8. Assessment of the Flathead Sole Stock in the Gulf of Alaska

By

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Executive Summary

Changes in the Input Data

- 1) The fishery catch and length compositions for 2008 and 2009 (through Sept. 26, 2009) were incorporated in the model.
- 2) The 2007 fishery catch and length compositions were updated.
- 3) The 2009 GOA groundfish survey biomass estimate and length composition data were added to the model. Survey biomass decreased from 280,990 t in 2007 to 225,377 t in 2009. Survey biomass estimates and length compositions were recalculated for all survey years.
- 4) Age compositions from the 1990, 1999, and 2007 groundfish surveys were added to the model.

Changes in the Assessment Model

Estimable scaling offset parameters for male selectivity (relative to asymptotic female selectivity) were incorporated into the assessment model for both fishery and survey selectivities. As a consequence, the fishing mortality experienced by fully-selected males may now differ from that experienced by fully-selected females. Fishing mortality is reported relative to fully-selected females.

Changes in the Assessment Results

- 1. The preferred model configuration incorporates the new option for male selectivity scaling parameters.
- 2. Based on the preferred model, the recommended ABC, based on an $F_{40\%}$ harvest level of 0.371, is 52,721 t for 2010 and 54,865 t for 2011.
- 3. The OFL, based on an $F_{35\%}$ harvest level of 0.481, is 65,567 t for 2010 and 68,206 t for 2011.
- 4. Projected female spawning biomass is estimated at 124,674 t for 2010 and 128,585 t for 2011.
- 5. Total biomass (age 3+) is estimated at 370,332 t for 2010 and 367,217 t for 2011.

The area apportionments corresponding to the recommended ABCs from the preferred model are:

| | Western | Central | West | Southeast | Grand |
|---------------|---------|---------|---------|-----------|--------|
| | Gulf | Gulf | Yakutat | Outside | Total |
| apportionment | 35.5% | 57.2% | 4.2% | 3.1% | 100.0% |
| 2010 ABC (t) | 18,741 | 30,155 | 2,212 | 1,613 | 52,721 |
| 2011 ABC (t) | 19,503 | 31,381 | 2,302 | 1,679 | 54,865 |

A summary of important reference values from the preferred model for this assessment, relative to the 2008 SAFE projections, is as follows:

| Orrentitu | 2009 Assessment | 2009 Assessment | 2008 Assessment | 2008 Assessment |
|-----------------------------|--------------------------|--------------------------|--------------------------|--------------------------|
| Quantity | Recommendations for 2010 | Recommendations for 2011 | Recommendations for 2010 | Recommendations for 2009 |
| Tier | 3a | 3a | 3a | 3a |
| age 3+ biomass (t) | 370,332 | 367,217 | 322,714 | 323,937 |
| Female spawning biomass (t) | 124,674 | 128,585 | 109,441 | 111,463 |
| ABC (t) | 52,721 | 54,865 | 47,652 | 46,464 |
| OFL (t) | 65,567 | 68,206 | 59,349 | 57,911 |
| $F_{ABC} = F_{40\%}$ | 0.371 | 0.371 | 0.380 | 0.380 |
| $F_{OFL} = F_{35\%}$ | 0.481 | 0.481 | 0.494 | 0.494 |

SSC Comments Specific to the Flathead Sole Assessments

SSC Comments on Assessments in General

SSC request: The SSC requested that the next round of assessments consider the possible use of ADF&G bottom trawl survey data to expand the spatial and depth coverage.

Author response: The current assessment model can not accommodate surveys from multiple sources. We are developing a new assessment model that will incorporate surveys from multiple sources as one of its new features. When completed, this new model will allow us to explore the utility of using the ADF&G bottom trawl survey data in future assessments.

Introduction

Flathead sole (*Hippoglossoides elassodon*) are distributed from northern California, off Point Reyes, northward along the west coast of North America and throughout the GOA and the BS, the Kuril Islands, and possibly the Okhotsk Sea (Hart 1973). They occur primarily on mixed mud and sand bottoms (Norcross et al., 1997; McConnaughey and Smith, 2000) in depths < 300 m (Stark and Clausen, 1995). The flathead sole distribution overlaps with the similar-appearing Bering flounder (*Hippoglossoides robustus*) in the northern half of the Bering Sea and the Sea of Okhotsk (Hart, 1973), but not in the Gulf of Alaska.

Adults exhibit a benthic lifestyle and occupy separate winter spawning and summertime feeding distributions on the EBS shelf and in the GOA. From over-winter grounds near the shelf margins, adults begin a migration onto the mid and outer continental shelf in April or May each year for feeding. The spawning period may range from as early as January but is known to occur in March and April, primarily in deeper waters near the margins of the continental shelf. Eggs are large (2.75 to 3.75 mm) and females have egg counts ranging from about 72,000 (20 cm fish) to almost 600,000 (38 cm fish). Eggs hatch in 9 to 20 days depending on incubation temperatures within the range of 2.4 to 9.8°C and have been found in ichthyoplankton sampling on the southern portion of the BS shelf in April and May (Waldron 1981). Larvae absorb the yolk sac in 6 to 17 days, but the extent of their distribution is unknown. Nearshore sampling indicates that newly settled larvae are in the 40 to 50 mm size range (Norcross et al. 1996). Fifty percent of flathead sole females in the GOA are mature at 8.7 years, or at about 33 cm (Stark, 2004). Juveniles less than age 2 have not been found with the adult population and probably remain in shallow nearshore nursery areas.

Fishery

Flathead sole in the Gulf of Alaska are caught in a directed fishery using bottom trawl gear. Typically 25 or fewer shore-based catcher vessels from 58-125' participate in this fishery, as do 5 catcher-processor vessels (90-130'). Fishing seasons are driven by seasonal halibut PSC apportionments, with approximately 7 months of fishing occurring between January and November. Catches of flathead sole occur only in the Western and Central management areas in the gulf (statistical areas 610 and 620 + 630, respectively). Recruitment to the fishery begins at about age 3.

Historically, catches of flathead sole have exhibited decadal-scale trends (Table 8.1, Fig. 8.1). From a high of ~2000 t in 1980, annual catches declined steadily to a low of ~150 t in 1986 but thereupon increased steadily, reaching a high of ~3100 t in 1996. Catches subsequently declined over the next three years, reaching a low of ~900 t in 1999, followed by an increasing trend through 2008, when the catch reached its highest level ever (3,419 t). As of Sept. 26, catch in 2009 was 2,740 t and is expected to be similar to that in 2008 by year's end (3,398 t).

Based on observer data, the majority of the flathead sole catch in the Gulf of Alaska is taken in the Shelikof Strait and on the Albatross Bank near Kodiak Island, as well as near Unimak Island (Figure 8.2). The spatial pattern of catches has been reasonably consistent over the past three years. Most of the catch is taken in the first and second quarters of the year (Figure 8.3).

Annual catches of flathead sole have been well below TACs in recent years, although the population appears to be capable of supporting higher exploitation rates (Table 8.2a). Limits on flathead sole catches are driven by within-season closures of the directed fishery due to restrictions on halibut PSC, not by attainment of the TAC (Table 8.2b). Recognizing this, TACs have typically been set much lower than the recommended ABC. Prior to 2003, flathead sole was a Tier 5 species and ABC's were based on natural mortality rates. Following the development and adoption of an age-structured assessment model in 2003,

ABCs for flathead sole in the Gulf of Alaska almost doubled from 2002 to 2003, from ~23,000 to 42,000 t. TACs, however, increased only moderately as a result.

Flathead sole are also caught in the pursuit of other species as bycatch. They are caught in the Pacific cod, bottom pollock and other flatfish fisheries and are caught with these species in the flathead soledirected fishery. The gross retention rate for flathead sole over all fisheries has been 87% or larger since 2005 (Table 8.2a).

Data

Fishery Data

This assessment used fishery catches from 1984 through 26 September, 2009 (Table 8.1, Fig. 8.1), as well as estimates of the proportion of individuals caught by length group and sex for the years 1985-2009 (as of Sept. 26; Tables 8.3a, b). Sample sizes for the size compositions are shown in Table 8.4a. Age composition data from the fishery is not currently used in the assessment model.

Survey Data

Because flathead sole are often taken incidentally in target fisheries for other species, CPUE from commercial fisheries seldom reflects trends in abundance for this species. It is therefore necessary to use fishery-independent survey data to assess the condition of this stock.

This assessment used estimates of total biomass for flathead sole in the Gulf of Alaska from triennial (1984-1999) and biennial (2001-2009) groundfish surveys conducted by the Alaska Fisheries Science Center's Resource Assessment and Conservation Engineering (RACE) division to provide an index of population abundance (Table 8.5, Figure 8.4). Although survey depth coverage has been inconsistent for depth strata > 500 m, the fraction of the flathead sole stock occurring in these depth strata is miniscule (Table 8.6), so we have not attempted to correct the survey estimates of total biomass for missing depth strata. In addition, the 2001 survey estimate did not sample the eastern section of the Gulf. We estimated the average fraction of stock biomass occurring in the unsampled area from the 1993, 1996 and 1999 surveys (~11%) and assigned a corresponding availability factor of 0.9 to the 2001 survey to correct for the missing area (Table 8.5). Since 1984, survey estimates of total biomass have fluctuated about a mean of ~220,000 t with no apparent trend. Estimated total biomass was ~225,000 t in 2009, a 20% decrease from the 2007 survey estimate of ~280,000 t (the largest in the time series) but a 6% increase over the 2005 estimate of ~213,000 t.

Estimates of the total number of individuals by length group from each RACE GOA groundfish survey (Table 8.7) were also incorporated into the assessment, as were estimates of total population numbers-at-age (Table 8.8). Survey age compositions were available for 1984, 1990, 1993, 1996, 1999, 2003, 2005 and 2007. Because age compositions were calculated from age-length data using the corresponding size compositions, size compositions were de-weighted in the model likelihood for years where age composition data was available to avoid double counting. Survey size composition data was fully weighted in the model likelihood for years when age compositions were unavailable (1987, 2001 and 2009). Sample sizes for the survey size and age compositions are given in Table 8.4b.

Data on individual growth was incorporated in the assessment using sex-specific age-length transition matrices (Table 8.9a, b). These matrices were also used in previous assessments (Stockhausen et al., 2005 and 2007). Sex-specific weight-at-age relationships and female maturity schedules used in previous assessments (Stockhausen et al., 2005 and 2007) were also used in this assessment (Table 8.10).

To summarize, the following data was incorporated in the assessment:

| Source | type | years |
|-----------|-----------------------|-------------------------|
| Fishery | catch | 1984-2009 |
| r ishei y | size compositions | 1985-2009 |
| | hiomass | 1984-1999 (triennial); |
| | bioinass | 2001-2009 (biennial) |
| G | ai-a aa muu aaiti ama | 1984-1999 (triennial); |
| Survey | size compositions | 2001-2009 (biennial) |
| | aga compositions | 1984, 1990, 1993, 1996, |
| | age compositions | 1999, 2003, 2005, 2007 |

Analytic Approach

Model structure

The assessment was conducted using a split-sex, age-structured model with parameters evaluated in a maximum likelihood context. The model structure (Appendix A) was developed following Fournier and Archibald's (1982) methods, with many similarities to Methot (1990). We implemented the model using automatic differentiation software developed as a set of libraries under C++ (ADModel Builder). ADModel Builder can estimate a large number of parameters in a non-linear model using automatic differentiation software extended from Greiwank and Corliss (1991) and developed into C++ class libraries. This software provides the derivative calculations needed for finding the minimum of an objective function via a quasi-Newton function minimization routine (e.g., Press et al. 1992). It also gives simple and rapid access to these routines and provides the ability to estimate the variance-covariance matrix for all parameters of interest.

This year, we expanded the options for normalizing fishery and survey selectivity curves in the model. Previously, sex-specific selectivity curves (for both fisheries and surveys) were normalized to the maximum (unnormalized) value for female selectivity. In this assessment, we added options to estimate the maximum selectivity for males relative to females for either fisheries or surveys (or both). The maximum selectivity for females is still set to 1 and fishing mortality values are relative to fully-selected females. Thus, selectivity curves are now calculated in the following manner:

$$s_{F}^{N}(a) = s_{F}^{U}(a) / \max\{s_{F}^{U}(a)\}\$$

$$s_{M}^{N}(a) = [s_{M}^{U}(a) / \max\{s_{F}^{U}(a)\}] \cdot e$$

where $s_F^N(a)$ is the normalized selectivity curve for females as a function of age, $s_F^U(a)$ is the corresponding unnormalized curve, $s_M^N(a)$ and $s_M^U(a)$ are the corresponding curves for males, and *r* is the log-scale parameter for the relative scale between males and females. The previous scheme for normalizing selectivities is obtained if *r* is set to 0 and not estimated.

The current assessment model covers 1984-2009. Age classes included in the model run from age 3 to 20. Age at recruitment was set at 3 years in the model due to the small number of fish caught at younger ages. The oldest age class in the model, age 20, serves as a plus group in the model; the typical maximum age of flathead sole based on otolith age determinations has been estimated at 25 years (Turnock et al., 2003a). Details of the population dynamics and estimation equations, description of variables and likelihood components are presented in Appendix A (Tables A.1, A.2, and A.3). Model parameters that are typically fixed are presented in Table A.4. A total of 81 parameters were estimated in the final model (Table A.5).

Parameters estimated independently

Model parameters related to natural mortality, growth, weight, maturity and survey catchability were fixed in the final model (Table A.4).

Natural mortality

As in the previous assessment (Stockhausen et al., 2007), natural mortality (M) was fixed at 0.2 yr⁻¹ for both sexes in all age classes. This value was based on a maximum observed age for flathead sole of 22 years (Spencer et al., 1999). Although maximum observed age has increased to 31 years in the Bering Sea, a preliminary analysis of independent estimates of natural mortality for BSAI flathead sole is not inconsistent with continued use of this value (Stockhausen, unpublished data).

Growth

Individual growth was incorporated in the model using sex-specific age-length transition matrices (Table 8.9). These were identical to those used in the previous assessment (Stockhausen et al., 2007). In terms of the von Bertalanffy growth equation, L_{inf} was estimated at 44.37 cm for females and 37.36 cm for males (Figure 8.6a). The length at age 2 (L_2) was estimated at 10.17 cm for males and 13.25 cm for females. The growth parameter k was estimated at 0.157 for females and 0.204 for males. Length at age t was modeled as:

$$L_t = L_{inf} + (L_2 - L_{inf})e^{-k(t-2)}$$

Weight at length

The weight-length relationship used for flathead sole was identical to that used in the previous assessment (Stockhausen et al., 2007): $W = 0.00428 L^{3.2298}$ for both sexes combined (weight in grams and length in centimeters). Weight-at-age (Table 8.10, Figure 8.6b) was estimated using the mean length-at-age and the weight-length relationship.

Maturity

The maturity schedule for Gulf of Alaska flathead sole was estimated using histological analysis of ovaries collected in January 1999 (Stark, 2004; Table 8.10, Figure 8.6c). A total of 180 samples were analyzed for estimation of age at maturity. Size at 50% mature was estimated to be 33.3 cm with a slope of 0.52 cm^{-1} from a sample of 208 fish. Age at 50% mature was 8.74 years with a slope of 0.773 yr⁻¹. Size at 50% mature was estimated at 32.0 cm for Bering Sea flathead sole (not significantly different from the GOA results), however, age at 50% mature was 9.7 due to slower growth in the Bering sea.

Survey catchability

Based on results from the 2003 assessment (Turnock et al., 2003a), which indicated that estimating survey catchability was problematic, we fixed overall survey catchability (Q in Table A.1) in the model to a value of 1.

