribution to a total log-likelihood function, and parameter values that minimize the negative log-likelihood are selected.

The log-likelihoods of the age compositions were modeled with a multinomial distribution. The log of the multinomial function (excluding constant terms) is

$$
n \sum_{t, a} p_{t, a} \ln \left(\hat{p}_{t, a}\right)
$$

where $n_{t}$ is the number of fish aged, and $p$ and $\hat{p}$ are the observed and estimated age proportion at age.
The log-likelihood of the survey biomass was modeled with a lognormal distribution:

$$
\lambda_{2} \sum_{t}\left(\ln \left(\text { obs_biom }_{t}\right)-\ln \left(\text { pred_biom }_{t}\right)\right)^{2} / 2 * c v(t)^{2}
$$

where $o b s_{\_}$biom $_{t}$ and pred_biom $_{t}$ are the observed and predicted survey biomass at time $t, c v(t)$ is the coefficient of variation of observed biomass in year $t$, and $\lambda_{2}$ is a weighting factor.
The predicted survey biomass for a given year is

$$
q_{-} s r v * \sum_{a} \operatorname{selsr}_{a}\left(\bar{N}_{a} * w t_{a}\right)
$$

where $\operatorname{selsrv}_{a}$ is the survey selectivity at age and $w_{a}$ is the population weight at age.
The log-likelihood of the catch biomass was modeled with a lognormal distribution:

$$
\lambda_{3} \sum_{t}\left(\ln \left(o b s_{-} c a t_{t}\right)-\ln \left(\text { pred_cat } t_{t}\right)\right)^{2}
$$

where $o b s_{-} c a t_{t}$ and pred_cat $t_{t}$ are the observed and predicted catch. Because the catch biomass is generally thought to be observed with higher precision than other variables, $\lambda_{3}$ is given a very high value (hence low variance in the total catch estimate) so as to fit the catch biomass nearly exactly. This can be accomplished by varying the $F$ levels, and the deviations in $F$ are not included in the overall likelihood function. The overall likelihood function (excluding the catch component) is

$$
\lambda_{1}\left(\sum_{t} \varepsilon_{t}+\sum_{a} \gamma_{a}\right)+n \sum_{t, a} p_{t, a} \ln \left(\hat{p}_{t, a}\right)+\lambda_{2} \sum_{t}\left(\ln \left(o b s_{-} \text {biom }_{t}\right)-\ln \left(\text { pred_biom }_{t}\right)\right)^{2} / 2 * c v(t)^{2}
$$

For the model run in this analysis, $\lambda_{1}, \lambda_{2}$, and $\lambda_{3}$ were assigned weights of 1,1 , and 500 , respectively. The value for age composition sample size, $n$, was set to 200 for surveys and 50 for the fishery. The likelihood function was maximized by varying the following parameters:

| Parameter type | Number |
| :--- | :---: |
| 1) fishing mortality mean $(\mu)$ | 1 |
| 2) fishing mortality deviations $\left(\varepsilon_{t}\right)$ | 44 |
| 3) recruitment mean | 1 |
| 4) recruitment deviations $\left(\nu_{t}\right)$ including initial yr | 65 |
| 5) fishery selectivity patterns both sexes | 4 |
| 6) survey selectivity patterns both sexes | 4 |
| Total parameters | 119 |

Finally, a Monte Carlo Markov Chain (MCMC) algorithm was used to obtain estimates of parameter uncertainty (Gelman et al. 1995). One and a half million MCMC simulations were conducted, with every 1,000 th sample saved for the sample from the posterior distribution. Ninety-five percent confidence intervals were produced as the values corresponding to the $5^{\text {th }}$ and $95^{\text {th }}$ percentiles of the MCMC evaluation. For this assessment, confidence intervals on female spawning biomass, total biomass and age three recruitment are presented.

## Results

## Model Evaluation

Retrospective analysis of the past 10 years of female spawning biomass estimates does not indicate a pattern of concern regarding misspecification of the model as all trajectories follow the same trend (Fig. 10.8). Survey estimates in 2012 and 2015 were more variable relative to the time-series (high in 2012 and low in 2015) but did not contribute to an undesirable pattern. Mohn's evaluation statistic was calculated at -0.01 .

## Time-Series Results

Using the survey catchability value of 1.2 , the stock assessment model (Model 2011_1) estimates that the total Alaska plaice biomass (ages 3+) increased from $461,364 \mathrm{t}$ in 1975 to a peak of $751,025 \mathrm{t}$ in 1984 (Figure 10.9, Table 10.9). Beginning in 1984, the total biomass steadily declined to 540,516 t by 2003 before increasing again to $557,393 \mathrm{t}$ in 2006. The model estimates a slow decrease thereafter to $431,739 \mathrm{t}$ in 2019. In contrast to previous assessments, the 2021 model run estimates a slight increase in total biomass after $2019(455,187 \mathrm{t})$. The estimated survey biomass also shows a slow decline since a peak value estimated in 1984 (Figure 10.10). The female spawning biomass (FSB) has also been very stable, declining slowly, since a peak in 1985 (Table 10.9 and Figure 10.11). This decline continues in 2021, and FSB is projected to be $141,838 \mathrm{~T}$ in 2022 relative to a $\mathrm{B}_{40 \%}$ value of $114,635 \mathrm{t}$. These patterns can be explained by model estimates of strong recruitment in recent years (Table 10.9 and Figure 10.12); these recruits contribute to the total biomass but are not yet mature.

As in past assessments, fitting fishery observations was de-emphasized by lowering the input sample sizes from 200 to 50. This contributed in part to producing estimates of $50 \%$ fishery selectivity at about 10 years for females and 9 for males (Figure 10.13, Table 10.10). The fits to the trawl survey age and length compositions are shown in Figures 10.14 and 10.15 and the fit to the fishery age and length compositions are shown in Figures 10.16 and 10.17.

The modest annual changes in stock biomass are primarily a function of recruitment variability, as fishing pressure has been light. The fully selected fishing mortality estimates show a maximum value of 0.14 in 1988, and the average annual F has averaged 0.04 from 1975-2019 (Table 10.11, Fig.10.18). Estimated numbers-at-age are shown in Table 10.12. The posterior distribution of the 2021 female spawning biomass estimate is shown in Figure 10.19.

## Harvest Recommendations

## Amendment 56 Reference Points

The reference fishing mortality rate for Alaska plaice is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Estimates of $B_{40 \%}, F_{40 \%}$, and $S P R_{40 \%}$ were obtained from a spawner-perrecruit analysis. Assuming that the average recruitment from 1977-2018 year classes estimated in this assessment represents a reliable estimate of equilibrium recruitment, then an estimate of $B_{40 \%}$ is calculated as the product of $S P R_{40 \%}$ * equilibrium recruits ( $=206,554 \mathrm{t}$ ). The 2022 female spawning biomass is estimated at $141,838 \mathrm{t}$. Since reliable estimates of 2022 spawning biomass ( $B$ ), $B_{40 \%}, F_{40 \%}$, and $F_{35 \%}$ exist and $B>B_{40 \%}(141,838 \mathrm{t}>114,635 \mathrm{t}$ ), Alaska plaice reference fishing mortality is defined in tier 3a of Amendment 56. For this tier, $F_{A B C}$ is constrained to be $\leq F_{40 \%}$, and $F_{O F L}$ is defined as $F_{35 \%}$. The values of these quantities are:

$$
2022 \text { SSB estimate }(B)=141,838 \mathrm{t}
$$

| $B_{40 \%}$ | $=$ | $114,635 \mathrm{t}$ |
| :--- | :--- | :--- |
| $F_{40 \%}$ | $=$ | 0.14 |
| $F_{A B C}$ | $=$ | 0.14 |
| $F_{35 \%}=$ | 0.17 |  |
| $F_{\text {OFL }}$ | $=$ | 0.17 |

## Specification of OFL and Maximum Permissible ABC

The estimated catch level for year 2022 associated with the overfishing level of $\mathrm{F}=0.17$ is $39,305 \mathrm{t}$. The 2022 recommended ABC associated with $\boldsymbol{F}_{A B C}$ of $\mathbf{0 . 1 4}$ is $\mathbf{3 2 , 6 9 7} \mathbf{t}$. Projections of Alaska plaice female spawning biomass (described below) from a harvest rate equal to the average fishing mortality rate of the past five years indicate that the female spawning stock may increase through 2030 and decline slightly thereafter (Fig. 10.20).

The relative trajectories of female spawning biomass and F (Figure 10.21) indicate that Alaska plaice are not experiencing overfishing, are not overfished, and are not approaching overfished condition.

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2021 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2022 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2022. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2020, are as follows ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC , so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2020 recommended in the assessment to the max $F_{A B C}$ for 2020. (Rationale: When $F_{A B C}$ is set at a value below max $F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to the 2015-2020 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 4: In all future years, the upper bound on $F_{A B C}$ is set at $F_{60 \%}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

The recommended $F_{A B C}$ and the maximum $F_{A B C}$ are equivalent in this assessment, and five-year projections of the mean Alaska plaice harvest and spawning stock biomass for the remaining four scenarios are shown in Table 10.14.

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether the Alaska plaice stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follows (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2020 under this scenario, then the stock is not overfished.)

Scenario 7: In 2020 and 2021, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2032 under this scenario, then the stock is not approaching an overfished condition.)

The results of these two scenarios indicate that Alaska plaice are neither overfished nor approaching an overfished condition. With regard to assessing the current stock level, the expected stock size in the year 2019 of scenario 6 is well above its $B_{35 \%}$ value of $116,600 \mathrm{t}$. With regard to whether the stock is likely to be in an overfished condition in the near future, the expected stock size in the year 2032 of scenario 7 is also greater than its $B_{35 \%}$ value. Figure 10.21 shows the relationship between the estimated time-series of female spawning biomass and fishing mortality and the tier 3 harvest control rule for Alaska plaice.

## Risk table and ABC Recommendation

The following template is used to complete the risk table:

|  | Assessment-related <br> considerations | Population <br> dynamics <br> considerations | Environmental/ecosystem <br> considerations | Fishery Performance |
| :--- | :--- | :--- | :--- | :--- |
| Level 1: | Typical to | Stock trends are <br> typical for the <br> Normal | No apparent <br> environmental/ecosystem <br> moderately <br> increased <br> uncertainty/minor <br> stock; recent <br> recruitment is | No apparent <br> fishery/resource-use <br> performance and/or <br> behavior concerns |
|  | unresolved issues |  |  |  |
| within normal |  |  |  |  |
| in assessment. | range. |  |  |  |
| Level 2: | Substantially | Stock trends are <br> unusual; <br> abundance <br> increasing or | Some indicators <br> showing adverse signals <br> relevant to the stock but the | Some indicators <br> showing adverse signals <br> but the pattern is not |
| increaned <br> concerns | increased <br> assessment |  |  |  |


|  | uncertainty/ unresolved issues. | decreasing faster than has been seen recently, or recruitment pattern is atypical. | pattern is not consistent across all indicators. | consistent across all indicators |
| :---: | :---: | :---: | :---: | :---: |
| Level 3: <br> Major <br> Concern | Major problems with the stock assessment; very poor fits to data; high level of uncertainty; strong retrospective bias. | Stock trends are highly unusual; very rapid changes in stock abundance, or highly atypical recruitment patterns. | Multiple indicators showing consistent adverse signals a) across the same trophic level as the stock, and/or b) up or down trophic levels (i.e., predators and prey of the stock) | Multiple indicators showing consistent adverse signals a) across different sectors, and/or <br> b) different gear types |
| Level 4: Extreme concern | Severe problems with the stock assessment; severe retrospective bias. Assessment considered unreliable. | Stock trends are unprecedented; More rapid changes in stock abundance than have ever been seen previously, or a very long stretch of poor recruitment compared to previous patterns. | Extreme anomalies in multiple ecosystem indicators that are highly likely to impact the stock; Potential for cascading effects on other ecosystem components | Extreme anomalies in multiple performance indicators that are highly likely to impact the stock |

The table is applied by evaluating the severity of four types of considerations that could be used to support a scientific recommendation to reduce the ABC from the maximum permissible. These considerations are stock assessment considerations, population dynamics considerations, environmental/ecosystem considerations, and fishery performance. Examples of the types of concerns that might be relevant include the following:

1. "Assessment considerations-data-inputs: biased ages, skipped surveys, lack of fisheryindependent trend data; model fits: poor fits to fits to fishery or survey data, inability to simultaneously fit multiple data inputs; model performance: poor model convergence, multiple minima in the likelihood surface, parameters hitting bounds; estimation uncertainty: poorlyestimated but influential year classes; retrospective bias in biomass estimates.
2. "Population dynamics considerations-decreasing biomass trend, poor recent recruitment, inability of the stock to rebuild, abrupt increase or decrease in stock abundance.
3. "Environmental/ecosystem considerations-adverse trends in environmental/ecosystem indicators, ecosystem model results, decreases in ecosystem productivity, decreases in prey abundance or availability, increases or increases in predator abundance or productivity.
4. "Fishery performance-fishery CPUE is showing a contrasting pattern from the stock biomass trend, unusual spatial pattern of fishing, changes in the percent of TAC taken, changes in the duration of fishery openings.

## Assessment related considerations

BSAI Alaska plaice have been assessed annually from bottom trawl surveys conducted on the EBS shelf from 1982-2021, with one skipped year in 2020 due to the coronavirus pandemic). Survey and fishery age compositions are derived from otoliths collected during the surveys and the fishery and are available one year after collection for the assessment. The assessment model exhibits good fits to all compositional and abundance data and converges to a single minima in the likelihood surface. Fishery length compositions are fit in most years (instead of age composition). Lack of fit to compositional data from mis-ageing has not been a concern. Recruitment estimates track strong year classes that are consistent with the data. Retrospective analysis of the past 10 years of female spawning biomass estimates from the current assessment model does not indicate a pattern of concern regarding misspecification of the model. Survey estimates in 2012 and 2015 were more variable relative to the time-series (high in 2012 and lowest yet observed in 2015). Mohn's evaluation statistic was calculated at -0.01 .

