# Chapter 2: Assessment of the Pacific Cod Stock in the Eastern Bering Sea and Aleutian Islands Area 

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## EXECUTIVE SUMMARY

## Summary of Major Changes

Relative to the November edition of last year's BSAI SAFE report, the following substantive changes have been made in the Pacific cod stock assessment.

Changes in the Input Data

1) Catch data for 2006 were updated, and preliminary catch data for 2007 were incorporated.
2) Commercial fishery size composition data were recompiled for the years 1990-2006 and incorporated, and preliminary size composition data from the 2007 commercial fisheries were incorporated.
3) Size composition data from the 2007 EBS shelf bottom trawl survey were incorporated.
4) The biomass estimate from the 2007 EBS shelf bottom trawl survey was incorporated into several preliminary models (the 2007 estimate of $423,703 \mathrm{t}$ was down about $18 \%$ from the 2006 estimate, and is the all-time low in the time series).
5) The numeric abundance estimates from the 1979-2007 EBS shelf bottom trawl surveys were incorporated into the final models (the 2007 estimate of $713,374,144$ fish was up about $86 \%$ from the 2006 estimate).
6) Age composition data from the 1995 and 2006 EBS shelf bottom trawl surveys were incorporated into some of the models.
7) Seasonal catch per unit effort (CPUE) data for the trawl, longline, and pot fisheries from 19912007 were incorporated.
8) Catch rates of Pacific cod from the 1998-2007 International Pacific Halibut Commission (IPHC) longline surveys were incorporated.
9) Pacific cod size composition data from the 2007 IPHC longline survey were incorporated.

## Changes in the Assessment Model

Many changes have been made in the stock assessment model since last year's assessment. Some of these are described in the report of a technical workshop held in April of this year (Thompson and Conners 2007). For example, the base model developed for the technical workshop differed with respect to last year's author-recommended model in the following respects:

| Feature | Last year's assessment | Technical workshop base model |
| :--- | :--- | :--- |
| Software | SS2 version 1.23d | SS2 verson 2.00c |
| Natural mortality rate $(M)$ | Fixed at 0.37 | Estimated internally |
| Length-at-age parameters | Estimated externally | Estimated internally |
| Pre-shift median recruitment | Estimated iteratively | Estimated internally |
| First year included in model | 1964 | 1976 |
| First year of current regime | 1977 | 1976 |
| No. initial year classes estimated | None (all set at equilibrium) | 10 |
| Selectivity pattern | 4-parameter double normal | 6-parameter double normal |
| Bounds on log catchability | Essentially unbounded | Low=2×ln(0.75), high=0 |
| Bounds on selectivity parameters | Essentially unbounded | Most have low=-10, high=10 |
| Form of prior distributions | All normal | Normal and symmetric beta |

Dozens of other models were also considered at the workshop.
Further changes were made between the technical workshop and this year's preliminary SAFE Report (Thompson et al. 2007). The following were among the ways in which Model 1 from the preliminary SAFE Report differed from the base model developed for the technical workshop:

| Feature | Technical workshop base model | Model 1 from preliminary SAFE |
| :---: | :---: | :---: |
| Software | SS2 version 2.00c | SS2 version 2.00i |
| Selectivities forced to be asymptotic | None | January-May trawl fishery |
| Time-varying selectivity | Fishery selectivities constant within blocks, surveys constant | Fishery and survey selectivities variable across years |
| Time-varying length at age 1 | Constant | Variable across years |
| Std. dev. of log-scale recruitment deviations ( $\sigma_{R}$ ) | $\sigma_{R}$ set iteratively | $\sigma_{R}=0.6$ |
| No. initial year classes estimated | 10 | 3 |
| Variability in length at age | $\mathrm{CV}=$ function of length at age | SD = function of age |
| Slope trawl survey data | Included | Excluded |
| Fishery CPUE data | Excluded | Included for comparison only |
| Form of age data | Marginal age compositions | Age-at-length compositions |
| Form of prior distributions | Normal and symmetric beta | All uniform |
| Bounds on log catchability | Low $=2 \times \ln (0.75)$, high $=0$ | Low=-2, high=2 |
| Bound parameters, if any | Parameters did not approach bounds due to priors | Fixed at bound (i.e., taken out of the estimation process) |

Three other models were also considered in the preliminary SAFE Report.
Relative to Model 1 from the preliminary SAFE Report, the following changes have been made in the base model presented in this assessment (Model 1):

| Feature | Model 1 from preliminary SAFE | New base model (Model 1) |
| :--- | :--- | :--- |
| Natural mortality rate | Estimated internally | Fixed at 0.34 |
| Form of age data | Age-at-length compositions | Marginal age compositions |
| Basis of maturity schedule | Length | Age |
| Basis of trawl survey selectivity | Length | Age |
| Seasonal structure | Partially seasonal | Fully seasonal |
| Selectivities forced to be | January-May trawl fishery | Jun-Aug trawl, Sep-Dec trawl, |
| asymptotic |  | Jun-Aug longline, Sep-Dec pot |
| Time-varying selectivity; std. | Fisheries and surveys variable | Fisheries constant, surveys |
| dev. of selectivity deviations $\left(\sigma_{S}\right)$ | across years; $\sigma_{S}=0.4$ | variable across years; $\sigma_{S}=0.2$ |
| Time-varying length at age 1 | Variable across years | Constant across years |
| Variability in length at age | SD = function of age | SD = function of length at age |
| Survey abundance units | Biomass | Numbers |
| Multinomial sample size; records | Square root of actual sample size; | Scaled bootstrap harmonic mean; |
| with small actual sample sizes | all records used | records with small N excluded |
| Std. dev. of log-scale recruitment | $\sigma_{R}=0.6$ | $\sigma_{R}$ set iteratively |
| deviations ( $\sigma_{R}$ ) |  |  |
| First year included in model | 1976 | 1977 |
| Starting year of regime shift | 1976 | 1977 |
| Parameters with large std. dev. | Left free | Fixed at their respective MLEs |

The reference to multinomial sample size in the table above refers to the fact that, as in previous assessments, size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery, and season within the year. In an attempt to move toward a more statistically based specification, this year a bootstrap analysis of the available fishery length data from 1990-2006 was undertaken.

Three other models are also presented in this assessment.

## Changes in Assessment Results

Under the base model, the following changes in assessment results were obtained:

1) The projected 2008 female spawning biomass for the BSAI stock is $398,000 \mathrm{t}$, up about $30 \%$ from last year's estimate for 2007 and up about $50 \%$ from last year's $F_{A B C}$ projection for 2008. However, the 2008 level is estimated to be only $29 \%$ of equilibrium unexploited female spawning biomass, compared to $38 \%$ estimated last year for 2007 and $33 \%$ projected last year for 2008.
2) The projected 2008 age $3+$ biomass for the BSAI stock is $1,080,000 \mathrm{t}$, up about $13 \%$ from last year's estimate for 2007.
3) The recommended 2008 ABC for the BSAI stock is $150,000 \mathrm{t}$, down about $15 \%$ from the actual 2007 ABC and up about $15 \%$ from last year's $F_{A B C}$ projection for 2008. The recommended preliminary 2009 ABC for the BSAI stock is $162,000 \mathrm{t}$.
4) The estimated 2008 OFL for the BSAI stock is $176,000 \mathrm{t}$, down about $15 \%$ from the actual 2007 OFL and up about $14 \%$ from last year's $F_{A B C}$ projection for 2008. The estimated preliminary 2009 OFL for the BSAI stock is $190,000 \mathrm{t}$.

## Responses to Comments from the SSC

## SSC Comments Specific to the Pacific Cod Assessments

From the December, 2006 minutes:
"With regard to the longline data, the SSC suggests excluding them from future assessments."
The above comment refers to use of Pacific cod CPUE data from the NMFS longline survey, which were included in some of the models presented in last year's assessment (though not in the model recommended by the authors and adopted by the Plan Team and SSC). These data have been excluded from this year's assessment.

Also from the December, 2006 minutes:
"Following the Plan Team meeting, a potential problem with the model fit was pointed out by an external reviewer and the author provided revised model B1 output on very short notice. The revised model provided a slightly improved fit overall. The largest differences between the revised and original model B1 results were in the fits to trawl survey age composition (worse fit in the revised model) and trawl survey size-at-age data (improved fit). While the differences in model fit are relatively minor, the model resulted in substantially higher estimates of biomass and the implied maximum ABC.
"The SSC is concerned that the revised model results did not receive any review by the Plan Team and that the apparent volatility of the model requires further investigation. The large difference in estimated biomass and the pattern of differences in likelihood components between the two model fits suggests that the model may be unstable or that two very different solutions provide very similar overall fits. Potential problems with the model configuration should be fully evaluated. All of the models examined by the author this year, including the revised model, suggest a series of poor recruitments from at least 2000 to 2004, and a decreasing trend in biomass that is projected to continue as these year classes enter the fishery. Therefore, the pattern of decreasing biomass in recent years and into the future appears to be a robust result.
"To resolve uncertainties within the assessment model, the SSC recommends that the AFSC conduct a workshop with the authors. In particular, the workshop should explore the following issues with regard to both the GOA and BSAI assessments:

- Estimation of growth inside the model versus the use of externally estimated length-at-age and weight-at-length parameters (with variances)
- Model convergence sensitivity to different weights assigned to the log priors and data components
- Model fit to contrasting models that fix $Q$ and estimate $M$ and alternatively fix $M$ and estimate Q.
- To fully explore the parameter space (and model fit), a suite of models incorporating fixed values for $M$ and $Q$ for a matrix of plausible values could also be explored.
- Consider a simpler logistic form for the survey selectivity and estimability of descending parameters for survey and fishery selectivity."

Following this request, the Alaska Fisheries Science Center convened a public workshop to examine various technical issues pertaining to the assessments for Pacific cod in the Bering Sea, Aleutian Islands, and Gulf of Alaska. The workshop took place at the Seattle lab of the Alaska Fisheries Science Center during the dates of April 24-25, 2007. A total of 44 people participated in the workshop. SSC chair Pat Livingston served as chair of the workshop and Liz Conners served as rapporteur. Many alternative models, for both the Bering Sea and Gulf of Alaska
stocks, were presented at the workshop, and several other models were developed during the workshop. Workshop participants contributed a total of 40 suggestions for the authors to consider in developing this year's stock assessments. A full report of the workshop was provided by Thompson and Conners (2007).

From the June, 2007 minutes:
"The SSC also reviewed the Pacific cod workshop report that evaluated stock assessment models in both the BS and GOA. The SSC received public comment from Kenny Down (Alaska Frontier Company). The SSC commends Dr. Grant Thompson (AFSC) for his excellent work and thanks the AFSC for conducting the workshop. The SSC looks forward to presentation of results of additional model simulations in October, 2007."

A preliminary stock assessment was produced in September of this year (Thompson et al. 2007). It included results of four new models developed in response to suggestions made at the technical workshop.

From the October, 2007 minutes:
"The SSC suggests that the analysts consider the following in some of the models brought forward in December:
i. "One or more model fits in which the value of natural mortality (M) is fixed. We are skeptical of model estimates of $M$, including the previous fixed value $M=0.37$. Purely for purposes of comparison we would like to see one fit with $M=0.37$. We would suggest that the author investigate the possibility of choosing a different fixed value based on life history theory (i.e., the value of $M$ for which the observed growth and maturity schedules are optimal).
ii. "Plots of the empirical length-at-age distributions calculated by keying out the survey length distributions using the length-stratified survey age readings. These empirical length-at-age frequencies must sum to the observed survey length frequencies, including the strong modes that the model fits fail to predict. This exercise may reveal differences between the empirical and estimated length-at-age distributions that will shed some light on the apparent inconsistencies between the age and length data."

Models 1 and 2 in the present assessment address the suggestions made in (i) above. Regarding (ii), none of the models presented in this assessment exhibit significant inconsistencies between the estimated mean lengths at age and the modes from the long-term average survey size compositions.

## SSC Comments on Assessments in General

From the December, 2006 minutes:
"The SSC appreciates the addition of phase-plane diagrams to most stock assessments and reiterates interest in these diagrams for all stock assessments in which it is possible to do so using standardized axes (i.e., X axis of B/Btarget; and Y axis of Fcatch/FOFL), formatted relative to harvest control rules. In addition, values from the most recent year should be provided annually by the assessment authors to the plan team."

Recalling that the December, 2005 SSC minutes identified the axes as " $F / F_{35 \%}$ versus $B / B_{35 \%}$," the quantity Btarget in the December, 2006 minutes is interpreted here to mean $B_{35 \%}$ and the quantity FOFL in the December, 2006 minutes is interpreted here to mean $F_{35 \%}$. The requested phaseplane diagram appears in Figure 2.9. Values from the most recent year (2007) are: relative biomass $=0.932$ and relative fishing mortality $=0.618$.

## INTRODUCTION

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $63^{\circ} \mathrm{N}$ latitude. Pacific cod is distributed widely over the eastern Bering Sea (EBS) as well as in the Aleutian Islands (AI) area. The resource in these two areas (BSAI) is managed as a single unit. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and Gulf of Alaska (GOA). Although at least one previous genetic study (Grant et al. 1987) failed to show significant evidence of stock structure within these areas, current genetic research underway at the Alaska Fisheries Science Center may soon shed additional light on the issue of stock structure of Pacific cod within the BSAI (M. Canino, AFSC, pers. commun.). Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the EBS or AI areas.