Parameters estimated conditionally

A total of 81 parameters were estimated in the final model (Table A.5). These consisted primarily of parameters on the recruitment of flathead sole to the population (44 parameters total, including ones determining the initial age composition) and values related to annual fishing mortality (27 parameters total). The separable age-component of fishing mortality was modeled using ascending logistic functions estimated separately for males and females (5 parameters total). The same approach was also used to estimate relative age-specific survey catchability (5 parameters total).

Annual recruitment to the age 3 year class was parameterized in the model using one parameter for the log-scale mean recruitment and 43 parameters for the annual log-scale deviation from the mean. Recruitments were estimated back to 1967 to provide an initial age distribution for the model in its

starting year (1984). In an analogous fashion, fully-recruited fishing mortality was parameterized in the model using one parameter for the log-scale mean and 26 parameters for the annual log-scale deviation from the mean.

Parameters in the model were selected based on minimizing an objective function equivalent to a negative log-likelihood function, hence the parameter estimates are maximum likelihood estimates. Components that contribute to the overall negative log likelihood include those related to observed fishery catches, fishery size compositions, survey biomass estimates, survey size compositions, survey age composition, and recruitment deviations (Table A.3). The observed fishery catch was assumed to have a lognormal error structure, as was estimated survey biomass. The recruitment deviation parameters were incorporated directly into the overall likelihood via three components: "early" recruitment, "ordinary" recruitment and "late" recruitment (Table A.3). The "early" recruitment component incorporated deviations from 1967 to 1983, i.e. those that determined the initial model age structure and were thus uninformed by contemporaneous catch data. The "ordinary" recruitment component incorporated deviations from 1984-2006, while the "late" recruitment component incorporated deviations from 2007-2009. "Late" recruitments are weighted separately in the likelihood from "ordinary" recruitments because there is generally little data to constrain recruitment estimates for the final few years in the model. This partitioning does not reflect any assumptions regarding changes in productivity with time: All three components were formulated assuming a lognormal error structure. The size and age compositions were assumed to be drawn from different sex-specific multinomial distributions. If this assumption were strictly correct, then the number of individuals contributing to each composition would be the appropriate corresponding sample size. However, because fish of the same size and age tend to be found together, size and age compositions tend to be overdispersed with respect to actual multinomial distributions. Also, the use of high sample sizes can lead to numerical problems in estimating the model parameters. Previous experience indicates that using a uniform sample size of 200 for compositions with more than 200 individuals provides an adequately simple solution to the problem of assigning sample sizes. Thus, a sample size of 200 was used for all compositions used in the likelihood (all age compositions, as well as size compositions from years with no corresponding age compositions).

Different weights can be assigned to each likelihood component to increase or decrease the relative degree of model fit to the data underlying the respective component; a larger weight induces a closer fit to a given likelihood component. Typically, a relatively large weight (e.g., 30) is applied to the catch component while smaller weights (e.g., 1) are applied to the survey biomass, recruitment, and size and age composition components. This reflects a belief that total catch data are reasonably well known (smaller variance) than the other types of data. For the recruitment components, larger weights applied to a component force the deviations contributing to that component closer to zero (and thus force recruitment closer to the geometric mean over the years that contribute to the component). The weights used in this assessment are given in Table 8.11.

Model evaluation

Several alternative model configurations were considered in a previous assessment (Stockhausen et al., 2005). Here, we took the model configuration selected in that assessment as a base case. As an alternative model, we allowed the model to estimate the relative scaling parameter for male selectivity for both the fishery and the survey. For both models, we assigned a weight of 30 to the catch-specific likelihood component and weights of 1 to all other likelihood components (Table 8.11). Initial values for the estimable parameters were set as listed in Table 8.12. To test whether resulting model solutions were indeed global, rather than local, maximum on the likelihood surface, we started the two model cases using several different parameter sets. All runs for a given case converged to the same final solution, providing evidence that the original solution was indeed the global maximum.

Fishery and survey selectivity functions for both model cases are illustrated in Figure 8.7. Ignoring the issue of scaling for the moment, the resulting functions are very similar for the two cases. The age by which fish are selected at 95% of their asymptotic rate in the fishery is 13.5 yrs for females and 13.0 yrs for males in the base case. In the alternative case, females reach 95% selectivity at a slightly younger age (13.0 yrs) while males reach 95% of their asymptotic rate at a slightly older age (13.5 yrs). For the survey, the age by which females are selected at 95% of their asymptotic rate is 9.8 yrs in the base case and 10.76 yrs in the alternative case. However, the log-scale male selectivity scaling parameters for both the fishery and survey are both different from 0 (the base case value) in the alternative model (0.159 for the fishery, -0.235 for the survey). As a result, asymptotic selectivity for males was slightly lower (21%) in the survey than that for females and higher in the fishery (17%). Somerton et al. (2007) showed that gear selectivity for flathead sole in the survey increases logistically with size. Because males reach a smaller asymptotic size than females, one would thus expect that age-specific survey selectivity for older males would be somewhat smaller than that for females of similar age.

Further comparison of the results from the two model cases are shown for several variables of interest in Fig. 8.8. Estimates for total biomass, spawning biomass and recruitment were consistently higher in the alternative case when compared with the base case, although the discrepancy was somewhat larger for total biomass and somewhat smaller for recruitment. This appears to be a consequence of the alternative model estimate for the survey male selectivity scale parameter being less than one (negative on a log scale). Estimates for survey biomass (not shown) are nearly identical for both models. When the male scaling parameter for the survey is less than 1, the underlying population must be larger to result in the same estimated survey biomass. Because the alternative and base models result in the same estimated survey biomass, the underlying population must be larger in the case of the alternative model to offset the fact that the survey in the alternative model is not "seeing" all the fish that the survey in the base case sees.

In contrast with the population estimates, estimates for fishing mortality (relative to older females) were consistently higher in the base model than in the alternative model. This may be either, to first order, a consequence of the value of the male scaling parameter for the fishery *or* for the survey. In the latter case, as we have already discussed, a negative (log-scale) estimate for the survey scaling parameter results in higher population biomass estimates. Because both models are constrained to closely fit the observed catch, estimates of fishing mortality from the alternative model will be smaller than those from the base model simply because the total population size is larger in the alternative model. In the case of the fishery scaling parameter, the alternative model estimated a positive (log-scale) value for that parameter, indicating that more (male) fish would be caught in the alternative model for the same value of fishing mortality as were caught in the base model (for the same population size). Because both models were constrained to fit the observed catch history, this could be achieved in the alternative model at lower fishing mortality than in the base model, since population sizes were similar. The results we obtained from the alternative model probably represent contributions from both these factors.

Likelihood profiles for the fishery and survey selectivity parameters were calculated for both model cases and profiles for individual selectivity parameters were visually compared (Fig. 8.9). In general, the profiles for individual parameters overlap to some extent between the two cases. The widths (i.e., standard deviations) of the profiles tend to be slightly larger for the alternative case, compared with the base case. It is clear from the profiles for the scaling parameters, though, that the estimated parameters are significantly different from 0, indicating that male and female asymptotic selectivities are not identical (as assumed in the base model). In addition, the overall fit to the data is about 10 likelihood units better in the alternative model than the base model (Table 8.13). While the base model fits survey biomass slightly better than the alternative model (~0.5 units), the alternative model fits the fishery size compositions (~3.5 units), the survey size compositions (~0.9 units), and the survey age compositions $(\sim 5.1 \text{ units})$ better than base model. As such, we have selected the alternative model as the preferred model to use for population projection, evaluation of harvest alternatives and status determination, and reference value calculation. However, we also provide a complementary summary table with reference values calculated using the base model at the end of the text portion of the chapter.

Final parameter estimates

The parameter estimates, based on the preferred alternative model, considered final for this assessment are given in Table 8.14 for all model parameters.

Schedules implied by parameter estimates

The estimated relative scaling parameter for male selectivity was significantly different from 0 for both the fishery and the survey (Figure 8.7). Asymptotic male selectivity was 21% smaller than female selectivity for the survey, while it was 17% larger for the fishery. The estimated selectivity curves for the fishery and survey indicate that the fishery generally catches older flathead sole than the survey (Figure 8.7). For the fishery, age at 95% selection was 13.0 for females and 13.5 for males. For the survey, the ages at 95% selection were younger: 10.8 for females and 8.2 for males.

Results

As expected, the accepted model (the alternative model) estimates of fishery catch closely matched the observed values (Table 8.15 and Figure 8.10). The model did not fit the fishery size compositions nearly as well, although its performance appeared to be reasonably good in most years (Figures 8.11 and 8.12 for females and males, respectively). Fits to the fishery size compositions were poorest when the observed size composition was dominated by a single size class and thus sharply peaked (e.g., 1987 in Figure 8.11). The smoothing inherent in using an age-length transition matrix to convert age classes to size classes precludes close fits to peaked size compositions.

The model did not fit observed survey biomass values as closely as it did the catch (Table 8.15 and Figure 8.13), but model estimates of survey biomass fell outside the 95% confidence intervals of the actual surveys for only two out of eleven survey years (1984 and 2001) so the fit was deemed satisfactory. As with the fishery size compositions, model fits to the survey size compositions were poorest when the observed size compositions were sharply peaked, but still generally reasonable (Figures 8.14 and 8.15). Finally, the model also fit the survey age compositions reasonably well (Figures 8.16 and 8.17).

The model also estimates other population variables of interest, such as time series of total biomass, spawning biomass, recruitment and fully-selected fishing mortality. In this assessment, total biomass is represented by age 3+ biomass whereas spawning biomass is female spawning biomass. Model estimates of age 3+ biomass increased moderately from 246,000 t in 1984 to 299,000 t in 1996 and 1997, then declined to a low of 293,000 t in 2000 and subsequently rose steadily in recent years to achieve their highest level in 2009 at 372,000 t (Table 8.16 and Figure 8.18). The estimated age 3+ biomass in this assessment is higher than that estimated in both the 2005 and 2007 assessments (Table 8.16, Figure 8.18). The estimated female spawning biomass is quite similar to that from the 2007 and 2005 assessments, but is slightly higher (4%, on average).

Model estimates of annual recruitment (age 3 numbers) ranged from a low of 180,000,000 individuals in 1999 to highs of 413,000,000 in 2002 and 411,000,000 in 2006 (Table 8.17 and Figure 8.19). Prior to 2000, recruitment was generally below the long-term average (278,000,000), while it has generally been higher since 2000. In 2009, recruitment was estimated below the long-term average, but this is expected because of the structure of the recruitment likelihood. Results from the current assessment are generally similar to those estimated in the 2007 assessment (Table 8.17, Figure 8.19). The only dramatic change

has been to revise the 2004 recruitment (2001 year class) from 167,000,000 individuals to 382,000,000. This is a result of the more complete entrance into the survey by this year class in the current survey.

A control rule plot showing the temporal trajectory of estimated fishing mortality and spawning biomass indicates that the GOA flathead sole stock has not been overfished nor has overfishing occurred (Figure 8.20).

Projections and Harvest Alternatives

The reference fishing mortality rate for flathead sole is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Estimates of $F_{40\%}$, $F_{35\%}$, and $SPR_{40\%}$ were obtained from a spawner-per-recruit analysis. Assuming that the average recruitment from the 1981-2007 year classes estimated in this assessment represents a reliable estimate of equilibrium recruitment, then an estimate of $B_{40\%}$ is calculated as the product of $SPR_{40\%}$ times the equilibrium number of recruits; this quantity is 49,899 t. The 2009 spawning stock biomass is estimated at 120,000 t. Since reliable estimates of the 2009 spawning biomass (*B*), $B_{40\%}$, $F_{40\%}$, and $F_{35\%}$ exist and $B > B_{40\%}$ (120,000 t > 49,899 t), the flathead sole reference fishing mortality is defined in Tier 3a.

For this tier, F_{ABC} is constrained to be $\leq F_{40\%}$, and F_{OFL} is defined to be $F_{35\%}$. The values of these quantities are:

| estimated 2009 SSB | = | 120,070 t |
|--------------------|--------|-----------|
| B 40% | = | 49,899 t |
| $F_{40\%}$ | = | 0.371 |
| F _{ABC} | \leq | 0.371 |
| B 35% | = | 43,661 t |
| F 35% | = | 0.481 |
| F _{OFL} | = | 0.481 |

Because the flathead sole stock has not been overfished in recent years and the stock biomass is relatively high, it is not recommended to adjust F_{ABC} downward from its upper bound; thus, the year 2010 recommended ABC associated with F_{ABC} of 0.371, is 52,721 t. The fishing mortality associated with overfishing (F_{OFL}) is 0.481. The corresponding OFL for 2010 is 65,567 t.

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2009 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2010 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2009. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This

projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2010, are as follow ("max F_{ABC} " refers to the maximum permissible value of F_{ABC} under Amendment 56):

Scenario 1: In all future years, F is set equal to max F_{ABC} . (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, F is set equal to a constant fraction of max F_{ABC} , where this fraction is equal to the ratio of the F_{ABC} value for 2010 recommended in the assessment to the max F_{ABC} for 2010. (Rationale: When F_{ABC} is set at a value below max F_{ABC} , it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, F is set equal to 50% of max F_{ABC} . (Rationale: This scenario provides a likely lower bound on F_{ABC} that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, F is set equal to the 2005-2009 average F. (Rationale: For some stocks, TAC can be well below ABC, and recent average F may provide a better indicator of F_{TAC} than F_{ABC} .)

Scenario 5: In all future years, *F* is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

The recommended F_{ABC} and the maximum F_{ABC} are equivalent in this assessment, so scenarios 1 and 2 yield identical results. The 12-year projections of the mean harvest, spawning stock biomass and fishing mortality for the five scenarios are shown in Tables 8.18-20.

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether the flathead sole stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follows (for Tier 3 stocks, the MSY level is defined as $B_{35\%}$):

Scenario 6: In all future years, F is set equal to F_{OFL} . (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2010, then the stock is not overfished.)

Scenario 7: In 2010 and 2011, F is set equal to max F_{ABC} , and in all subsequent years, F is set equal to F_{OFL} . (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2022 under this scenario, then the stock is not approaching an overfished condition.)

The results of these two scenarios indicate that the flathead sole are not overfished and is not approaching an overfished condition. With regard to assessing the current stock level, the expected stock size in the year 2010 of scenario 6 is 124,674 t, almost 3 times $B_{35\%}$ (43,661 t). Thus the stock is not currently overfished. With regard to whether the stock is approaching an overfished condition, the expected spawning stock size in the year 2022 of scenario 7 (45,825 t) is greater than $B_{35\%}$; thus, the stock is not approaching an overfished condition.

Estimating an ABC and OFL for 2011 is somewhat problematic as these values depend on the catch that will be taken in 2010. The actual catch taken in the GOA flathead sole fishery has been substantially smaller than the TAC for the past several years, but the catch has been rising steadily since 1999 (Figure 8.1). The year end 2009 catch was predicted to be 3,398 t, almost as much as in 2008 (3,419 t; the largest catch in the time series). Thus, we assumed that a reasonable estimate of the catch to be taken in 2010 was the same as that taken in 2008. Using these values and the estimated population size at the start of 2009 from the model, we projected the stock ahead through 2009-2010 and calculated the ABC and OFL for 2011. The estimated ABC for 2011 is 54,865 t while the estimated OFL is 68,206. Total biomass for 2011 is estimated at 367,217 t, while female spawning biomass is estimated at 128,585.