## Population dynamics considerations

The female spawning biomass is projected to remain at levels well-above the $\mathrm{B}_{40 \%}$ value. The above average recruitment in 1998 and the recent increase from 2008-2013 is the result of above average year classes spawned in 2001 and 2002 that contributed to the high level of mature biomass. The female spawning biomass trend is similar to the total biomass trend with a peak level estimated in 1985 and a slow decline thereafter that continues to the present. Fishing pressure on Alaska plaice has been light as they are mostly caught as bycatch in the yellowfin sole fishery. Fishing mortality estimates have averaged 0.04 from 1975-2020, well below ABC levels. The present biomass is estimated at $58 \%$ of the peak 1985 level and is at 1.7 times the level of $\mathrm{B}_{\text {MSY }}$. Projections indicate that the FSB will remain wellabove the $\mathrm{B}_{\mathrm{MSY}}$ level through 2028. Population dynamics are not a concern for this assessment.

## Environmental/ecosystem considerations

Environmental processes: Beginning in approximately 2014, the eastern Bering Sea (EBS) entered a warm phase of unprecedented duration. The EBS remains in this warm phase, though to a lesser degree compared to the extreme years of 2018 and 2019. Through summer 2021, satellite observations of SST exceeded one standard deviation above the long term average for much of the past year. Sea ice formation in fall of 2020 was delayed due to residual warmth in the system, which has become the 'new normal' in this protracted warm phase. While the areal extent of sea ice was closer to the pre-2014 levels than at any point in the last 7 years, ice thickness differed between the northern (thicker ice) and southern (thinner/no ice) shelves due to opposing prevailing winds. The summer 2021 cold pool remained significantly reduced in area, and its southern boundary was shifted northwestward (Siddon, 2021). Summer bottom temperatures varied spatially over the shelf. Near-average conditions were present over the SEBS, while the NBS had a very warm inner domain (i.e., Norton Sound) and a small cold pool over the middle domain to the southwest of St. Lawrence Island (Rohan and Barnett, 2021).

Alaska plaice contain a glycol-protein that works to inhibit ice crystal formation in the blood, indicating this species may tolerate colder bottom water temperature. NOAA AFSC bottom trawl surveys conducted in the northern Bering Sea in 2010 and 2017-2019 showed a widespread distribution there. Between 2019 and 2021, the condition of Alaska plaice (as measured by length-weight residuals) declined in both the SEBS and NBS. Within the SEBS, condition was negative across strata (Rohan and Prohaska, 2021). Impacts to the overall population structure remain unknown, but sampling in the northern Bering Sea in 2017 for a NPRB project indicated differential age-at-maturity and size-at-age compared to Alaska plaice in the eastern Bering Sea.

Multiple ecosystem 'red flags' occurred in the NBS this year: crab population declines, salmon run failures in the Arctic-Yukon-Kuskokwim region, and seabird die-offs combined with low colony attendance and poor reproductive success. Whether a single or suite of mechanisms can be identified to explain these coincident events, the common thread in these collapses is the marine environment in the

NBS. Concerns about the food web dynamics and carrying capacity in the NBS have existed since 2018, highlighted by the gray whale Unusual Mortality Event and short-tailed shearwater mass mortality event.

The dominant prey of Alaska plaice are polychaete worms and clam siphons. Direct measurements of infaunal abundance trends are not available, however, abundance trends of motile epifauna that also consume infauna (i.e., indirect measurements) are quantified from the bottom trawl survey. The biomass of motile epifauna peaked in 2017 and remains above their long term mean in 2021. Trends in motile epifauna biomass indicate benthic productivity, which suggests that sufficient prey may have been available for Alaska plaice over the southeastern Bering Sea shelf. Brittle stars, sea stars, and other echinoderms account for $50 \%$ of this guild and these groups are well above their long term means. Crab within this functional group, including hermit crabs, king crabs, tanner crab, and snow crab are all below their long term means. (Whitehouse, 2021).

Predators of Alaska plaice include Pacific cod, Pacific halibut, and yellowfin sole. Pacific cod and Pacific halibut are included in the apex predator guild. In 2021, the biomass of apex predators was below their long term mean. The trend in the apex predator guild is largely driven by Pacific cod, whose recent (20162021) mean biomass is below their long term mean. In addition to a decrease in overall biomass, the spatial distribution of Pacific cod may provide a potential refuge from predation in the inner domain. Conversely, the spatial distribution of the relative abundance of Pacific halibut overlaps with that of Alaska plaice and may represent increased predation pressure. Yellowfin sole is included in the benthic forage guild, whose biomass is at the lowest level over the times series (Whitehouse, 2021), and may indicate a decrease in predation pressure.

Competitors for Alaska plaice prey resources include other benthic foragers, like northern rock sole and flathead sole, included in the benthic foragers guild. Trends in benthic forager biomass suggest a reduction in prey competition.

## Summary for Environmental/Ecosystem considerations:

- Near-average bottom temperatures were present over the SEBS, while the NBS had a very warm inner domain but a small cold pool over the middle domain to the southwest of St. Lawrence Island.
- In 2021, fish condition declined in both the SEBS and NBS relative to 2019.
- Concerns about the food web dynamics and carrying capacity in the NBS have existed since 2018 and may reflect poor feeding conditions in the northern Bering Sea.
- Sufficient prey may have been available over the southern shelf based on trends in motile epifauna.
- Predation pressure may be mixed; a decrease in Pacific cod biomass and potential refuge from predation in the inner domain may be countered by the spatial overlap with Pacific halibut in the inner domain of the SEBS. Declines in the benthic forager guild, including Yellowfin sole, may also indicate a decrease in predation pressure.
- Trends in benthic forager biomass over the SEBS suggest a reduction in prey competition.

Together, the most recent data available suggest an ecosystem risk Level 1 - Normal: "No apparent environmental/ecosystem concerns."

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## Fishery performance

Because Alaska plaice are a non-target stock, fishery performance indicators (e.g. CPUE) are not good indicators of population status. Overall, there are no fishery concerns regarding Alaska plaice.

Summary and ABC recommendations

| Assessment-related <br> considerations | Population dynamics <br> considerations | Environmental/ecosystem <br> considerations | Fishery performance <br> considerations |
| :--- | :--- | :--- | :--- |
| Level 1: Only minor, <br> low level of concern | Level 1: Stock trends <br> are typical for the stock <br> and expected given <br> stock dynamics; recent <br> recruitment is within <br> the normal range. | Level 1:No apparent <br> environmental/ecosystem <br> concerns | Level 1: Normal |

The overall score of level 1 suggests that setting the ABC below the maximum permissible is not warranted.

## Ecosystem considerations

## Ecosystem Effects on the stock

1) Prey availability/abundance trends

The feeding habits of juvenile Alaska plaice are relatively unknown, although the larvae are relatively large at hatching ( 5.85 mm ) with more advanced development than other flatfish (Pertseva-Ostroumova 1961).

For adult fish, Zhang (1987) found that the diet consisted primarily of polychaetes and amphipods regardless of size. For fish under 30 cm , polychaetes contributed $63 \%$ of the total diet with sipunculids (marine worms) and amphipods contributing $21.7 \%$ and $11.6 \%$, respectively. For fish over 30 cm , polychaetes contributed $75.2 \%$ of the total diet with amphipods and echiurans (marine worms) contributing $6.7 \%$ and $5.7 \%$, respectively. Similar results were in stomach sampling from 1993-1996, with polychaetes and marine worms composing the majority of the Alaska plaice diet (Lang et al. 2003). McConnaughy and Smith (2000) contrasted the food habits of several flatfish between areas of high and low CPUE, using aggregated data from 1982 to 1994. For Alaska plaice, the diets were nearly identical
with $76.5 \%$ of the diet composed of polychaetes and unsegmented coelomate worms in the high CPUE areas as compared to $83.1 \%$ in the low CPUE areas.

Alaska plaice diet from 99 stomachs sampled in 2000


## 2) Predator population trends

Alaska plaice contribute a relatively small portion of the diets of Pacific cod, Pacific halibut, and yellowfin sole as compared with other flatfish. Total consumption estimates of Alaska plaice from 1993 to 1996 ranged from 0 t in 1996 to 574 t in 1994 (Lang et al. 2003). Consumption by yellowfin sole is upon fish $<2 \mathrm{~cm}$ whereas consumption by Pacific halibut is upon fish > 19 cm (Lang et al. 2003).
3) Changes in habitat quality

The habitats occupied by Alaska plaice are influenced by temperature, which has shown considerable variation in the eastern Bering Sea in recent years. For example, the timing of spawning and advection to nursery areas are expected to be affected by environmental variation. Musienko (1970) reported that spawning occurs immediately after the ice melt, with peak spawning occurring at water temperatures from -1.53 to 4.11. In 1999, one of the coldest years in the eastern Bering Sea, the distribution was shifted further to the southeast than it was during 1998-2002. However, in 2003, one of the warmest years in the EBS, the distribution was shifted further to the southeast than observed in 1999.

## Fishery effects on the ecosystem

Alaska plaice are not a targeted species and are harvested in a variety of fisheries in the BSAI area. Since 2002, when single-species management for Alaska plaice was initiated, harvest estimates by fishery are available. Most Alaska plaice are harvested by the yellowfin sole fishery, accounting for over $80 \%$ of the Alaska plaice catch since 2002. Flathead sole, rock sole, and Pacific cod fisheries make up the remainder of the catch. The ecosystem effects of the yellowfin sole fishery can be found with the yellowfin sole assessment in this SAFE document.

Due to the minimal consumption estimates of Alaska plaice (Lang et al. 2003) by other groundfish predators, the yellowfin sole fishery does not have a significant impact upon those species preying upon Alaska plaice. Additionally, the relatively light fishing mortality rates experienced by Alaska plaice are not expected to have significant impacts on the size structure of the population or the maturity and
fecundity at age. It is not known what effects the fishery may have on the maturity-at-age of Alaska plaice but is it is expected to be minimal given the results of the histological maturity study completed in 2015 (TenBrink and Wilderbuer 2015). The yellowfin sole fishery, however, does contribute substantially to the total discards in the EBS, as indicated by the discarding of Alaska plaice discussed in this assessment, and general discards within this fishery discussed in the yellowfin sole assessment.

## Data Gaps and Research Priorities

Authors suggest a genetic study on Alaska plaice stock structure throughout their range in the Bering Sea and AI.

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## Tables

Table 10.1. Harvest (t) of Alaska plaice from 1977-2021. *2021 data includes catch through October 17, 2021.

|  | TAC | ABC | Catch |  | TAC | ABC | Catch |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1977 | NA | NA | 2,589 | 2003 | 10,000 | 137,000 | 9,978 |
| 1978 | NA | NA | 10,420 | 2004 | 10,000 | 203,000 | 7,888 |
| 1979 | NA | NA | 13,672 | 2005 | 8,000 | 189,000 | 11,194 |
| 1980 | NA | NA | 6,902 | 2006 | 8,000 | 188,000 | 17,318 |
| 1981 | NA | NA | 8,653 | 2007 | 15,000 | 183,000 | 19,522 |
| 1982 | NA | NA | 6,811 | 2008 | 50,000 | 217,000 | 17,376 |
| 1983 | NA | NA | 10,766 | 2009 | 50,000 | 232,000 | 13,944 |
| 1984 | NA | NA | 18,982 | 2010 | 50,000 | 224,000 | 16,165 |
| 1985 | NA | NA | 24,888 | 2011 | 16,000 | 65,100 | 23,656 |
| 1986 | NA | NA | 46,519 | 2012 | 24,000 | 53,400 | 16,612 |
| 1987 | NA | NA | 18,567 | 2013 | 20,000 | 55,200 | 23,523 |
| 1988 | NA | NA | 61,638 | 2014 | 24,500 | 55,100 | 19,447 |
| 1989 | NA | NA | 14,134 | 2015 | 18,500 | 44,900 | 14,614 |
| 1990 | NA | NA | 10,926 | 2016 | 14,500 | 41,000 | 13,384 |
| 1991 | NA | NA | 15,003 | 2017 | 13,000 | 36,000 | 16,491 |
| 1992 | NA | NA | 18,074 | 2018 | 16,100 | 34,590 | 23,340 |
| 1993 | NA | NA | 13,846 | 2019 | 18,000 | 33,600 | 16,251 |
| 1994 | NA | NA | 10,882 | 2020 | 17,000 | 31,600 | 20,078 |
| 1995 | NA | NA | 19,172 | 2021 | 24,500 | 31,657 | $15,411 *$ |
| 1996 | NA | NA | 16,096 |  |  |  |  |
| 1997 | NA | NA | 21,236 |  |  |  |  |
| 1998 | NA | NA | 14,296 |  |  |  |  |
| 1999 | NA | NA | 13,997 |  |  |  |  |
| 2000 | NA | NA | 14,487 |  |  |  |  |
| 2001 | NA | NA | 8,685 |  |  |  |  |
| 2002 | NA | NA | 12,176 |  |  |  |  |

Table 10.2 Discarded and retained BSAI Alaska plaice catch ( t ) for 2003-2021, from the NMFS Alaska regional office catch accounting system. *2021 data includes catch through October 17, 2021.

|  |  |  |  | $\%$ <br> retal |
| ---: | ---: | ---: | ---: | ---: |
| 2003 | 9,322 | 351 | 9,673 | $3.6 \%$ |
| discarded | retained | retained |  |  |
| 2004 | 7,481 | 408 | 7,888 | $5.2 \%$ |
| 2005 | 10,400 | 794 | 11,194 | $7.1 \%$ |
| 2007 | 14,755 | 2,563 | 17,318 | $14.8 \%$ |
| 2008 | 15,576 | 3,946 | 19,522 | $20.2 \%$ |
| 2009 | 9,330 | 8,047 | 17,377 | $46.3 \%$ |
| 2010 | 5,061 | 8,882 | 13,944 | $63.7 \%$ |
| 2011 | 7,196 | 10,321 | 16,164 | $63.8 \%$ |
| 2012 | 3,588 | 13,023 | 23,655 | $69.6 \%$ |
| 2013 | 9,052 | 14,471 | 23,522 | $78.4 \%$ |
| 2014 | 3,700 | 15,748 | 19,447 | $61.5 \%$ |
| 2015 | 1,231 | 13,383 | 14,614 | $91.6 \%$ |
| 2016 | 2,070 | 11,314 | 13,384 | $84.5 \%$ |
| 2017 | 1,953 | 14,538 | 16,491 | $88.2 \%$ |
| 2018 | 2,017 | 21,323 | 23,340 | $91.4 \%$ |
| 2019 | 608 | 15,643 | 16,251 | $96.3 \%$ |
| 2020 | 1,472 | 18,606 | 20,078 | $92.7 \%$ |
| $2021^{*}$ | 1,227 | 13,807 | 15,035 | $91.8 \%$ |