## FISHERY

Catches of Pacific cod taken in the EBS, AI, and BSAI for the periods 1964-1980 and 1981-2007 are shown in Tables 2.1a and 2.1b, 2.2a and 2.2b, and 2.3a and 2.3b, respectively. The catches in Tables 2.1a, 2.2a, and 2.3a are broken down by year and fleet sector (foreign, joint venture, domestic annual processing), while the catches in Tables 2.1b, 2.2b, and 2.3 b are broken down by gear type as well. During the early 1960s, a Japanese longline fishery harvested BSAI Pacific cod for the frozen fish market. Beginning in 1964, the Japanese trawl fishery for walleye pollock (Theragra chalcogramma) expanded and cod became an important bycatch species and an occasional target species when high concentrations were detected during pollock operations. By the time that the Magnuson Fishery Conservation and Management Act went into effect in 1977, foreign catches of Pacific cod had consistently been in the 30,000-70,000 $t$ range for a full decade. In 1981, a U.S. domestic trawl fishery and several joint venture fisheries began operations in the BSAI. The foreign and joint venture sectors dominated catches through 1988, but by 1989 the domestic sector was dominant and by 1991 the foreign and joint venture sectors had been displaced entirely. Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components. Figures 2.1a-2.1c show areas in which sampled hauls or sets for each of the three main gear types (trawl, longline, and pot) were concentrated during JanuaryMay, June-August, and September-December, 2006. Figures 2.1d-2.1e show the corresponding information for January-May and June-August, 2007 (preliminary data). To create these figures, the EEZ off Alaska was divided into $20 \mathrm{~km} \times 20 \mathrm{~km}$ squares. For each gear type, a square is shaded if more than two hauls/sets containing Pacific cod were sampled in it during the respective season and year.

The history of acceptable biological catch (ABC) and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate (i.e., all-gear, combined area) commercial catches in Table 2.4. From 1980 through 2007, TAC averaged about $79 \%$ of ABC, and aggregate commercial catch averaged about $89 \%$ of TAC. In 10 of these 28 years (37\%), TAC equaled ABC exactly, and in 5 of these 27 years (19\%), catch exceeded TAC (by an average of 4\%). In 2007, TAC was set at $97 \%$ of ABC to account for a small, State-managed fishery inside State of Alaska waters. Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3) changes in the stock assessment model. For example, in the assessments for fishery years 1980 through 2007, seven different assessment models were used (Table 2.4). All assessments from 1993 through 2004 used the Stock Synthesis 1 modeling software with primarily length-based data, albeit with some changes in model structure from time to time. The assessment was migrated to Stock Synthesis 2 in 2005, and several changes have been made to the model within the SS2 framework each year since then. Historically, the great majority of the BSAI catch has come from the EBS area. During the most recent complete five-year period (2002-2006), the EBS accounted for an average of about $86 \%$ of the BSAI catch.

The catches shown in Tables 2.1b, 2.2b, 2.3b, and 2.4 include estimated discards. Discard rates of Pacific cod in the various EBS and AI target fisheries are shown for each year 1991-2002 in Table 2.5a and for each year 2003-2004 in Table 2.5b. Values for 2005-2007 have not yet been tabulated.

Seasons for the Pacific cod fisheries are defined in 50 CFR $\S 679.23(5)$ as follows:
(i) Hook-and-line gear. Subject to other provisions of this part, directed fishing for CDQ and nonCDQ Pacific cod with vessels equal to or greater than $60 \mathrm{ft}(18.3 \mathrm{~m})$ LOA using hook-and-line gear is authorized only during the following two seasons:
(A) A season. From 0001 hours, A.l.t., Jan. 1 through 1200 hours, A.l.t., June 10; and
(B) B season. From 1200 hours, A.l.t., June 10 through 2400 hours, A.l.t., Dec. 31.
(ii) Trawl gear. Subject to other provisions of this part, directed fishing for CDQ and non-CDQ Pacific cod with trawl gear in the BSAI is authorized only during the following three seasons:
(A) A season. From 1200 hours, A.l.t., Jan. 20 through 1200 hours, A.l.t., Apr. 1;
(B) B season. From 1200 hours, A.l.t., Apr. 1 through 1200 hours, A.l.t., June 10; and
(C) C season. From 1200 hours, A.l.t., June 10 through 1200 hours, A.l.t., Nov. 1.
(iii) Pot gear. Subject to other provisions of this part, non-CDQ directed fishing for Pacific cod with vessels equal to or greater than $60 \mathrm{ft}(18.3 \mathrm{~m}) \mathrm{LOA}$ using pot gear in the BSAI is authorized only during the following two seasons:
(A) A season. From 0001 hours, A.l.t., January 1 through 1200 hours, A.l.t., June 10; and
(B) B season. From 1200 hours, A.l.t., September 1 through 2400 hours, A.l.t., Dec. 31.
(iv) Jig gear. Subject to other provisions of this part, directed fishing for CDQ and non-CDQ Pacific cod with jig gear is authorized only during the following three seasons:
(A) A season. From 0001 hours, A.l.t., Jan. 1 through 1200 hours, A.l.t., Apr. 30;
(B) B season. From 1200 hours, A.l.t., Apr. 30 through 1200 hours, A.l.t., Aug. 31; and
(C) C season. From 1200 hours, A.l.t., Aug. 31 through 2400 hours, A.l.t., Dec. 31.

Under Amendment 85, 10.7\% of the TAC is allocated to the CDQ fisheries. The remaining $89.3 \%$ is allocated as follows:

|  | Percentage |  |
| :--- | ---: | ---: |
| Sector | non-CDQ TAC | overall TAC |
| Jig vessels | 1.4 | 1.250 |
| Hook-and-line/pot catcher vessels $<60 \mathrm{ft}$. LOA | 2.0 | 1.786 |
| Hook-and-line/pot catcher vessels $\geq 60 \mathrm{ft} . \mathrm{LOA}$ | 0.2 | 0.179 |
| Hook-and-line catcher-processors | 48.7 | 43.489 |
| Pot catcher vessels $>60$ ft. LOA | 8.4 | 7.501 |
| Pot catcher-processors | 1.5 | 1.340 |
| AFA trawl catcher-processors | 2.3 | 2.054 |
| Non-AFA trawl cathcer-processors | 13.4 | 11.966 |
| Trawl catcher vessels | 22.1 | 19.735 |
| Total | 100.0 | 89.300 |

Amendment 85 further apportions the above allocations (in percent) by season as follows:

| Gear Type | A Season | B Season | C Season |
| :---: | :---: | :---: | :---: |
| CDQ trawl | 60 | 20 | 20 |
| CDQ trawl catcher vessels | 70 | 10 | 20 |
| CDQ trawl catcher-processors | 50 | 30 | 20 |
| Non-CDQ trawl catcher vessels | 74 | 11 | 15 |
| Non-CDQ trawl catcher-processors | 75 | 25 | 0 |
| CDQ hook-and-line catcher-processors, and hook-and-line |  |  |  |
| catcher vessels $\geq 60 \mathrm{ft}$. LOA | 60 | 40 | n/a |
| Non-CDQ hook-and-line catcher-processors, hook-and-line catcher vessels $\geq 60 \mathrm{ft}$. LOA, pot catcher-processors, and pot |  |  |  |
| catcher vessels $\geq 60 \mathrm{ft}$. LOA | 51 | 49 | n/a |
| CDQ jig vessels | 40 | 20 | 40 |
| Non-CDQ jig vessels | 60 | 20 | 20 |
| All other nontrawl vessels | --- no | onal allo | e ---------- |

## DATA

This section describes data used in the current stock assessment models. It does not attempt to summarize all available data pertaining to Pacific cod in the BSAI.

## Commercial Catch Data

## Catch Biomass

Catches taken in the EBS for the period 1977-2007 are shown in Table 2.6. Catches for the years 19771980 may not include discards. Catches in these tables are broken down by the three main gear types and intra-annual periods consisting of the months January-May, June-August, and September-December. This particular division, which was suggested by participants in the EBS fishery, is intended to reflect actual intra-annual differences in fleet operation (e.g., fishing operations during the spawning period may be different than at other times of year). In years for which estimates of the distribution by gear or period were not available, proxies based on other years' distributions were used.

## Catch Size Composition

Fishery size compositions are presently available, by gear, for at least one gear type in every year from 1977 through the first part of 2007. For ease of representation and analysis, length frequency data for Pacific cod can usefully be grouped according to the following set of 25 intervals or "bins," with the upper and lower boundaries shown in cm:

| BinNumber: | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| LowerBound: | 9 | 12 | 15 | 18 | 21 | 24 | 27 | 30 | 33 | 36 | 39 | 42 | 45 | 50 | 55 | 60 | 65 | 70 | 75 | 80 | 85 | 90 | 95 | 100 | 105 |
| UpperBound: | 11 | 14 | 17 | 20 | 23 | 26 | 29 | 32 | 35 | 38 | 41 | 44 | 49 | 54 | 59 | 64 | 69 | 74 | 79 | 84 | 89 | 94 | 99 | 104 | 110 |

The collections of relative length frequencies are shown by year and size bin for the trawl fishery in Tables 2.7a, 2.7b, and 2.7c; the longline fishery in Tables 2.8a, 2.8b, and 2.8c; and the pot fishery in Tables 2.9a, 2.9b, and 2.9c.

## Catch Per Unit Effort

Catch per unit effort are available by gear and season for the years 1991-2007 and are shown below (units are $\mathrm{kg} / \mathrm{hr}$ for trawl gear, $\mathrm{kg} / \mathrm{hook}$ for longline gear, and $\mathrm{kg} /$ pot for pot gear):

|  | Trawl |  |  |  | Longline |  |  | Pot |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| Year | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 |  |
| 1991 | 58.85 | 49.23 | 24.40 | 1.02 | 0.71 | 0.55 |  | 68.53 | 103.16 |  |
| 1992 | 49.11 | 104.89 | 30.40 | 0.80 | 0.50 | 0.49 | 76.14 | 49.20 | 26.94 |  |
| 1993 | 51.00 | 50.84 | 106.54 | 0.66 | 0.35 |  | 85.53 |  |  |  |
| 1994 | 51.41 | 45.97 | 51.01 | 0.73 |  | 0.58 | 86.03 |  | 97.17 |  |
| 1995 | 62.04 | 57.44 | 62.76 | 0.86 |  | 0.60 | 85.19 | 69.59 | 52.18 |  |
| 1996 | 35.70 | 36.59 | 32.98 | 0.81 |  | 0.54 | 67.67 | 53.19 | 49.54 |  |
| 1997 | 51.20 | 32.77 | 75.73 | 0.87 |  | 0.58 | 76.71 | 47.20 | 46.56 |  |
| 1998 | 36.26 | 27.95 | 43.34 | 0.74 |  | 0.44 | 63.17 | 46.65 | 32.70 |  |
| 1999 | 37.54 | 16.67 | 20.80 | 0.68 | 0.46 | 0.50 | 53.98 | 40.13 | 37.50 |  |
| 2000 | 32.73 | 14.17 | 22.40 | 0.68 | 0.49 | 0.40 | 51.31 |  |  |  |
| 2001 | 22.42 | 45.65 | 14.42 | 0.56 | 0.44 | 0.41 | 70.30 |  | 46.11 |  |
| 2002 | 29.70 | 31.72 | 16.28 | 0.68 | 0.39 | 0.37 | 67.55 |  | 44.93 |  |
| 2003 | 26.91 | 33.46 | 21.86 | 0.52 | 0.35 | 0.35 | 73.86 |  | 58.16 |  |
| 2004 | 50.06 | 29.76 | 16.12 | 0.56 | 0.34 | 0.36 | 76.48 |  | 51.93 |  |
| 2005 | 45.06 | 21.12 |  | 0.64 | 0.36 | 0.35 | 86.93 |  | 46.12 |  |
| 2006 | 41.58 | 25.74 |  | 0.76 | 0.43 | 0.37 | 87.16 |  | 52.40 |  |
| 2007 | 37.82 | 47.98 |  | 0.73 | 0.48 |  | 64.63 |  | 64.24 |  |

## Survey Data

## EBS Shelf Bottom Trawl Survey

The relative size compositions from bottom trawl surveys of the EBS shelf conducted by the Alaska Fisheries Science Center since 1979 are shown in Tables 2.10a for the years 1979-1981 and 2.10b for the years 1982-2007, using the same length bins defined above for the commercial catch size compositions. The survey is shown as two separate time series because of a gear change that was instituted in 1982.
Following a decade-long hiatus in production ageing of Pacific cod, the Age and Growth Unit of the Alaska Fisheries Science Center began ageing samples of Pacific cod from the EBS shelf bottom trawl surveys a few years ago (Roberson 2001, Roberson et al. 2005). To date, the otolith collections from the 1994-2006 surveys have been read. The relative age compositions from these surveys are shown in Table 2.11. The number of fish aged for each of these years is shown below:

| Year: | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| $\mathrm{~N}:$ | 715 | 599 | 252 | 719 | 635 | 860 | 864 | 950 | 947 | 1360 | 1040 | 609 | 1301 |

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl surveys are shown in Table 2.12a (1979-1981) and 2.12b (1982-2007), together with their respective standard errors. Upper and lower $95 \%$ confidence intervals are also shown for the biomass estimates. Survey results indicate that biomass increased steadily from 1978 through 1983, and then remained relatively constant from 1983 through 1988. The highest biomass ever observed by the survey was the 1994 estimate of $1,368,120 \mathrm{t}$. Following the high observation in 1994, the survey biomass estimate declined steadily through 1998. The survey biomass estimates remained in the 596,000-619,000 t range from 2002 through 2005. The 2006 biomass estimate was $517,698 \mathrm{t}$, a $14 \%$ drop from the 2005 value and the second lowest estimate in the time series. This was followed by another drop in 2007, when the survey produced a biomass estimate of $423,703 \mathrm{t}$, an $18 \%$ drop from the 2006 value. The 2007 value is the lowest estimate in the time series. However, the 2007 estimate of numeric abundance moved in the opposite direction, an $86 \%$ increase over the 2006 estimate and the highest numeric abundance since 2001.