Area allocation of harvests

TAC's for flathead sole in the Gulf of Alaska are divided among four smaller management areas (Western, Central, West Yakutat and Southeast Outside). As in the previous assessment, the area-specific ABC's for flathead sole in the GOA are divided up over the four management areas by applying the fraction of the most recent survey biomass estimated for each area (relative to the total over all areas) to the 2010 and 2011 ABC's. The area-specific allocations for 2010 and 2011 are:

| | Western Gulf | Central Gulf | West Yakutat | Southeast Outside | Grand Total |
|---------------|-----------------|-----------------|-----------------|----------------------|----------------|
| apportionment | 35.5% | 57.2% | 4.2% | 3.1% | 100.0% |
| 2010 ABC (t) | 18,741 | 30,155 | 2,212 | 1,613 | 52,721 |
| 2011 ABC (t) | 19,503 | 31,381 | 2,302 | 1,678 | 54,865 |

Ecosystem Considerations

Ecosystem effects on the stock

Prey availability/abundance trends

Based on results from an ecosystem model for the Gulf of Alaska (Aydin et al., 2007), flathead sole in the Gulf of Alaska occupy an intermediate trophic level as both juvenile and adults (Fig. 8.21). Pandalid shrimp and brittle stars were the most important prey for adult flathead sole in the Gulf of Alaska (64% by weight in sampled stomachs; Yang and Nelson, 2000; Fig. 8.22a), while euphausids and mysids constituted the most important prey items for juvenile flathead sole (Fig. 8.22b)... Other major prey items included polychaetes, mollusks, bivalves and hermit crabs for both juveniles and adults. Commercially important species that were consumed included age-0 Tanner crab (3%) and age-0 walleye pollock (< 0.5% by weight). Little to no information is available to assess trends in abundance for the major benthic prey species of flathead sole.

Predator population trends

Important predators on flathead sole include arrowtooth flounder, walleye pollock, Pacific cod, and other groundfish (Fig. 8.23). Pacific cod and Pacific halibut are the major predators on adults, while arrowtooth flounder, sculpins, walleye pollock and Pacific cod are the major predators on juveniles. The flatfish-directed fishery constitutes the third-largest known source of mortality on flathead sole adults. However, the largest component of mortality on adults is unexplained.

Arrowtooth flounder are currently the most abundant groundfish in the Gulf of Alaska, and have steadily increased in abundance since the early 1970's (Turnock et al., 2003b). The abundance of walleye pollock has declined rather steadily since the early 1990's, but recent evidence suggests the stock may be starting to increase again (Dorn et al., 2004). Pacific cod abundance in the Gulf of Alaska has been declining since 1990 (Thompson et al., 2004). Although the continued increase in abundance of arrowtooth

flounder is cause for some concern, the abundance of flathead sole has actually increased in recent years. Predation by arrowtooth may be limiting the potential rate of increase of flathead sole under current conditions, but it does not appear to represent a threat to the stock.

Fishery effects on ecosystem

Catches of flathead sole have been concentrated in several areas in the Gulf of Alaska over the past few years (Figure 8.2). These areas include Shelikof Straight, Portlock Bank and Davidson Bank. The ecosystem effects of this spatial concentration of fishing activity are unknown.

Prohibited species such as halibut, salmon, and crab are also taken to some extent in the flathead soledirected fishery (Table 8.21). In 2009 thus far, the overall prohibited species catch (PSC) for halibut was almost 52,000 kg halibut—a decrease from the 2008 catch of almost 92,000 kg but larger than the 2007 and 2006 catches (approximately 27,000 and 37,000 kg, respectively). The PSC for crab in the directed fishery is mainly Bairdi tanner crab, with catches sometimes fluctuating by factors of 3-4 between years. The PSC for crab thus far in the 2009 directed fishery was approximately 7,000 Bairdi tanner crab, similar to that caught in 2008. The PSC for salmon in the directed fishery is mainly Chinook, with 118 individuals caught in 2009. No individuals were caught in the two previous years.

Over the past four years, the flathead sole-directed fishery caught more arrowtooth flounder than any other non-prohibited species, including flathead sole (Table 8.22). Flathead sole was the second most-caught species in the directed fishery. Only small amounts of arrowtooth were retained (typically 10%), while generally more than 90% of flathead sole was retained. Pacific cod was the third most-caught species, with retention rates typically greater than 90%.

Effects of discards and offal production on the ecosystem are unknown for the flathead sole fishery.

Data gaps and research priorities

The AFSC's Age and Growth Program has made substantial progress in processing survey age data for flathead sole in the Gulf of Alaska. While this information has been incorporated in the current stock assessment in the form of survey age compositions, age information also enters the assessment in the form of age-length conversion matrices estimated outside the assessment model. The matrices currently used in the assessment are now several years old. One of our goals for the next assessment is to use the newly-available age data to revise growth schedules for GOA flathead and reassess these age-length conversion matrices. In addition, we anticipate incorporating such estimation directly into the assessment model, rather than performing it outside the model. This approach will also allow us to incorporate ageing error into the model structure.

Although the AFSC's Age and Growth Program has made substantial progress in processing survey age data for flathead sole in the Gulf of Alaska, the amount of fishery age data is almost nonexistent. Additional age data (both survey and fishery) should improve future stock assessments by allowing improved estimates of individual growth and age-length transition matrices, and by filling in missing years with age composition data.

Further modeling research should address the use of length-based approaches to fishery and survey selectivity in the assessment model, as well as alternative forms for the selectivity function. The utility of potential environmental predictors of recruitment (e.g., temperature) should also be investigated. We will also revisit the estimates used for natural mortality in the model.

Summary Preferred model

| Freienrea model. | | | |
|---------------------------------|-----------|---------|--|
| Tier | 3a | | |
| Reference mortality rates | | | |
| - M | 0.2 | | |
| $F_{35\%}$ | 0.481 | | |
| $F_{40\%}$ | 0.371 | | |
| Equilibrium female spawning | biomass | | |
| | 124,747 t | | |
| $B_{40\%}$ | 49,899 t | | |
| $B_{35\%}$ | 43,661 t | | |
| Fishing rates | | | |
| For | 0.481 | | |
| F_{ABC} (maximum permissible) | 0.371 | | |
| F_{ABC} (recommended) | 0.371 | | |
| Projected biomass | 2010 | 2011 | |
| Age $3+$ biomass (t) | 370,332 | 367,217 | |
| Female spawning biomass (t) | 124,674 | 128,585 | |
| | | | |
| Harvest limits | 2010 | 2011 | |
| OFL (t) | 65,567 | 68,206 | |
| ABC (maximum permissible; | 52,721 | 54,865 | |
| ABC (recommended; t) | 52,721 | 54,865 | |
| | | | |

| Base model. | | |
|---------------------------------|-----------|---------|
| Tier | 3a | |
| Reference mortality rates | | |
| M | 0.2 | |
| $F_{35\%}$ | 0.530 | |
| $F_{40\%}$ | 0.406 | |
| Equilibrium female spawning l | oiomass | |
| $B_{100\%}$ | 111,884 t | |
| $B_{40\%}$ | 44,754 t | - |
| B 33% | 39,159 t | |
| Fishing rates | | |
| F OFL | 0.530 | _ |
| F_{ABC} (maximum permissible) | 0.406 | |
| F ABC (recommended) | 0.406 | |
| Projected biomass | 2010 | 2011 |
| Age 3+ biomass (t) | 328,862 | 325,922 |
| Female spawning biomass (t) | 110,387 | 113,717 |
| Harvest limits | 2010 | 2011 |
| OFL (t) | 59.295 | 61.601 |
| ABC (maximum permissible; | 47,422 | 49,286 |
| ABC (recommended: t) | 47,422 | 49.286 |

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Tables

| | total catch |
|------|--------------|
| year | (t) |
| 1978 | 452 |
| 1979 | 165 |
| 1980 | 2,068 |
| 1981 | 1,070 |
| 1982 | 1,368 |
| 1983 | 1,080 |
| 1984 | 549 |
| 1985 | 320 |
| 1986 | 147 |
| 1987 | 151 |
| 1988 | 520 |
| 1989 | 747 |
| 1990 | 1,447 |
| 1991 | 1,717 |
| 1992 | 2,034 |
| 1993 | 2,366 |
| 1994 | 2,580 |
| 1995 | 2,181 |
| 1996 | 3,107 |
| 1997 | 2,446 |
| 1998 | 1,742 |
| 1999 | 900 |
| 2000 | 1,547 |
| 2001 | 1,911 |
| 2002 | 2,145 |
| 2003 | 2,425 |
| 2004 | 2,390 |
| 2005 | 2,530 |
| 2006 | 3,134 |
| 2007 | 3,163 |
| 2008 | 3,419 |
| 2009 | 2,740 |

Table 8.1. Annual catch of flathead sole in the Gulf of Alaska, from 1978 to 2009. 2009 catch is through Sept. 26, 2009.

| Voor | Author | | | OFL (t) | Total Catch | % |
|------|---------|---------|---------|---------|--------------------|----------|
| Iear | ABC (t) | ADC (l) | TAC (I) | OFL (I) | (t) | Retained |
| 1995 | | 28,790 | 9,740 | 31,557 | 2,181 | |
| 1996 | | 52,270 | 9,740 | 31,557 | 3,107 | |
| 1997 | | 26,110 | 9,040 | 34,010 | 2,446 | |
| 1998 | | 26,110 | 9,040 | 34,010 | 1,742 | |
| 1999 | | 26,010 | 9,040 | 34,010 | 900 | |
| 2000 | | 26,270 | 9,060 | 34,210 | 1,547 | |
| 2001 | | 26,270 | 9,060 | 34,210 | 1,911 | |
| 2002 | 22,684 | 22,690 | 9,280 | 29,530 | 2,145 | |
| 2003 | 41,402 | 41,390 | 11,150 | 51,560 | 2,425 | 88 |
| 2004 | 51,721 | 51,270 | 10,880 | 64,750 | 2,390 | 80 |
| 2005 | 36,247 | 45,100 | 10,390 | 56,500 | 2,530 | 87 |
| 2006 | 37,820 | 37,820 | 9,077 | 47,003 | 3,134 | 89 |
| 2007 | 39,110 | 39,110 | 9,148 | 48,658 | 3,163 | 89 |
| 2008 | 44,735 | 44,735 | 11,054 | 55,787 | 3,419 | 90 |
| 2009 | 46,464 | 46,464 | 11,181 | 57,911 | 2,740 | 96 |

Table 8.2a. Time series of recent reference points (ABC, OFL), TACs, total catch and retention rates for flathead sole.

| I Cal | Dates | Status |
|-------|-----------------|-------------------------|
| 2005 | Jan 20 | open |
| | Aug 19 | halibut bycatch status |
| | Sep 1 | open |
| | Sep 4 | halibut bycatch status |
| 2006 | Jan 20 | open |
| | Feb 23 | halibut bycatch status |
| | Feb 27 | open |
| | Iun 10 | halibut bycatch status |
| | Jul 1 | open |
| | Son 1 | balibut bugatah status |
| | Sep 1 | hanout bycatch status |
| | Spe 6 | open |
| | Sep 6 | halibut bycatch status |
| | Sep 20 | open |
| | Spe 20 | halibut bycatch status |
| | Sep 25 | open |
| | Sep 25 | halibut bycatch status |
| | Oct 1 | open |
| | Oct 8 | halibut bycatch status |
| 2007 | Jan 20 | open |
| | Jun 4 | halibut bycatch status |
| | Jul 1 | open |
| | Aug 10 | halibut bycatch status |
| | Sep 1 | open |
| | Sep 1 | halibut bycatch status |
| | Sep 6 | open |
| | Sep 6 | halibut bycatch status |
| | Sep 0 Sep 11 | open |
| | Sep 11 | halibut bycatch status |
| | Sep 21 | open |
| | Sep 23 | halibut bycatch status |
| | Oct 1 | open |
| | Oct 8 | halibut bycatch status |
| | Oct 10 | open |
| | Oct 15 | halibut bycatch status |
| | Oct 22 | open |
| 2008 | Jon 20 | open |
| 2008 | Jall 20 | A 80 yessels subject to |
| | Inn 22 | Add vessels subject to |
| | Jall 25 | sideboard mints. nanout |
| | | A 80 waggala subject to |
| | Jan 29 | A80 vessels subject to |
| | N/ 10 | sideboard limits: open |
| | Mar 10 | halibut bycatch status |
| | Mar 21 | open |
| | May 21 | halibut bycatch status |
| | Jul I | open |
| | Aug 7 | halibut bycatch status |
| | Sep 1 | open |
| | Sep 3 | halibut bycatch status |
| | Sep 10 | open |
| | Sep 11 | halibut bycatch status |
| | Oct 1 | open |
| | Nov 6 | halibut bycatch status |
| | Nov 16 | open |
| 2009 | Jan 20 | open |
| | Sep 2 | halibut bycatch status |
| | Oct 1 | open |

 Table 8.2b. Status of flathead sole fishery in recent years.

 Year Dates Status

Table 8.3a. Annual fishery length compositions for female flathead sole. The 2009 composition is based on observer reports through Sept. 26. Fishery length compositions are normalized to 1 over both sexes.