Table 10.3. Alaska plaice sample sizes from the BSAI fishery, 2008-2020.

| year | hauls w/ AK <br> plaice | hauls w/ <br> lengths | lengths | hauls w/ <br> otoliths | otoliths | otoliths <br> aged |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2008 | 11,741 | 1,641 | 7,494 | 329 | 381 | 0 |
| 2009 | 9,176 | 1,950 | 8,795 | 412 | 443 | 0 |
| 2010 | 9,743 | 1,810 | 8,781 | 344 | 398 | 0 |
| 2011 | 9,914 | 2,800 | 14,328 | 545 | 686 | 0 |
| 2012 | 9,782 | 2,962 | 13,611 | 548 | 600 | 0 |
| 2013 | 11,026 | 3,469 | 16,646 | 649 | 787 | 0 |
| 2014 | 8,217 | 1,900 | 14,366 | 607 | 714 | 0 |
| 2015 | 11,263 | 2,501 | 11,924 | 475 | 577 | 0 |
| 2016 | 13,469 | 1,704 | 12,273 | 495 | 581 | 0 |
| 2017 | 12,353 | 2,999 | 14,464 | 594 | 667 | 0 |
| 2018 | 13,618 | 4,461 | 24,917 | 859 | 1,155 | 0 |
| 2019 | 18,333 | 4,324 | 21,113 | NA | NA | 0 |
| 2020 | 14,019 | 3,001 | 16,403 | NA | NA | 0 |

Table 10.4. Non-commercial catches ( t ) of Alaska plaice in the BSAI, 2010-2019. As of the SAFE publication, 2020 data were not yet available.

|  | NMFS area |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :--- |
|  | 508 | 509 | 512 | 513 | 514 | 516 | 517 | 519 | 521 | 524 |  |
| 2010 | 0.03 | 1.43 | 1.43 | 2.66 | 13.74 | 1.42 | 0.18 | 0.00 | 0.57 | 6.17 | 27.64 |
| 2011 | 0.02 | 1.84 | 1.29 | 2.57 | 9.62 | 1.59 | 0.00 | 0.00 | 0.35 | 1.19 | 18.47 |
| 2012 | 0.02 | 1.41 | 1.40 | 3.55 | 9.95 | 1.30 | 0.12 | 0.00 | 0.63 | 1.76 | 20.15 |
| 2013 | 0.00 | 2.86 | 0.82 | 4.07 | 6.41 | 1.67 | 0.01 | 0.00 | 0.54 | 0.79 | 17.18 |
| 2014 | 0.01 | 1.49 | 0.99 | 3.61 | 6.11 | 0.96 | 0.00 | 0.00 | 0.70 | 1.66 | 15.53 |
| 2015 | 0.03 | 0.84 | 0.74 | 2.92 | 5.54 | 0.60 | 0.00 | 0.00 | 0.59 | 1.20 | 12.46 |
| 2016 | 0.02 | 1.26 | 0.83 | 3.72 | 4.69 | 0.56 | 0.01 | 0.00 | 1.85 | 1.99 | 14.92 |
| 2017 | 0.03 | 1.67 | 0.91 | 3.19 | 13.08 | 0.94 | 0.03 | 0.00 | 2.82 | 6.09 | 28.76 |
| 2018 | 0.02 | 1.56 | 0.54 | 3.33 | 5.86 | 0.66 | 0.01 | 0.00 | 3.84 | 3.59 | 19.43 |
| 2019 | 0.04 | 1.18 | 0.82 | 2.65 | 8.79 | 0.76 | 0.01 | 0.01 | 2.88 | 7.11 | 24.24 |

Table 10.5. Estimated biomass, $95 \%$ confidence intervals and standard deviations ( t ) of Alaska plaice from the eastern Bering Sea shelf trawl survey, 1982-2021. No survey occurred in 2020 due to the coronavirus pandemic.

|  | biomass (t) | std. deviation | lower C.I. | upper C.I. |
| :---: | :---: | :---: | :---: | :---: |
| 1982 | 716,020 | 64,856 | 587,605 | 844,434 |
| 1983 | 651,434 | 58,712 | 535,183 | 767,685 |
| 1984 | 769,540 | 112,631 | 541,913 | 997,168 |
| 1985 | 579,978 | 61,006 | 457,966 | 701,990 |
| 1986 | 548,626 | 62,608 | 423,411 | 673,842 |
| 1987 | 547,867 | 55,866 | 437,253 | 658,482 |
| 1988 | 676,860 | 137,491 | 404,628 | 949,092 |
| 1989 | 515,039 | 57,013 | 402,154 | 627,925 |
| 1990 | 495,346 | 46,557 | 403,163 | 587,530 |
| 1991 | 534,274 | 50,503 | 433,268 | 635,280 |
| 1992 | 516,518 | 55,630 | 406,370 | 626,665 |
| 1993 | 516,126 | 50,553 | 416,031 | 616,222 |
| 1994 | 623,314 | 53,293 | 517,794 | 728,834 |
| 1995 | 554,850 | 63,028 | 430,055 | 679,645 |
| 1996 | 532,322 | 67,555 | 398,563 | 666,082 |
| 1997 | 632,145 | 71,474 | 490,625 | 773,664 |
| 1998 | 455,904 | 58,691 | 338,523 | 573,285 |
| 1999 | 480,514 | 40,346 | 400,628 | 560,399 |
| 2000 | 446,101 | 67,613 | 309,456 | 582,746 |
| 2001 | 546,224 | 68,497 | 410,600 | 681,848 |
| 2002 | 425,663 | 53,533 | 318,598 | 532,728 |
| 2003 | 462,038 | 95,866 | 270,307 | 653,769 |
| 2004 | 480,961 | 63,022 | 356,177 | 605,744 |
| 2005 | 507,713 | 55,471 | 397,880 | 617,546 |
| 2006 | 641,642 | 83,064 | 475,514 | 807,771 |
| 2007 | 422,986 | 37,452 | 348,832 | 497,140 |
| 2008 | 509,303 | 47,430 | 415,391 | 603,215 |
| 2009 | 529,699 | 50,359 | 429,988 | 629,410 |
| 2010 | 498,117 | 46,866 | 405,323 | 590,912 |
| 2011 | 519,578 | 72,781 | 374,015 | 665,141 |
| 2012 | 581,896 | 83,432 | 415,033 | 748,759 |
| 2013 | 505,583 | 65,596 | 375,703 | 635,464 |
| 2014 | 451,624 | 48,850 | 354,901 | 548,347 |
| 2015 | 355,640 | 38,641 | 279,132 | 432,149 |
| 2016 | 425,217 | 41,191 | 343,659 | 506,775 |
| 2017 | 491,050 | 52,458 | 387,182 | 594,918 |
| 2018 | 419,509 | 37,223 | 345,807 | 493,212 |
| 2019 | 368,787 | 29,038 | 311,292 | 426,282 |
| 2021 | 333,830 | 28,647 | 277,110 | 390,551 |

Table 10.6a. Alaska plaice population female numbers at age (millions) estimated from the NMFS Bering Sea groundfish surveys and age readings of sampled fish.

| females | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.0 | 0.0 | 2.8 | 11.3 | 43.8 | 37.3 | 60.1 | 31.8 | 60.2 | 54.2 | 24.7 | 75.0 | 0.0 | 10.1 | 4.4 | 5.7 | 18.5 | 13.9 | 32.9 | 0.0 | 11.2 | 0.0 | 4.4 |
| 1992 | 0.0 | 0.0 | 4.1 | 4.7 | 7.4 | 31.6 | 27.8 | 20.1 | 34.3 | 24.0 | 15.7 | 25.0 | 21.7 | 20.7 | 22.9 | 17.5 | 19.8 | 10.3 | 7.8 | 2.3 | 10.6 | 10.3 | 8.7 |
| 1993 | 0.0 | 0.0 | 5.5 | 14.9 | 30.2 | 42.1 | 53.7 | 5.6 | 2.4 | 25.2 | 42.7 | 26.6 | 38.8 | 24.9 | 16.1 | 12.1 | 12.1 | 4.0 | 9.7 | 0.0 | 3.6 | 4.5 | 12.3 |
| 1994 | 0.0 | 0.0 | 7.7 | 14.8 | 45.1 | 38.8 | 21.5 | 45.2 | 16.5 | 11.3 | 55.3 | 11.7 | 50.0 | 21.5 | 31.3 | 21.8 | 8.3 | 13.7 | 0.0 | 2.0 | 0.0 | 0.0 | 30.3 |
| 1995 | 0.0 | 0.0 | 10.0 | 31.4 | 32.8 | 47.2 | 34.3 | 16.8 | 23.4 | 16.6 | 10.1 | 30.1 | 30.4 | 27.4 | 19.1 | 19.7 | 21.3 | 9.3 | 14.2 | 16.1 | 4.2 | 0.0 | 26.7 |
| 1998 | 0.0 | 0.9 | 3.7 | 9.8 | 35.7 | 37.2 | 58.5 | 28.4 | 39.9 | 43.0 | 17.6 | 24.7 | 14.5 | 10.5 | 19.6 | 8.8 | 6.1 | 6.5 | 11.9 | 6.1 | 3.3 | 3.4 | 6.2 |
| 2000 | 0.0 | 0.1 | 3.9 | 3.9 | 22.2 | 27.1 | 53.2 | 26.8 | 33.9 | 18.9 | 21.0 | 15.9 | 13.8 | 16.9 | 14.2 | 11.5 | 23.6 | 9.9 | 18.4 | 3.8 | 5.7 | 6.9 | 26.6 |
| 2001 | 0.0 | 0.0 | 4.1 | 9.5 | 13.6 | 45.0 | 21.6 | 83.3 | 34.9 | 45.0 | 15.2 | 16.2 | 10.9 | 18.9 | 18.5 | 8.6 | 9.8 | 21.1 | 15.9 | 8.9 | 3.2 | 3.8 | 23.5 |
| 2002 | 0.0 | 0.0 | 2.8 | 8.0 | 17.8 | 16.0 | 31.4 | 15.6 | 39.4 | 17.1 | 30.6 | 15.4 | 15.1 | 22.1 | 4.7 | 14.0 | 4.9 | 4.9 | 7.5 | 12.0 | 11.4 | 5.0 | 27.7 |
| 2005 | 0.9 | 2.0 | 13.2 | 23.1 | 34.2 | 31.5 | 30.8 | 27.9 | 23.6 | 27.9 | 32.3 | 19.0 | 21.8 | 17.9 | 10.7 | 9.7 | 5.8 | 9.2 | 3.5 | 3.1 | 3.6 | 0.0 | 31. |
| 2006 | 0.3 | 4.4 | 50.8 | 25.1 | 60.1 | 60.1 | 45.5 | 31.0 | 22.2 | 12.2 | 27.7 | 27.1 | 25.6 | 8.4 | 18.6 | 14.6 | 19.3 | 8.1 | 7.4 | 13.2 | 6.2 | 7.6 | 9.3 |
| 2007 | 0.0 | 4.0 | 43.5 | 56.5 | 36.0 | 25.1 | 20.2 | 25.4 | 28.7 | 16.8 | 18.2 | 17.7 | 9.1 | 22.6 | 7.9 | 8.2 | 6.2 | 9.6 | 0.8 | 8.3 | 0.0 | 5.1 | 23.4 |
| 2008 | 0.0 | 0.0 | 12.3 | 46.1 | 60.0 | 42.7 | 21.6 | 33.9 | 33.0 | 26.4 | 10.8 | 14.1 | 26.2 | 25.4 | 21.0 | 5.9 | 1.4 | 8.7 | 9.5 | 10.6 | 7.4 | 1.5 | 19.3 |
| 2009 | 0.0 | 0.5 | 9.9 | 14.3 | 89.0 | 61.3 | 24.4 | 36.1 | 26.6 | 17.6 | 15.9 | 12.0 | 18.5 | 16.5 | 19.5 | 21.3 | 10.6 | 9.3 | 9.8 | 7.1 | 8.8 | 6.2 | 11.6 |
| 2010 | 0.0 | 0.0 | 4.6 | 10.4 | 16.1 | 85.2 | 56.0 | 28.9 | 29.6 | 26.8 | 13.4 | 13.3 | 17.4 | 8.6 | 21.0 | 17.6 | 14.6 | 9.8 | 10.6 | 2.4 | 10.2 | 4.5 | 17.6 |
| 2011 | 0.0 | 0.0 | 0.6 | 21.0 | 34.4 | 31.6 | 73.7 | 60.3 | 24.6 | 16.2 | 26.2 | 8.6 | 9.7 | 14.1 | 4.2 | 19.3 | 13.6 | 12.0 | 8.9 | 7.1 | 9.8 | 6.7 | 20.1 |
| 2012 | 0.0 | 0.0 | 1.4 | 10.0 | 19.6 | 36.9 | 40.0 | 62.8 | 55. | 41.9 | 22.9 | 29.8 | 10.6 | 10.4 | 11.7 | 16.5 | 25.2 | 22.4 | 8.2 | 14.7 | 15.5 | 6.3 | 31.1 |
| 2013 | 0.0 | 0.0 | 3.5 | 8.9 | 12.7 | 37.3 | 33.8 | 21.3 | 49.3 | 38.5 | 43.3 | 27.8 | 14.0 | 11.8 | 3.7 | 9.2 | 7.1 | 6.8 | 5.1 | 8.4 | 9.7 | 7.0 | 22.4 |
| 2014 | 0.0 | 0.7 | 2.3 | 7.2 | 20.6 | 17.1 | 28.7 | 38.5 | 30.4 | 43.4 | 29.1 | 7.6 | 16.4 | 10.7 | 10.2 | 8.2 | 12.8 | 3.9 | 8.1 | 11.6 | 6.3 | 4.3 | 21.6 |
| 2016 | 0.0 | 0.0 | 2.9 | 7.9 | 17.2 | 14.9 | 20.7 | 8.5 | 35.3 | 11.8 | 18.5 | 37.4 | 19.1 | 8.9 | 5.2 | 12.4 | 14.1 | 3.0 | 3.4 | 1.7 | 13.6 | 7.0 | 23.6 |
| 2017 | 0.9 | 1.1 | 10.4 | 12.6 | 14.2 | 53.7 | 14.6 | 26.9 | 18.3 | 18.8 | 15.8 | 9.7 | 15.0 | 28.0 | 9.3 | 12.6 | 12.6 | 5.0 | 0.5 | 4.1 | 3.3 | 6.7 | 25.5 |
| 2018 | 0.0 | 9.6 | 5.3 | 11.0 | 6.6 | 19.2 | 44.7 | 6.1 | 29.5 | 6.3 | 17.9 | 10.8 | 4.2 | 18.0 | 14.2 | 9.5 | 11.4 | 3.8 | 4.3 | 7.3 | 0.7 | 2.9 | 26.5 |
| 2019 | 0.2 | 9.3 | 26.6 | 11.5 | 8.1 | 12.4 | 20.6 | 27.2 | 15.7 | 16.3 | 5.7 | 9.1 | 11.1 | 12.2 | 10.5 | 4.1 | 6.2 | 4.2 | 12.4 | 5.4 | 0.3 | 1.1 | 10.8 |