## EBS Slope Bottom Trawl Survey

The Alaska Fisheries Science Center conducted bottom trawl surveys of the EBS slope in 2002 and 2004. The biomass estimates and standard errors from the 2002 and 2004 surveys are shown below (all figures are in $t$ ):

| Year | Biomass | Standard Error |
| ---: | ---: | ---: |
| 2002 | 7511 | 1944 |
| 2004 | 5756 | 968 |

Because the survey estimates of Pacific cod biomass on the slope are so small (on the order of $1 \%$ of the shelf biomass estimates), the slope survey data are not used in the BSAI Pacific cod assessment.

## Aleutian Bottom Trawl Survey

Biomass estimates for the Aleutian Islands region were derived from U.S.-Japan cooperative bottom trawl surveys conducted during the summers of 1980, 1983, and 1986, and by U.S. bottom trawl surveys of the same area in 1991, 1994, 1997, 2000, 2002, 2004, and 2006. These surveys covered both the Aleutian management area ( 170 degrees east to 170 degrees west) and a portion of the Bering Sea management area ("Southern Bering Sea") not covered by the EBS shelf bottom trawl surveys. The time series of biomass estimates from the overall Aleutian survey area are shown together with their sum below (all estimates are in t :

| Year | Survey Type | Aleutian Survey Area |
| :---: | :---: | ---: |
| 1980 | U.S.-Japan | 148,272 |
| 1983 | U.S.-Japan | 215,755 |
| 1986 | U.S.-Japan | 255,072 |
| 1991 | U.S. | 191,049 |
| 1994 | U.S. | 184,068 |
| 1997 | U.S. | 83,416 |
| 2000 | U.S. | 136,028 |
| 2002 | U.S. | 82,970 |
| 2004 | U.S. | 114,161 |
| 2006 | U.S. | 92,526 |

For many years, the assessments of Pacific cod in the BSAI used a weighted average formed from EBS and Aleutian survey biomass estimates to provide a conversion factor which was used to translate model projections of EBS catch and biomass into BSAI equivalents. Prior to the 2004 assessment, the weighted average was based on the sums of the biomass estimates from the EBS shelf and AI survey biomass time series. However, in December of 2003 the SSC requested that alternative methods of estimating relative biomass between the EBS and AI be explored. Following a presentation of some possible alternatives, the SSC recommended that an approach based on a simple Kalman filter be used (SSC Minutes, October, 2004). In the 2006 assessment, the Kalman filter approach was applied to the updated (through 2006) time series, indicating that the best estimate of the current biomass distribution is $84 \% \mathrm{EBS}$ and $16 \% \mathrm{AI}$ (the previous proportions were $85 \%$ and $15 \%$, respectively).

## IPHC Longline Survey

The International Pacific Halibut Commission (IPHC) conducts an annual longline survey designed to estimate the relative abundance of Pacific halibut (Hippoglossus stenolepis). The survey also takes Pacific cod incidentally. The CPUE time series (number of Pacific cod per hook) since 1998 is as follows:

| 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 0.160 | 0.104 | 0.112 | 0.103 | 0.076 | 0.096 | 0.104 | 0.101 | 0.116 | 0.107 |

Pacific cod length composition data have not been taken historically in the IPHC survey. However, during this year's survey, IPHC staff made a special effort to collect Pacific cod size composition data. A total of 2785 lengths were collected, distributed among length bins as follows (no Pacific cod were recorded below length bin 12):

| Bin: | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| No.: | 2 | 11 | 55 | 238 | 599 | 720 | 440 | 240 | 144 | 122 | 99 | 60 | 39 | 16 |

## ANALYTIC APPROACH

## Model Structure

## History of Model Structures Developed Under Stock Synthesis 1 and 2

Beginning with the 1993 SAFE report (Thompson and Methot 1993) and continuing through the 2004 SAFE report (Thompson and Dorn 2004), a model using the Stock Synthesis 1 (SS1) assessment program (Methot 1986, 1990, 1998, 2000) and based largely on length-structured data formed the primary analytical tool used to assess the EBS Pacific cod stock. It should be emphasized that the model has always been intended to assess only the EBS portion of the BSAI stock. Conversion of model estimates of EBS biomass and catch to BSAI equivalents has traditionally been accomplished by application of an expansion factor based on the relative survey biomasses between the EBS and AI.

SS1 is a program that used the parameters of a set of equations governing the assumed dynamics of the stock (the "model parameters") as surrogates for the parameters of statistical distributions from which the data are assumed to be drawn (the "distribution parameters"), and varies the model parameters systematically in the direction of increasing likelihood until a maximum is reached. The overall likelihood is the product of the likelihoods for each of the model components. In part because the overall likelihood can be a very small number, SS1 uses the logarithm of the likelihood as the objective function. Each likelihood component is associated with a set of data assumed to be drawn from statistical distributions of the same general form (e.g., multinomial, lognormal, etc.). Typically, likelihood components are associated with data sets such as catch size (or age) composition, survey size (or age) composition, and survey abundance (either biomass or numbers, either relative or absolute).

SS1 permits each data time series to be divided into multiple segments, resulting in a separate set of parameter estimates for each segment. The EBS Pacific cod assessments, for example, have usually divided the shelf bottom trawl survey size composition time series into pre-1982 and post-1981 segments to account for the effects of a change in the trawl survey gear instituted in 1982. Also, to account for possible differences in selectivity between the mostly foreign (also joint venture) and mostly domestic fisheries, the fishery size composition time series have traditionally been split into pre-1989 and post1988 segments.

In the EBS Pacific cod model, each year has traditionally been partitioned into three seasons: JanuaryMay, June-August, and September-December (these seasonal boundaries were suggested by industry participants). Four fisheries have traditionally been defined: The January-May trawl fishery, the JuneDecember trawl fishery, the longline fishery, and the pot fishery.

Following a series of modifications from 1993 through 1997, the base model for EBS Pacific cod remained completely unchanged from 1997 through 2001. During the late 1990s, a number of attempts were made to estimate the natural mortality rate $M$ and the shelf bottom trawl survey catchability coefficient $Q$, but these were not particularly successful and the Plan Team and SSC always opted to
retain the base model in which $M$ and $Q$ were fixed at their traditional values of 0.37 and 1.0 , respectively.

A minor modification of the base model was suggested by the SSC in 2001, namely, that consideration be given to dividing the domestic era into pre-2000 and post-1999 segments. This modification was tested in the 2002 assessment (Thompson and Dorn 2002), where it was found to result in a statistically significant improvement in the model's ability to fit the data. In the 2004 assessment (Thompson and Dorn 2004), further modifications were made to the base model. The 2004 model included a set of selectivity parameters for the EBS slope bottom trawl survey and added new likelihood components for the age compositions and length-at-age data from the 1998-2003 EBS shelf bottom trawl surveys and the size composition and biomass data from the 2002 and 2004 EBS slope bottom trawl surveys. Incorporation of age data and slope survey data had been suggested by the SSC (SSC minutes, December 2003).

A major change took place in the 2005 assessment (Thompson and Dorn 2005), as the model was migrated to the newly developed Stock Synthesis 2 (SS2) program, which makes use of the ADMB modeling architecture (Fournier 2005) currently used in most age-structured assessments of BSAI and GOA groundfish. The move to SS2 facilitated improved estimation of model parameters as well as statistical characterization of the uncertainty associated with parameter estimates and derived quantities such as spawning biomass. Technical details of SS2 were described by Methot $(2005,2007)$.

The 2006 assessment (Thompson et al. 2006) explored alternative functional forms for selectivity, use of Pacific cod incidental catch data from the NMFS sablefish longline survey, and the influence of prior distributions. After the assessment was completed, an external reviewer discovered that the authors’ recommended model, which was also accepted by the Plan Team as the preferred model, had converged at a local rather than global minimum. This led the SSC to request that a technical workshop be held to explore alternative models.

## Review of Technical Workshop and Preliminary SAFE Report

Many changes have been made in the stock assessment model since last year's assessment. Some of these are described in the report of the technical workshop requested by the SSC, which was held in April of this year (Thompson and Conners 2007). For example, the base model developed for the technical workshop differed with respect to last year's author-recommended model in the following respects:

| Feature | Last year's assessment | Technical workshop base model |
| :--- | :--- | :--- |
| Software | SS2 version 1.23d | SS2 verson 2.00c |
| Natural mortality rate $(M)$ | Fixed at 0.37 | Estimated internally |
| Length-at-age parameters | Estimated externally | Estimated internally |
| Pre-shift median recruitment | Estimated iteratively | Estimated internally |
| First year included in model | 1964 | 1976 |
| First year of current regime | 1977 | 1976 |
| No. initial year classes estimated | None (all set at equilibrium) | 10 |
| Selectivity pattern | 4-parameter double normal | 6-parameter double normal |
| Bounds on log catchability | Essentially unbounded | Low=2×ln(0.75), high=0 |
| Bounds on selectivity parameters | Essentially unbounded | Most have low=-10, high=10 |
| Form of prior distributions | All normal | Normal and symmetric beta |

Dozens of other models were also considered at the workshop. Most of these were developed in advance of the workshop to address specific requests made by the SSC pertaining to catchability, natural mortality, estimation of growth parameters inside the model, use of asymptotic selectivity, and use of prior
distributions. Other models were developed during the workshop itself in response to suggestions made by workshop participants.

Further changes were made between the technical workshop and this year's preliminary SAFE Report (Thompson et al. 2007). Much of the effort in developing these models was directed toward ensuring that the estimation routine had reached the global maximum in the absence of informative prior distributions. Generally, it was necessary to choose starting values for model parameters that were close to the final (estimated) values.

Four models were presented in the preliminary SAFE Report. The following were among the ways in which Model 1 from the preliminary SAFE Report differed from the base model developed for the technical workshop:

| Feature | Technical workshop base model | Model 1 from preliminary SAFE |
| :---: | :---: | :---: |
| Software | SS2 version 2.00c | SS2 version 2.00i |
| Selectivities forced to be asymptotic | None | January-May trawl fishery |
| Time-varying selectivity | Fishery selectivities constant within blocks, surveys constant | Fishery and survey selectivities variable across years |
| Time-varying length at age 1 | Constant | Variable across years |
| Std. dev. of log-scale recruitment deviations ( $\sigma_{R}$ ) | $\sigma_{R}$ set iteratively | $\sigma_{R}=0.6$ |
| No. initial year classes estimated | 10 | 3 |
| Variability in length at age | $\mathrm{CV}=$ function of length at age | SD = function of age |
| Slope trawl survey data | Included | Excluded |
| Fishery CPUE data | Excluded | Included for comparison only |
| Form of age data | Marginal age compositions | Age-at-length compositions |
| Form of prior distributions | Normal and symmetric beta | All uniform |
| Bounds on log catchability | Low $=2 \times \ln (0.75)$, high $=0$ | Low=-2, high=2 |
| Bound parameters, if any | Parameters did not approach bounds due to priors | Fixed at bound (i.e., taken out of the estimation process) |

The other three models in the preliminary SAFE Report considered the effects of increasing natural mortality as a function of age, giving a large emphasis to fitting the longline fishery CPUE data, starting the model in 1982 rather than 1976, ignoring the age data, and iteratively adjusting the average input multinomial sample sizes and root-mean-squared-errors of the abundance indices.

The use of age-at-length composition data rather than marginal age compositions in three of the models presented in the preliminary SAFE Report was a response in part to a finding from the technical workshop. In the models presented at the technical workshop (which used marginal age compositions), the mean lengths estimated for the first few ages seemed to be distinctly different from the first few modes in the long-term average size composition from the bottom trawl survey. However, use of age-atlength composition data failed to eliminate this discrepancy.

## Model Structures Considered in This Year's Assessment

Four models are presented in this assessment. Relative to Model 1 from the preliminary SAFE Report, the following changes have been made in the base model presented in this assessment (Model 1):

| Feature | Model 1 from preliminary SAFE | New base model (Model 1) |
| :--- | :--- | :--- |
| Natural mortality rate | Estimated internally | Fixed at 0.34 |
| Form of age data | Age-at-length compositions | Marginal age compositions |
| Basis of maturity schedule | Length | Age |
| Basis of trawl survey selectivity | Length | Age |
| Seasonal structure | Partially seasonal | Fully seasonal |
| Selectivities forced to be | January-May trawl fishery | Jun-Aug trawl, Sep-Dec trawl, |
| asymptotic |  | Jun-Aug longline, Sep-Dec pot |
| Time-varying selectivity; std. | Fisheries and surveys variable | Fisheries constant, surveys |
| dev. of selectivity deviations $\left(\sigma_{S}\right)$ | across years; $\sigma_{S}=0.4$ | variable across years; $\sigma_{S}=0.2$ |
| Time-varying length at age 1 | Variable across years | Constant across years |
| Variability in length at age | SD = function of age | SD = function of length at age |
| Survey abundance units | Biomass | Numbers |
| Multinomial sample size; records | Square root of actual sample size; | Scaled bootstrap harmonic mean; |
| with small actual sample sizes | all records used | records with small N excluded |
| Std. dev. of log-scale recruitment | $\sigma_{R}=0.6$ | $\sigma_{R}$ set iteratively |
| deviations $\left(\sigma_{R}\right)$ |  |  |
| First year included in model | 1976 | 1977 |
| Starting year of regime shift | 1976 | 1977 |
| Parameters with large std. dev. | Left free | Fixed at their respective MLEs |

Model 1 was developed partly in response to requests from the SSC for inclusion of a model in which the natural mortality rate was based on other life history parameters (see below) and for inclusion of a model in which marginal age compositions are used. The decision to set the standard deviation of time-varying selectivity parameters at 0.2 was based on Francis et al. (2003).