| | Length c | utpoints (| (cm) | | | | | | | | | | | | | | | |
|------|----------|------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| year | 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1985 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0100 | 0.0558 | 0.0657 | 0.0817 | 0.1135 | 0.0837 | 0.0478 | 0.0219 | 0.0060 | 0.0020 | 0.0000 |
| 1986 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0129 | 0.0065 | 0.0774 | 0.0839 | 0.0903 | 0.0645 | 0.0581 | 0.0194 | 0.0129 | 0.0129 | 0.0000 | 0.0000 |
| 1987 | 0.0000 | 0.0000 | 0.0000 | 0.0058 | 0.0000 | 0.0058 | 0.0116 | 0.0465 | 0.1047 | 0.1395 | 0.2558 | 0.0698 | 0.0349 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1988 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0005 | 0.0032 | 0.0053 | 0.0069 | 0.0354 | 0.0994 | 0.1274 | 0.1332 | 0.1142 | 0.0840 | 0.0418 | 0.0143 | 0.0026 |
| 1989 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0256 | 0.0233 | 0.0653 | 0.0956 | 0.0979 | 0.0443 | 0.0140 | 0.0093 | 0.0023 |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0201 | 0.1409 | 0.0940 | 0.0940 | 0.0403 | 0.0470 | 0.0000 | 0.0000 | 0.0000 |
| 1991 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0011 | 0.0027 | 0.0106 | 0.0217 | 0.0342 | 0.0422 | 0.0601 | 0.0973 | 0.0927 | 0.0589 | 0.0270 | 0.0084 | 0.0027 | 0.0015 |
| 1992 | 0.0000 | 0.0000 | 0.0008 | 0.0011 | 0.0049 | 0.0081 | 0.0111 | 0.0266 | 0.0356 | 0.0465 | 0.0630 | 0.0723 | 0.0603 | 0.0418 | 0.0231 | 0.0242 | 0.0155 | 0.0133 |
| 1993 | 0.0011 | 0.0006 | 0.0008 | 0.0011 | 0.0037 | 0.0065 | 0.0034 | 0.0056 | 0.0115 | 0.0213 | 0.0399 | 0.0590 | 0.0581 | 0.0528 | 0.0427 | 0.0371 | 0.0298 | 0.0247 |
| 1994 | 0.0000 | 0.0000 | 0.0005 | 0.0029 | 0.0067 | 0.0100 | 0.0257 | 0.0371 | 0.0413 | 0.0689 | 0.0660 | 0.0760 | 0.0698 | 0.0570 | 0.0299 | 0.0247 | 0.0138 | 0.0280 |
| 1995 | 0.0000 | 0.0000 | 0.0004 | 0.0004 | 0.0015 | 0.0015 | 0.0062 | 0.0128 | 0.0438 | 0.0601 | 0.0841 | 0.0934 | 0.0790 | 0.0353 | 0.0236 | 0.0120 | 0.0124 | 0.0294 |
| 1996 | 0.0000 | 0.0004 | 0.0015 | 0.0030 | 0.0045 | 0.0056 | 0.0054 | 0.0112 | 0.0244 | 0.0337 | 0.0595 | 0.0752 | 0.0802 | 0.0646 | 0.0432 | 0.0277 | 0.0192 | 0.0368 |
| 1997 | 0.0005 | 0.0005 | 0.0010 | 0.0017 | 0.0050 | 0.0084 | 0.0109 | 0.0226 | 0.0278 | 0.0533 | 0.0670 | 0.0875 | 0.0794 | 0.0461 | 0.0263 | 0.0174 | 0.0107 | 0.0099 |
| 1998 | 0.0000 | 0.0002 | 0.0004 | 0.0004 | 0.0011 | 0.0026 | 0.0046 | 0.0124 | 0.0221 | 0.0322 | 0.0575 | 0.0822 | 0.0877 | 0.0655 | 0.0373 | 0.0254 | 0.0159 | 0.0227 |
| 1999 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0080 | 0.0000 | 0.0240 | 0.0400 | 0.0480 | 0.1040 | 0.1360 | 0.0800 | 0.0560 | 0.0080 | 0.0400 | 0.0160 |
| 2000 | 0.0000 | 0.0000 | 0.0007 | 0.0014 | 0.0007 | 0.0036 | 0.0080 | 0.0174 | 0.0282 | 0.0477 | 0.0745 | 0.0788 | 0.0665 | 0.0427 | 0.0398 | 0.0188 | 0.0123 | 0.0072 |
| 2001 | 0.0000 | 0.0000.0 | 0.0000 | 0.0000 | 0.0008 | 0.0025 | 0.0016 | 0.0098 | 0.0263 | 0.0279 | 0.0443 | 0.0541 | 0.0705 | 0.0582 | 0.0533 | 0.0377 | 0.0336 | 0.0410 |
| 2002 | 0.0000 | 0.0008 | 0.0023 | 0.0008 | 0.0023 | 0.0039 | 0.0124 | 0.0202 | 0.0419 | 0.0489 | 0.0559 | 0.0761 | 0.0730 | 0.0621 | 0.0466 | 0.0272 | 0.0101 | 0.0163 |
| 2003 | 0.0004 | 0.0000 | 0.0004 | 0.0000 | 0.0000 | 0.0028 | 0.0040 | 0.0048 | 0.0132 | 0.0227 | 0.0279 | 0.0450 | 0.0630 | 0.0570 | 0.0514 | 0.0315 | 0.0211 | 0.0191 |
| 2004 | 0.0000 | 0.0000 | 0.0000 | 0.0007 | 0.0021 | 0.0057 | 0.0136 | 0.0107 | 0.0264 | 0.0314 | 0.0507 | 0.0600 | 0.0700 | 0.0771 | 0.0707 | 0.0457 | 0.0271 | 0.0300 |
| 2005 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0023 | 0.0087 | 0.0110 | 0.0140 | 0.0308 | 0.0343 | 0.0483 | 0.0797 | 0.0837 | 0.0802 | 0.0500 | 0.0209 | 0.0140 | 0.0116 |
| 2006 | 0.0005 | 0.0005 | 0.0016 | 0.0011 | 0.0038 | 0.0055 | 0.0093 | 0.0132 | 0.0241 | 0.0346 | 0.0538 | 0.0757 | 0.0686 | 0.0669 | 0.0461 | 0.0252 | 0.0148 | 0.0115 |
| 2007 | 0.0000 | 0.0000 | 0.0000 | 0.0012 | 0.0043 | 0.0031 | 0.0160 | 0.0185 | 0.0314 | 0.0579 | 0.0672 | 0.0702 | 0.0598 | 0.0333 | 0.0333 | 0.0197 | 0.0222 | 0.0327 |
| 2008 | 0.0000 | 0.0007 | 0.0042 | 0.0023 | 0.0021 | 0.0105 | 0.0064 | 0.0168 | 0.0300 | 0.0320 | 0.0579 | 0.0915 | 0.0785 | 0.0686 | 0.0315 | 0.0242 | 0.0189 | 0.0219 |
| 2009 | 0.0000 | 0.0000 | 0.0005 | 0.0024 | 0.0021 | 0.0087 | 0.0117 | 0.0263 | 0.0249 | 0.0283 | 0.0549 | 0.0756 | 0.0654 | 0.0526 | 0.0483 | 0.0190 | 0.0137 | 0.0280 |

Table 8.3b. Annual fishery length compositions for male flathead sole. The 2009 composition is based on observer reports through Sept. 26. Fishery length compositions are normalized to 1 over both sexes.

| gth | Cut | points (| cm) | | | | | | | | | | | | | | | |
|-------------------|------------|---------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|--------|
| 14 16 | 16 | | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 | 48 |
| $0 \qquad 0$ | 0 | | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 0000 0.0000 0.0 | 0.0000 0.0 | 0.0 | 0000 | 0.0020 | 0.0020 | 0.0000 | 0.0139 | 0.0677 | 0.1335 | 0.1653 | 0.0916 | 0.0259 | 0.0080 | 0.0020 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 0000 0.0000 0.0 | 0.0000.C | 0.0 | 0000 | 0.0065 | 0.0000 | 0.0000 | 0.0129 | 0.0645 | 0.1355 | 0.1548 | 0.0645 | 0.0452 | 0.0516 | 0.0258 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 0000 0.0000 0.000 | 0.0000.0 | 0.0 | 0000 | 0.0058 | 0.0000 | 0.0233 | 0.0233 | 0.0640 | 0.1047 | 0.0930 | 0.0116 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 0000 0.0000 0.0 | 0.0000.0 | 0.0 | 0000 | 0.0000 | 0.0000 | 0.0000 | 0.0042 | 0.0143 | 0.0560 | 0.0957 | 0.0809 | 0.0476 | 0.0233 | 0.0069 | 0.0021 | 0.0011 | 0.0000 | 0.0000 |
| 0000 0.0000 0.000 | 0.0000 0.0 | 0.0 | 0000 | 0.0000 | 0.0000 | 0.0070 | 0.0210 | 0.0932 | 0.0653 | 0.2051 | 0.1096 | 0.0373 | 0.0396 | 0.0280 | 0.0140 | 0.0023 | 0.0000 | 0.0000 |
| 0000 0.0000 0.0 | 0.0000.0 | 0.0 | 0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0336 | 0.0537 | 0.2349 | 0.1678 | 0.0470 | 0.0268 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 004 0.0004 0. | 0.0004 0. | 0 | 0008 | 0.0004 | 0.0072 | 0.0125 | 0.0300 | 0.0391 | 0.0802 | 0.1513 | 0.1365 | 0.0563 | 0.0148 | 0.0061 | 0.0015 | 0.0011 | 0.0004 | 0.0000 |
| 005 0.0038 0. | 0.0038 0. | 0 | 0027 | 0.0092 | 0.0144 | 0.0231 | 0.0348 | 0.0598 | 0.1049 | 0.1176 | 0.0989 | 0.0497 | 0.0201 | 0.0084 | 0.0035 | 0.0003 | 0.0000 | 0.0000 |
| 008 0.0011 0 | 0.0011 0. | 0 | 0014 | 0.0031 | 0.0067 | 0.0129 | 0.0244 | 0.0385 | 0.0845 | 0.1289 | 0.1457 | 0.0935 | 0.0337 | 0.0183 | 0.0031 | 0.0011 | 0.0003 | 0.0022 |
| 0 0000.0 0000 | 0 0000.0 | 0 | 0000 | 0.0014 | 0.0076 | 0.0157 | 0.0219 | 0.0323 | 0.0732 | 0.1116 | 0.0893 | 0.0413 | 0.0166 | 0.0043 | 0.0067 | 0.0043 | 0.0010 | 0.0147 |
| 0 0000.0 0000 | 0 0000.0 | 0 | .0019 | 0.0019 | 0.0050 | 0.0097 | 0.0151 | 0.0480 | 0.0821 | 0.1069 | 0.1031 | 0.0632 | 0.0260 | 0.0163 | 0.0136 | 0.0070 | 0.0027 | 0.0015 |
| 0004 0.0006 0 | 0.0006 0 | 0 | 0024 | 0.0056 | 0.0065 | 0.0080 | 0.0115 | 0.0292 | 0.0659 | 0.1036 | 0.1094 | 0.0692 | 0.0456 | 0.0277 | 0.0117 | 0.0054 | 0.0004 | 0.0004 |
| 0005 0.0007 0 | 0.0007 0 | 0 | .0022 | 0.0050 | 0.0067 | 0.0156 | 0.0188 | 0.0419 | 0.0804 | 0.1114 | 0.1029 | 0.0779 | 0.0362 | 0.0126 | 0.0032 | 0.0015 | 0.0005 | 0.0060 |
| 0004 0.0000 0 | 0 0000.0 | 0 | .0013 | 0.0029 | 0.0040 | 0.0075 | 0.0159 | 0.0390 | 0.0672 | 0.1120 | 0.1149 | 0.0813 | 0.0383 | 0.0159 | 0.0139 | 0.0071 | 0.0042 | 0.0040 |
| 0 0000.0 0000 | 0 0000.0 | 0 | 0000 | 0.0000 | 0.0000 | 0.0000 | 0.0160 | 0.0160 | 0.0560 | 0.0880 | 0.1040 | 0.0720 | 0.0320 | 0.0480 | 0.0080 | 0.0000 | 0.0000 | 0.0000 |
| 0000 0.0007 0 | 0.0007 0 | 0 | 0000 | 0.0036 | 0.0072 | 0.0181 | 0.0296 | 0.0557 | 0.0701 | 0.0940 | 0.1048 | 0.0745 | 0.0419 | 0.0268 | 0.0166 | 0.0043 | 0.0029 | 0.0007 |
| 000 0.0016 0 | 0.0016 0 | 0 | .0008 | 0.0033 | 0.0025 | 0.0074 | 0.0107 | 0.0230 | 0.0607 | 0.1066 | 0.1025 | 0.0968 | 0.0648 | 0.0320 | 0.0098 | 0.0074 | 0.0057 | 0.0025 |
| 000 0.0016 0 | 0.0016 0 | \circ | 0000. | 0.0031 | 0.0039 | 0.0062 | 0.0148 | 0.0349 | 0.0745 | 0.1017 | 0.1071 | 0.0668 | 0.0404 | 0.0264 | 0.0062 | 0.0078 | 0.0016 | 0.0023 |
| 000 0.0004 (|).0004 (| $\overline{}$ | 0.0004 | 0.0024 | 0.0076 | 0.0112 | 0.0239 | 0.0418 | 0.0737 | 0.0961 | 0.1419 | 0.1172 | 0.0697 | 0.0271 | 0.0175 | 0.0032 | 0.0008 | 0.0008 |
| 0000.0 0000 | 0000.0 | | 0.0000 | 0.0014 | 0.0043 | 0.0086 | 0.0286 | 0.0328 | 0.0600 | 0.0785 | 0.0871 | 0.0749 | 0.0500 | 0.0350 | 0.0086 | 0.0071 | 0.0000 | 0.0014 |
| 0000.0 0000 | 0000.0 | - | 0.0006 | 0.0006 | 0.0035 | 0.0052 | 0.0163 | 0.0372 | 0.0837 | 0.1041 | 0.0971 | 0.0919 | 0.0401 | 0.0174 | 0.0035 | 0.0023 | 0.0041 | 0.0029 |
| 0000.0 0000 | 0000.0 | $\overline{}$ | 0.0005 | 0.0027 | 0.0033 | 0.0082 | 0.0148 | 0.0477 | 0.0987 | 0.1240 | 0.1064 | 0.0724 | 0.0384 | 0.0176 | 0.0049 | 0.0016 | 0.0011 | 0.0005 |
| 000 0.0012 | 0.0012 | - | 0.0037 | 0.0074 | 0.0117 | 0.0154 | 0.0259 | 0.0376 | 0.0733 | 0.0998 | 0.0844 | 0.0776 | 0.0462 | 0.0179 | 0.0117 | 0.0074 | 0.0055 | 0.0025 |
| 0004 0.0000 | 0000.(| | 0.0034 | 0.0025 | 0.0038 | 0.0096 | 0.0204 | 0.0396 | 0.0480 | 0.1102 | 0.1234 | 0.0899 | 0.0279 | 0.0151 | 0.0042 | 0.0015 | 0.0019 | 0.0001 |
| 0000 0.0000 | 0000.0 | \circ | 0.0026 | 0.0042 | 0.0092 | 0.0149 | 0.0304 | 0.0389 | 0.0658 | 0.1130 | 0.1223 | 0.0835 | 0.0278 | 0.0155 | 0.0087 | 0.0000 | 0.0011 | 0.0001 |

| | | Size com | positions | |
|------|-------|----------|-----------|-------|
| year | | total | o - | _ |
| | hauls | indiv.s | females | males |
| 1990 | 3 | 274 | 65 | 84 |
| 1991 | 48 | 4301 | 1213 | 1418 |
| 1992 | 77 | 4958 | 1650 | 2034 |
| 1993 | 55 | 4801 | 1425 | 2140 |
| 1994 | 56 | 4089 | 1175 | 930 |
| 1995 | 46 | 2818 | 1280 | 1301 |
| 1996 | 174 | 11207 | 2297 | 2330 |
| 1997 | 72 | 4827 | 1926 | 2113 |
| 1998 | 128 | 6509 | 2569 | 2896 |
| 1999 | 7 | 130 | 70 | 55 |
| 2000 | 111 | 1464 | 667 | 796 |
| 2001 | 80 | 1446 | 664 | 757 |
| 2002 | 86 | 1326 | 645 | 643 |
| 2003 | 168 | 2592 | 920 | 1609 |
| 2004 | 79 | 1590 | 816 | 765 |
| 2005 | 118 | 1838 | 882 | 947 |
| 2006 | 124 | 1872 | 835 | 990 |
| 2007 | 122 | 1830 | 840 | 985 |
| 2008 | 100 | 1628 | 815 | 798 |
| 2009 | 72 | 1249 | 556 | 679 |

Table 8.4a. Sample sizes the domestic fishery.