Table 10.6b. Alaska plaice population male numbers at age (millions) estimated from the NMFS Bering Sea groundfish surveys and age readings of sampled fish.

| males | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1988 | 0.0 | 0.0 | 3.2 | 6.0 | 35.9 | 35.5 | 45.8 | 31.0 | 98.7 | 16.5 | 6.1 | 27.8 | 14.6 | 9.5 | 4.1 | 27.7 | 11.1 | 23.6 | 25.5 | 12.6 | 0.0 | 5.0 | 15.3 |
| 1992 | 0.0 | 5.2 | 16.4 | 1.2 | 22.3 | 28.9 | 18.7 | 21.6 | 45.1 | 24.7 | 20.8 | 19.5 | 10.6 | 9.9 | 13.3 | 29.2 | 10.4 | 11.6 | 5.3 | 5.4 | 3.0 | 14.0 | 5.2 |
| 1993 | 0.0 | 0.0 | 2.9 | 36.8 | 14.8 | 25.5 | 43.7 | 15.2 | 17.7 | 34.2 | 42.9 | 6.1 | 12.1 | 15.7 | 12.0 | 4.4 | 5.6 | 0.0 | 26.5 | 0.0 | 9.4 | 26.6 | 24.6 |
| 1994 | 0.2 | 2.0 | 13.6 | 13.1 | 57.6 | 61.5 | 15.2 | 30.2 | 21.3 | 14.8 | 57.3 | 47.0 | 31.0 | 45.4 | 24.3 | 9.7 | 0.0 | 23.5 | 1.7 | 0.0 | 0.0 | 2.8 | 20.6 |
| 1995 | 0.0 | 0.0 | 0.0 | 28.6 | 20.5 | 84.9 | 21.0 | 17.6 | 39.0 | 17.4 | 20.1 | 17.2 | 27.5 | 28.1 | 20.7 | 11.6 | 8.5 | 1.3 | 5.7 | 14.6 | 1.5 | 4.4 | 16.0 |
| 1998 | 0.0 | 0.3 | 5.1 | 22.1 | 37.9 | 34.1 | 51.3 | 31.6 | 26.4 | 27.2 | 11.5 | 18.0 | 15.0 | 12.0 | 11.9 | 5.4 | 6.6 | 7.4 | 3.0 | 3.1 | 1.9 | 3.4 | 0.1 |
| 2000 | 0.0 | 0.0 | 9.0 | 1.0 | 20.9 | 20.9 | 75.6 | 44.6 | 27.8 | 30.2 | 21.6 | 16.4 | 3.3 | 12.3 | 6.8 | 6.6 | 29.6 | 21.3 | 11.7 | 0.0 | 7.0 | 20.0 | 18.8 |
| 2001 | 0.0 | 0.0 | 1.7 | 17.1 | 6.4 | 71.1 | 46.7 | 58.7 | 26.3 | 53.3 | 23.1 | 71.1 | 5.4 | 13.4 | 35.7 | 21.3 | 9.2 | 21.6 | 9.1 | 2.2 | 5.0 | 2.2 | 15.2 |
| 2002 | 0.0 | 0.0 | 1.1 | 10.0 | 14.4 | 20.6 | 27.9 | 51.9 | 36.2 | 19.7 | 30.5 | 12.7 | 29.3 | 4.9 | 4.0 | 23.9 | 1.5 | 5.5 | 9.1 | 11.7 | 1.6 | 1.1 | 7.7 |
| 2005 | 0.6 | 4.1 | 10.1 | 31.9 | 23.1 | 49.9 | 14.4 | 42.6 | 18.5 | 32.4 | 40.8 | 21.6 | 10.4 | 13.2 | 8.6 | 5.9 | 1.3 | 5.3 | 6.5 | 8.7 | 0.0 | 2.1 | 3.8 |
| 2006 | 0.1 | 9.8 | 49.8 | 27.2 | 61.9 | 68.4 | 48.6 | 25.5 | 21.7 | 33.5 | 23.0 | 52.3 | 35.5 | 8.3 | 20.9 | 0.0 | 11.3 | 6.7 | 0.0 | 1.4 | 0.0 | 0.0 | 19.5 |
| 2007 | 1.6 | 4.0 | 39.2 | 63.3 | 46.7 | 18.9 | 21.2 | 41.5 | 37.0 | 6.9 | 12.8 | 20.2 | 20.9 | 28.1 | 16.1 | 2.3 | 2.3 | 9.6 | 2.8 | 1.0 | 3.6 | 3.9 | 3.2 |
| 2008 | 0.0 | 0.0 | 6.7 | 87.2 | 60.3 | 14.5 | 29.6 | 55.2 | 13.5 | 33.0 | 15.6 | 13.0 | 23.6 | 37.3 | 12.3 | 28.5 | 15.4 | 3.8 | 11.9 | 19.5 | 10.5 | 3.0 | 4.6 |
| 2009 | 0.0 | 2.9 | 5.4 | 12.3 | 93.4 | 84.1 | 71.9 | 39.9 | 23.1 | 25.6 | 11.5 | 39.2 | 19.2 | 26.5 | 16.4 | 18.7 | 26.2 | 12.7 | 9.5 | 12.4 | 3.4 | 2.6 | 14.3 |
| 2010 | 0.0 | 0.5 | 6.6 | 17.0 | 31.7 | 61.5 | 65.0 | 40.4 | 48.4 | 35.7 | 30.2 | 24.5 | 11.0 | 5.3 | 13.8 | 12.9 | 46.7 | 15.5 | 6.5 | 17.0 | 8.0 | 1.6 | 27.6 |
| 2011 | 0.0 | 1.1 | 1.4 | 17.5 | 44.2 | 26.8 | 57.4 | 64.0 | 22.5 | 33.9 | 32.2 | 11.7 | 13.3 | 15.2 | 21.5 | 24.2 | 12.3 | 24.5 | 12.3 | 9.4 | 5.3 | 10.9 | 14. |
| 2012 | 0.0 | 0.0 | 7.4 | 3.6 | 39.9 | 62.2 | 25.9 | 88.4 | 50.2 | 33.7 | 20.9 | 12.9 | 9.2 | 12.6 | 12.1 | 7.3 | 12.0 | 20.0 | 11.5 | 16.2 | 11.9 | 2.7 | 15.4 |
| 2013 | 0.0 | 0.0 | 1.3 | 7.1 | 21.6 | 47.1 | 35.2 | 26.8 | 51.8 | 73.9 | 32.2 | 16.9 | 15.8 | 13.9 | 4.7 | 17.9 | 12.5 | 14.0 | 5.7 | 1.1 | 1.9 | 2.2 | 16.9 |
| 2014 | 0.0 | 0.0 | 1.5 | 0.5 | 28.1 | 22.3 | 52.8 | 32.2 | 14.9 | 46.2 | 5.8 | 15.4 | 9.2 | 9.1 | 6.1 | 25.4 | 2.3 | 4.2 | 10.5 | 3.6 | 1.5 | 2.8 | 21.8 |
| 2016 | 0.4 | 1.3 | 2.7 | 5.4 | 23.9 | 7.4 | 11.7 | 22.9 | 17.4 | 31.2 | 12.5 | 28.8 | 15.6 | 11.7 | 18.4 | 9.5 | 9.7 | 4.7 | 0.3 | 3.4 | 10.9 | 9.3 | 10.9 |
| 2017 | 4.0 | 1.6 | 5.8 | 7.4 | 21.4 | 53.6 | 21.5 | 25.2 | 8.5 | 30.9 | 17.5 | 8.5 | 28.3 | 18.7 | 13.9 | 13.0 | 8.7 | 10.8 | 2.1 | 3.5 | 1.8 | 3.9 | 31.5 |
| 2018 | 1.0 | 13.8 | 10.9 | 27.3 | 13.6 | 15.2 | 58.9 | 14.2 | 2.3 | 6.8 | 8.1 | 23.1 | 14.7 | 20.6 | 13.8 | 4.9 | 5.0 | 1.1 | 12.3 | 4.3 | 0.9 | 2.3 | 16.8 |
| 2019 | 0.7 | 4.1 | 38.5 | 11.6 | 12.8 | 10.3 | 21.9 | 15.8 | 17.4 | 14.0 | 10.0 | 7.7 | 11.2 | 11.4 | 16.2 | 13.4 | 9.1 | 3.9 | 1.0 | 0.0 | 4.0 | 1.5 | 1.6 |

Table 10.7. Alaska plaice sample sizes from the BSAI trawl survey. The hauls columns refer to the number (Num.) of hauls from which either lengths or aged otoliths were obtained. No survey was performed in 2020 due to the coronavirus pandemic.

| Year | Total <br> Hauls | Hauls <br> w/Lengths | Num. <br> lengths | Hauls <br> w/otoliths | Hauls <br> w/ages | Num. <br> otoliths | Num. <br> ages |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 334 | 152 | 14274 | 27 | 27 | 298 | 298 |
| 1983 | 353 | 118 | 11624 |  |  |  |  |
| 1984 | 355 | 151 | 14026 | 32 |  | 457 |  |
| 1985 | 357 | 168 | 10914 | 24 |  | 430 |  |
| 1986 | 354 | 236 | 12349 |  |  |  |  |
| 1987 | 357 | 172 | 8533 |  |  |  |  |
| 1988 | 373 | 170 | 7079 | 10 | 10 | 284 | 284 |
| 1989 | 374 | 207 | 7741 |  |  |  |  |
| 1990 | 371 | 215 | 7739 | 10 |  | 228 |  |
| 1991 | 372 | 235 | 8163 |  |  |  |  |
| 1992 | 356 | 219 | 7584 | 10 | 10 | 311 | 311 |
| 1993 | 375 | 241 | 8365 | 4 | 4 | 183 | 183 |
| 1994 | 375 | 248 | 9299 | 6 | 6 | 228 | 228 |
| 1995 | 376 | 252 | 9919 | 11 | 11 | 287 | 285 |
| 1996 | 375 | 254 | 10186 | 5 |  | 250 |  |
| 1997 | 376 | 248 | 10143 | 3 |  | 82 |  |
| 1998 | 375 | 281 | 10101 | 14 | 14 | 420 | 416 |
| 1999 | 373 | 268 | 13024 | 13 |  | 297 |  |
| 2000 | 372 | 250 | 9803 | 16 | 16 | 368 | 359 |
| 2001 | 375 | 261 | 10990 | 16 | 16 | 339 | 335 |
| 2002 | 375 | 251 | 8409 | 24 | 24 | 359 | 355 |
| 2003 | 376 | 252 | 8343 | 15 |  | 320 |  |
| 2004 | 375 | 262 | 8578 | 17 |  | 325 |  |
| 2005 | 373 | 262 | 9284 | 20 | 20 | 341 | 337 |
| 2006 | 376 | 255 | 12097 | 18 | 18 | 362 | 362 |
| 2007 | 376 | 261 | 11729 | 43 | 42 | 343 | 335 |
| 2008 | 375 | 252 | 12804 | 35 | 35 | 342 | 338 |
| 2009 | 376 | 233 | 13547 | 68 | 68 | 620 | 590 |
| 2010 | 376 | 225 | 11366 | 60 | 51 | 627 | 448 |
| 2011 | 376 | 236 | 11514 | 59 | 59 | 571 | 560 |
| 2012 | 376 | 240 | 10399 | 62 | 62 | 484 | 475 |
| 2013 | 376 | 221 | 9705 | 69 | 69 | 544 | 537 |
| 2014 | 376 | 215 | 7296 | 51 | 51 | 502 | 490 |
| 2015 | 376 | 223 | 5989 |  |  |  |  |
| 2016 | 376 | 250 | 6312 | 56 | 56 | 488 | 472 |
| 2017 | 376 | 258 | 8065 | 70 | 70 | 556 | 552 |
| 2018 | 376 | 280 | 12038 | 60 | 59 | 472 | 463 |
| 2019 | 376 | 277 | 9071 | 61 | 60 | 525 | 517 |
| 2021 | 376 | 275 | 8530 | 65 |  | 522 |  |
|  |  |  |  |  |  |  |  |

Table 10.8 Estimated maturity at age for female Alaska plaice. Anatomical estimates were estimated by Zhang (1987). Histological estimates (TenBrink and Wilderbuer 2015) are used in the assessment.
$\left.\begin{array}{ccc}\hline & \text { proportion mature }\end{array}\right]$.