Model 2 is the same as Model 1 except that the natural mortality rate is fixed at 0.37 . This model is included in response to another SSC request (the SSC asked that this model be included "purely for purposes of comparison").

Model 3 is the same as Model 1 except that the natural mortality rate is estimated internally. This model is included to determine whether the available data are sufficient to permit estimation of this parameter.

Model 4 differs from Model 1 in several respects. It estimates the natural mortality rate internally, constrains survey selectivities to be asymptotic, ignores the age data, ignores the late-1970s regime shift, starts the model in 1982 instead of 1977, bases maturity on length rather than age, bases survey selectivity on length rather than age, allows more survey selectivity parameters to have annual deviations, sets the standard deviation of selectivity parameter deviations at 0.4 rather than 0.2 , and ignores the initial catch in estimating the initial fishing mortality rate. This model is included in response to public comment.

## Parameters Estimated Independently

## Natural Mortality

In the 1993 BSAI Pacific cod assessment (Thompson and Methot 1993), the natural mortality rate $M$ was estimated using SS1 at a value of 0.37 . Although attempts have been made to re-estimate $M$ in some years (during the late 1990s and, most recently, in the 2005 assessment (Thompson and Dorn 2005)), all models of the BSAI Pacific cod stock accepted by the Plan Team and SSC since 1993 have ultimately
retained a value of 0.37 for $M$, as have all subsequent assessments of the GOA Pacific cod stock (with one exception, in 1995). Other published estimates of $M$ for Pacific cod are shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

As noted above, two of the models in the present assessment estimate $M$ independently. Model 1 fixes $M$ at a value of 0.34, based on Equation 7 of Jensen (1996) and an age at maturity of 4.9 years (Stark 2007). Model 2 fixes $M$ at the traditional value of 0.37 . Models 3 and 4 estimate $M$ conditionally.

## Variability in Estimated Age

Variability in estimated age in SS2 is based on the standard deviation of estimated age. Weighted least squares regression was used in the 2005 and 2006 assessments to estimate a proportional relationship between standard deviation and age. The regression was not recomputed this year, because only two new years' worth of age data were available. The estimated proportionality is 0.103 (i.e, the standard deviation of estimated age was modeled as $0.103 \times$ age).

## Weight at Length

Parameters governing the allometric relationship between weight (kg) and length (cm) were re-estimated in the 2006 assessment by log-log regression from the same data used to estimate the parameters of the length-at-age relationship. The parameter values were: multiplicative constant $=3.86 \times 10^{-6}$, and exponent $=3.266$. These were not re-estimated in the present assessment.

## Maturity

A detailed history and evaluation of parameter values used to describe the maturity schedule for BSAI Pacific cod was presented in the 2005 assessment (Thompson and Dorn 2005). A length-based maturity schedule has been used for many years. The parameter values used for this schedule in the 2005 and 2006 assessments were set on the basis of a study by Stark (2007) at the following values: length at $50 \%$ maturity $=58 \mathrm{~cm}$ and slope of linearized logistic equation $=-0.132$. The same parameter values are used for Model 4 in the present assessment. However, recent changes in SS2 allow for use of either a lengthbased or an age-based maturity schedule. Models 1-3 in the present assessment use an age-based schedule with intercept $=4.9$ years and slope $=-0.965$ (Stark 2007). The use of an age-based rather than a length-based schedule follows a recommendation from the author of the maturity study from which the parameter values were taken (James Stark, Alaska Fisheries Science Center, personal communication).

## Parameters Estimated Conditionally

Parameters estimated conditionally (i.e., within individual SS2 runs, based on the data and the parameters estimated independently) include the natural mortality rate (Models 3 and 4), length-at-age parameters, parameters governing variability in length at age, log median recruitment, initial fishing mortality, survey catchability, selectivity parameters, annual recruitment deviations, and annual deviations in one (Models 1-3) or two (Model 4) parameters governing the ascending limb of the trawl survey selectivity schedule.

A new, "recommended" (Methot 2007) selectivity function has been implemented for the present assessment, as it was at the technical workshop and in the preliminary SAFE Report. One of the things that may have led to convergence problems with the 2006 Bering Sea assessment model was that the fourparameter double-normal selectivity function used in that assessment exhibited differentiability problems. The new form of the double-normal selectivity pattern is supposed to exhibit superior performance. As with the double-normal selectivity pattern used in last year's assessments, the new form is constructed from two underlying and rescaled normal distributions, with a horizontal line segment joining the two peaks. The new form uses the following six parameters:

1. Beginning of peak region (where the curve first reaches a value of 1.0)
2. Width of peak region (where the curve first departs from a value of 1.0)
3. Ascending "width" (equal to twice the variance of the underlying normal distribution)
4. Descending width
5. Initial selectivity (at minimum length/age; not used in old form)
6. Final selectivity (at maximum length/age; not used in old form)

All but the "beginning of peak region" parameter are transformed: The widths are log-transformed and the other parameters are logit-transformed.

For all parameters estimated within individual SS2 runs, the estimator used is the mode of the logarithm of the joint posterior distribution, which is in turn calculated as the sum of the logarithms of the parameter-specific prior distributions (see below) and the logarithm of the likelihood function.

In addition to the above, the full set of year-, season-, and gear-specific fishing mortality rates are also estimated conditionally, but not in the same sense as the above parameters. The fishing mortality rates are determined exactly rather than estimated statistically because SS2 assumes that the input total catch data are true values rather than estimates, so the fishing mortality rates can be computed algebraically given the other parameter values and the input catch data.

Uniform prior distributions were used for all parameters.

## Likelihood Components

All four models included likelihood components for trawl survey relative abundance, fishery and survey size composition, recruitment, and parameter deviations. In addition, Models 1-3 included likelihood components for age composition and initial catch (Model 4 did not attempt to fit either of these data types).

In SS2, emphasis factors are specified to determine which likelihood components receive the greatest attention during the parameter estimation process. As in previous assessments, each likelihood component in each model was given an emphasis of 1.0 in the present assessment.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery, and time period within the year. In the parameter estimation process, SS2 weights a given size composition observation (i.e., the size frequency distribution observed in a given year, gear/fishery, and period) according to the emphasis associated with the respective likelihood component and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which SS1 was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to compensate for
contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. For many years, previous Pacific cod assessments, assumed a multinomial sample size equal to the square root of the true length sample size, rather than the true length sample size itself. Given the true length sample sizes observed in the EBS Pacific cod data, this procedure tended to give values somewhat below 400 while still providing SS2 with usable information regarding the appropriate effort to devote to fitting individual length samples.

Although the "square root rule" for specifying multinomial sample sizes gives reasonable values, the rule itself is largely ad hoc. In an attempt to move toward a more statistically based specification, this year a bootstrap analysis of the available fishery length data from 1990-2006 was undertaken. The actual sample sizes are shown by year, gear, and season in Tabel 2.13a. The average actual sample size across all years, gears, and seasons is 37,375 . The harmonic mean sample sizes from the bootstrap analysis are shown in Table 2.13b. The harmonic means are smaller than the actual sample sizes (average $=8,244$ ), but still range well into the thousands. A multinomial sample size in the thousands would likely overemphasize the size composition data. As a compromise, it may be reasonable to scale the harmonic means proportionally. For example, Table 2.13c shows the harmonic means rescaled to achieve an average sample size of 300 . For comparison, the square roots of the actual sample sizes are shown in Table 2.13d, and the ratios of the rescaled harmonic means to the square roots are shown in Table 2.13e. In general, the rescaled harmonic means are greater than the square roots for the January-May trawl fishery and the longline fishery (all three seasons). The reverse is true for the June-August trawl fishery, the September-December trawl fishery, the June-August pot fishery, and the September-December pot fishery. The two methods result in fairly similar values for the January-May pot fishery. Overall, the rescaled harmonic means are larger than the square roots by a factor of about 2:1.

If the rescaled harmonic mean approach is adopted, the question remains of what to do about years not covered by the bootstrap analysis (2007 and pre-1990) and what to do about the survey samples. A possible solution is provided by noting the consistency of the ratios between the harmonic means (the raw harmonic means, not the rescaled harmonic means) and the actual sample sizes. These ratios are shown in Table 2.13f. For the years prior to 1999, the ratio is very consistently close to 0.16 , and for the years after 1998, the ratio is very consistently close to 0.34 . This consistency was used to specify the missing values as follows: For fishery data, the sample sizes for length compositions from years prior to 1999 were tentatively set at $16 \%$ of the actual sample size, and the sample sizes for length compositions from 2007 were tentatively set at $34 \%$ of the actual sample size. For the pre- 1982 trawl survey, length compositions were tentatively set at $16 \%$ of an assumed sample size of 10,000 . For the post-1981 trawl survey and IPHC survey length compositions, sample sizes were tentatively set at $34 \%$ of the actual sample size. Then, with sample sizes for fishery length compositions from 1990-2006 tentatively set at their bootstrap harmonic means (not rescaled), all sample sizes were adjusted proportionally so that the average was 300 . The resulted in the set of multinomial sample sizes shown in Table 2.14.

## Use of Age Composition Data in Parameter Estimation

Like the size composition data, the age composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery (in this case, the EBS shelf bottom trawl survey), and time period within the year (in this case, the June-August period). Input sample sizes for the multinomial distributions were computed by scaling the actual number of otoliths read in each year proportionally such that the average of the input sample sizes was equal to 300 .

To avoid double counting of the same data, Models 1-3 ignore length composition data from the EBS shelf bottom trawl surveys in years where age data are available.

## Use of Fishery CPUE and Survey Relative Abundance Data in Parameter Estimation

Fishery CPUE data are included in the models for comparative purposes only. Their respective catchabilities are estimated analytically, not statistically. The same is true for the relative abundance data from the IPHC longline survey.

For the trawl surveys, each year's survey abundance datum is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey abundance in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey abundance datum's standard error to the survey abundance datum itself serves as the distribution's coefficient of variation.

## Use of Recruitment Deviation "Data" in Parameter Estimation

The recruitment deviations likelihood component is different from traditional likelihoods because it does not involve "data" in the same sense that traditional likelihoods do. Instead, the log-scale recruitment deviation plays the role of the datum and the log-scale recruitment mean and $\sigma_{R}$ play the role of the parameters in a normal distribution, but, of course, all of these are treated as parameters by SS2 (although $\sigma_{R}$ is fixed).

## MODEL EVALUATION

As described above, four models are evaluated in the present assessment. Briefly, Model 1 fixes $M$ at a value of 0.34 based on life history theory, Model 2 fixes M at the traditional value of 0.37 , Model 3 estimates $M$ internally, and Model 4 differs with respect to the others in several ways.

All models appeared to converge successfully and the Hessian matrices from all models were positive definite. Once convergence appeared to be achieved, 50-100 additional runs were undertaken for each model with initial parameter values displaced from their converged values to provide additional assurance that another (better) solution did not exist.

## Overall Conclusions Common to All Models

Before choosing a preferred model, it is important to note that, in many respects, the descriptions of the stock provided by all of the models are, qualitatively at least, very similar. For example, Figure 2.2 compares the time series of numbers of age 0 fish (on a log scale) as estimated by all four models. The models are mostly in agreement as to the strongest year classes in the time series, particularly the 1977 year class (Model 4, which does not start the time series until 1979, does not estimate the strength of the 1977 year class). Models 1-3 agree that the 2001-2005 year classes currently appear to be weak (Model 4 shows 2001 and 2003-4 to be weak, with 2002 and 2005 just slightly above average). Figure 2.3 compares the time series of female spawning biomass as estimated by all four models. The overall shapes of all the estimated time series are again qualitatively similar, with the main difference being one of scale. All models show spawning biomass to have declined overall since about 1985, with recent short-term decline starting in about 2005.

## Comparing and Contrasting the Models

Tables 2.15a-c present summaries of some key results from the four models. In each row of these tables, the cell with the highest value is shaded green (light gray if the document is viewed in grayscale) and the cell with the lowest value is shaded pink (dark gray).

Table 2.15a is structured as follows:
Section 1: Parameter counts. This section enumerates the number of parameters, both fixed and estimated, by type. The number of free parameters ranges from 112-118, while the number of total (fixed and free) parameters ranges from 145-156.

Section 2: Aggregate likelihood components. In general, lower values are better than higher values. However, note that Model 4 does not use two of the aggregate likelihood components (age compositions and initial catch) and uses a different data set than the other three models, so values are not strictly comparable across models.

Section 3: Relative abundance likelihoods. The only likelihoods that are actually used in this section are the trawl survey likelihoods (pre-1982 and post-1981 in Models 1-3, post-1981 only in Model 4). The others are shown for comparative purposes only.

Section 4: Size composition likelihoods. The aggregate size composition likelihood is broken down by individual fishery/season and gear.
Table 2.15b is structured as follows:
Section 1: Life history and recruitment. This section contains the natural mortality rate, parameters governing the length-at-age relationship and variability in length at age, mean length at age for the trawl survey as measured in July for ages 1-3, median recruitment, and standard deviation of log recruits. The natural mortality rate ranged from 0.22 (Model 3) to 0.46 (Model 4). The mean lengths at ages 1-3 are presented so that they can be compared with the first three modes from the long-term average trawl survey size composition, which occur at 17,33 , and 45 cm . The standard deviation of $\log$ recruits was fixed at 0.78 for Models $1-3$, which is equal to the standard deviation of the estimated recruitment deviations from Model 1. The standard deviation of log recruits in Model 4 was fixed at 0.6.