Table 8.4 b. Sample sizes the groundfish survey.

| | biomass | | Size com | positions | | | Age com | positions | |
|------|-------------|-------|----------|-----------|-------|-------|---------|-----------|-------|
| year | | | total | | | | total | | |
| | total hauls | hauls | indiv.s | females | males | hauls | indiv.s | females | males |
| 1984 | 929 | 264 | 25316 | 13875 | 11291 | | 653 | 369 | 284 |
| 1987 | 783 | 197 | 27298 | 15931 | 11350 | | | | |
| 1990 | 708 | 286 | 24322 | 12939 | 11255 | 22 | 247 | 138 | 107 |
| 1993 | 775 | 364 | 26124 | 13592 | 12294 | 36 | 312 | 179 | 132 |
| 1996 | 807 | 417 | 21416 | 11086 | 9975 | 55 | 528 | 285 | 243 |
| 1999 | 764 | 389 | 16052 | 7941 | 8023 | 47 | 605 | 316 | 288 |
| 2001 | 489 | 245 | 11877 | 5962 | 5899 | | | | |
| 2003 | 809 | 434 | 25885 | 13279 | 12479 | 87 | 499 | 249 | 250 |
| 2005 | 839 | 413 | 23499 | 12501 | 10907 | 76 | 551 | 296 | 243 |
| 2007 | 820 | 411 | 25539 | 13563 | 11860 | 82 | 755 | 410 | 345 |
| 2009 | 823 | 454 | 21010 | 10304 | 10360 | | | | |

Table 8.5. Biomass estimates (t) by NPFMC regulatory area for GOA flathead sole from the NMFS bottom trawl surveys. Note that in 2001 the eastern GOA was not surveyed. This was accounted for in the assessment model by assuming, based on previous surveys, that availability for this year was was 0.9. The maximum depth stratum included in each survey is also noted.

| Year | Western Gulf | Central Gulf | West Yakutat | Southeast | Total | Std. Dev | Max Depth (m) |
|------|-----------------|--------------|-----------------|-----------|---------|----------|------------------|
| 1984 | 45,100 | 158,539 | 45,694 | 9 | 249,341 | 30,355 | 1000 |
| 1987 | 33,603 | 113,483 | 30,455 | 5 | 177,546 | 18,956 | 1000 |
| 1990 | 58,740 | 161,257 | 23,019 | 40 | 243,055 | 28,877 | 500 |
| 1993 | 57,871 | 113,976 | 16,720 | 124 | 188,690 | 24,486 | 500 |
| 1996 | 66,732 | 122,730 | 12,751 | 3,308 | 205,521 | 18,430 | 500 |
| 1999 | 49,636 | 139,356 | 15,115 | 3,482 | 207,590 | 24,404 | 1000 |
| 2001 | 68,164 | 85,430 | | | 153,594 | 18,300 | 500 |
| 2003 | 67,055 | 170,852 | 17,154 | 2,234 | 257,294 | 19,913 | 700 |
| 2005 | 59,458 | 142,043 | 11,400 | 312 | 213,213 | 16,944 | 1000 |
| 2007 | 78,361 | 176,529 | 21,430 | 3,970 | 280,290 | 23,778 | 1000 |
| 2009 | 80,115 | 128,910 | 9,458 | 6,894 | 225,377 | 25,041 | 1000 |

Table 8.6. Biomass estimates (t) by depth stratum for GOA flathead sole from the NMFS bottom trawl surveys. Note that in 2001 the eastern GOA was not surveyed.

| VOOR | | Dep | th range (m |) | |
|------|---------|---------|-------------|---------|------|
| year | 1-100 | 101-200 | 201-300 | 301-500 | >500 |
| 1984 | 118,974 | 121,791 | 8,571 | 5 | 0 |
| 1987 | 91,482 | 75,475 | 10,553 | 36 | 0 |
| 1990 | 157,014 | 76,306 | 9,713 | 22 | |
| 1993 | 113,072 | 65,143 | 10,278 | 198 | |
| 1996 | 119,657 | 78,545 | 7,270 | 50 | |
| 1999 | 145,347 | 58,641 | 3,581 | 14 | 8 |
| 2001 | 93,433 | 56,133 | 4,006 | 22 | |
| 2003 | 146,018 | 101,421 | 9,855 | 0 | 0 |
| 2005 | 114,895 | 92,869 | 5,297 | 151 | 0 |
| 2007 | 139,806 | 130,661 | 9,823 | 0 | 0 |
| 2009 | 138,824 | 80,395 | 6,157 | 0 | 0 |

| a) Fema | ales. | | | | | | | | | | | | | |
|---------|-------------------------|-----------------|--------|--------|--------|--------|--------|--------|--------|---------|--------|--------|--------|--------|
| year | Length bin cutpoi 14 | ints (cm) 16 | 18 | 20 | | 74 | 36 | 38 | 08 | čt | ۶t | УL | 38 | 40 |
| 1984 | 567 | 3,098 | 3,337 | 7,306 | 14,170 | 20,489 | 29,800 | 45,645 | 63,475 | 76,302 | 69,592 | 48,288 | 28,087 | 17,406 |
| 1987 | 883 | 2,085 | 2,838 | 4,499 | 6,056 | 12,189 | 17,593 | 25,161 | 36,188 | 50,279 | 54,302 | 45,327 | 26,573 | 14,468 |
| 1990 | 1,269 | 3,347 | 6,036 | 6,002 | 9,283 | 15,446 | 19,887 | 24,583 | 37,464 | 46,874 | 55,347 | 60,532 | 52,045 | 30,967 |
| 1993 | 2,584 | 5,010 | 7,744 | 12,730 | 13,216 | 15,423 | 20,372 | 21,820 | 25,208 | 31,170 | 41,270 | 43,396 | 36,634 | 23,356 |
| 1996 | 3,360 | 6,318 | 10,043 | 14,294 | 16,104 | 19,497 | 21,345 | 25,059 | 29,741 | 34,375 | 37,894 | 40,168 | 33,867 | 23,395 |
| 1999 | 2,895 | 2,940 | 5,795 | 8,718 | 10,041 | 16,134 | 18,845 | 21,287 | 25,158 | 28,741 | 31,886 | 35,669 | 31,739 | 27,829 |
| 2001 | 2,777 | 4,699 | 5,728 | 8,070 | 9,822 | 7,348 | 9,242 | 12,441 | 17,973 | 20,460 | 29,033 | 26,925 | 24,106 | 18,520 |
| 2003 | 3,066 | 6,647 | 10,771 | 15,271 | 20,514 | 25,006 | 23,932 | 24,520 | 28,685 | 35,373 | 46,891 | 47,205 | 42,591 | 35,524 |
| 2005 | 4,988 | 7,391 | 10,305 | 14,894 | 20,011 | 22,229 | 27,086 | 30,483 | 33,432 | 38,116 | 37,285 | 35,590 | 34,358 | 24,141 |
| 2007 | 2,429 | 6,105 | 10,258 | 20,784 | 19,669 | 18,962 | 23,767 | 25,095 | 35,366 | 40,488 | 50,423 | 51,276 | 44,433 | 33,155 |
| 2009 | 4,488 | 3,880 | 7,286 | 10,748 | 14,502 | 16,873 | 24,109 | 28,080 | 35,058 | 28,175 | 38,788 | 43,340 | 39,670 | 24,950 |
| b) Male | S. | | | | | | | | | | | | | |
| VOGF | Length bin cutpoi | ints (cm) | | | | | | | | | | | | |
| ycar | 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 |
| 1984 | 958 | 2,651 | 3,872 | 10,794 | 19,758 | 34,522 | 54,303 | 81,720 | 76,269 | 40,785 | 19,368 | 10,317 | 5,446 | 1,990 |
| 1987 | 1,257 | 2,623 | 3,876 | 5,448 | 6,445 | 15,112 | 26,507 | 48,973 | 53,591 | 33 ,389 | 14,119 | 7,290 | 3,550 | 1,731 |
| 1990 | 1,061 | 4,055 | 5,883 | 8,099 | 11,657 | 19,990 | 29,710 | 45,839 | 65,958 | 73,288 | 42,626 | 12,664 | 3,977 | 850 |
| 1993 | 2,205 | 5,315 | 9,757 | 12,897 | 16,987 | 23,213 | 29,095 | 39,372 | 50,735 | 54,631 | 36,488 | 12,636 | 5,513 | 2,599 |
| 1996 | 4,039 | 6,250 | 9,608 | 14,129 | 18,421 | 22,021 | 27,807 | 37,472 | 49,772 | 52,356 | 41,352 | 17,459 | 5,026 | 1,607 |
| 1999 | 2,484 | 4,313 | 7,246 | 11,893 | 17,227 | 21,067 | 30,364 | 42,405 | 59,243 | 60,992 | 49,672 | 24,469 | 7,967 | 1,647 |
| 2001 | 2,519 | 5,015 | 7,128 | 8,810 | 10,981 | 13,831 | 17,031 | 27,453 | 37,617 | 39,651 | 36,558 | 19,205 | 6,125 | 2,013 |
| 2003 | 4,634 | 6,574 | 11,065 | 17,329 | 24,994 | 31,230 | 36,233 | 41,029 | 54,997 | 57,972 | 53,126 | 33,017 | 14,061 | 4,857 |
| 2005 | 4,727 | 7,283 | 12,201 | 15,830 | 23,301 | 33,863 | 45,026 | 49,439 | 52,297 | 49,895 | 37,689 | 24,343 | 9,653 | 2,244 |
| 2007 | 4,193 | 6,756 | 13,904 | 23,942 | 25,572 | 25,987 | 33,840 | 43,611 | 53,832 | 57,060 | 48,576 | 34,409 | 14,457 | 5,924 |
| 2009 | 3,558 | 4,350 | 8,914 | 16,447 | 23,573 | 31,578 | 41,192 | 46,564 | 52,118 | 56,359 | 47,310 | 28,938 | 11,366 | 3,234 |

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| $ \begin{array}{cccccccccccccccccccccccccccccccccccc$ | 732 | 6,984 | 36,300 | 74,155 | 104,300 | 74,810 | 47,661 | 24,199 | 24,848 | 4,627 | 2,992 | 0 | 0 | 0 | 0 | 0 |
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| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | .746 | 18,705 | 18,484 | 22,728 | 23,396 | 24,017 | 49,392 | 25,997 | 22,142 | 19,556 | 17,817 | 4,674 | 10,333 | 10,345 | 5,432 | 758 |
| $\begin{array}{cccccccccccccccccccccccccccccccccccc$ | .180 | 36,747 | 26,716 | 45,246 | 32,697 | 20,360 | 22,297 | 24,929 | 16,811 | 17,244 | 14,740 | 6,557 | 12,507 | 2,794 | 4,049 | 803 |
| $ \begin{array}{cccccccccccccccccccccccccccccccccccc$ | 00 | 3 23,768 | 20,784 | 21,354 | 37,554 | 32,344 | 30,160 | 22,069 | 19,428 | 20,935 | 9,843 | 11,778 | 7,231 | 6,379 | 7,171 | 1,791 |
| 9 48,192 56,383 33,181 23,400 32,891 24,245 16,342 14,216 9,983 10,575 9,960 2 39,390 31,793 27,646 56,536 51,835 24,194 28,467 23,274 12,986 28,465 12,683 2 6 7 8 9 10 11 12 13 14 15 3 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 3 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 3 29,810 33,483 38,549 35,774 13,969 12,160 26,810 24,797 17,194 0 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 3 26,579 38,037 44,421 34,330 31,534 16,454 9,247 6,710 <td< td=""><td>59</td><td>2 50,233</td><td>52,481</td><td>13,806</td><td>37,912</td><td>43,306</td><td>50,772</td><td>16,791</td><td>14,290</td><td>10,785</td><td>24,386</td><td>3,205</td><td>2,332</td><td>382</td><td>4,405</td><td>4,587</td></td<> | 59 | 2 50,233 | 52,481 | 13,806 | 37,912 | 43,306 | 50,772 | 16,791 | 14,290 | 10,785 | 24,386 | 3,205 | 2,332 | 382 | 4,405 | 4,587 |
| 2 39,390 31,793 27,646 56,536 51,835 24,194 28,467 23,274 12,986 28,465 12,683 5 6 7 8 9 10 11 12 13 14 15 3 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 3 29,810 33,483 28,696 36,881 38,549 35,774 13,969 12,160 26,810 24,797 17,194 0 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 8 30,360 23,837 44,421 34,830 34,399 31,534 16,454 9,247 6,710 6,140 6,892 | F | 9 48,192 | 56,383 | 33,181 | 23,400 | 32,891 | 24,245 | 16,342 | 14,216 | 9,983 | 10,575 | 9,960 | 4,152 | 5,346 | 2,470 | 1,387 |
| 5 6 7 8 9 10 11 12 13 14 15 3 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 3 29,810 33,483 28,696 36,881 38,549 35,774 13,969 12,160 26,810 24,797 17,194 0 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,469 14,387 13,195 6,993 30,360 23,837 44,421 34,330 31,534 16,454 9,247 6,710 6,140 6,892 | 78 | 2 39,390 | 31,793 | 27,646 | 56,536 | 51,835 | 24,194 | 28,467 | 23,274 | 12,986 | 28,465 | 12,683 | 6,019 | 12,586 | 4,512 | 1,923 |
| 5 6 7 8 9 10 11 12 13 14 15 33 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 33 29,810 33,483 28,696 36,881 38,549 35,774 13,069 12,160 26,810 24,797 17,194 00 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 78 30,360 23,837 44,421 34,339 31,534 16,454 9,247 6,710 6,140 6,892 | | | | | | | | | | | | | | | | |
| 5 6 7 8 9 10 11 12 13 14 15 03 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 83 29,810 33,483 28,696 36,881 38,549 35,774 13,969 12,160 26,810 24,797 17,194 00 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 78 30,360 23,837 44,421 34,330 31,534 16,454 9,247 6,710 6,140 6,892 | | | | | | | | | | | | | | | | |
| 5 6 7 8 9 10 11 12 13 14 15 103 $29,061$ $41,741$ $48,344$ $96,634$ $61,205$ $16,899$ $21,343$ $9,159$ $1,421$ $4,745$ $2,773$ 583 $29,810$ $33,483$ $28,696$ $36,881$ $38,549$ $35,774$ $13,969$ $12,160$ $26,797$ $17,194$ 500 $26,579$ $38,034$ $21,547$ $48,187$ $19,3711$ $15,327$ $30,706$ $12,449$ $14,387$ $13,195$ $6,993$ 78 $30,360$ $23,837$ $44,421$ $34,399$ $31,534$ $16,454$ $9,247$ $6,140$ $6,140$ $6,892$ | | | | | | | | | | | | | | | | |
| 03 29,061 41,741 48,344 96,634 61,205 16,899 21,343 9,159 1,421 4,745 2,773 83 29,810 33,483 28,696 36,881 38,549 35,774 13,969 12,160 26,810 24,797 17,194 00 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 78 30,360 23,837 44,421 34,830 34,399 31,534 16,454 9,247 6,710 6,140 6,892 | | 5 | 9 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 |
| 83 29,810 33,483 28,696 36,881 38,549 35,774 13,969 12,160 26,810 24,797 17,194 00 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 78 30,360 23,837 44,421 34,830 34,399 31,534 16,454 9,247 6,710 6,140 6,892 | 4 | 03 29,061 | 41,741 | 48,344 | 96,634 | 61,205 | 16,899 | 21,343 | 9,159 | 1,421 | 4,745 | 2,773 | 0 | 0 | 0 | 0 |
| 00 26,579 38,034 21,547 48,187 19,371 15,327 30,706 12,449 14,387 13,195 6,993 78 30,360 23,837 44,421 34,830 34,399 31,534 16,454 9,247 6,710 6,140 6,892 | œ. | 83 29,810 | 33,483 | 28,696 | 36,881 | 38,549 | 35,774 | 13,969 | 12,160 | 26,810 | 24,797 | 17,194 | 4,452 | 0 | 3,360 | 0 |
| 78 30,360 23,837 44,421 34,830 34,399 31,534 16,454 9,247 6,710 6,140 6,892 | ~ | 300 26,579 | 38,034 | 21,547 | 48,187 | 19,371 | 15,327 | 30,706 | 12,449 | 14,387 | 13,195 | 6,993 | 5,117 | 1,902 | 0 | 68 |
| | 0 | 78 30,360 | 23,837 | 44,421 | 34,830 | 34,399 | 31,534 | 16,454 | 9,247 | 6,710 | 6,140 | 6,892 | 3,200 | 2,905 | 232 | 1,202 |
| 33 31,988 29,956 21,892 61,304 44,990 33,109 26,041 22,030 10,088 13,624 4,753 | 17 | 33 31,988 | 29,956 | 21,892 | 61,304 | 44,990 | 33,109 | 26,041 | 22,030 | 10,088 | 13,624 | 4,753 | 1,572 | 7,129 | 3,766 | 1,975 |
| 90 64,911 68,289 28,709 16,977 39,693 21,243 18,447 5,498 10,919 3,074 3,654 | 0 | 90 64,911 | 68,289 | 28,709 | 16,977 | 39,693 | 21,243 | 18,447 | 5,498 | 10,919 | 3,074 | 3,654 | 1,189 | 3,116 | 3,308 | 4,701 |
| 35 39,610 56,586 60,672 38,238 22,515 14,721 15,575 3,836 14,354 10,745 1,379 | | 35 39,610 | 56,586 | 60,672 | 38,238 | 22,515 | 14,721 | 15,575 | 3,836 | 14,354 | 10,745 | 1,379 | 6,296 | 1,724 | 2,006 | 2,560 |
| 02 42,624 50,946 38,777 53,200 51,102 11,551 9,996 12,398 8,488 20,029 5,047 | ล | 02 42,624 | 50,946 | 38,777 | 53,200 | 51,102 | 11,551 | 966,6 | 12,398 | 8,488 | 20,029 | 5,047 | 8,021 | 7,169 | 4,120 | 0 |

Table 8.8. Survey age compositions for flathead sole. Numbers are in 1000's of individuals.