Table 10.9a. Model estimates (1975-2004) of female spawning biomass, total biomass (ages 3+), and recruitment (age 3), with comparison to the 2019 model estimates. CV $=$ coefficient of variation for the 2021 results from MCMC.

|  | female spawning biomass <br> (t) |  |  | total biomass (t) |  |  | recruitment (millions) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2019 | 2021 | CV | 2019 | 2021 | CV | 2019 | 2021 | CV |
| 1975 | 121,067 | 123,353 | 0.08 | 459,143 | 461,364 | 0.04 | 276 | 277 | 0.16 |
| 1976 | 136,623 | 138,402 | 0.07 | 503,309 | 505,094 | 0.03 | 271 | 271 | 0.15 |
| 1977 | 162,595 | 163,847 | 0.06 | 556,371 | 557,972 | 0.03 | 512 | 516 | 0.09 |
| 1978 | 196,546 | 197,324 | 0.05 | 603,031 | 604,383 | 0.03 | 310 | 309 | 0.12 |
| 1979 | 227,682 | 228,086 | 0.04 | 638,112 | 639,282 | 0.02 | 281 | 281 | 0.11 |
| 1980 | 253,099 | 253,301 | 0.03 | 667,294 | 668,355 | 0.02 | 293 | 292 | 0.10 |
| 1981 | 274,910 | 275,042 | 0.03 | 695,898 | 696,844 | 0.02 | 202 | 202 | 0.11 |
| 1982 | 293,156 | 293,280 | 0.03 | 717,966 | 718,745 | 0.02 | 217 | 216 | 0.10 |
| 1983 | 312,114 | 312,294 | 0.03 | 737,654 | 738,319 | 0.02 | 235 | 235 | 0.09 |
| 1984 | 328,224 | 328,455 | 0.02 | 750,529 | 751,025 | 0.02 | 272 | 271 | 0.08 |
| 1985 | 336,915 | 337,149 | 0.02 | 744,387 | 744,688 | 0.02 | 122 | 121 | 0.13 |
| 1986 | 334,267 | 334,438 | 0.02 | 727,489 | 727,578 | 0.02 | 134 | 133 | 0.11 |
| 1987 | 321,777 | 321,888 | 0.02 | 688,943 | 688,885 | 0.02 | 231 | 230 | 0.08 |
| 1988 | 311,661 | 311,684 | 0.02 | 674,534 | 674,270 | 0.02 | 141 | 140 | 0.11 |
| 1989 | 288,277 | 288,207 | 0.02 | 615,934 | 615,546 | 0.02 | 188 | 187 | 0.09 |
| 1990 | 286,601 | 286,511 | 0.02 | 612,428 | 611,783 | 0.02 | 293 | 290 | 0.07 |
| 1991 | 284,096 | 283,964 | 0.02 | 607,897 | 606,958 | 0.02 | 171 | 168 | 0.10 |
| 1992 | 276,969 | 276,763 | 0.02 | 603,322 | 602,010 | 0.02 | 262 | 259 | 0.08 |
| 1993 | 268,008 | 267,716 | 0.02 | 594,715 | 593,092 | 0.02 | 212 | 210 | 0.08 |
| 1994 | 261,909 | 261,503 | 0.02 | 595,368 | 593,333 | 0.02 | 297 | 293 | 0.07 |
| 1995 | 257,129 | 256,572 | 0.02 | 597,945 | 595,458 | 0.02 | 227 | 224 | 0.08 |
| 1996 | 251,194 | 250,447 | 0.02 | 592,989 | 590,087 | 0.02 | 228 | 226 | 0.08 |
| 1997 | 247,943 | 246,973 | 0.02 | 586,857 | 583,575 | 0.02 | 122 | 120 | 0.11 |
| 1998 | 244,951 | 243,739 | 0.02 | 574,643 | 570,994 | 0.02 | 141 | 138 | 0.10 |
| 1999 | 246,131 | 244,694 | 0.02 | 567,965 | 564,022 | 0.02 | 142 | 140 | 0.10 |
| 2000 | 248,451 | 246,791 | 0.02 | 560,694 | 556,439 | 0.02 | 165 | 162 | 0.09 |
| 2001 | 251,654 | 249,793 | 0.02 | 552,737 | 548,377 | 0.02 | 185 | 186 | 0.09 |
| 2002 | 255,755 | 253,706 | 0.02 | 550,792 | 546,151 | 0.02 | 194 | 190 | 0.08 |
| 2003 | 255,366 | 253,154 | 0.02 | 545,305 | 540,516 | 0.02 | 196 | 195 | 0.08 |
| 2004 | 253,135 | 250,811 | 0.02 | 549,216 | 544,059 | 0.02 | 329 | 323 | 0.06 |

Table 10.9b. Model estimates (2005-2021) of female spawning biomass, total biomass (ages 3+), and recruitment (age 3), with comparison to the 2019 model estimates. $\mathrm{CV}=$ coefficient of variation for the 2021 results from MCMC

|  | female spawning biomass ( t ) |  |  | total biomass (t) |  |  | recruitment (millions) |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2019 | 2021 | CV | 2019 | 2021 | CV | 2019 | 2021 | CV |
| 2005 | 249,168 | 246,756 | 0.02 | 561,502 | 555,696 | 0.02 | 384 | 374 | 0.06 |
| 2006 | 242,629 | 240,130 | 0.02 | 563,329 | 557,393 | 0.02 | 144 | 147 | 0.10 |
| 2007 | 234,261 | 231,677 | 0.02 | 561,348 | 555,115 | 0.02 | 198 | 196 | 0.09 |
| 2008 | 227,649 | 224,973 | 0.03 | 557,172 | 550,375 | 0.02 | 190 | 181 | 0.09 |
| 2009 | 225,280 | 222,491 | 0.03 | 550,181 | 542,967 | 0.02 | 93 | 90 | 0.13 |
| 2010 | 227,957 | 224,970 | 0.03 | 544,706 | 537,444 | 0.02 | 112 | 115 | 0.12 |
| 2011 | 232,110 | 228,814 | 0.03 | 531,901 | 525,057 | 0.02 | 67 | 76 | 0.15 |
| 2012 | 234,044 | 230,469 | 0.03 | 514,202 | 507,461 | 0.02 | 192 | 189 | 0.10 |
| 2013 | 234,526 | 230,800 | 0.03 | 498,222 | 492,699 | 0.02 | 87 | 105 | 0.15 |
| 2014 | 228,355 | 224,523 | 0.03 | 472,739 | 468,205 | 0.03 | 75 | 79 | 0.19 |
| 2015 | 221,437 | 217,529 | 0.03 | 453,587 | 448,724 | 0.03 | 150 | 124 | 0.18 |
| 2016 | 213,914 | 210,119 | 0.03 | 438,967 | 433,499 | 0.03 | 134 | 115 | 0.23 |
| 2017 | 204,492 | 201,119 | 0.03 | 437,517 | 433,194 | 0.03 | 366 | 392 | 0.16 |
| 2018 | 192,433 | 189,703 | 0.03 | 426,497 | 431,118 | 0.04 | 119 | 292 | 0.32 |
| 2019 | 179,279 | 177,137 | 0.03 | 424,410 | 431,739 | 0.04 | 413 | 383 | 0.47 |
| 2020 |  | 167,786 | 0.04 |  | 439,490 | 0.05 |  | 238 | 0.64 |
| 2021 |  | 158,090 | 0.04 |  | 455,187 | 0.06 |  | 413 | 0.00 |

Table 10.10. Model estimates of age-specific Alaska plaice female and male selectivity from the fishery and the EBS shelf bottom trawl survey.

|  | fishery |  | survey |  |
| ---: | ---: | ---: | ---: | ---: |
|  | females | males | females | males |
| 3 | 0.007 | 0.018 | 0.017 | 0.015 |
| 4 | 0.016 | 0.032 | 0.043 | 0.043 |
| 5 | 0.036 | 0.056 | 0.105 | 0.117 |
| 6 | 0.077 | 0.095 | 0.234 | 0.281 |
| 7 | 0.156 | 0.158 | 0.444 | 0.536 |
| 8 | 0.293 | 0.250 | 0.676 | 0.773 |
| 9 | 0.480 | 0.371 | 0.845 | 0.909 |
| 10 | 0.673 | 0.512 | 0.935 | 0.967 |
| 11 | 0.821 | 0.651 | 0.974 | 0.989 |
| 12 | 0.911 | 0.768 | 0.990 | 0.996 |
| 13 | 0.958 | 0.855 | 0.996 | 0.999 |
| 14 | 0.981 | 0.913 | 0.999 | 1.000 |
| 15 | 0.991 | 0.949 | 0.999 | 1.000 |
| 16 | 0.996 | 0.971 | 1.000 | 1.000 |
| 17 | 0.998 | 0.983 | 1.000 | 1.000 |
| 18 | 0.999 | 0.990 | 1.000 | 1.000 |
| 19 | 1.000 | 0.995 | 1.000 | 1.000 |
| 20 | 1.000 | 0.997 | 1.000 | 1.000 |
| 21 | 1.000 | 0.998 | 1.000 | 1.000 |
| 22 | 1.000 | 0.999 | 1.000 | 1.000 |
| 23 | 1.000 | 0.999 | 1.000 | 1.000 |
| 24 | 1.000 | 1.000 | 1.000 | 1.000 |
| 25 | 1.000 | 1.000 | 1.000 | 1.000 |

Table 10.11. Assessment model estimates of full-selection ("full") and average ("ave") fishing mortality for female and male Alaska plaice in the BSAI.

|  | female |  |  | male |  |  |  | female |  |  | male |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | full |  | ave | full |  | ave |  | full |  | ave | full |  | ave |
| 1975 |  | 0.013 | 0.009 |  | 0.013 | 0.009 | 2000 |  | 0.039 | 0.028 |  | 0.039 | 0.026 |
| 1976 |  | 0.016 | 0.012 |  | 0.016 | 0.011 | 2001 |  | 0.023 | 0.017 |  | 0.023 | 0.016 |
| 1977 |  | 0.010 | 0.007 |  | 0.010 | 0.007 | 2002 |  | 0.032 | 0.023 |  | 0.032 | 0.022 |
| 1978 |  | 0.035 | 0.025 |  | 0.035 | 0.023 | 2003 |  | 0.027 | 0.019 |  | 0.027 | 0.018 |
| 1979 |  | 0.040 | 0.029 |  | 0.040 | 0.027 | 2004 |  | 0.021 | 0.015 |  | 0.021 | 0.014 |
| 1980 |  | 0.019 | 0.013 |  | 0.019 | 0.013 | 2005 |  | 0.030 | 0.022 |  | 0.030 | 0.020 |
| 1981 |  | 0.021 | 0.015 |  | 0.021 | 0.014 | 2006 |  | 0.048 | 0.034 |  | 0.048 | 0.032 |
| 1982 |  | 0.016 | 0.011 |  | 0.016 | 0.011 | 2007 |  | 0.056 | 0.040 |  | 0.056 | 0.038 |
| 1983 |  | 0.023 | 0.017 |  | 0.023 | 0.016 | 2008 |  | 0.051 | 0.036 |  | 0.051 | 0.034 |
| 1984 |  | 0.039 | 0.028 |  | 0.039 | 0.027 | 2009 |  | 0.041 | 0.029 |  | 0.041 | 0.028 |
| 1985 |  | 0.051 | 0.036 |  | 0.051 | 0.034 | 2010 |  | 0.047 | 0.034 |  | 0.047 | 0.032 |
| 1986 |  | 0.097 | 0.069 |  | 0.097 | 0.066 | 2011 |  | 0.069 | 0.049 |  | 0.069 | 0.046 |
| 1987 |  | 0.039 | 0.028 |  | 0.039 | 0.027 | 2012 |  | 0.048 | 0.035 |  | 0.048 | 0.033 |
| 1988 |  | 0.137 | 0.098 |  | 0.137 | 0.093 | 2013 |  | 0.069 | 0.049 |  | 0.069 | 0.047 |
| 1989 |  | 0.033 | 0.024 |  | 0.033 | 0.022 | 2014 |  | 0.059 | 0.042 |  | 0.059 | 0.040 |
| 1990 |  | 0.026 | 0.018 |  | 0.026 | 0.017 | 2015 |  | 0.046 | 0.032 |  | 0.046 | 0.031 |
| 1991 |  | 0.036 | 0.025 |  | 0.036 | 0.024 | 2016 |  | 0.043 | 0.031 |  | 0.043 | 0.029 |
| 1992 |  | 0.044 | 0.031 |  | 0.044 | 0.030 | 2017 |  | 0.055 | 0.039 |  | 0.055 | 0.037 |
| 1993 |  | 0.034 | 0.025 |  | 0.034 | 0.023 | 2018 |  | 0.083 | 0.059 |  | 0.083 | 0.056 |
| 1994 |  | 0.027 | 0.020 |  | 0.027 | 0.019 | 2019 |  | 0.061 | 0.044 |  | 0.061 | 0.041 |
| 1995 |  | 0.050 | 0.035 |  | 0.050 | 0.034 | 2020 |  | 0.079 | 0.057 |  | 0.079 | 0.054 |
| 1996 |  | 0.042 | 0.030 |  | 0.042 | 0.029 | 2021 |  | 0.069 | 0.049 |  | 0.069 | 0.047 |
| 1997 |  | 0.057 | 0.041 |  | 0.057 | 0.039 |  |  |  |  |  |  |  |
| 1998 |  | 0.039 | 0.028 |  | 0.039 | 0.026 |  |  |  |  |  |  |  |
| 1999 |  | 0.038 | 0.027 |  | 0.038 | 0.026 |  |  |  |  |  |  |  |

Table 10.12a. Estimated female numbers at age (millions) from the stock assessment model, 1975-1999.