Section 2: Catchability and selectivity. This section summarizes the estimated trawl survey catchabilities (on a log scale), the size/age at which selectivity first reaches a value of 1.0 and the selectivity at maximum size/age for each fishery/survey, and the average of the product of trawl survey catchability and trawl survey selectivity within the 60-81 cm range. Nichol et al. (2007) estimated that this quantity had an average value of 0.47 for 11 fish equipped with archival tags. For the four models, this quantity ranged from 0.56 (Model 1) to 1.08 (Model 3). Selectivities are also plotted for each fishery/season and survey for all four models in Figures 2.4a-d.

Table 2.15c is structured as follows:
Section 1: Examples of historical biomass. Because these are estimates rather than projections, and the data used in the model all pertain to the EBS rather than the BSAI, these values are for the EBS only. Total biomass, age 3+ biomass, and female spawning biomass are shown for three example years: 1977 (the starting year of all the models except Model 4), 1985 (the approximate year of peak spawning biomass in all of the models), and 2007 (the most recent year of data). In general, Model 1 tends to give the highest values and Model 3 the lowest.

Section 2: Projected biomass. These are projections for the overall BSAI stock. Total biomass, age 3+ biomass, and female spawning biomass are shown for 2008 and 2009. In general, Model 2 tends to give the highest values and Model 3 the lowest.

Section 3: Spawning biomass reference points. B100\%, B40\%, and B35\% are shown, along with the projected ratios of female spawning biomass to B100\% in 2008 and 2009. Model 3 gives the highest estimates of B100\%, B40\%, and B35\% and Model 4 the lowest. Conversely, Model 3 gives the lowest ratios and Model 4 the highest.

Section 4: ABC reference points. F40\%, FABC, ABC, and relative changes in ABC are shown for 2008 and 2009. For all of these quantities, Model 3 gives the lowest values and Model 4 the highest.

Section 5: OFL reference points. Analogous to the ABC reference points section.

Tables 2.16a-2.16c show the estimates and standard deviations for every parameter estimated by any of the models. A blank cell indicates that a parameter was fixed a priori, bound or had a high standard deviation (>10). Parameters that were bound or had high standard deviations were fixed and then removed from the estimation process. An entry of " $n / \mathrm{a}$ " means that a parameter is not applicable to a particular model.

Tables 2.17a and 2.17b provide alternative measures of how well the models are fitting the fishery CPUE and survey relative abundance data. Table 2.17a shows root mean squared errors (lower values are better) and Table 2.17b shows correlations between observed and estimated values. Generally, Models 1 and 3 give the lowest RMSEs and Model 4 the highest, while Model 3 gives the highest correlations and Models 2 and 4 the lowest. The correlations with the IPHC abundance data were particularly disappointing, with no model being able to achieve a correlation greater than -0.38 . Figures $2.5 a-\mathrm{d}$ plot observed versus estimated CPUE and relative abundance for all fisheries/seasons and surveys under all four models.

Tables 2.18a and 2.18b provide alternative measures of how well the models are fitting the size composition data (higher values are better). Table 2.18a compares median input sample size to median output sample size ("effective" sample size McAllister and Ianelli 1997), and Table 2.18b shows the same thing, but using means rather than medians. Note that Model 4 uses a different data set than the other three models, so results are not strictly comparable across models. This caveat notwithstanding, Model 4 tended to have the highest effective sample sizes and Model 2 the lowest.

For age composition data, the following table summarizes the input and output sample sizes (Model 4 does not use age composition data):

| Source: | Input | Model 1 | Model 2 | Model 3 |
| :--- | ---: | ---: | ---: | ---: |
| Median: | 309 | 123 | 111 | 157 |
| Mean: | 300 | 65 | 50 | 70 |

## Evaluation Criteria

Because all of the models seem to perform reasonably well in terms of fitting the data, the following criteria are therefore proposed for this year's assessment:

1) The model should assume or estimate a reasonable value for $M$.
2) The model should estimate mean trawl survey lengths for ages 1-3 that are close to the first three modes from the long-term average trawl survey size composition.
3) The model should estimate a reasonable average for the product of trawl survey catchability and trawl survey selectivity for the $60-81 \mathrm{~cm}$ size range.
It should be understood that the above criteria are not proposed as absolutes, but rather as useful guidelines for the present assessment while model structure is being refined.

## Selection of Final Model

Criterion \#1 argues against choosing Models 3 or 4. If the life history theory published by Jensen (1996) and the age of maturity published by Stark (2007) are accurate, $M$ should be close to be 0.34 . The values of $M$ estimated by Models $3(0.22)$ and $4(0.46)$ are sufficiently different from this value that they should not be adopted without further investigation. It may also be noted that the SSC has recently expressed skepticism about the traditional $M$ of 0.37 (Model 2), suggesting that it should be included in the present assessment "purely for purposes of comparison."

Criterion \#2 does not rule out any of the four models.

Criterion \#3 favors Model 1. The estimate obtained by Nichol et al. (2007) is 0.47 , and the value closest to that is obtained by Model 1. A bootstrap analysis of Nichol et al.'s data indicate that Model 1's estimate of 0.56 falls within the $95 \%$ confidence interval, but the other models' estimates do not.

By process of elimination, then, Model 1 is recommended as the preferred model.
Final Parameter Estimates and Associated Schedules
Final estimates of all statistically estimated parameters in Model 1 are shown in Tables 2.16a-c.
Estimates of year-, gear-, and season-specific fishing mortality rates from Model 1 are shown in Table 2.19.

Schedules of selectivity at length/age from Model 1 are shown in Table 2.20. As noted previously, these are plotted in Figure 2.4a.

Schedules of length at age and weight at age for the population, each fishery/season, and each survey from Model 1 are shown in Tables 2.21 and 2.22, respectively.

## RESULTS

## Definitions

The biomass estimates presented here will be defined in two ways: 1 ) age $3+$ biomass, consisting of the biomass of all fish aged three years or greater in January of a given year; and 2) spawning biomass, consisting of the biomass of all spawning females in a given year. The recruitment estimates presented here will be defined as numbers of age 0 fish in a given year. The fishing mortality rates presented here will be defined as full-selection, instantaneous fishing mortality rates expressed on a per annum scale.

## Biomass

Table 2.23 shows the time series of EBS (not expanded to BSAI) Pacific cod female spawning biomass for the years 1977-2007 as estimated last year and this year under Model 1. Both estimated time series are accompanied by their respective $95 \%$ confidence intervals.

The estimated time series of EBS age 3+ biomass and female spawning biomass from Model 1 are shown, together with the observed time series of trawl survey biomass (assuming a catchability of 1.0), in Figure 2.6. All three biomass trends show a declining trend for at least the last three years.

## Recruitment

Table 2.24 shows the time series of EBS (not expanded to BSAI) Pacific cod age 0 recruitment (1000s of fish) for the years 1977-2006 as estimated last year and this year under Model 1. Both estimated time series are accompanied by their respective $95 \%$ confidence intervals.

Model 1’s recruitment estimates for the entire time series (1977-2006) are shown in Figure 2.7, along with their respective $95 \%$ confidence intervals and the average for the current environmental regime. For the time series as a whole, the largest year class appears to have been the 1977 cohort. Other large cohorts include the 1978, 1982, 1984, 1989, 1992, 1996, and 1999 year classes. Of the 2001-2005 year classes, however, none have $95 \%$ confidence intervals that extend above the 1977-2006 average. One potential bright spot on the horizon is the 2006 year class, whose point estimate is currently the second highest in the time series. However, its confidence interval is very large, since the only data currently available to estimate its strength is the size composition data from the 2007 shelf trawl survey.

To date, it has not been possible to estimate a reliable stock-recruitment relationship for this stock. With the move to SS2, prospects for future estimation of such a relationship should improve, and one of the models developed for this year's technical workshop (Thompson and Conners 2007) included internal estimation of a stock-recruitment relationship. In the interim, Figure 2.8 is provided to give some indication of the possible relationship between stock and recruitment. The Ricker (1954) curve shown in this figure (fit by maximum likelihood, ignoring process error) is intended to be illustrative only, and is not recommended for management purposes.

## Exploitation

Table 2.25 shows the time series of EBS Pacific cod catch divided by age 3+ biomass for the years 19772007 as estimated last year and this year under Model 1.

Figure 2.9 plots the trajectory of relative fishing mortality and relative female spawning biomass from 1977 through 2007 based on Model 1, overlaid with the current harvest control rules (fishing mortality rates in the figure are standardized relative to $F_{35 \%}$ and biomasses are standardized relative to $B_{35 \%}$, per SSC request). The entire trajectory lies underneath the $F_{\text {OFL }}$ control rule except for the years 1977-1978. For the period since 1979, the entire trajectory also fell below the $\max F_{A B C}$ control rule, except for 1997, when the fishing mortality rate appears to have exceeded the retroactively calculated $m a x F_{A B C}$ by about 4\% It should also be noted that the current harvest control rules did not go into effect until 1999.

## PROJECTIONS AND HARVEST ALTERNATIVES

## Amendment 56 Reference Points

Amendment 56 to the BSAI Groundfish Fishery Management Plan (FMP) defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL ( $F_{O F L}$ ), the maximum permissible ABC, and the fishing mortality rate used to set the maximum permissible ABC. The fishing mortality rate used to set ABC ( $F_{A B C}$ ) may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the BSAI are managed under Tier 3 of Amendment 56. Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

$$
\begin{aligned}
& \text { 3a)Stock status: } B / B_{40 \%}>1 \\
& F_{\text {OFL }}=F_{35 \%} \\
& F_{\text {ABC }} \leq F_{40 \%} \\
& \text { 3b)Stock status: } 0.05<B / B_{40 \%} \leq 1 \\
& F_{\text {OFL }}=F_{35 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& F_{\text {ABC }} \leq F_{40 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& \text { 3c)Stock status: } \quad B / B_{40 \%} \leq 0.05 \\
& F_{\text {OFL }}=0 \\
& F_{A B C}=0
\end{aligned}
$$

Other useful biomass reference points which can be calculated using this assumption are $B_{100 \%}$ and $B_{35 \%}$, defined analogously to $B_{40 \%}$. These reference points are estimated as follows, based on Model 1:

| Reference point: | $B_{35 \%}$ | $B_{40 \%}$ | $B_{100 \%}$ |
| ---: | :---: | :---: | :---: |
| BSAI: | $473,000 \mathrm{t}$ | $540,000 \mathrm{t}$ | $1,350,000 \mathrm{t}$ |
| EBS: | $397,000 \mathrm{t}$ | $454,000 \mathrm{t}$ | $1,130,000 \mathrm{t}$ |

For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. For this assessment, the apportionment was based on Model 1's estimates of fishing mortality by gear for the five most recent complete years of data (2001-2006). The average fishing mortality rates for those years implied that total fishing mortality was divided among the three main gear types according to the following percentages: trawl $28.1 \%$, longline $60.7 \%$, and pot $0.112 \%$. This apportionment results in estimates of $F_{35 \%}$ and $F_{40 \%}$ equal to 0.37 and 0.31 , respectively.

## Specification of OFL and Maximum Permissible ABC

BSAI spawning biomass for 2008 is estimated by Model 1 at a value of $398,000 \mathrm{t}$. This is about $11 \%$ below the BSAI $B_{40 \%}$ value of $540,000 \mathrm{t}$, thereby placing Pacific cod in sub-tier "b" of Tier 3. Given this, Model 1 estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2007 and 2008 as follows:

| Quantity | Overfishing Level | Maximum Permissible ABC |
| :--- | ---: | ---: |
| 2008 BSAI catch: | $176,000 \mathrm{t}$ | $150,000 \mathrm{t}$ |
| 2009 BSAI catch | $190,000 \mathrm{t}$ | $162,000 \mathrm{t}$ |
| 2008 Fishing mortality: | 0.26 | 0.22 |
| 2009 Fishing mortality: | 0.26 | 0.22 |

The age 3+ biomass estimates for 2008 and 2009 from Model 1 are 1,080,000 t and 1,420,000 t.

## ABC Recommendation

## Review of Past Approaches

BSAI Pacific cod ABCs for the years 1998-2002 were based on a harvest strategy that attempted to address some of the statistical uncertainty in the assessment model, namely the uncertainty surrounding parameters the natural mortality rate $M$ and survey catchability $Q$ (Thompson and Dorn 1997, 1998, 1999). For the 2001-2002 ABCs, the strategy was simplified by assuming that the ratio between the recommended $F_{A B C}$ and $F_{40 \%}$ estimate given in the 1999 assessment ( 0.87 ) was an appropriate factor by which to multiply the current maximum permissible $F_{A B C}$ to obtain a recommended $F_{A B C}$ (Thompson and Dorn 2001). For the 2003 and 2004 ABCs, concerns regarding the performance of the assessment model led to a decision that kept ABC constant at the 2002 level of $223,000 \mathrm{t}$, well below the maximum permissible level estimated in the respective assessments (Thompson and Dorn 2002, 2003). In the 2004 assessment (Thompson and Dorn 2004), the maximum permissible value for the 2005 ABC was estimated to be 227,000 t, only slightly higher than the 2003-2004 ABCs of 223,000 t. Because the 2003-2004 "constant catch" ABCs were intended to provide a precautionary alternative to the model's maximum permissible ABCs, it seemed appropriate in the 2004 assessment to consider another method for recommending ABC . This method was based on a consideration of the mean-variance tradeoff associated with future catches predicted by the standard projection model, and resulted in a 2005 ABC of 206,000 t. In the 2005 assessment, the Plan Team and SSC selected a model that resulted in a maximum permissible ABC of $194,000 \mathrm{t}$, which was adopted as the 2006 ABC. Similarly, the maximum permissible ABC was selected in the 2006 assessment, giving an ABC of 176,000 t.