Table 8.9a. Age-length transition matrices for female flathead sole. Values at a row/column combination correspond to the fraction of individuals at the age indicated by the row that fall into the length group indicated by the column.

| | length cutpt. | s (cm) | | | | | | | | | | | | | | | |
|----|----------------|--------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|
| ge | 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 |
| ., | 3 0.154 | 0.396 | 0.348 | 0.094 | 0.008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0.004 | 0.038 | 0.175 | 0.348 | 0.302 | 0.113 | 0.018 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| ~, | 0 | 0.002 | 0.018 | 0.09 | 0.235 | 0.321 | 0.229 | 0.086 | 0.017 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| | 0 9 | 0 | 0.002 | 0.013 | 0.063 | 0.174 | 0.281 | 0.265 | 0.146 | 0.047 | 0.009 | 0.001 | 0 | 0 | 0 | 0 | 0 |
| | 0 | 0 | 0 | 0.002 | 0.013 | 0.056 | 0.15 | 0.251 | 0.263 | 0.173 | 0.071 | 0.018 | 0.003 | 0 | 0 | 0 | 0 |
| ~ | 0 | 0 | 0 | 0 | 0.003 | 0.016 | 0.061 | 0.148 | 0.237 | 0.251 | 0.175 | 0.08 | 0.024 | 0.005 | 0.001 | 0 | 0 |
| Ο, | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.023 | 0.074 | 0.159 | 0.235 | 0.236 | 0.162 | 0.075 | 0.024 | 0.005 | 0.001 | 0 |
| 7 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.035 | 0.095 | 0.179 | 0.236 | 0.217 | 0.139 | 0.063 | 0.02 | 0.004 | 0.001 |
| - | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.004 | 0.017 | 0.055 | 0.125 | 0.202 | 0.233 | 0.19 | 0.111 | 0.046 | 0.013 | 0.003 |
| 1. | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.032 | 0.085 | 0.161 | 0.221 | 0.218 | 0.155 | 0.079 | 0.029 | 0.008 |
| ÷ | 3 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.019 | 0.057 | 0.125 | 0.196 | 0.225 | 0.188 | 0.114 | 0.05 | 0.016 |
| 1 | • | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.012 | 0.04 | 0.096 | 0.169 | 0.219 | 0.208 | 0.144 | 0.073 | 0.027 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.008 | 0.028 | 0.074 | 0.144 | 0.206 | 0.217 | 0.169 | 0.096 | 0.04 |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.02 | 0.058 | 0.122 | 0.191 | 0.22 | 0.187 | 0.117 | 0.054 |
| 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.004 | 0.015 | 0.046 | 0.104 | 0.175 | 0.218 | 0.2 | 0.136 | 0.068 |
| 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.011 | 0.037 | 0.09 | 0.161 | 0.213 | 0.209 | 0.151 | 0.081 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.03 | 0.078 | 0.148 | 0.207 | 0.215 | 0.164 | 0.093 |
| 2(| 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.025 | 0.069 | 0.137 | 0.201 | 0.218 | 0.175 | 0.103 |

Table 8.9b. Age-length transition matrices for male flathead sole. Values at a row/column combination correspond to the fraction of individuals at the age indicated by the row that fall into the length group indicated by the column.

| | length cutpt: | s (cm) | | | | | | | | | | | | | | | |
|-----|---------------|---------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|-------|----|
| age | 14 | , 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 |
| ñ | 0.706 | 0.265 | 0.029 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.038 | 0.205 | 0.405 | 0.279 | 0.067 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 | 0.001 | 0.015 | 0.095 | 0.274 | 0.353 | 0.204 | 0.052 | 0.006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0.001 | 0.011 | 0.067 | 0.205 | 0.323 | 0.26 | 0.108 | 0.023 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0.001 | 0.013 | 0.066 | 0.188 | 0.298 | 0.263 | 0.13 | 0.036 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 |
| 80 | 0 | 0 | 0 | 0.003 | 0.019 | 0.081 | 0.199 | 0.288 | 0.245 | 0.122 | 0.036 | 0.006 | 0.001 | 0 | 0 | 0 | 0 |
| 0 | 0 | 0 | 0 | 0.001 | 0.006 | 0.033 | 0.112 | 0.227 | 0.281 | 0.212 | 0.097 | 0.027 | 0.005 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0.002 | 0.014 | 0.06 | 0.158 | 0.257 | 0.261 | 0.165 | 0.065 | 0.016 | 0.002 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.033 | 0.106 | 0.213 | 0.271 | 0.217 | 0.11 | 0.035 | 0.007 | 0.001 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.019 | 0.071 | 0.17 | 0.258 | 0.248 | 0.153 | 0.06 | 0.015 | 0.002 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.012 | 0.049 | 0.135 | 0.236 | 0.263 | 0.188 | 0.086 | 0.025 | 0.005 | 0.001 | 0 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.035 | 0.108 | 0.212 | 0.267 | 0.214 | 0.11 | 0.036 | 0.008 | 0.001 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.026 | 0.088 | 0.191 | 0.264 | 0.234 | 0.132 | 0.047 | 0.011 | 0.002 | 0 |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.019 | 0.072 | 0.172 | 0.259 | 0.248 | 0.151 | 0.058 | 0.014 | 0.002 | 0 |
| 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.015 | 0.061 | 0.156 | 0.253 | 0.258 | 0.167 | 0.068 | 0.017 | 0.003 | 0 |
| 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.012 | 0.052 | 0.143 | 0.246 | 0.266 | 0.18 | 0.076 | 0.02 | 0.003 | 0 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.009 | 0.045 | 0.132 | 0.241 | 0.272 | 0.191 | 0.083 | 0.022 | 0.004 | 0 |
| 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.008 | 0.039 | 0.122 | 0.235 | 0.278 | 0.201 | 0.089 | 0.024 | 0.004 | 0 |

| | Lengt | h (cm) | Weigh | nt (kg) | Maturity |
|-----|-------|---------|-------|---------|----------|
| Age | Males | Females | Males | Females | ogive |
| 3 | 15.2 | 17.8 | 0.03 | 0.05 | 0.0117 |
| 4 | 19.3 | 21.6 | 0.06 | 0.09 | 0.0251 |
| 5 | 22.6 | 24.9 | 0.10 | 0.14 | 0.0527 |
| 6 | 25.3 | 27.8 | 0.15 | 0.20 | 0.1076 |
| 7 | 27.6 | 30.2 | 0.20 | 0.26 | 0.2072 |
| 8 | 29.4 | 32.2 | 0.24 | 0.32 | 0.3615 |
| 9 | 30.8 | 34.0 | 0.28 | 0.38 | 0.5508 |
| 10 | 32.0 | 35.5 | 0.32 | 0.44 | 0.7265 |
| 11 | 33.0 | 36.8 | 0.35 | 0.49 | 0.8520 |
| 12 | 33.8 | 37.9 | 0.37 | 0.54 | 0.9257 |
| 13 | 34.5 | 38.8 | 0.40 | 0.58 | 0.9643 |
| 14 | 35.0 | 39.6 | 0.42 | 0.62 | 0.9832 |
| 15 | 35.4 | 40.3 | 0.43 | 0.66 | 0.9922 |
| 16 | 35.8 | 40.9 | 0.45 | 0.69 | 0.9964 |
| 17 | 36.1 | 41.4 | 0.46 | 0.72 | 0.9983 |
| 18 | 36.3 | 41.8 | 0.47 | 0.74 | 0.9992 |
| 19 | 36.5 | 42.2 | 0.48 | 0.76 | 0.9996 |
| 20 | 36.7 | 42.5 | 0.48 | 0.83 | 0.9998 |

Table 8.10. Age-specific schedules for flathead sole in the Gulf of Alaska. Maturity ogive is based on Stark (2004).

| | Fishery | | | Survey | | | | I | Recruitmen | t | |
|--------------------|-----------------------|--------------|------------|------------------|-----------------|--------|-----|--------|------------|--------|--------|
| catch | length composition | biomass | le comp | ngth ositions | age composit | ions | | early | ordinary | late | |
| 30 | 1 | 1 | | 1 | 1 | | | 1 | 1 | 1 | |
| Table 8.1 | 2. Initial paran | neter values | for all | model ca | ises. | | | | | | |
| $\ln R_{\circ}$ | 17 | | | | | | | | | | |
| τ_t | | | | | 19 | 967-20 | 07: | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| - | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | |
| $\overline{\ln F}$ | 0 | | | | | | | | | | |
| ε, | | 198 | 34-2007: | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.00 | 000 | 0.0000 | 0.0000 | 0.0000 | |
| Fishery Sele | ectivity | | | | | | | | | | |
| · | females | males | | | | | | | | | |
| slope | 0.4 | 0.4 | | | | | | | | | |
| A ₅₀ | 5 | 5 | | | | | | | | | |
| scale par. | | 0 | | | | | | | | | |
| Survey Sele | ctivity | | | | | | | | | | |
| | females | males | | | | | | | | | |
| slope | 0.8 | 0.4 | | | | | | | | | |
| A ₅₀ | 4 | 4 | | | | | | | | | |
| scale par. | | 0 | | | | | | | | | |

| | Table 8.11. | Likelihood r | nultiplier | settings f | or all | model | cases. |
|--|-------------|--------------|------------|------------|--------|-------|--------|
|--|-------------|--------------|------------|------------|--------|-------|--------|

Table 8.13. Comparison of likelihood components for the base case and Alternative 1 models. Highlighted values are at least 0.5 log-likelihood units larger than the corresponding component from the other model.

| likelihood | Case | | | | |
|--------------------------|-----------|---------------|--|--|--|
| component | base | Alternative 1 | | | |
| ordinary recruitment | 15.737 | 15.7209 | | | |
| "late" recruitment | 0.343677 | 0.356455 | | | |
| "early" recruitment | 12.4452 | 12.3782 | | | |
| fishery catch | 0.0175164 | 0.0170343 | | | |
| fishery size composition | 545.242 | 541.772 | | | |
| survey biomass | 14.245 | 14.6922 | | | |
| survey size composition | 34.0818 | 33.1606 | | | |
| survey age composition | 147.314 | 142.214 | | | |

| $\ln R_{o}$ | 18.413415 | | | | | | | | | |
|--------------------|-----------|--------------|------------|---------|---------|------------|---------|---------|---------|---------|
| τ_t | | | | | 1 | 1967-2009: | -1.6349 | -0.8287 | -0.8996 | -0.9693 |
| | -1.0271 | -1.0748 | -1.1125 | -1.1407 | -0.4489 | -0.8817 | -0.0217 | 0.3868 | 0.9026 | 0.7113 |
| | 0.1486 | 0.1754 | 0.1344 | 0.1698 | 0.1793 | 0.0248 | 0.1004 | 0.3508 | 0.0913 | 0.1693 |
| | 0.2070 | 0.4533 | 0.1653 | 0.3409 | 0.1858 | 0.2231 | 0.1685 | -0.0546 | -0.0970 | 0.3280 |
| | 0.6279 | 0.7322 | 0.4791 | 0.6547 | 0.6061 | 0.7284 | 0.5712 | 0.1738 | 0.0014 | |
| $\overline{\ln F}$ | -4.515926 | | | | | | | | | |
| ε, | | | 1984-2009: | -0.2622 | -0.9900 | -1.9309 | -2.0551 | -0.9499 | -0.6492 | -0.0260 |
| - 1 | 0.1327 | 0.2953 | 0.4449 | 0.5319 | 0.3657 | 0.7022 | 0.4683 | 0.1314 | -0.5235 | -0.0146 |
| | 0.1825 | 0.2941 | 0.4177 | 0.4161 | 0.4742 | 0.6835 | 0.6802 | 0.7209 | 0.4595 | |
| Fishery Sele | ectivity | | | | | | | | | |
| | females | males | | | | | | | | |
| slope | 0.9206304 | 10.17185706 | | | | | | | | |
| A ₅₀ | 9.761787 | 0 | | | | | | | | |
| scale par. | | 0.158825192 | | | | | | | | |
| Survey Sele | ctivity | | | | | | | | | |
| | females | males | | | | | | | | |
| slope | 0.6389232 | 0.912693831 | | | | | | | | |
| A_{50} | 6.1550805 | 4.997881639 | | | | | | | | |
| scale par. | | -0.235182306 | | | | | | | | |

Table 8.14. Final parameter estimates from the preferred model (Alternative 1 Model).

| NOOR | | catch (t) | | survey biomass (t) | | | | |
|------|-----------|-----------|----------|--------------------|---------|----------|--|--|
| year | estimated | std dev | observed | estimated | std dev | observed | | |
| 1984 | 556 | 71 | 549 | 165,550 | 10,165 | 249,341 | | |
| 1985 | 331 | 42 | 320 | 180,350 | 10,260 | | | |
| 1986 | 157 | 20 | 147 | 191,230 | 10,198 | I | | |
| 1987 | 162 | 21 | 151 | 198,900 | 10,050 | 177,546 | | |
| 1988 | 538 | 68 | 520 | 204,160 | 9,850 | | | |
| 1989 | 767 | 98 | 747 | 207,340 | 9,615 | | | |
| 1990 | 1,466 | 187 | 1,447 | 209,470 | 9,364 | 243,055 | | |
| 1991 | 1,733 | 222 | 1,717 | 210,590 | 9,104 | | | |
| 1992 | 2,040 | 261 | 2,034 | 211,600 | 8,850 | | | |
| 1993 | 2,358 | 301 | 2,366 | 212,450 | 8,605 | 188,690 | | |
| 1994 | 2,560 | 327 | 2,580 | 213,740 | 8,392 | | | |
| 1995 | 2,166 | 276 | 2,181 | 215,340 | 8,207 | | | |
| 1996 | 3,035 | 386 | 3,107 | 217,530 | 8,050 | 205,521 | | |
| 1997 | 2,406 | 306 | 2,446 | 218,500 | 7,908 | ł | | |
| 1998 | 1,733 | 220 | 1,742 | 219,180 | 7,776 | | | |
| 1999 | 916 | 116 | 900 | 219,370 | 7,655 | 207,590 | | |
| 2000 | 1,547 | 197 | 1,547 | 219,200 | 7,551 | | | |
| 2001 | 1,900 | 241 | 1,911 | 196,270 | 6,735 | 170,745 | | |
| 2002 | 2,127 | 270 | 2,145 | 217,860 | 7,485 | İ | | |
| 2003 | 2,392 | 304 | 2,425 | 219,990 | 7,608 | 257,294 | | |
| 2004 | 2,362 | 300 | 2,390 | 225,360 | 7,954 | | | |
| 2005 | 2,477 | 314 | 2,530 | 233,510 | 8,597 | 213,221 | | |
| 2006 | 3,045 | 386 | 3,134 | 243,230 | 9,569 | | | |
| 2007 | 3,085 | 391 | 3,163 | 253,140 | 10,883 | 280,290 | | |
| 2008 | 3,335 | 423 | 3,419 | 262,590 | 12,552 | | | |
| 2009 | 2,698 | 343 | 2,740 | 270,060 | 14,573 | 225,377 | | |

Table 8.15. Estimated catch and survey biomass from the preferred model.