| female | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 75 | 138.6 | 115.4 | 127.4 | 17 | 152 | 99.0 | 25.1 | 23.3 | 14.2 | 11.1 | 10.1 | 8.6 | 7.7 | 6.9 | 6. | 5.2 | 4.7 | 4.3 | 3.9 | 3.6 | 3.3 | 3.1 | 8.4 |
| 1976 | 13 | 12 | 10 | 11 | 15 | 13 | 86.6 |  | 20.3 |  | 9.6 | 8.7 | 7.5 | 6.7 | 6.0 | 5.3 | 4.5 | 4.1 | 3.7 | 3.4 | 3.1 | 2.9 | 9.9 |
| 1977 | 257.0 | 119.5 | 106.8 | 8.9 | 98.1 | 134.0 | 116.8 | 75.5 | 19 | 17.6 | 10.6 | 8.3 | 7.6 | 6.5 | 5.8 | 5.2 | 4.6 | . 9 | . 5 | 3.2 | 2.9 | . 7 | 11.0 |
| 1978 | 155.0 | 225.7 | 104.9 | 3.8 | 78.0 | 86.0 | 117.4 | 102.1 | 65.8 | 16.6 | 15.3 | 9.3 | 7.2 | 6.6 | 5.6 | 5.0 | 4.5 | 4.0 | 3.4 | 3. | 2.8 | 2.6 | 11.9 |
| 1979 | 140. | 136. | 19 | 92 | 82.1 | 68.1 | 74.7 | 10 | 87.6 | 56.2 | 14.1 | 13.0 | 7.9 | 6.1 | 5.6 | 4.8 | 4.3 | 3.8 | 3.4 | 2.9 | 2.6 | 2.3 | 12.3 |
| 1980 | 145 | 12 | 11 | 17 | 80.5 | 71.7 | 59.1 | 64.4 | 86.6 | 74.4 | 47.5 | 11.9 | 11.0 | 6.6 | 5.2 | 4.7 | 4.0 | 6 | 3. 2 | . 9 | 2. | 2.2 | 12.3 |
| 1981 | 100 | 128 | 10 | 10 | 15 | 70.5 |  |  |  | 74.9 |  | 41.0 |  | 9.5 | 5.7 | 4.5 | 4.1 | 3.5 | 3.1 | . 8 | 2.5 | . 1 | 12.5 |
| 1982 | 108.4 | 88.6 | 112 | 94.9 | 91.9 | 133.3 | 61.5 |  |  | . 2 | 64.5 | 55.3 |  | 8.8 | 8.1 | 4.9 | 3.8 | 3.5 | . 0 | 2.7 | 2. | 2.1 | 12.6 |
| 1983 | 117. | 95 | 77.8 | 98.7 | 83.3 | . 5 | 116.5 | 53.6 | 47 | . 6 | 41.7 | 55.8 | 47.8 | 30.5 | 7.6 | 7.0 | 4.2 | 3.3 | 3.0 | 2.6 | 2.3 | 2.1 | 12.7 |
| 19 | 135 | 103 | 83.5 | 68 | 86 | 72. | 70.2 | 101. | 46. | 40.7 | 33.2 | 35.8 | 47.9 | 41.0 | 26.2 | 6.6 | 6.0 | 3.6 | 2.9 | 2.6 | 2.2 | 2.0 | 12. |
| 1985 | 60.5 | 118.9 | 90 | 73 | 59 | 75. | 63. | 60.5 | 86 | 39 | 34 | 28. | 30.2 | 40.4 | 34.6 | 22. | 5.5 | 5.1 | 3.1 | 2.4 | 2.2 | . 9 | 12.4 |
| 1986 | 66. | 53.1 | 104.3 | 79.2 | 64.1 | 52.0 | 65.3 | 54.2 | 51.3 | 72.9 | 33 | 28 | 23.5 | 25.3 | 33.8 | 28 | 18.4 | 4.6 | 4.2 | 2.6 | 2.0 | 1.8 | 11.9 |
| 1987 | 115.0 | 58.5 | 46.6 | 91.3 | 69.0 | 55.4 | 44.4 | 54.7 | 44.6 | 41.6 | 58.6 | 26.4 | 23.0 | 18.7 | 20.1 | 26.9 | 23 | 14.7 | 3.7 | 3.4 | 2.0 | 1.6 | 10.9 |
| 1988 | 70 | 10 |  | 40.8 | 79.9 | 60.2 | 48.1 | 38.3 | 46.8 | 37.9 |  | 49.5 | 22.3 | 19.5 |  | 17.0 | 22.7 | , 4 | 12.4 | 3.1 | 2.9 | 1.7 | 10.6 |
| 1989 | 93. |  | 88.5 | 44.8 | 35.5 | 68.7 | 50.8 | 39.5 | 30 | 36.7 | 29.4 | 27.1 | 38. | 17.1 | 14.9 | 12.1 | 13 | 17.4 | 14 | 9.5 | 2.4 | 2. | 9.4 |
| 1 | 14 | 82.1 |  | 77.6 | 39.3 | 31.0 | 59.7 | 3.9 | 3 | 26. | 31.3 | 25 | 23 | 32.3 | 14.5 | 12.7 | 10.3 | 11.1 | 14.8 | 12.6 | 8.1 | 2.0 | 9.8 |
| 1991 | 84.4 | 127.3 | 72. | 47 | 68.0 | 34. | 27. | 51 | 37 | 29.2 | 22 | 26 | 21 | 19.7 | 27.6 | 12 | 10.8 | 8.8 | 9.5 | 12.6 | 10.8 | 6.9 | 10 |
| 1992 | 129.6 | 74.1 | 111.7 | 63.2 | 41.5 | 59. | 29.9 | 23 | 44 | 32. | 24 | 19 | 22.7 | 18.1 | 16.7 | 23.4 | 10.6 | 9.2 | 7.5 | 8.0 | 10.7 | 9.2 | 14 |
| 1993 | 105. | 113. | 65.0 | 97.9 | 55.3 | 36.2 | 51.5 | 25.7 | 19.9 | 37.6 | 27.3 | 20 | 16 | 19.1 | 15.2 | 14 | 19.7 | 8.9 | 7.7 | 6.3 | 6.7 | 9.0 | 19. |
| 1994 | 147. | 92. |  |  | 85. | 48.3 |  |  | 22.0 |  |  | 23.2 | 17.7 | 13.6 | 16.2 | 12 | 11.9 | 16.7 | 7.5 | 6.5 | 5.3 | 5.7 | 24 |
| 1995 | 112.4 | 129. | 80. | 87.6 | 50.0 | 75.0 | 42.1 | 27.3 | 38. | 18.9 | 14.5 | 27 | 19. | 15.2 | 11.6 | 13.8 | 11.0 | 10.2 | 14.3 | 6.4 | 5.6 | 4.5 | 25 |
| 1996 | 112.9 | 98. | 113.2 | 70. | 76 | 43.6 | 4.9 | 36.1 | 23 | 32.3 | 15.9 | 12.2 | 22 | 16.6 | 12.7 | 9.7 | 11.6 | 9.2 | 8.5 | 11.9 | 5.4 | 4.7 | 25 |
| 1997 | 60.3 | 99.1 | 86.6 | 99.3 | 62.0 | 66.8 | 37.8 | 55.8 | 30.8 | 19.7 | 27.3 | 13. | 10.2 | 19.3 | 13.9 | 10.7 | 8.2 | 9.7 | 7.8 | 7.2 | 10.0 | 4.5 | 25. |
| 1998 | 69.4 | 52.9 | 86.9 | 75.9 | 86.8 | 54.0 | 57.7 | 32.3 | 47.2 | 25.8 | 16.4 | 22.7 | 11.1 | 8.5 | 16.0 | 11.6 | 8.8 | 6.8 | 8.1 | 6.4 | 5.9 | 8.3 | 24.7 |
| 1999 | 70.4 | 60.9 | 46.4 | 76.2 | 66.4 | 75.7 | 46.9 | 49.7 | 27.6 | 40.1 | 21.9 | 13.9 | 19.2 | 9.4 | 7.2 | 13.5 | 9.8 | 7.5 | 5.7 | 6.8 | 5.4 | 5.0 | 27.9 |

Table 10.12b. Estimated female numbers at age (millions) from the stock assessment model, 2000-2021.

| female | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 00 | 81. | 61. | 53.5 | 40 | 66.7 | 58.0 | 65.8 | 40.4 | 42.6 | 23.5 | 34.1 | 18.5 | 11 | 16.2 | 7.9 | . 1 | 11.4 | 8.3 | 6.3 | 4.8 | 5.8 | 6 | 27.8 |
| 001 | 93.0 | 1.2 | 4.3 | 46.9 | 5.7 | 58.3 | 50.3 | 56.7 | 34 | 36.2 | 19.9 | 28.8 | 15.7 | 9.9 | 13.7 | 6.7 | 5.1 | 9.6 | 7.0 | 5.3 | 4.1 | 9 | 27.4 |
| 2002 | 95.3 | 81.6 | 62.5 | 47.6 | 1.1 | 31.2 | 50.8 | 43.7 | 49.0 | 29.8 | 31.1 | 17.1 | 24.7 | 13.4 | 8.5 | 11.8 | 5.8 | 4.4 | 8.3 | 6.0 | 4.6 | 3.5 | 27.7 |
| 003 | 97.5 | 83.7 | 71.6 | 54.8 | 1.7 | 35.9 | 27 | 43.9 | 37.6 | 1.9 | 25.4 | 26.5 | 14 | 21.0 | 11.4 | 7.2 | 10.0 | . 9 | . 7 | 7.0 | 5.1 | 9 | . 5 |
| 2004 | 161.2 | 85.6 | 73.4 | 62.8 | 8.0 | 36.5 | 31.3 | 23.5 | 37 | 3 | 35.9 | 21.7 | 22.7 | 12.4 | 18.0 | 9.8 | 6.2 | 8.5 | 4.2 | 3.2 | 6.0 | 4 | 26.0 |
| 2005 | 187.0 | 141.5 | 75.1 | 64.4 | 55.1 | 42.0 | 31.8 | 27.2 | 20 | 32.7 | 7 8 | 30.9 | . 7 | 19.5 | . 7 | 15.5 | 8.4 | 5.3 | 7.3 | 3.6 | 2.8 | 5.2 | 26.1 |
| 2006 | 73.6 | 164.2 | 124.2 | 65.9 | 6.5 | 48.1 | 36.6 | 27.5 | 23 | 17.4 | 7.9 | 23.7 | 26.3 | 15.9 | 16.6 | . 1 | 13.2 | 7.2 | 4.5 | 6.3 | 3.1 | 2.3 | 26.6 |
| 2007 | 98.1 | 64.6 | 144.0 | 108.9 | 57.6 | 49.2 | 41.7 | 31. | 23.4 | 19.7 | 14.7 | 23.4 | 19.9 | 22.1 | 13.3 | 13.9 | 7.6 | 11.0 | 6.0 | 3.8 | 5.2 | 2. | 24 |
| 2008 | 91.0 | 86.1 | 56.7 | 126.2 | 95.2 | 50.2 | 42.5 | 35. | 26.6 | 19.6 | 16.5 | 12.2 | 19.5 | 16.5 | 18.3 | 11.1 | 11.6 | 6.3 | 9.2 | 5.0 | 3.1 | 4 | 22 |
| 2009 | 44.8 | 79.9 | 75.5 | 49.7 | 110.4 | 82. | 43 | 36 | 30.2 | 22. | 16. | 13 | 10.2 | 16.3 | 13.8 | 15.3 | 9.3 | 9.6 | 5.3 | 7.6 | 4. | 2.6 | 22.2 |
| 2010 | 57.9 | 39.3 | 70.1 | 6.2 | 43.5 | 6.3 | 72.0 | 37.4 | 31. | 25. | 8.9 | 13.9 | 11 | 8.6 | 13.7 | 11.6 | 12.9 | 7.8 | 8.1 | 4.5 | 6.4 | 3.5 | 21.0 |
| 2011 | 38.1 | 50.8 | 34.5 | 61.4 | 58.0 | 37.9 | 83.4 | 61.8 | 31.8 | 26. | 21.6 | 15.9 | 11 | 9.7 | 7.2 | 11.5 | 9.7 | 10.8 | 6.5 | 6.8 | 3.7 | 5.4 | 20.5 |
| 2012 | 94.2 | 33.5 | 44.6 | 30.2 | 53.7 | 50. | 32 | 70 | 51.8 | 26.4 | 21.7 | 17.8 | 13.0 | 9.6 | 8.0 | 5.9 | 9.4 | 8.0 | 8.9 | 5.4 | 5.6 | 3.1 | 21.2 |
| 2013 | 52.5 | 82.7 | 29.4 | 39.1 | 26.4 | 46.8 | 43.6 | 28.0 | 60.2 | 43.7 | 22.2 | 18.2 | 14.9 | 10.9 | 8.0 | 6.7 | 4.9 | 7.9 | 6.7 | 7.4 | 4.5 | 4.7 | 20 |
| 2014 | 39.7 | 46. | 72.6 | 25. | 34. | 23.0 | 40.2 | 37. | 23.5 | 50.0 | 36.0 | 18.2 | 14.9 | 12.2 | 8.9 | 6.6 | 5.5 | 4.1 | 6.5 | 5.5 | 6. | 3.7 | 20.5 |
| 2015 | 62.3 | 34.8 | 40.4 | 63. | 22.5 | 29.7 | 19.8 | 34. | 31.2 | 19.6 | 41.6 | 29.9 | 15.1 | 12.4 | 10.1 | 7.4 | 5.4 | 4.5 | 3.4 | 5.3 | 4. | 5.0 | 20.0 |
| 2016 | 57.1 | 54.7 | 30.5 | 35.4 | 55.6 | 9.6 | 25.7 | 17 | 29. | 26. | 16.5 | 35.0 | 25 | 12 | 10.4 | 8.5 | 6.2 | 4.6 | 3.8 | 2.8 | 4. | 3.8 | 21.0 |
| 2017 | 196.0 | 50.1 | 8.0 | 26. | 31.0 | 8.5 | 17.0 | 22 | 1 | 4.8 | 22.3 | 13.9 | 29.4 | 21 | 10.7 | 8.7 | 7.1 | 5.2 | 3.8 | 3.2 | 2. | 3.8 | 20.8 |
| 2018 | 140.5 | 172.0 | 44.0 | . 1 | 23.4 | 27.0 | 41.9 | 14.5 | 18.7 | 12.2 | 20.7 | 18.6 | 11.6 | 24.5 | 17.6 | 8.9 | 7.2 | 5.9 | 4.3 | 3.2 | 2.7 | 2.0 | 20.5 |
| 2019 | 205.1 | 123.3 | 150.9 | 38.5 | 36.7 | 20.3 | 23.1 | 35.4 | 12.1 | 15.4 | 9.9 | 16.8 | 15.0 | 9.4 | 19.8 | 14.2 | 7.2 | 5.9 | 4.8 | 3.5 | 2.6 | 2.2 | 18.1 |
| 2020 | 107.4 | 180.0 | 108.2 | 132.2 | 33.6 | 31.9 | 17.5 | 19.7 | 29.8 | 10.1 | 12.8 | 8.2 | 13.9 | 12.4 | 7.7 | 16.3 | 11.7 | 5.9 | 4.8 | 4.0 | 2.9 | 2.1 | 16.8 |
| 2021 | 206.4 | 94.3 | 157.8 | 94.7 | 115.4 | 29.2 | 27.4 | 14.8 | 16.4 | 24.5 | 8.2 | 10.4 | 6.7 | 11.3 | 10.1 | 6.3 | 13.3 | 9.5 | 4.8 | 3.9 | 3.2 | 2.4 | 15.3 |