## Recommendation for 2008

Based on Model 1, the maximum permissible ABC (Tier 3b) for 2008 is $150,000 \mathrm{t}$. This would constitute a $15 \%$ decrease from the 2007 value of 176,000 , roughly paralleling the decrease in the trawl survey biomass estimate of $18 \%$. Because the stock is in Tier 3 b , added precaution is already built into the maximum ABC computation. Therefore, $150,000 \mathrm{t}$ is the recommended ABC for 2008. For comparison, a Tier 5 computation (using a BSAI biomass estimate of $516,000 \mathrm{t}$ based on the most recent EBS and AI bottom trawl surveys) would set the maximum permissible 2008 ABC at $132,000 \mathrm{t}$ if $M$ is assumed to be 0.34 and $143,000 \mathrm{t}$ if $M$ is assumed to be 0.37 .

## Area Allocation of Harvests

At present, ABC of BSAI Pacific cod is not allocated by area. However, the Council is presently considering the possibility of specifying separate harvests in the EBS and AI.

## Standard Harvest and Recruitment Scenarios and Projection Methodology

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with an estimated vector of 2007 numbers at age. This vector is then projected forward to the beginning of 2008 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2007. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2008, are as follow ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2007 recommended in the assessment to the max $F_{A B C}$ for 2007. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, the upper bound on $F_{A B C}$ is set at $F_{60 \%}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2002-2006 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{\text {TAC }}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be 1) above its MSY level in 2008 or 2) above $1 / 2$ of its MSY level in 2008 and above its MSY level in 2018 under this scenario, then the stock is not overfished.)

Scenario 7: In 2008 and 2009, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2020 under this scenario, then the stock is not approaching an overfished condition.)

## Projections and Status Determination

## Scenario Projections and Two-Year Ahead Overfishing Level

Projections corresponding to the standard scenarios are shown for Model 1 in Tables 2.26-2.31 (Table 2.26 combines scenarios 1 and 2 , which are redundant).

In addition to the seven standard harvest scenarios, Amendments 48/48 to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2008, it does not provide the best estimate of OFL for 2009, because the mean 2008 catch under Scenario 6 is predicated on the 2008 catch being equal to the 2008 OFL, whereas the actual 2008 catch will likely be less than the 2008 OFL. Table 2.15c contains the appropriate one- and two-year ahead projections for both ABC and OFL under any of the four models considered in the present assessment.

## Status Determination

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock overfished? This depends on the stock's estimated spawning biomass in 2008:
a. If spawning biomass for 2008 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2008 is estimated to be above $B_{35 \%}$ the stock is above its MSST.
c. If spawning biomass for 2008 is estimated to be above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 2.30). If the mean spawning biomass for 2018 is below $B_{35 \%}$, the stock is below its MSST. Otherwise, the stock is above its MSST.

Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7 (Table 2.31):
a. If the mean spawning biomass for 2010 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2010 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c. If the mean spawning biomass for 2010 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2020. If the mean spawning biomass for 2020 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

Based on the above criteria and Tables 2.30 and 2.31, the stock is not overfished and is not approaching an overfished condition.

## ECOSYSTEM CONSIDERATIONS

Attachment 2.1 contains a summary of new results from ecosystem models on the role of Pacific Cod in the Eastern Bering Sea and Aleutian Islands ecosystems. The material in the present section is largely unchanged from last year's assessment.

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Boldt (ed.), 2005). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g., Hare and Mantua 2000). In the present assessment, an attempt was made to estimate the change in median recruitment of EBS Pacific cod associated with the 1977 regime shift. According to Model 1, pre-1977 median recruitment was only about $20 \%$ of post-1976 median recruitment. Establishing a link between environment and recruitment within a particular regime is more difficult. In the 2004 assessment (Thompson and Dorn 2004), for example, the correlations between age 1 recruits spawned since 1977 and monthly values of the Pacific Decadal Oscillation (Mantua et al. 1997) were computed and found to be very weak.

The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), Lang et al. (2003), Westrheim (1996), and Yang (2004). The composition of Pacific cod prey varies to some extent by time and area. In terms of percent occurrence, some of the most important items in the diet of Pacific cod in the BSAI and GOA have been polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, some of the most important dietary items have been euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, some of the most important dietary items have been walleye pollock, fishery offal, yellowfin sole, and crustaceans. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include Pacific cod, halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

## Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Bycatch of Nontarget and "Other" Species

Bycatch of nontarget species and members of the "other species" group are shown in the following set of tables (for the 2003-2005 tables, the "hook and line" gear type includes both longline and jig gear): Tables 2.32a and 2.32b show bycatch for the EBS Pacific cod trawl fishery in 1997-2002 and 2003-2005, respectively. Tables 2.33a and 2.33b show bycatch for the EBS Pacific cod longline fishery in 1997-2002 and the EBS Pacific cod hook and line fishery in 2003-2005, respectively. Tables 2.34 a and 2.34 b show bycatch for the EBS Pacific cod pot fishery in 1997-2002 and 2003-2005, respectively. Tables 2.35a and 2.35b show bycatch for the AI Pacific cod trawl fishery in 1997-2002 and 2003-2005, respectively. Tables 2.36a and 2.36b show bycatch for the AI Pacific cod longline fishery in 1997-2002 and the AI Pacific cod hook and line fishery in 2003-2005, respectively. Tables 2.37 shows bycatch for the AI Pacific cod pot fishery in 1997-2002 (no data exist for this fishery in 2003-2005).

It is not clear how much bycatch of a particular species constitutes "too much" in the context of ecosystem concerns. As a first step toward possible prioritization of future investigation into this question, it might be reasonable to focus on those species groups for which a Pacific cod fishery had a bycatch in excess of $100 t$ and accounted for more than $10 \%$ of the total bycatch in at least two of the three most recent years. This criterion results in the following list of impacted species groups (an "X" indicates that the criterion was met for that area/species/gear combination).

| Area | Species group | Trawl | Hook and Line |
| :---: | :---: | :---: | :---: |
| EBS | Grenadier |  | X |
| EBS | Large sculpins | X | X |
| EBS | Misc. fish | X |  |
| EBS | Other sculpins |  | X |
| EBS | Shark |  | X |
| EBS | Skate |  | X |
| AI | Skate |  | X |

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery operates to some extent in the same geographic areas used by Steller sea lion as foraging grounds (Livingston (ed.), 2002).

The Fisheries Interaction Team of the Alaska Fisheries Science Center has been engaged in research to determine the effectiveness of recent management measures designed to mitigate the impacts of the Pacific cod fisheries (among others) on Steller sea lions. Results from studies conducted in 2002-2003 were summarized by Conners et al. (2004). These studies included a tagging feasibility study, which may evolve into an ongoing research effort capable of providing information on the extent and rate to which Pacific cod move in and out of various portions of Steller sea lion critical habitat. Nearly 6,000 cod with spaghetti tags were released, of which approximately 1,000 had been returned as of September, 2003.

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the hook and line fishery for Pacific cod (Tables 2.33b and 2.36b). Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross
(Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft . LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed sets was as follows:

| Gear | BS | AI | GOA |
| :--- | ---: | ---: | ---: |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513,517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort were dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.

Impacts of the Pacific cod fisheries on essential fish habitat were further analyzed in an environmental impact statement by NMFS (2005).

## Data Gaps and Research Priorities

Understanding of the above ecosystem considerations would be improved if future research were directed toward closing certain data gaps. Such research would have several foci, including the following: 1) ecology of the Pacific cod stock, including spatial dynamics, trophic and other interspecific relationships, and the relationship between climate and recruitment; 2) behavior of the Pacific cod fishery, including spatial dynamics; 3) determinants of trawl survey catchability and selectivity; 4) ecology of species taken as bycatch in the Pacific cod fisheries, including estimation of biomass, carrying capacity, and resilience; and 5) ecology of species that interact with Pacific cod, including estimation of biomass, carrying capacity, and resilience.

## SUMMARY

The major results of the Pacific cod stock assessment are summarized in Table 2.38.

## ACKNOWLEDGMENTS

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Table 2.1a—Summary of 1964-1980 catches (t) of Pacific cod in the Eastern Bering Sea by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

| Eastern Bering Sea only: |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: |
| Year | Foreign | Joint Venture | Domestic | Total |
| 1964 | 13408 | 0 | 0 | 13408 |
| 1965 | 14719 | 0 | 0 | 14719 |
| 1966 | 18200 | 0 | 0 | 18200 |
| 1967 | 32064 | 0 | 0 | 32064 |
| 1968 | 57902 | 0 | 0 | 57902 |
| 1969 | 50351 | 0 | 0 | 50351 |
| 1970 | 70094 | 0 | 0 | 70094 |
| 1971 | 43054 | 0 | 0 | 43054 |
| 1972 | 42905 | 0 | 0 | 42905 |
| 1973 | 53386 | 0 | 0 | 53386 |
| 1974 | 62462 | 0 | 0 | 62462 |
| 1975 | 51551 | 0 | 0 | 51551 |
| 1976 | 50481 | 0 | 0 | 50481 |
| 1977 | 33335 | 0 | 0 | 33335 |
| 1978 | 42512 | 0 | 31 | 42543 |
| 1979 | 32981 | 8370 | 780 | 33761 |
| 1980 | 35058 |  | 2433 | 45861 |

Table 2.1b—Summary of 1981-2007 catches (t) of Pacific cod in the Eastern Bering Sea by fleet sector and gear type. All catches include discards. LLine = longline, Subt. = sector subtotal. Catches for 2007 are through early October.

Eastern Bering Sea only:

|  | Foreign |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |
| 1981 | 30347 | 5851 | 36198 | 7410 | 7410 | 12884 | 1 | 0 | 14 | 12899 | 56507 |  |  |
| 1982 | 23037 | 3142 | 26179 | 9312 | 9312 | 23893 | 5 | 0 | 1715 | 25613 | 61104 |  |  |
| 1983 | 32790 | 6445 | 39235 | 9662 | 9662 | 45310 | 4 | 21 | 569 | 45904 | 94801 |  |  |
| 1984 | 30592 | 26642 | 57234 | 24382 | 24382 | 43274 | 8 | 0 | 205 | 43487 | 125103 |  |  |
| 1985 | 19596 | 36742 | 56338 | 35634 | 35634 | 51425 | 50 | 0 | 0 | 51475 | 143447 |  |  |
| 1986 | 13292 | 26563 | 39855 | 57827 | 57827 | 37646 | 48 | 62 | 167 | 37923 | 135605 |  |  |
| 1987 | 7718 | 47028 | 54746 | 47722 | 47722 | 46039 | 1395 | 1 | 0 | 47435 | 149903 |  |  |
| 1988 | 0 | 0 | 0 | 106592 | 106592 | 93706 | 2474 | 299 | 0 | 96479 | 203071 |  |  |
| 1989 | 0 | 0 | 0 | 44612 | 44612 | 119631 | 13935 | 145 | 0 | 133711 | 178323 |  |  |
| 1990 | 0 | 0 | 0 | 8078 | 8078 | 115493 | 47114 | 1382 | 0 | 163989 | 172067 |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 129392 | 76734 | 3343 | 0 | 209469 | 209469 |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 77259 | 80174 | 7512 | 33 | 164978 | 164978 |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 81790 | 49295 | 2098 | 2 | 133185 | 133185 |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 84931 | 78566 | 8037 | 730 | 172264 | 172264 |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 110956 | 97665 | 19275 | 599 | 228496 | 228496 |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 91910 | 88882 | 28006 | 267 | 209064 | 209064 |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 93924 | 117008 | 21493 | 173 | 232598 | 232598 |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 60780 | 84323 | 13232 | 192 | 158526 | 158526 |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 51902 | 81463 | 12399 | 100 | 145865 | 145865 |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 53815 | 81640 | 15849 | 68 | 151372 | 151372 |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 35655 | 90360 | 16385 | 52 | 142452 | 142452 |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 51065 | 100269 | 15051 | 166 | 166552 | 166552 |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 47580 | 106967 | 21957 | 155 | 176659 | 176659 |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 57784 | 109692 | 17238 | 231 | 184945 | 184945 |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 52604 | 112994 | 17104 | 104 | 182807 | 182807 |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 53202 | 95485 | 18957 | 81 | 167725 | 167725 |  |  |
| 2007 | 0 | 0 | 0 | 0 | 0 | 45107 | 74338 | 16903 | 82 | 136430 | 136430 |  |  |

Table 2.2a-Summary of 1964-1980 catches (t) of Pacific cod in the Aleutian Islands region by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

Aleutian Islands region only:

| Year | Foreign | Joint Venture | Domestic | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1964 | 241 | 0 | 0 | 241 |
| 1965 | 451 | 0 | 0 | 451 |
| 1966 | 154 | 0 | 0 | 154 |
| 1967 | 293 | 0 | 0 | 293 |
| 1968 | 289 | 0 | 0 | 289 |
| 1969 | 220 | 0 | 0 | 220 |
| 1970 | 283 | 0 | 0 | 283 |
| 1971 | 2078 | 0 | 0 | 2078 |
| 1972 | 435 | 0 | 0 | 435 |
| 1973 | 977 | 0 | 0 | 977 |
| 1974 | 1379 | 0 | 0 | 1379 |
| 1975 | 2838 | 0 | 0 | 2838 |
| 1976 | 4190 | 0 | 0 | 4190 |
| 1977 | 3262 | 0 | 0 | 3262 |
| 1978 | 3295 | 0 | 0 | 3295 |
| 1979 | 5593 | 0 | 0 | 5593 |
| 1980 | 5788 | 0 | 0 | 5788 |

Table 2.2b—Summary of 1981-2007 catches (t) of Pacific cod in the Aleutian Islands region by fleet sector and gear type. All catches include discards. LLine = longline, Subt. = sector subtotal. Catches for 2007 are through early October.