| | | Age 3+ Biomass (1000's t) | | | | | | Female Spawning Stock Biomass (1000's t) | | | | |
|------|------------|---------------------------|-----------|---------|-----------|---------|-----------|--|-----------|---------|-----------|---------|
| year | 2009 Asses | ssment | 2007 Asse | ssment | 2005 Asse | ssment | 2009 Asse | essment | 2007 Asse | ssment | 2005 Asse | ssment |
| | mean | std dev | mean | std dev | mean | std dev | mean | std dev | mean | std dev | mean | std dev |
| 1984 | 246 | 15 | 244 | 13 | 248 | 13 | 59 | 4 | 61 | 4 | 65 | 4 |
| 1985 | 259 | 15 | 254 | 13 | 256 | 13 | 70 | 5 | 73 | 4 | 76 | 4 |
| 1986 | 267 | 15 | 262 | 13 | 263 | 13 | 81 | 5 | 83 | 5 | 85 | 5 |
| 1987 | 273 | 15 | 266 | 12 | 266 | 12 | 89 | 5 | 90 | 5 | 91 | 5 |
| 1988 | 279 | 15 | 271 | 12 | 270 | 12 | 95 | 6 | 94 | 5 | 95 | 5 |
| 1989 | 283 | 15 | 274 | 12 | 271 | 12 | 98 | 6 | 96 | 5 | 97 | 5 |
| 1990 | 285 | 15 | 275 | 12 | 271 | 11 | 100 | 6 | 97 | 5 | 97 | 5 |
| 1991 | 287 | 15 | 276 | 11 | 271 | 11 | 101 | 6 | 97 | 5 | 97 | 4 |
| 1992 | 290 | 15 | 280 | 11 | 274 | 11 | 101 | 5 | 97 | 4 | 97 | 4 |
| 1993 | 292 | 15 | 281 | 11 | 273 | 11 | 101 | 5 | 97 | 4 | 96 | 4 |
| 1994 | 295 | 15 | 282 | 11 | 272 | 11 | 101 | 5 | 97 | 4 | 95 | 4 |
| 1995 | 297 | 15 | 283 | 11 | 272 | 11 | 101 | 5 | 97 | 4 | 95 | 4 |
| 1996 | 299 | 15 | 284 | 11 | 272 | 11 | 102 | 5 | 98 | 4 | 95 | 4 |
| 1997 | 299 | 15 | 283 | 11 | 269 | 11 | 103 | 5 | 98 | 4 | 95 | 4 |
| 1998 | 297 | 15 | 279 | 11 | 265 | 11 | 103 | 5 | 99 | 4 | 95 | 4 |
| 1999 | 294 | 15 | 274 | 11 | 258 | 11 | 104 | 5 | 99 | 4 | 95 | 4 |
| 2000 | 293 | 15 | 275 | 11 | 259 | 11 | 106 | 5 | 100 | 4 | 95 | 4 |
| 2001 | 297 | 15 | 279 | 12 | 262 | 12 | 106 | 5 | 100 | 4 | 94 | 4 |
| 2002 | 305 | 16 | 288 | 13 | 269 | 14 | 106 | 5 | 99 | 4 | 93 | 4 |
| 2003 | 314 | 17 | 297 | 14 | 280 | 16 | 105 | 5 | 98 | 4 | 92 | 4 |
| 2004 | 326 | 19 | 302 | 16 | 286 | 18 | 104 | 6 | 97 | 4 | 91 | 4 |
| 2005 | 338 | 21 | 308 | 18 | 292 | 20 | 105 | 6 | 98 | 4 | 91 | 5 |
| 2006 | 352 | 23 | 320 | 21 | | | 107 | 6 | 100 | 5 | | |
| 2007 | 365 | 26 | 322 | 24 | | | 110 | 6 | 103 | 5 | | |
| 2008 | 371 | 28 | | | | | 115 | 7 | | | | |
| 2009 | 372 | 31 | | | | | 120 | 8 | | ļ | | |

Table 8.16. Estimated age 3+ population biomass and female spawning biomass from the preferred model.

| Veen | 2009 Assessment (| millions) | 2007 Assessmen | nt (millions) | 2005 Assessmen | nt (millions) |
|------|-------------------|-----------|----------------|---------------|----------------|---------------|
| rear | Mean | Std Dev | Mean | Std Dev | Mean | Std Dev |
| 1984 | 235 | 35 | 165 | 36 | 163 | 35 |
| 1985 | 238 | 32 | 247 | 43 | 241 | 42 |
| 1986 | 204 | 29 | 239 | 39 | 233 | 38 |
| 1987 | 220 | 30 | 180 | 32 | 175 | 32 |
| 1988 | 282 | 36 | 269 | 39 | 259 | 38 |
| 1989 | 218 | 31 | 211 | 34 | 201 | 33 |
| 1990 | 235 | 31 | 224 | 34 | 212 | 33 |
| 1991 | 244 | 33 | 238 | 36 | 222 | 34 |
| 1992 | 312 | 36 | 326 | 42 | 305 | 40 |
| 1993 | 234 | 30 | 188 | 33 | 175 | 31 |
| 1994 | 279 | 35 | 215 | 38 | 200 | 36 |
| 1995 | 239 | 31 | 272 | 42 | 253 | 39 |
| 1996 | 248 | 32 | 228 | 38 | 211 | 36 |
| 1997 | 235 | 32 | 212 | 39 | 193 | 36 |
| 1998 | 188 | 28 | 154 | 34 | 140 | 31 |
| 1999 | 180 | 28 | 133 | 32 | 121 | 29 |
| 2000 | 276 | 37 | 351 | 54 | 320 | 52 |
| 2001 | 372 | 44 | 349 | 57 | 327 | 57 |
| 2002 | 413 | 51 | 366 | 69 | 359 | 73 |
| 2003 | 321 | 47 | 337 | 75 | 352 | 86 |
| 2004 | 382 | 61 | 167 | 80 | 192 | 96 |
| 2005 | 364 | 66 | 302 | 114 | 242 | 105 |
| 2006 | 411 | 90 | 447 | 174 | | |
| 2007 | 352 | 102 | 148 | 113 | | |
| 2008 | 236 | 136 | | | | |
| 2009 | 199 | 111 | | | | |

Table 8.17. Estimated age 3 recruitment from the preferred model.

| | | | | Catch (t) | | | |
|------|------------|------------|------------|------------|------------|------------|------------|
| year | scenario 1 | scenario 2 | scenario 3 | scenario 4 | scenario 5 | scenario 6 | scenario 7 |
| 2009 | 3,398 | 3,398 | 3,398 | 3,398 | 3,398 | 3,398 | 3,398 |
| 2010 | 52,721 | 52,721 | 28,326 | 3,281 | 0 | 65,567 | 52,721 |
| 2011 | 42,609 | 42,609 | 26,106 | 3,421 | 0 | 49,204 | 42,609 |
| 2012 | 35,805 | 35,805 | 24,289 | 3,537 | 0 | 39,256 | 44,645 |
| 2013 | 30,981 | 30,981 | 22,684 | 3,615 | 0 | 32,851 | 36,072 |
| 2014 | 27,100 | 27,100 | 21,058 | 3,633 | 0 | 28,096 | 29,964 |
| 2015 | 23,807 | 23,807 | 19,367 | 3,589 | 0 | 24,159 | 25,377 |
| 2016 | 21,359 | 21,359 | 17,834 | 3,505 | 0 | 20,486 | 21,323 |
| 2017 | 19,818 | 19,818 | 16,704 | 3,421 | 0 | 19,234 | 19,543 |
| 2018 | 19,080 | 19,080 | 15,995 | 3,354 | 0 | 19,199 | 19,290 |
| 2019 | 18,854 | 18,854 | 15,571 | 3,305 | 0 | 19,448 | 19,459 |
| 2020 | 18,810 | 18,810 | 15,306 | 3,266 | 0 | 19,638 | 19,626 |
| 2021 | 18,804 | 18,804 | 15,133 | 3,235 | 0 | 19,722 | 19,708 |
| 2022 | 18,799 | 18,799 | 15,013 | 3,209 | 0 | 19,748 | 19,738 |

Table 8.18. Projected catch (t) for the seven projection scenarios.

Table 8.19. Female spawning biomass (t) for the seven projection scenarios. The values of $B_{40\%}$ and $B_{35\%}$ are 49,899 t and 43,661 t, respectively.

| _ | | | | | | | | |
|---|------|------------|------------|------------|-------------|------------|-------------|------------|
| | | | | Female s | pawning bio | omass (t) | | |
| | year | scenario 1 | scenario 2 | scenario 3 | scenario 4 | scenario 5 | scena rio 6 | scenario 7 |
| | 2009 | 120,066 | 120,066 | 120,066 | 120,066 | 120,066 | 120,066 | 120,066 |
| | 2010 | 124,674 | 124,674 | 124,674 | 124,674 | 124,674 | 124,674 | 124,674 |
| | 2011 | 103,337 | 103,337 | 115,805 | 128,656 | 130,343 | 96,800 | 103,337 |
| | 2012 | 88,143 | 88,143 | 107,918 | 131,249 | 134,538 | 78,865 | 88,143 |
| | 2013 | 76,554 | 76,554 | 100,299 | 131,997 | 136,772 | 66,444 | 72,063 |
| | 2014 | 67,154 | 67,154 | 92,705 | 130,867 | 136,982 | 57,145 | 60,446 |
| | 2015 | 59,860 | 59,860 | 85,705 | 128,523 | 135,793 | 50,445 | 52,327 |
| | 2016 | 55,020 | 55,020 | 80,127 | 125,929 | 134,144 | 46,522 | 47,485 |
| | 2017 | 52,397 | 52,397 | 76,264 | 123,727 | 132,687 | 45,336 | 45,706 |
| | 2018 | 51,249 | 51,249 | 73,776 | 122,008 | 131,544 | 45,328 | 45,448 |
| | 2019 | 50,827 | 50,827 | 72,171 | 120,639 | 130,612 | 45,563 | 45,586 |
| | 2020 | 50,669 | 50,669 | 71,084 | 119,439 | 129,727 | 45,730 | 45,723 |
| | 2021 | 50,598 | 50,598 | 70,349 | 118,486 | 129,017 | 45,803 | 45,793 |
| | 2022 | 50,564 | 50,564 | 69,841 | 117,678 | 128,382 | 45,833 | 45,825 |

| | | Fishing mortality | | | | | | | | | |
|------|------------|-------------------|------------|------------|------------|------------|------------|--|--|--|--|
| year | scenario 1 | scenario 2 | scenario 3 | scenario 4 | scenario 5 | scenario 6 | scenario 7 | | | | |
| 2009 | 0.0218 | 0.0218 | 0.0218 | 0.0218 | 0.0218 | 0.0218 | 0.0218 | | | | |
| 2010 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4812 | 0.3713 | | | | |
| 2011 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4812 | 0.3713 | | | | |
| 2012 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4812 | 0.4812 | | | | |
| 2013 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4812 | 0.4812 | | | | |
| 2014 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4812 | 0.4812 | | | | |
| 2015 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4769 | 0.4810 | | | | |
| 2016 | 0.3713 | 0.3713 | 0.1857 | 0.0201 | 0.0000 | 0.4454 | 0.4539 | | | | |
| 2017 | 0.3681 | 0.3681 | 0.1857 | 0.0201 | 0.0000 | 0.4337 | 0.4371 | | | | |
| 2018 | 0.3642 | 0.3642 | 0.1857 | 0.0201 | 0.0000 | 0.4336 | 0.4346 | | | | |
| 2019 | 0.3628 | 0.3628 | 0.1857 | 0.0201 | 0.0000 | 0.4359 | 0.4361 | | | | |
| 2020 | 0.3626 | 0.3626 | 0.1857 | 0.0201 | 0.0000 | 0.4377 | 0.4376 | | | | |
| 2021 | 0.3627 | 0.3627 | 0.1857 | 0.0201 | 0.0000 | 0.4383 | 0.4382 | | | | |
| 2022 | 0.3627 | 0.3627 | 0.1857 | 0.0201 | 0.0000 | 0.4385 | 0.4384 | | | | |

Table 8.20. Fishing mortality for the seven projection scenarios.

| | Halibut | Salmon (#'s) | | | HalibutSalmon (#'s)Crab (#'s) | | | | | | |
|------|---------|--------------|-----------------|-------|-------------------------------|------------------|----------|-----------|----------------|--------|--|
| year | (kg) | Chinook | non- Chinook | Total | Opilio Tanner | Bairdi Tanner | Red King | Blue King | Golden King | Total | |
| 2003 | 203,807 | 612 | 19 | 631 | 174 | 17,330 | 0 | 0 | 533 | 18,037 | |
| 2004 | 101,755 | 1,389 | 90 | 1,479 | 0 | 7,275 | 0 | 0 | 0 | 7,275 | |
| 2005 | 52,798 | 16 | 0 | 16 | 0 | 32,471 | 0 | 0 | 0 | 32,471 | |
| 2006 | 36,528 | 56 | 0 | 56 | 0 | 25,884 | 0 | 0 | 0 | 25,884 | |
| 2007 | 27,029 | 0 | 0 | 0 | 0 | 254 | 0 | 0 | 0 | 254 | |
| 2008 | 91,959 | 0 | 0 | 0 | 272 | 7,077 | 0 | 0 | 0 | 7,349 | |
| 2009 | 51,777 | 118 | 0 | 118 | 0 | 7,073 | 0 | 0 | 0 | 7,073 | |

Table 8.21. Prohibited species catch (PSC) in the flathead sole target fishery.

Table 8.22. Catch of non-prohibited species in the flathead sole target fishery. The species accounting for the two largest totals are highlighted.

| | 20 |)09 | 20 | 08 | 20 |)07 | 20 | 2006 | |
|--------------------------------------|-----------|------------|-----------|------------|-----------|------------|-----------|------------|--|
| Species | total (t) | % retained | |
| Atka mackerel | 18 | 99% | 3 | 98% | 36 | 71% | 17 | 84% | |
| arrowtooth flounder | 779 | 7% | 801 | 21% | 723 | 10% | 839 | 10% | |
| Dover sole and turbot | 1 | 100% | 4 | 98% | 1 | 0% | 3 | 80% | |
| flathead sole | 367 | 97% | 572 | 92% | 423 | 90% | 522 | 82% | |
| northern rockfish | 1 | . 89% | 0 | 100% | 2 | . 0% | 2 | 0% | |
| all sculpins, sharks, squid, octopus | 6 | , 78% | 14 | 74% | 35 | 0% | 16 | 0% | |
| pacific cod | 108 | 94% | 125 | 84% | 131 | 90% | 38 | 92% | |
| pelagic shelf rockfish | 1 | 82% | 2 | 100% | 2 | . 0% | 0 | 100% | |
| pollock | 57 | 94% | 45 | 97% | 27 | 99% | 33 | 94% | |
| POP | 2 | . 6% | 2 | 2% | 11 | 13% | 4 | 75% | |
| rex sole | 77 | 86% | 86 | 98% | 110 | 98% | 68 | 93% | |
| rougheye | 2 | . 16% | 0 | 42% | 0 | 100% | 2 | 14% | |
| other rockfish complex | 0 | 0% | 2 | 53% | 0 | 99% | 0 | 99% | |
| sablefish | 8 | , 98% | 1 | 61% | 4 | 100% | 4 | 87% | |
| shallow water flatfish | 56 | 97% | 41 | 98% | 26 | 95% | 29 | 27% | |
| shortraker | 2 | . 97% | 0 | 0% | 0 | 0% | 7 | 71% | |
| thornyhead | 5 | 100% | 0 | 100% | 7 | 100% | 6 | 94% | |
| unidentified skate | 9 | 52% | 5 | 28% | 20 | 64% | 0 | 0% | |
| big skate | 39 | 94% | 66 | 84% | 23 | 99% | 30 | 64% | |
| longnose skate | 12 | . 95% | 11 | 81% | 13 | 19% | 11 | 55% | |

Figures



Figure 8.1. Fishery catches for GOA flathead sole, 1984-2009 (as of Sept. 26, 2009).





Figure 8.3. Spatial patterns of fishery catches for GOA flathead sole from the first three quarters of 2009.