Table 10.13a. Estimated male numbers at age (millions) from the stock assessment model, 1975-1999.

| m | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1975 | 138.6 | 115.4 | 127 | 174.5 | 152.5 | 9.0 | 25.1 | . 3 | 14.2 | 11 | 10 | 8.6 | 7.7 | 6. | 6.1 | 5.2 | 4.7 | 4.3 | 3.9 | 3.6 | 3.3 | 3.1 | 8.4 |
| 1976 | 136.1 | 121.7 | 101 | 111.8 | 153.0 | 133.7 | 86.6 | 2.0 | 20.3 | 12.3 | 9.7 | 8.8 | 7.5 | 6.7 | 6.0 | 5.3 |  | 4.1 | 3.7 | 3.4 | 3.1 | 2.9 | 9.9 |
| 1977 | 257.0 | 119.5 | 106.8 | 8.9 | 98.0 | 134.0 | 116.9 | 75.6 | 19.1 | 17.7 | 10.7 | 8.4 | 7.6 | 6.5 | 5.8 | 5.2 | 4.6 | 3.9 | 3.5 | 3.2 | 2.9 | 2.7 | 11.0 |
| 1978 | 155.0 | 225.7 | 104.9 | 93.7 | 8.0 | 85.9 | 117.4 | 102.3 | 6.1 | 16 | 15.4 | 9.3 | 7.3 | 6.6 | 5.6 | 5.0 | 4.5 | 4.0 | 3.4 | 3.1 | 2.8 | 2.6 | 11.9 |
| 197 |  | 136 | 197.9 | 91.9 |  |  |  |  |  |  |  | 13.1 | 7.9 | 6.2 |  |  |  | 3.8 |  | 2.9 | 2.6 | 2.3 | 12.3 |
| 1980 | 145 | 123 | 119.3 | 173.4 | 80.4 |  | 59.2 |  |  |  |  |  | 11.1 | 6.7 | 5.2 | 4.7 | 4.0 | 3.6 | 3.2 | 2.9 | 2.4 | 2.2 | 12.3 |
| 1981 | 100.9 | 128.0 | 108. | 104 | 152.0 | 70 | 62.6 | 51.6 | 56.3 | 75 | 65.3 | 41.7 | 10.4 | 9.6 | 5.8 | 4.5 | 4.1 | 3.5 | 3.1 | 2.8 | 2.5 | 2.1 | 2.5 |
| 1982 | 108.4 | 88.5 | 112.3 | 94.8 | 1.7 | 133.0 | 61. | 4.5 | 44.8 | 48.7 | 65 | 56.3 | 35.9 | 9.0 | 8.2 | 5.0 | 3.9 | 3.5 | 3.0 | 2.7 | 2.4 | 2.1 | 12.6 |
| 1983 | 117.3 | 95.1 | 77 | 98.6 | 83.1 | 80.3 | 116.4 | 53.7 | 47.5 | 39 | 42 | 56.8 | 48.8 | 31.1 | 7.8 | 7.1 | 4.3 | 3.3 | 3. | 2. | 2.3 | 2. | 12. |
| 198 | 135.4 | 102.9 | 83. | 68.2 | 86.4 | 72 | 70. | 101.3 | 46.6 | 41 | 33.6 | 36 | 48.9 | 41.9 | 26.7 | 6.7 | 6.1 | 3.7 | 2.9 | 2.6 | 2.2 | 2.0 | 12. |
| 1985 | 60.5 | 118.8 | 90.3 | 73.1 | 59.6 | 75.4 | 63.2 | 60.7 | 87.2 | 39 | 35.0 | 28.5 | 30.8 | 41.3 | 35.4 | 22 | 5.6 | 5.2 | 3.1 | 2.4 | 2.2 | 1.9 | 12.4 |
| 1986 | 66. |  | 104.2 | 79.0 | 63.9 | 51.9 | 65.3 | 54.5 | 51.9 | 74.1 |  | 29.4 | 23.9 | 25.8 | 34.5 | 29 | 18.8 | 4.7 | 4.3 | 2.6 | 2.0 | 1.8 | 11.9 |
| 1987 | 115.0 | 58 | 46.5 | 91. | 68.8 | 55.3 | 44.5 | 55.3 | 45.5 |  | 60.4 | 27.2 | 23.6 | 19.2 | 20.6 |  | 23.6 | 15.0 | 3.8 | 3. | 2.1 | 1.6 | 0.9 |
| 1988 | 70.2 | 100.9 | 51.2 | 40.7 | 79.6 | 60.0 | 48.1 | 8.5 | . 6 | 39.0 | 36.5 |  | 23.1 | 20.0 | .2 |  | 23 | 19.9 | 12.7 | 3.2 | 2.9 | . 7 | 10.6 |
| 1989 | 93.6 | 61.5 | 88.2 | 44.6 | 35. | 68. | 50.9 | 40. | 31. | 38 | 30 | 28 | 39.7 | 17.8 | 15 | 12 | 13 | 17.8 | 15.2 | 9.7 | 2. | 2.2 | 9.4 |
| 1990 | 145.0 | 82. | 54. | 77.3 | 9. | 30.8 | 59. | 44. | 4.6 | 27 | 32 | 26 | 24.3 | 33.8 | 15. | 13 | 10.6 | 11.3 | 15.2 | 13.0 | 8.2 | 2. | 9.9 |
| 199 | 84.4 | 127.3 | 72.0 | 47.3 | 67.7 | 34.2 | 26.9 | 1.8 | 38.3 | 29.9 | 23 | 28 | 22.6 | 20.8 | 28.9 | 12 | 11.2 | 9.0 | 9.7 | 13.0 | 11. | 7. | 10.2 |
| 1992 | 129. | 74.1 | 111 | 63.1 | 41.4 | 59.2 | 29.7 | 23. | . 7 | 32.8 | 25.5 | 19.9 | 23 | 19.1 | 17.6 | 24 | 11 | 9.5 | 7. | 8.2 | 11. | 9. | 14 |
| 1993 | 10 | 113 |  | 97.8 | 55.2 | 36.1 |  | 25.7 | 20.0 |  | 27.9 | 21.6 | 16. | 20. | 16.1 | 14 | 20 | 9.2 | 8.0 | 6.4 | 6.9 | 9.2 | 0. |
| 1994 | 14 | 92.1 | 99.7 | 56.9 |  | 48.2 | 31.4 |  | 22.2 | 17.2 |  | 23.8 | 18 | 14.2 | 17.1 |  | 12 | 17.5 | 7.8 | 6.8 | 5.5 | 5.9 | 25. |
| 1995 | 112.4 | 129.0 | 80.8 | 87.5 | . 9 | 74.8 | 42.0 | 27.3 | 38.6 | 19.1 | 14 | 28. | 20.4 | 15.7 | 12.2 | 14.6 | 11.7 | 10.8 | 15.0 | 6.7 | 5.8 | 4.7 | 26 |
| 1996 | 112.9 | 98.6 | 113.1 | 70.8 | 76.4 | 43.4 | 64.9 | 36.2 | 23.4 | 32.8 | 16.2 | 12.4 | 23.5 | 17.1 | 13.2 | 10.2 | 12.2 | 9.8 | 9.0 | 12.5 | 5.6 | 4.8 | 25.9 |
| 1997 | 60.3 | 99.1 | 86. | 99.1 | 61.9 | 66.7 | 37.7 | 56.1 | 31.1 | 20.0 | 27.9 | 13.7 | 10.5 | 19.8 | 14.4 | 11.1 | 8.6 | 10.3 | 8.2 | 7.6 | 10.5 | 4.7 | 25.9 |
| 1998 | 69.4 | 52.9 | 86. | 75.7 | 86.5 | 53.9 | 57.7 | 32.5 | 47.8 | 26.4 | 16.8 | 23 | 11.4 | 8.7 | 16.4 | 11.9 | 9.2 | 7.1 | 8.5 | 6.8 | 6.3 | 8.7 | 25.4 |
| 1999 | 70. | 60.9 | 46. | 76. | 66.2 | 75.5 | 46.8 | 50.0 | 27.9 | 41.0 | 22.5 | 14.3 | 19.8 | 9.7 | 7.4 | 13.9 | 10.1 | 7.8 | 6.0 | 7.2 | 5.8 | 5.3 | 28.8 |

Table 10.13b. Estimated male numbers at age (millions) from the stock assessment model, 2000-2021.

| male | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 | 81.1 | 61.8 | 53.4 | 40.6 | 66.6 | 57.8 | 65.7 | 40.5 | 43.0 | 23.9 | 34.9 | 19.1 | 12.1 | 16.7 | 8.2 | 6.2 | 11.8 | 8.5 | 6.6 | 5.1 | 6.1 | 4.9 | 28.8 |
| 2001 | 93.0 | 71.2 | 54.2 | 46.8 | 35.6 | 58.1 | 50.3 | 56.8 | 34.9 | 36.8 | 20.4 | 29.7 | 16.2 | 10.2 | 14.1 | 6.9 | 5.3 | 9.9 | 7.2 | 5.6 | 4.3 | 5.1 | 28.5 |
| 2002 | 95.3 | 81.6 | 62.4 | 47.5 | 41.0 | 31.1 | 50.7 | 43.8 | 49.3 | 30.2 | 31.8 | 17.6 | 25.5 | 13.9 | 8.8 | 12.1 | 5.9 | 4.5 | 8.5 | 6.2 | 4.8 | 3.7 | 28.8 |
| 2003 | 97.5 | 83.6 | 71.6 | 54.7 | 41.6 | 35.8 | 27.1 | 44.0 | 37.8 | 42.4 | 25.9 | 27.1 | 15.0 | 21.7 | 11.8 | 7.5 | 10.3 | 5.0 | 3.9 | 7.2 | 5.3 | 4.0 | 27.6 |
| 2004 | 161.2 | 85.6 | 73.4 | 62.8 | 47.9 | 36.4 | 31.2 | 23.6 | 38.1 | 32.6 | 36.5 | 22.2 | 23.3 | 12.8 | 18.6 | 10.1 | 6.4 | 8.8 | 4.3 | 3.3 | 6.2 | 4.5 | 27.1 |
| 2005 | 187.0 | 141.5 | 75.1 | 64.4 | 55.0 | 42.0 | 31.8 | 27.2 | 20.5 | 33.0 | 28.2 | 31.5 | 19.1 | 20.0 | 11.0 | 16.0 | 8.7 | 5.5 | 7.6 | 3.7 | 2.8 | 5.3 | 27.2 |
| 2006 | 73.6 | 164.1 | 124.1 | 65.8 | 56.4 | 48.1 | 36.6 | 27.6 | 23.5 | 17.6 | 28.3 | 24.1 | 26.9 | 16.3 | 17.1 | 9.4 | 13.6 | 7.4 | 4.7 | 6.5 | 3.2 | 2.4 | 27.7 |
| 2007 | 98.1 | 64.6 | 143.9 | 108.7 | 57.5 | 49.1 | 41.7 | 31.5 | 23.6 | 20.0 | 14.9 | 23.9 | 20.3 | 22.6 | 13.7 | 14.3 | 7.9 | 11.4 | 6.2 | 3.9 | 5.4 | 2.6 | 25.2 |
| 2008 | 91.0 | 86.1 | 56.6 | 125.9 | 94.9 | 50.1 | 42.5 | 35.9 | 26.9 | 20.0 | 16.9 | 12.5 | 19.9 | 16.9 | 18.8 | 11.4 | 11.9 | 6.5 | 9.5 | 5.2 | 3.3 | 4.5 | 23.1 |
| 2009 | 44.8 | 79.8 | 75.4 | 49.5 | 110.1 | 82.7 | 43.4 | 36.7 | 30.7 | 22.9 | 16.9 | 14.2 | 10.5 | 16.7 | 14.1 | 15.7 | 9.5 | 9.9 | 5.5 | 7.9 | 4.3 | 2.7 | 23.1 |
| 2010 | 57.9 | 39.3 | 70.0 | 66.1 | 43.3 | 96.0 | 71.9 | 37.5 | 31.5 | 26.2 | 19.5 | 14.3 | 12.0 | 8.8 | 14.1 | 11.9 | 13.2 | 8.0 | 8.4 | 4.6 | 6.7 | 3.6 | 21.7 |
| 2011 | 38.1 | 50.8 | 34.5 | 61.3 | 57.8 | 37.8 | 83.3 | 62.0 | 32.2 | 26.8 | 22.2 | 16.4 | 12.1 | 10.1 | 7.4 | 11.8 | 10.0 | 11.1 | 6.7 | 7.0 | 3.9 | 5.6 | 21.2 |
| 2012 | 94.2 | 33.4 | 44.5 | 30.2 | 53.5 | 50.2 | 32.6 | 71.3 | 52.6 | 27.0 | 22.4 | 18.4 | 13.5 | 9.9 | 8.3 | 6.1 | 9.7 | 8.2 | 9.1 | 5.5 | 5.7 | 3.2 | 22.0 |
| 2013 | 52.5 | 82.7 | 29.3 | 39.0 | 26.4 | 46.6 | 43.5 | 28.1 | 61.1 | 44.7 | 22.9 | 18.8 | 15.5 | 11.4 | 8.3 | 6.9 | 5.1 | 8.1 | 6.8 | 7.6 | 4.6 | 4.8 | 21.0 |
| 2014 | 39.7 | 46.0 | 72.4 | 25.6 | 34.0 | 22.9 | 40.2 | 37.3 | 23.8 | 51.3 | 37.2 | 18.9 | 15.5 | 12.7 | 9.3 | 6.8 | 5.7 | 4.2 | 6.6 | 5.6 | 6.2 | 3.8 | 21.2 |
| 2015 | 62.3 | 34.8 | 40.3 | 63.4 | 22.4 | 29.6 | 19.8 | 34.6 | 31.7 | 20.1 | 43.0 | 31.1 | 15.7 | 12.9 | 10.5 | 7.7 | 5.7 | 4.7 | 3.5 | 5.5 | 4.6 | 5.2 | 20.7 |
| 2016 | 57.1 | 54.7 | 30.5 | 35.3 | 55.4 | 19.5 | 25.7 | 17.1 | 29.6 | 27.1 | 17.1 | 36.4 | 26.2 | 13.2 | 10.8 | 8.9 | 6.5 | 4.7 | 3.9 | 2.9 | 4.6 | 3.9 | 21.7 |
| 2017 | 196.0 | 50.1 | 48.0 | 26.7 | 30.9 | 48.3 | 17.0 | 22.2 | 14.7 | 25.3 | 23.0 | 14.5 | 30.7 | 22.1 | 11.2 | 9.1 | 7.4 | 5.5 | 4.0 | 3.3 | 2.4 | 3.9 | 21.5 |
| 2018 | 140.5 | 171.9 | 43.9 | 42.0 | 23.3 | 26.9 | 41.9 | 14.6 | 19.0 | 12.4 | 21.3 | 19.3 | 12.1 | 25.6 | 18.4 | 9.3 | 7.6 | 6.2 | 4.5 | 3.3 | 2.8 | 2.0 | 21.1 |
| 2019 | 205.1 | 123.2 | 150.6 | 38.4 | 36.6 | 20.2 | 23.1 | 35.6 | 12.3 | 15.8 | 10.3 | 17.4 | 15.7 | 9.8 | 20.7 | 14.9 | 7.5 | 6.1 | 5.0 | 3.7 | 2.7 | 2.2 | 18.7 |
| 2020 | 107.4 | 179.9 | 108.0 | 131.8 | 33.5 | 31.8 | 17.5 | 19.9 | 30.3 | 10.4 | 13.2 | 8.5 | 14.5 | 13.0 | 8.1 | 17.1 | 12.3 | 6.2 | 5.1 | 4.1 | 3.0 | 2.2 | 17.3 |
| 2021 | 206.4 | 94.2 | 157.5 | 94.4 | 114.8 | 29.0 | 27.4 | 14.9 | 16.7 | 25.3 | 8.6 | 10.8 | 7.0 | 11.8 | 10.6 | 6.6 | 13.9 | 10.0 | 5.0 | 4.1 | 3.4 | 2.5 | 15.8 |

Table 10.14. Projections of spawning biomass ( t ), catch $(\mathrm{t})$, and fishing mortality rate $(\mathrm{F})$ for each of the several scenarios. The values of B40\% and B35\% are 114,635 t and $100,306 \mathrm{t}$, respectively. $\mathrm{SSB}=$ female spawning biomass.