Aleutian Islands region only:

|  | Foreign |  |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |  |  |
| 1981 | 2680 | 235 | 2915 | 1749 | 1749 | 2744 | 26 | 0 | 0 | 2770 | 7434 |  |  |  |  |
| 1982 | 1520 | 476 | 1996 | 4280 | 4280 | 2121 | 0 | 0 | 0 | 2121 | 8397 |  |  |  |  |
| 1983 | 1869 | 402 | 2271 | 4700 | 4700 | 1459 | 0 | 0 | 0 | 1459 | 8430 |  |  |  |  |
| 1984 | 473 | 804 | 1277 | 6390 | 6390 | 314 | 0 | 0 | 0 | 314 | 7981 |  |  |  |  |
| 1985 | 10 | 829 | 839 | 5638 | 5638 | 460 | 0 | 0 | 0 | 460 | 6937 |  |  |  |  |
| 1986 | 5 | 0 | 5 | 6115 | 6115 | 784 | 1 | 1 | 0 | 786 | 6906 |  |  |  |  |
| 1987 | 0 | 0 | 0 | 10435 | 10435 | 2662 | 22 | 88 | 0 | 2772 | 13207 |  |  |  |  |
| 1988 | 0 | 0 | 0 | 3300 | 3300 | 1698 | 137 | 30 | 0 | 1865 | 5165 |  |  |  |  |
| 1989 | 0 | 0 | 0 | 6 | 6 | 4233 | 284 | 19 | 0 | 4536 | 4542 |  |  |  |  |
| 1990 | 0 | 0 | 0 | 0 | 0 | 6932 | 602 | 7 | 0 | 7541 | 7541 |  |  |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 3414 | 3203 | 3180 | 0 | 9797 | 9797 |  |  |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 14558 | 22108 | 6317 | 84 | 43068 | 43068 |  |  |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 17312 | 16860 | 0 | 33 | 34204 | 34204 |  |  |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 14382 | 7009 | 147 | 0 | 21539 | 21539 |  |  |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 10574 | 4935 | 1024 | 0 | 16534 | 16534 |  |  |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 21179 | 5819 | 4611 | 0 | 31609 | 31609 |  |  |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 17349 | 7151 | 575 | 89 | 25164 | 25164 |  |  |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 20531 | 13771 | 424 | 0 | 34726 | 34726 |  |  |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 16437 | 7874 | 3750 | 69 | 28130 | 28130 |  |  |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 20362 | 16183 | 3107 | 33 | 39684 | 39684 |  |  |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 15826 | 17817 | 544 | 19 | 34207 | 34207 |  |  |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 27929 | 2865 | 7 | 0 | 30801 | 30801 |  |  |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 31478 | 974 | 2 | 0 | 32455 | 32455 |  |  |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 25766 | 3099 | 0 | 0 | 28865 | 28865 |  |  |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 19613 | 3001 | 0 | 13 | 22627 | 22627 |  |  |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 20054 | 3552 | 567 | 8 | 24181 | 24181 |  |  |  |  |
| 2007 | 0 | 0 | 0 | 0 | 0 | 28456 | 4635 | 626 | 7 | 33724 | 33724 |  |  |  |  |

Table 2.3a-Summary of 1964-1980 catches (t) of Pacific cod in the combined Eastern Bering Sea and Aleutian Islands region by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

Eastern Bering Sea and Aleutian Islands region combined:

| Year | Foreign | Joint Venture | Domestic | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1964 | 13649 | 0 | 0 | 13649 |
| 1965 | 15170 | 0 | 0 | 15170 |
| 1966 | 18354 | 0 | 0 | 18354 |
| 1967 | 32357 | 0 | 0 | 32357 |
| 1968 | 58191 | 0 | 0 | 58191 |
| 1969 | 50571 | 0 | 0 | 50571 |
| 1970 | 70377 | 0 | 0 | 70377 |
| 1971 | 45132 | 0 | 0 | 45132 |
| 1972 | 43340 | 0 | 0 | 43340 |
| 1973 | 54363 | 0 | 0 | 54363 |
| 1974 | 63841 | 0 | 0 | 63841 |
| 1975 | 54389 | 0 | 0 | 54389 |
| 1976 | 54671 | 0 | 0 | 54671 |
| 1977 | 36597 | 0 | 0 | 36597 |
| 1978 | 45807 | 0 | 31 | 45838 |
| 1979 | 38574 | 0370 | 780 | 39354 |
| 1980 | 40846 |  | 2433 | 51649 |

Table 2.3b—Summary of 1981-2006 catches ( t ) of Pacific cod in the combined Eastern Bering Sea and Aleutian Islands region by fleet sector and gear type. All catches include discards. LLine = longline, Subt. = sector subtotal. Catches for 2006 are through early October.

Eastern Bering Sea and Aleutian Islands region combined:

|  | Foreign |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |
| 1981 | 33027 | 6086 | 39113 | 9159 | 9159 | 15628 | 27 | 0 | 14 | 15669 | 63941 |  |  |
| 1982 | 24557 | 3618 | 28175 | 13592 | 13592 | 26014 | 5 | 0 | 1715 | 27734 | 69501 |  |  |
| 1983 | 34659 | 6847 | 41506 | 14362 | 14362 | 46769 | 4 | 21 | 569 | 47363 | 103231 |  |  |
| 1984 | 31065 | 27446 | 58511 | 30772 | 30772 | 43588 | 8 | 0 | 205 | 43801 | 133084 |  |  |
| 1985 | 19606 | 37571 | 57177 | 41272 | 41272 | 51885 | 50 | 0 | 0 | 51935 | 150384 |  |  |
| 1986 | 13297 | 26563 | 39860 | 63942 | 63942 | 38430 | 49 | 63 | 167 | 38709 | 142511 |  |  |
| 1987 | 7718 | 47028 | 54746 | 58157 | 58157 | 48701 | 1417 | 89 | 0 | 50207 | 163110 |  |  |
| 1988 | 0 | 0 | 0 | 109892 | 109892 | 95404 | 2611 | 329 | 0 | 98344 | 208236 |  |  |
| 1989 | 0 | 0 | 0 | 44618 | 44618 | 123864 | 14219 | 164 | 0 | 138247 | 182865 |  |  |
| 1990 | 0 | 0 | 0 | 8078 | 8078 | 122425 | 47716 | 1389 | 0 | 171530 | 179608 |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 132806 | 79937 | 6523 | 0 | 219266 | 219266 |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 91818 | 102282 | 13829 | 117 | 208046 | 208046 |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 99102 | 66155 | 2098 | 35 | 167389 | 167389 |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 99313 | 85575 | 8184 | 730 | 193802 | 193802 |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 121530 | 102600 | 20299 | 599 | 245029 | 245029 |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 113089 | 94701 | 32617 | 267 | 240673 | 240673 |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 111273 | 124159 | 22068 | 262 | 257762 | 257762 |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 81310 | 98094 | 13657 | 192 | 193253 | 193253 |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 68339 | 89337 | 16150 | 169 | 173995 | 173995 |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 74177 | 97823 | 18956 | 101 | 191056 | 191056 |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 51482 | 108177 | 16929 | 71 | 176659 | 176659 |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 78994 | 103134 | 15058 | 166 | 197352 | 197352 |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 79059 | 107941 | 21959 | 156 | 209114 | 209114 |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 83550 | 112790 | 17239 | 231 | 213810 | 213810 |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 72217 | 115995 | 17104 | 117 | 205434 | 205434 |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 73256 | 99037 | 19524 | 89 | 191906 | 191906 |  |  |
| 2007 | 0 | 0 | 0 | 0 | 0 | 73564 | 78973 | 17529 | 88 | 170154 | 170154 |  |  |

Table 2.4-History of Pacific cod ABC, TAC, total BSAI catch, and type of stock assessment model used to recommend ABC. Catch for 2007 is current through early October. "SS1" refers to Stock Synthesis 1 and "SS2" refers to Stock Synthesis 2. Each cell in the "Stock Assessment Model" column lists the type of model used to recommend the ABC in the corresponding row, meaning that the model was produced in the year previous to the one listed in the corresponding row.

| Year | ABC | TAC | Catch | Stock assessment model (from previous year) |
| ---: | ---: | ---: | ---: | ---: |
| 1980 | 148,000 | 70,700 | 45,947 | projection of 1979 survey numbers at age |
| 1981 | 160,000 | 78,700 | 63,941 | projection of 1979 survey numbers at age |
| 1982 | 168,000 | 78,700 | 69,501 | projection of 1979 survey numbers at age |
| 1983 | 298,200 | 120,000 | 103,231 | projection of 1979 survey numbers at age |
| 1984 | 291,300 | 210,000 | 133,084 | projection of 1979 survey numbers at age |
| 1985 | 347,400 | 220,000 | 150,384 | projection of 1979-1985 survey numbers at age |
| 1986 | 249,300 | 229,000 | 142,511 | separable age-structured model |
| 1987 | 400,000 | 280,000 | 163,110 | separable age-structured model |
| 1988 | 385,300 | 200,000 | 208,236 | separable age-structured model |
| 1989 | 370,600 | 230,681 | 182,865 | separable age-structured model |
| 1990 | 417,000 | 227,000 | 179,608 | separable age-structured model |
| 1991 | 229,000 | 229,000 | 219,266 | separable age-structured model |
| 1992 | 182,000 | 182,000 | 208,046 | SS1 model (age-based data) |
| 1993 | 164,500 | 164,500 | 167,389 | SS1 model (length-based data) |
| 1994 | 191,000 | 191,000 | 193,802 | SS1 model (length-based data) |
| 1995 | 328,000 | 250,000 | 245,029 | SS1 model (length-based data) |
| 1996 | 305,000 | 270,000 | 240,673 | SS1 model (length-based data) |
| 1997 | 30,000 | 270,000 | 257,762 | SS1 model (length-based data) |
| 1998 | 210,000 | 210,000 | 193,253 | SS1 model (length-based data) |
| 1999 | 177,000 | 177,000 | 173,995 | SS1 model (length-based data) |
| 2000 | 193,000 | 193,000 | 191,056 | SS1 model (length-based data) |
| 2001 | 188,000 | 188,000 | 176,659 | SS1 model (length-based data) |
| 2002 | 223,000 | 200,000 | 19,352 | SS1 model (length-based data) |
| 2003 | 223,000 | 207,500 | 20,114 | SS1 model (length-based data) |
| 2004 | 223,000 | 215,500 | 213,810 | SS1 model (length-based data) |
| 2005 | 206,000 | 206,000 | 164,404 | SS1 model (length- and age-based data) |
| 2006 | 194,000 | 194,000 | 191,906 | SS2 model (length- and age-based data) |
| 2007 | 176,000 | 170,720 | 170,154 | SS2 model (length- and age-based data) |

Table 2.5a—Pacific cod discard rates by area, target species/group, and year for the period 1991-2002 (see Table 2.5b for the period 2003-2004). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

| Eastern Bering Sea |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Target species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Arrowtooth flounder | 0.61 | 0.00 | 0.94 |  | 0.66 | 0.08 | 0.07 | 1.00 | 1.00 | 0.99 | 1.00 | 0.22 |
| Atka mackerel | 1.00 |  | 0.70 | 1.00 |  | 0.23 |  | 0.51 | 0.00 | 0.00 | 1.00 |  |
| Flathead sole |  |  |  |  | 0.39 | 0.58 | 0.10 | 0.75 | 0.87 | 0.75 | 0.00 | 1.00 |
| Greenland turbot | 0.01 | 0.00 | 0.12 | 0.04 | 0.35 | 0.09 | 0.03 | 0.04 | 0.13 | 0.10 | 0.01 | 0.18 |
| Other flatfish | 0.63 | 0.31 | 0.47 | 0.88 | 0.22 | 0.28 | 0.91 | 0.28 | 0.33 | 0.32 | 0.00 | 0.00 |
| Other species | 0.04 | 0.99 | 0.38 |  | 1.00 | 1.00 | 0.01 | 0.95 | 0.07 | 0.92 | 0.08 | 0.00 |
| Pacific cod | 0.03 | 0.04 | 0.08 | 0.06 | 0.07 | 0.04 | 0.03 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 |
| Pollock | 0.70 | 0.85 | 0.73 | 0.68 | 0.21 | 0.41 | 0.24 | 0.42 | 0.49 | 0.68 | 0.84 | 0.52 |
| Rock sole | 1.00 | 0.00 | 0.08 | 0.87 | 0.25 | 0.90 |  | 1.00 | 0.02 | 0.16 | 1.00 | 1.00 |
| Rockfish | 1.00 | 0.00 | 0.89 | 0.01 | 0.84 | 0.69 | 0.16 |  | 0.00 | 0.03 | 0.00 | 0.00 |
| Sablefish | 0.00 | 0.12 | 0.42 | 0.40 | 0.96 | 0.94 | 0.78 | 0.93 | 0.61 | 0.98 | 0.12 | 0.48 |
| Unknown | 0.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.49 | 0.04 | 0.02 |  |  |
| Yellowfin sole |  | 0.74 | 0.72 | 0.50 | 0.08 | 1.00 | 0.24 | 0.77 | 0.50 | 0.60 | 0.39 | 0.77 |
| All targets | 0.03 | 0.04 | 0.08 | 0.06 | 0.07 | 0.04 | 0.03 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 |
| Aleutian Islands |  |  |  |  |  |  |  |  |  |  |  |  |
| Target species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Arrowtooth flounder | 1.00 |  |  |  |  |  |  |  |  |  | 0.00 | 0.00 |
| Atka mackerel |  |  |  |  |  |  |  | 1.00 |  | 1.00 | 1.00 | 1.00 |
| Flathead sole |  | 0.35 |  |  |  |  |  |  |  |  |  |  |
| Greenland turbot | 0.11 | 0.00 | 0.73 | 0.58 | 0.40 | 0.89 | 0.04 | 0.01 | 0.18 | 0.40 | 0.00 | 0.00 |
| Other species |  | 1.00 |  |  | 0.00 |  |  |  | 0.14 | 0.08 | 0.00 | 0.06 |
| Pacific cod | 0.02 | 0.03 | 0.12 | 0.09 | 0.04 | 0.04 | 0.05 | 0.02 | 0.02 | 0.02 | 0.01 | 0.02 |
| Pollock | 0.76 | 0.00 | 0.29 | 0.00 | 0.47 | 0.74 | 0.75 | 0.61 | 0.00 |  |  |  |
| Rock sole |  |  | 0.00 |  |  |  |  |  |  |  |  |  |
| Rockfish | 0.83 |  | 0.75 | 0.28 | 0.18 | 0.80 | 0.91 | 1.00 | 0.64 | 0.12 | 0.22 | 0.03 |
| Sablefish | 1.00 | 0.04 | 0.49 | 0.52 | 0.97 | 0.53 | 0.70 | 0.88 | 0.51 | 0.31 | 0.06 | 0.76 |
| Unknown | 0.09 |  |  |  | 1.00 | 1.00 |  | 0.03 |  | 1.00 | 1.00 |  |
| All targets | 0.04 | 0.03 | 0.12 | 0.09 | 0.12 | 0.04 | 0.06 | 0.02 | 0.02 | 0.02 | 0.01 | 0.02 |