Figure 8.4. GOA survey biomass for flathead sole. Error bars represent 95% lognormal confidence intervals. The GOA survey did not include the eastern gulf in 2001. The value shown here for 2001 has been corrected to account for this (see text).



Figure 8.5. Spatial patterns of CPUE for flathead sole in the GOA groundfish surveys for 2005, 2007 and 2009.

a) Length-at-age.











Figure 8.6. Age-specific chedules for GOA flathead sole: females solid line, males dotted line.



Figure 8.7. Comparison of selectivity functions from: a) the base case (left) and b) alternative 1 (right). Survey selectivities are plotted in red with a dotted line, fishery selectivities are plotted in black with asolid line. Male selectivity functions are plotted with a triangle symbol, female selectivity functions are plotted with a triangle symbol.



Figure 8.8. Further comparison of model results between the base case and the alternative using: a) estimated total biomass (upper left), b) estimated spawning biomass (upper right), c) recruitment (middle left), d) annual fishing mortality (middle right), e) $F_{40\%}$ and $F_{35\%}$,(lower left), and f) ABC, OFL, virgin biomass (B0), $B_{40\%}$, and $B_{35\%}$ (lower right).



Figure 8.9. Comparison of likelihood profiles for fishery and survey selectivity-related parameters from the base case (dashed line) and the alternative case (solid line). "a50" denotes the parameter for the age at which the unscaled logistic function is 50%. "scale parameter" denotes the log-scale offset for scaling male selectivity relative to asymptotic female selectivity.



Figure 8.10. Predicted and observed annual catches for GOA flathead sole from the preferred model. Predicted catch = solid line, observed catch = dotted line with circles.



Length (cm)

Figure 8.11. Fit to female GOA flathead sole fishery length composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.



Length (cm)

Figure 8.12. Fit to male GOA flathead sole fishery length composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.



Fig. 8.13. Predicted and observed survey biomass for GOA flathead sole from the preferred model. Predicted survey biomass = triangles, observed survey biomass = circles (error bars are approximate lognormal 95% confidence intervals).



Figure 8.14. Fit to the female GOA flathead sole survey length composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.



Figure 8.15. Fit to the male GOA flathead sole survey length composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.



Figure 8.16. Fit to the female survey GOA flathead sole age composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.



Figure 8.17. Fit to the male survey GOA flathead sole age composition data from the preferred model. Dashed lines represent the model prediction, solid lines represent the data.





Figure 8.18. Upper: Estimated age 3+ biomass (circles) and female spawning biomass (triangles) for GOA flathead sole. Error bars are approximate lognormal 95% confidence intervals. Lower: Comparison of total biomass (dark blue) and spawning biomass (light blue) estimates from the 2009, 2007, and 2005 assessments.



Figure 8.19. Upper: Estimated age 3 recruitments of GOA rex sole with approximate 95% lognormal confidence intervals. Horizontal line is mean recruitment. Lower: Comparison of recruitment estimates from the 2009, 2007, and 2005 assessments.



Figure 8.20. Control rule plot of estimated fishing mortality versus estimated female spawning biomass for GOA flathead sole from the preferred model. F_{OFL} = solid line, F_{maxABC} = dashed line.



Figure 8.21a. Gulf of Alaska food web from the GOA ecosystem model (Aydin et al., 2007) highlighting adult flathead sole links to predators (blue boxes and lines) and prey (green boxes and lines). Box size reflects relative standing stock biomass.



Figure 8.21b. Gulf of Alaska food web from the GOA ecosystem model (Aydin et al., 2007) highlighting juveile flathead sole links to predators (blue boxes and lines) and prey (green boxes and lines). Box size reflects relative standing stock biomass.



Figure 8.22a. Diet composition for Gulf of Alaska adult flathead sole from the GOA ecosystem model (Aydin et al., 2007).



Figure 8.22b. Diet composition for Gulf of Alaska juvenile flathead sole from the GOA ecosystem model (Aydin et al., 2007).



Figure 8.23a. Decomposition of natural mortality for Gulf of Alaska adult flathead sole from the GOA ecosystem model (Aydin et al., 2007).



Figure 8.23b. Decomposition of natural mortality for Gulf of Alaska juvenile flathead sole from the GOA ecosystem model (Aydin et al., 2007).

Appendix A.

| Quantity | Definition | | | | | | |
|--------------------------------|---|--|--|--|--|--|--|
| Т | number of years in the model. | | | | | | |
| Α | number of age classes (18). | | | | | | |
| L | number of length classes (18). | | | | | | |
| T _{min} | model start year (1984). | | | | | | |
| T_{max} | assessment year (2009). | | | | | | |
| t | time index. | | | | | | |
| а | age index $(1 \le a \le A; a = 1 \text{ corresponds to age at recruitment}).$ | | | | | | |
| x | sex index $(1 \le x \le 2; 1 = \text{female}, 2 = \text{male})$. | | | | | | |
| l | length index ($1 \le l \le L$; $l=1$ corresponds to minimum length class). | | | | | | |
| $\{t^{S}\}$ | set of years for which survey biomass data is available. | | | | | | |
| $\{t^{F,A}\}$ | set of years for which fishery age composition data is available. | | | | | | |
| $\{t^{F,L}\}$ | set of years for which fishery length composition data is available. | | | | | | |
| $\{t^{S,A}\}$ | set of years for which survey age composition data is available. | | | | | | |
| $\{t^{S,L}\}$ | set of years for which survey length composition data is available. | | | | | | |
| $L^{x}_{l,a}$ | elements of length-age conversion matrix (proportion of sex x fish in age class a that are in length class l). (fixed) | | | | | | |
| $W_{x,a}$ | mean body weight (kg) of sex x fish in age group a . (fixed) | | | | | | |
| ϕ_{a} | proportion of females mature at age <i>a</i> . (fixed) | | | | | | |
| $\overline{\ln R_0}$ | mean value of log-transformed recruitment. (estimable) | | | | | | |
| $	au_t$ | recruitment deviation in year t. (estimable) | | | | | | |
| M_x | instantaneous natural mortality rate. (fixed) | | | | | | |
| ln <i>F</i> | mean value of log-transformed fishing mortality. (estimable) | | | | | | |
| \mathcal{E}_t | deviations in fishing mortality rate in year <i>t</i> . (estimable) | | | | | | |
| R_t | recruitment in year <i>t</i> . | | | | | | |
| $N_{t,x,a}$ | number of fish of sex x and age class a in year t. | | | | | | |
| $C_{t,x,a}$ | catch (number) of fish of sex x and age class a in year t. | | | | | | |
| $p^{F,A}_{t,x,a}$ | proportion of the total catch in year t that is sex x and in age class a. | | | | | | |
| $p^{F,L}_{t,x,l}$ | proportion of the total catch in year <i>t</i> that is sex <i>x</i> and in length class <i>l</i> . | | | | | | |
| $p^{S,A}_{t,x,a}$ | proportion of the survey biomass in year <i>t</i> that is sex <i>x</i> and in age group a. | | | | | | |
| $p^{S,L}_{t,x,l}$ | proportion of the survey biomass in year <i>t</i> that is sex <i>x</i> and in age group a. | | | | | | |
| C_t | total catch (yield) in tons in year <i>t</i> . | | | | | | |
| $F_{t,x,a}$ | instantaneous fishing mortality rate for sex x and age group a in year t. | | | | | | |
| $Z_{t,x,a}$ | instantaneous total mortality for sex x and age group a in year t. | | | | | | |
| s ^{FU} _{x,a} | unnormalized fishery selectivity for sex <i>x</i> and age group <i>a</i> . | | | | | | |
| s ^{SU} _{x,a} | unnormalized survey selectivity for sex <i>x</i> and age group <i>a</i> . | | | | | | |
| s ^{FN} _{x,a} | normalized fishery selectivity for sex x and age group a. | | | | | | |
| $s^{SN}_{x,a}$ | normalized survey selectivity for sex x and age group a. | | | | | | |

Table A.1. List of quantities and their definitions as used in the model.

| Equation | Description |
|---|--|
| $\tau_t \sim N(0, \sigma_R^2)$ | Random deviate associated with recruitment. |
| $N_{t,x,1} = R_t = \exp\left(\overline{\ln R_0} + \tau_t\right)$ | Recruitment (assumed equal for males and females). |
| $N_{t+1,x,a+1} = N_{t,x,a} e^{-Z_{t,x,a}}$ | Numbers at age. |
| $N_{t+1,x,A} = N_{t,x,A-1}e^{-Z_{t,x,A-1}} + N_{t,x,A}e^{-Z_{t,x,A}}$ | Numbers in "plus" group. |
| $C_{t,x,a} = \frac{F_{t,x,a}}{Z_{t,x,a}} (1 - e^{-Z_{t,x,a}}) N_{t,x,a}$ | Catch at age (in numbers caught). |
| $C_{t} = \sum_{x=1}^{2} \sum_{a=1}^{A} w_{x,a} C_{t,x,a}$ | Total catch in tons (i.e., yield). |
| $FSB_t = \sum_{a=1}^A w_{1,a} \phi_a N_{t,1,a}$ | Female spawning biomass. |
| $Z_{t,x,a} = F_{t,x,a} + M$ | Total mortality. |
| $F_{t,x,a} = s_{x,a}^{F} \cdot \exp\left(\overline{\ln F} + \varepsilon_{t}\right)$ | Fishing mortality. |
| $\varepsilon_t \sim N(0, \sigma_F^2)$ | Random deviate associated with fishing mortality. |
| $s_{x,a}^{FU} = \frac{1}{1 + e^{(-b_x^F(age - 50A_x^F))}}$ | Unnormalized fishery selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x,a}^{SU} = \frac{1}{1 + e^{(-b_x^S(age_{-50}A_x^S))}}$ | Unnormalized survey selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x,a}^{FN} = \exp(r_x^F) \frac{s_{x,a}^{FU}}{\max\{s_{1,a}^{FU}\}}$ | Normalized fishery selectivity. $r^{F_{l}} \equiv 0$. |
| $s_{x,a}^{SN} = \exp(r_x^S) \frac{s_{x,a}^{SU}}{\max\{s_{1,a}^{SU}\}}$ | Normalized survey selectivity. $r^{S_{I}} \equiv 0$. |
| $N^{s}_{t,x,a} = Q s^{s}_{x,a} N_{t,x,a}$ | Survey numbers for sex x , age a at time t . |
| $SB_{t} = \sum_{x=1}^{2} \sum_{a=1}^{A} w_{x,a} N^{S}_{t,x,a}$ | Total survey biomass. |
| $p_{t,x,a}^{F,A} = C_{t,x,a} / \sum_{x=1}^{2} \sum_{a=1}^{A} C_{t,x,a}$ | Proportion at age in the catch. |
| $p_{t,x,l}^{F,L} = \sum_{a=1}^{A} L_{l,a}^{x} \cdot p_{t,x,a}^{F,A}$ | Proportion at length in the catch. |
| $p_{t,x,a}^{S,A} = N_{t,x,a}^{S} / \sum_{x=1}^{2} \sum_{a=1}^{A} N_{t,x,a}^{S}$ | Proportion at age in the survey. |
| $p_{t,x,l}^{S,L} = \sum_{a=1}^{A} L_{l,a}^{x} \cdot p_{t,x,a}^{S,A}$ | Proportion at length in the survey. |

Table A.2. Model equations describing the model populations dynamics.

| Component | Description | |
|---|--|--|
| $\sum_{t=1}^{T} \left[\log(C_t^{obs}) - \log(C_t) \right]^2$ | Catch; assumes a lognormal distribution. | |
| $\sum_{t \in \{t^{F,A}\}} \sum_{x=1}^{2} \sum_{a=1}^{A} n_t^{samp} \cdot p_{t,x,a}^{F,A,obs} \cdot \log(p_{t,x,a}^{F,A}) - \text{offset}$ | Fishery age composition; assumes a multinomial distribution. Observed sample size is n_t^{samp} . | |
| $\sum_{t \in \{t^{F,L}\}} \sum_{x=1}^{2} \sum_{l=1}^{L} n_{t}^{samp} \cdot p_{t,x,l}^{F,L,obs} \cdot \log(p_{t,x,l}^{F,L}) - \text{offset}$ | Fishery length composition; assumes a multinomial distribution. Observed sample size is n_t^{samp} . | |
| $\sum_{t \in \{t^{F,A}\}} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{samp} \cdot p_{t,x,a}^{S,A,obs} \cdot \log(p_{t,x,a}^{S,A}) - \text{offset}$ | Survey age composition; assumes a multinomial distribution. Observed sample size is n_t^{samp} . | |
| $\sum_{t \in \{t^{F,L}\}} \sum_{x=1}^{2} \sum_{l=1}^{L} n_t^{samp} \cdot p_{t,x,l}^{S,L,obs} \cdot \log(p_{t,x,l}^{S,L}) - \text{offset}$ | Survey length composition; uses a multinomial distribution. Observed sample size is n_t^{samp} . | |
| offset = $\sum_{t} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{samp} \cdot p_{t,x,a}^{obs} \cdot \log(p_{t,x,a}^{obs}))$ | The offset constants for age composition components are calculated from the observed proportions and the sample sizes. A similar formula is used for length composition component offsets. | |
| $\sum_{t \in [t^{S}]} \left[\frac{\log \left[\frac{SB_{t}^{obs}}{SB_{t}} \right]}{\sqrt{2} \cdot s.d.(\log(SB_{t}^{obs}))} \right]^{2}$ | Survey biomass; assumes a lognormal distribution. | |
| $\sum_{t=T_{\min}}^{T_{\max}-3} (\tau_t)^2$ | Recruitment; assumes a lognormal distribution, since τ_t is on a log scale. | |
| $\sum_{t=T_{\max}-2}^{T_{\max}} (\tau_t)^2$ | "Late" recruitment; assumes a lognormal distribution, since τ_t is on a log scale. | |
| $\sum_{t=T_{\min}-A+1}^{T_{\min}-1} (\tau_t)^2$ | "Early" recruitment; assumes a lognormal distribution, since τ_t is on a log scale. Determines age composition at starting year of model. | |

Table A.3. Likelihood components.

| Tuble 71.1. Turumeters fixed in the model. | | |
|--|--|--|
| Parameter | Description | |
| $M_x = 0.2$ | sex-specific natural mortality rate. | |
| Q = 1.0 | survey catchability. | |
| $L^{x}_{l,a}$ | sex-specific length-at-age conversion matrix. | |
| $W_{x,a}$ | sex-specific weight-at-age. | |
| ϕ_a | proportion of females mature at age <i>a</i> . | |

Table A.4. Parameters fixed in the model.

Table A.5. Parameters estimated in the model. A total of 81 parameters were estimated.

| Parameter | Subscript range | Total no. of Parameters | Description |
|-------------------------------|---------------------------------------|----------------------------|--|
| $\ln(R_0)$ | NA | 1 | natural log of the geometric mean value of age 3 recruitment. |
| $	au_t$ | $T_{\min} - A + 1 \le t \le T_{\max}$ | 43 | log-scale recruitment deviation in year <i>t</i> . |
| $\ln(f_0)$ | NA | 1 | natural log of the geometric mean value of fishing mortality. |
| \mathcal{E}_t | $T_{\min} \leq t \leq T_{\max}$ | 26 | log-scale deviations in fishing mortality rate in year <i>t</i> . |
| r_2^F | NA | 1 | scaling from female to male fishery selectivity (log-scale). |
| b^{F}_{x} , 50 A^{F}_{x} | 1≤ <i>x</i> ≤2 | 4 | sex-specific selectivity parameters (slope and age at 50% selected) for the fishery. |
| r_{2}^{S} | <i>S</i> =1 | 1 | scaling from female to male survey selectivity (log-scale). |
| $b_{x}^{s}, {}_{50}A_{x}^{s}$ | 1≤x≤2 S=1 | 4 | sex-specific selectivity parameters (slope and age at 50% selected) for the survey. |