## Scenarios 1 and 2

Maximum ABC harvest permissible

|  | SSB | catch | F |
| :--- | :--- | :--- | :--- |
| 2021 | 143,242 | 16,760 | 0.07 |
| 2022 | 142,024 | 18,584 | 0.08 |
| 2023 | 145,641 | 18,584 | 0.08 |
| 2024 | 151,806 | 34,564 | 0.14 |
| 2025 | 155,036 | 34,893 | 0.14 |
| 2026 | 158,047 | 35,396 | 0.14 |
| 2027 | 159,648 | 35,697 | 0.14 |
| 2028 | 158,761 | 35,567 | 0.14 |
| 2029 | 155,227 | 34,947 | 0.14 |
| 2030 | 150,022 | 33,957 | 0.14 |
| 2031 | 144,309 | 32,807 | 0.14 |
| 2032 | 139,001 | 31,677 | 0.14 |
| 2033 | 134,500 | 30,660 | 0.14 |
| 2034 | 130,828 | 29,754 | 0.14 |

## Scenario 4

Upper bound on ABC is $\mathrm{F} 60 \%$

|  | SSB | catch | F |
| :--- | :--- | :--- | :--- |
| 2021 | 143,242 | 16,760 | 0.07 |
| 2022 | 142,024 | 18,584 | 0.08 |
| 2023 | 145,641 | 18,584 | 0.08 |
| 2024 | 154,109 | 16,666 | 0.07 |
| 2025 | 166,097 | 17,748 | 0.07 |
| 2026 | 177,696 | 18,888 | 0.07 |
| 2027 | 187,562 | 19,901 | 0.07 |
| 2028 | 194,339 | 20,657 | 0.07 |
| 2029 | 197,557 | 21,099 | 0.07 |
| 2030 | 197,941 | 21,256 | 0.07 |
| 2031 | 196,562 | 21,213 | 0.07 |
| 2032 | 194,456 | 21,063 | 0.07 |
| 2033 | 192,213 | 20,870 | 0.07 |
| 2034 | 190,065 | 20,666 | 0.07 |

## Scenario 3

Harvest at average $F$ over the past 5 years

|  | SSB | catch | F |
| :--- | :---: | :---: | :---: |
| 2021 | 143,242 | 16,760 | 0.07 |
| 2022 | 142,024 | 18,584 | 0.08 |
| 2023 | 145,641 | 18,584 | 0.08 |
| 2024 | 154,659 | 12,277 | 0.05 |
| 2025 | 168,844 | 13,243 | 0.05 |
| 2026 | 182,757 | 14,258 | 0.05 |
| 2027 | 195,000 | 15,186 | 0.05 |
| 2028 | 204,132 | 15,925 | 0.05 |
| 2029 | 209,579 | 16,428 | 0.05 |
| 2030 | 211,971 | 16,706 | 0.05 |
| 2031 | 212,320 | 16,819 | 0.05 |
| 2032 | 211,657 | 16,832 | 0.05 |
| 2033 | 210,592 | 16,793 | 0.05 |
| 2034 | 209,397 | 16,728 | 0.05 |

Scenario 5
No fishing

|  | SSB | catch | F |
| :---: | :---: | ---: | :---: |
| 2021 | 143,242 | 16,760 | 0.07 |
| 2022 | 142,024 | 18,584 | 0.08 |
| 2023 | 145,641 | 18,584 | 0.08 |
| 2024 | 156,171 | 0 | 0.00 |
| 2025 | 176,604 | 0 | 0.00 |
| 2026 | 197,432 | 0 | 0.00 |
| 2027 | 217,128 | 0 | 0.00 |
| 2028 | 233,999 | 0 | 0.00 |
| 2029 | 247,148 | 0 | 0.00 |
| 2030 | 256,889 | 0 | 0.00 |
| 2031 | 263,987 | 0 | 0.00 |
| 2032 | 269,385 | 0 | 0.00 |
| 2033 | 273,680 | 0 | 0.00 |
| 2034 | 277,186 | 0 | 0.00 |

Table 10.14 continued. Projections of spawning biomass ( $1,000 \mathrm{~s} t$ ), catch ( $1,000 \mathrm{~s} \mathrm{t}$ ), and fishing mortality rate for each of the several scenarios. The values of B40\% and B35\% are $114,635 \mathrm{t}$ and $100,306 \mathrm{t}$, respectively. $\mathrm{SSB}=$ female spawning biomass.

| Scenario 6 <br> Determination of overfishing |  |  |  | Scenario 7 <br> Determination of whether Alaska plaice are approaching an overfished condition |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SSB | catch | F |  | SSB | catch | F |
| 2021 | 143,242 | 16,760 | 0.07 | 2021 | 143,242 | 16,760 | 0.07 |
| 2022 | 139,227 | 39,305 | 0.17 | 2022 | 140,136 | 32,697 | 0.14 |
| 2023 | 133,128 | 37,157 | 0.17 | 2023 | 137,071 | 31,612 | 0.14 |
| 2024 | 132,757 | 36,557 | 0.17 | 2024 | 138,338 | 38,104 | 0.17 |
| 2025 | 135,463 | 36,836 | 0.17 | 2025 | 140,190 | 38,146 | 0.17 |
| 2026 | 138,062 | 37,345 | 0.17 | 2026 | 141,998 | 38,442 | 0.17 |
| 2027 | 139,362 | 37,628 | 0.17 | 2027 | 142,574 | 38,532 | 0.17 |
| 2028 | 138,287 | 37,413 | 0.17 | 2028 | 140,862 | 38,145 | 0.17 |
| 2029 | 134,714 | 36,636 | 0.17 | 2029 | 136,747 | 37,221 | 0.17 |
| 2030 | 129,664 | 35,452 | 0.17 | 2030 | 131,249 | 35,918 | 0.17 |
| 2031 | 124,318 | 34,026 | 0.17 | 2031 | 125,536 | 34,420 | 0.17 |
| 2032 | 119,595 | 32,465 | 0.17 | 2032 | 120,506 | 32,806 | 0.17 |
| 2033 | 115,905 | 31,052 | 0.17 | 2033 | 116,564 | 31,322 | 0.17 |
| 2034 | 113,196 | 29,912 | 0.16 | 2034 | 113,659 | 30,113 | 0.16 |

Table 10.15. Non-target species catch (t) when Alaska plaice were the fishery target, 2017-2021. The 2021 data are through October 17, 2021.

| species/group | 2017 | 2018 | 2019 | 2020 | 2021 |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Benthic urochordata | 0.30 | 1.8 | 0.34 | 0.05 | 26.2 |
| Birds - Northern Fulmar | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Bivalves | 0.00 | 0.02 | 0.06 | 0.00 | 0.06 |
| Brittle star unidentified | 0.00 | 0.01 | 0.02 | 0.00 | 0.09 |
| Capelin | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Corals and bryozoans | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |
| Eelpouts | 0.02 | 0.54 | 0.00 | 0.00 | 0.08 |
| Eulachon | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Greenlings | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Hermit crab unidentified | 0.00 | 0.04 | 0.01 | 0.00 | 0.12 |
| Invertebrate unidentified | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 |
| Misc crabs | 0.06 | 0.43 | 0.09 | 0.00 | 0.26 |
| Misc crustaceans | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Misc fish | 0.09 | 0.25 | 0.12 | 0.28 | 2.8 |
| Misc inverts (worms etc) | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Other osmerids | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |
| Pacific Sand lance | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Pacific Sandfish | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Pandalid shrimp | 0.00 | 0.01 | 0.01 | 0.00 | 0.00 |
| Polychaete unidentified | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 |
| Saffron Cod | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |
| Sculpin | 0.00 | 0.00 | 0.00 | 0.00 | 45.5 |
| Scypho jellies | 0.01 | 0.28 | 0.38 | 0.63 | 2.77 |
| Sea anemone unidentified | 0.01 | 0.01 | 0.01 | 0.00 | 0.19 |
| Sea pens whips | 0.00 | 0.06 | 0.00 | 0.00 | 0.00 |
| Sea star | 4.2 | 65.0 | 18.6 | 1.1 | 105.1 |
| Snails | 0.02 | 0.21 | 0.04 | 0.01 | 1.67 |
| Sponge unidentified | 0.00 | 0.07 | 0.14 | 0.00 | 0.02 |
| Stichaeidae | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| urchins dollars cucumbers | 0.00 | 0.03 | 0.00 | 0.00 | 0.02 |
| total | 4.70 | 68.85 | 19.85 | 2.11 | 184.93 |

Figures


Figure 10.1. Catches (t) of Alaska plaice in the Bering Sea and Aleutian Islands, 1975-2021. The 2021 data are shown as of October 17. Data are from the Alaska Regional Office. Red line indicates the recommended 2022 ABC.


Figure 10.2. Estimated survey biomass ( t ) and $95 \%$ confidence intervals from NMFS bottom trawl surveys in the eastern Bering Sea (EBS; 1987-2021) and northern Bering Sea (2010, 2017, 2019, 2021). No survey was conducted in the EBS during 2020 due to the coronavirus pandemic.


Figure 10.3. Eastern and northern Bering Sea survey CPUE (kg/ha) of Alaska plaice in 2010, 2017, 2019, and 2021. Data are courtesy of Lyle Britt, AFSC RACE.


Figure 10.4. Comparison of bottom trawl survey length compositions from the eastern Bering Sea (blue) and northern Bering Sea (yellow) in four different years, both sexes combined.

## M profile for Alaska plaice



Figure 10.5. Stock assessment model fit (in terms of $-\log$ (likelihood)) to a range of male and female natural mortality values.


Figure 10.6. Length-at-age relationships for Alaska plaice used in the assessment model.


Figure 10.7. Weight-at-age relationships for Alaska plaice used in the assessment model.


Figure 10.8. Retrospective plot of female spawning biomass (t) from 2012 to 2021. Mohn's test statistic $\rho=-0.01$.


Figure 10.9. Estimated beginning year total biomass of Alaska plaice from the assessment model. 95\% percent confidence intervals are from MCMC integration.


Figure 10.10. Observed (open circles) and model-predicted (solid red line) survey biomass of Alaska plaice. Error bars indicate $95 \%$ confidence intervals from the design-based survey.


Figure 10.11. Model estimates of Alaska plaice female spawning biomass (t) with estimates of $\mathrm{B}_{35 \%}$ (red line) and $\mathrm{B}_{40 \%}$ (green line). Ninety-five percent credible intervals are from MCMC integration.


Figure 10.12. Estimated recruitment (age 3) for Alaska plaice. 95\% credible intervals are from MCMC integration.


Figure 10.13. Model estimates of survey and fishery selectivity.


Figure 10.14. Survey age composition (solid line = observed, dotted line $=$ predicted, females in red above x axis, males in blue below x axis).


Figure 10.14. Survey age composition (continued).


Figure 10.14. Survey age composition (continued).


Figure 10.15, Survey length composition by year (solid line = observed, dotted line = predicted, females in red above x axis, males in blue below x axis).)


Figure 10.15. Survey length composition (continued).


Figure 10.16. Fishery age composition by year (solid line $=$ observed, dotted line $=$ predicted, females above x axis, males below x axis).


Figure 10.17. Fishery length composition by year (solid line = observed, dotted line = predicted, females above x axis, males below x axis).


Figure 10.17. Fishery length composition (continued).


Figure 10.18. Model estimates of full-selection fishing mortality for Alaska plaice.


Figure 10.19. Posterior distribution of the 2021 estimate of female spawning biomass ( t ) from mcmc integration. $\mathrm{B}_{40 \%}$ and $\mathrm{B}_{35 \%}$ are indicated in the plot.


Figure 10.20. Model projection of Alaska plaice at the harvest rate of the average of the past five years (i.e., Alternative 3) using the estimated 2021 numbers-at-age from the stock assessment model for the starting point.


Figure 10.21. Phase-plane diagram of the relative trajectories of female spawning biomass and full-selection fishing mortality. Horizontal axis contains model-estimated female spawning biomass relative to B35\%; vertical axis contains model-estimated full-selection fishing mortality relative to $\mathrm{F} 35 \%$. Dashed black lines indicate the relevant limit reference points; red line indicates the ABC control rule. Fuchsia-filled square is the current-year (2021) values; yellow fill indicates projected 2022 \& 2023 values.