Table 2.5b—Pacific cod discard rates by area, target species/group, and year for the period 2003-2004 (see Table 2.5a for the period 1991-2002; note that the IFQ halibut target does not exist in Table 2.5a). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

|  | Eastern Bering Sea |  | Aleutian Islands |  |
| :--- | :---: | :---: | :---: | :---: |
| Target species/group | 2003 | 2004 | 2003 | 2004 |
| Arrowtooth flounder | 0.01 | 0.00 |  |  |
| Atka mackerel | 0.02 | 0.00 | 0.03 | 0.02 |
| Flathead sole | 0.00 | 0.02 |  |  |
| Greenland turbot | 0.07 | 0.05 | 0.00 | 0.38 |
| IFQ halibut | 0.28 | 0.28 | 0.58 |  |
| Other flattish | 0.02 | 0.00 | 0.00 |  |
| Other species | 0.02 | 0.04 | 0.01 | 0.01 |
| Pacific cod | 0.01 | 0.01 |  |  |
| Pollock | 0.00 | 0.02 | 0.11 | 0.02 |
| Rock sole | 0.08 | 0.03 | 0.00 | 0.06 |
| Rockfish | 0.00 | 0.00 | 0.37 |  |
| Sablefish | 0.44 | 0.03 |  |  |
| Unknown | 0.06 | 0.02 | 0.01 |  |
| Yellowfin sole | 0.02 | 0.01 | 0.01 |  |

Table 2.6—EBS catch ( t ) of Pacific cod by year, gear, and period for the years 1977-2007. Season 3 catch values for 2007 are extrapolations based on the previous year's catch. Because direct estimates of gear- and period-specific catches are not available for the years 1977-1980, the figures shown here are estimates derived by distributing each year's total catch according to the average proportion observed for each gear/period combination during the years 1981-1988.

| Year | Trawl Fishery |  |  |  | Longline Fishery |  |  |  | Pot Fishery |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
|  | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 |  |  |
| 1977 | 14935 | 6139 | 6858 | 1851 | 260 | 3292 | 0 | 0 | 0 |  |  |
| 1978 | 19710 | 8101 | 9051 | 2443 | 343 | 4344 | 0 | 0 | 0 |  |  |
| 1979 | 16131 | 6630 | 7407 | 1999 | 281 | 3555 | 0 | 0 | 0 |  |  |
| 1980 | 18387 | 7558 | 8444 | 2279 | 320 | 4053 | 0 | 0 | 0 |  |  |
| 1981 | 15067 | 14087 | 21486 | 1286 | 624 | 3942 | 0 | 0 | 0 |  |  |
| 1982 | 21742 | 18151 | 16348 | 363 | 475 | 2308 | 0 | 0 | 0 |  |  |
| 1983 | 40757 | 24300 | 22705 | 2941 | 748 | 2756 | 0 | 0 | 0 |  |  |
| 1984 | 48237 | 24964 | 25045 | 5012 | 2128 | 19508 | 0 | 0 | 0 |  |  |
| 1985 | 55673 | 28673 | 22310 | 13703 | 1710 | 21379 | 0 | 0 | 0 |  |  |
| 1986 | 59786 | 26598 | 22382 | 8895 | 438 | 17278 | 0 | 0 | 0 |  |  |
| 1987 | 64413 | 15604 | 21462 | 20947 | 723 | 26752 | 0 | 0 | 0 |  |  |
| 1988 | 127470 | 25662 | 47166 | 444 | 646 | 1385 | 90 | 51 | 160 |  |  |
| 1989 | 127459 | 16986 | 19798 | 3810 | 4968 | 5157 | 33 | 63 | 49 |  |  |
| 1990 | 101645 | 11402 | 10524 | 13171 | 16643 | 17299 | 0 | 986 | 395 |  |  |
| 1991 | 107979 | 15549 | 5863 | 25470 | 21472 | 29792 | 12 | 1042 | 2288 |  |  |
| 1992 | 59460 | 11840 | 5959 | 49696 | 24201 | 6276 | 2622 | 4632 | 258 |  |  |
| 1993 | 67148 | 5362 | 9280 | 49244 | 27 | 23 | 2073 | 24 | 0 |  |  |
| 1994 | 61009 | 5806 | 18115 | 57968 | 13 | 20585 | 4923 | 0 | 3113 |  |  |
| 1995 | 90366 | 8543 | 12047 | 68458 | 26 | 29180 | 12484 | 3469 | 3322 |  |  |
| 1996 | 78194 | 3126 | 10590 | 62011 | 26 | 26845 | 18143 | 6401 | 3462 |  |  |
| 1997 | 81313 | 3927 | 8684 | 70676 | 43 | 46290 | 14584 | 3576 | 3333 |  |  |
| 1998 | 45008 | 5603 | 10169 | 54234 | 18 | 30071 | 9022 | 2779 | 1432 |  |  |
| 1999 | 44904 | 3312 | 3686 | 55180 | 1923 | 24360 | 9346 | 1001 | 2052 |  |  |
| 2000 | 44508 | 4578 | 4730 | 40180 | 1375 | 40086 | 15742 | 0 | 107 |  |  |
| 2001 | 22849 | 7025 | 5781 | 38368 | 6700 | 45291 | 11645 | 442 | 4298 |  |  |
| 2002 | 37008 | 9554 | 4503 | 50024 | 12132 | 38113 | 10852 | 401 | 3799 |  |  |
| 2003 | 34515 | 9986 | 3079 | 53156 | 11032 | 42773 | 15452 | 74 | 6586 |  |  |
| 2004 | 42181 | 12407 | 3197 | 56050 | 10459 | 43183 | 12560 | 521 | 4388 |  |  |
| 2005 | 45014 | 6664 | 926 | 53556 | 12773 | 46665 | 12147 | 0 | 4957 |  |  |
| 2006 | 46045 | 6124 | 1033 | 51079 | 14598 | 29808 | 14265 | 0 | 4692 |  |  |
| 2007 | 35403 | 8753 | 1033 | 44206 | 12810 | 29808 | 12256 | 18 | 4692 |  |  |

Table 2.7a-Length frequencies for the January-May trawl fishery by length bin

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 3 | 16 | 19 | 73 | 220 | 103 | 29 | 19 | 13 | 4 | 5 | 4 | 0 | 1 | 2 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 45 | 94 | 204 | 315 | 329 | 77 | 122 | 147 | 144 | 37 | 5 | 4 | 3 | 1 | 1 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 36 | 75 | 235 | 635 | 1014 | 1560 | 1038 | 971 | 714 | 497 | 632 | 485 | 197 | 86 | 49 | 17 | 5 | 2 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 6 | 58 | 113 | 64 | 73 | 294 | 386 | 518 | 729 | 731 | 534 | 241 | 104 | 51 | 41 | 21 | 3 | 3 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 50 | 154 | 93 | 95 | 176 | 492 | 758 | 1626 | 2344 | 2071 | 1307 | 644 | 211 | 77 | 36 | 21 | 12 | 6 |
| 1984 | 0 | 1 | 2 | 1 | 0 | 15 | 194 | 401 | 367 | 220 | 105 | 223 | 709 | 779 | 1264 | 2262 | 3195 | 2930 | 2027 | 1039 | 434 | 144 | 24 | 13 | 2 |
| 1985 | 0 | 0 | 2 | 0 | 4 | 0 | 2 | 39 | 116 | 257 | 720 | 1752 | 2234 | 1079 | 1388 | 2440 | 4999 | 5563 | 4288 | 2630 | 1385 | 594 | 221 | 67 | 23 |
| 1986 | 0 | 4 | 16 | 8 | 34 | 60 | 118 | 249 | 635 | 761 | 683 | 783 | 2228 | 3560 | 3287 | 2095 | 2631 | 3469 | 3357 | 2442 | 1346 | 454 | 168 | 58 | 17 |
| 1987 | 0 | 0 | 3 | 13 | 15 | 58 | 192 | 440 | 477 | 592 | 1161 | 2054 | 3898 | 2890 | 3326 | 5470 | 5461 | 4306 | 3650 | 3106 | 1953 | 1076 | 440 | 198 | 63 |
| 1988 | 1 | 0 | 1 | 1 | 6 | 29 | 92 | 580 | 1448 | 1956 | 2185 | 4311 | 11135 | 10599 | 10194 | 9103 | 10096 | 12012 | 10395 | 5807 | 3010 | 1686 | 814 | 346 | 92 |
| 1989 | 0 | 0 | 3 | 3 | 1 | 0 | 28 | 217 | 494 | 795 | 720 | 954 | 3110 | 4341 | 4654 | 5664 | 7033 | 8561 | 8246 | 6265 | 3826 | 1867 | 919 | 388 | 144 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 11 | 93 | 214 | 284 | 269 | 232 | 203 | 416 | 853 | 1482 | 2458 | 3274 | 3396 | 3059 | 2109 | 1365 | 738 | 424 | 161 | 52 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 14 | 128 | 335 | 393 | 367 | 389 | 604 | 2129 | 2128 | 1770 | 2416 | 3307 | 3528 | 3007 | 2104 | 1371 | 761 | 403 | 192 | 66 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 9 | 52 | 156 | 323 | 382 | 568 | 1077 | 2241 | 1742 | 1545 | 1531 | 1753 | 1532 | 1384 | 1024 | 682 | 409 | 230 | 102 | 44 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 10 | 93 | 428 | 617 | 658 | 1718 | 2987 | 4493 | 3792 | 3576 | 2542 | 1640 | 1288 | 1041 | 759 | 505 | 316 | 182 | 78 | 35 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 13 | 136 | 457 | 789 | 664 | 398 | 626 | 2039 | 2917 | 2912 | 2322 | 2297 | 1901 | 1170 | 699 | 424 | 240 | 140 | 69 | 34 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 20 | 71 | 127 | 163 | 303 | 1181 | 2663 | 4198 | 2714 | 3176 | 3669 | 3894 | 3045 | 1763 | 1022 | 624 | 345 | 175 | 92 | 32 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 7 | 76 | 224 | 277 | 232 | 232 | 507 | 1862 | 3157 | 2940 | 2095 | 2323 | 2488 | 1957 | 1292 | 733 | 441 | 227 | 116 | 59 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 8 | 76 | 296 | 564 | 574 | 439 | 503 | 1842 | 2099 | 2798 | 3872 | 3840 | 2762 | 1612 | 1010 | 641 | 342 | 169 | 75 | 31 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 11 | 106 | 204 | 191 | 144 | 125 | 191 | 697 | 831 | 920 | 1389 | 1982 | 2110 | 1283 | 646 | 312 | 183 | 102 | 45 | 20 |
| 1999 | 1 | 0 | 0 | 0 | 0 | 1 | 36 | 143 | 134 | 119 | 347 | 847 | 1669 | 1011 | 1038 | 1292 | 1673 | 1697 | 1218 | 781 | 384 | 190 | 77 | 36 | 17 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 1 | 21 | 61 | 54 | 83 | 180 | 336 | 950 | 1383 | 1491 | 1376 | 1361 | 1405 | 1104 | 761 | 466 | 259 | 135 | 63 | 28 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 10 | 29 | 54 | 37 | 59 | 306 | 487 | 646 | 918 | 972 | 783 | 497 | 358 | 215 | 137 | 61 | 30 | 13 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 4 | 34 | 148 | 255 | 253 | 221 | 261 | 749 | 860 | 906 | 1494 | 1912 | 1672 | 959 | 440 | 211 | 97 | 45 | 19 | 10 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 31 | 95 | 128 | 139 | 246 | 670 | 703 | 760 | 989 | 1290 | 1466 | 1049 | 622 | 308 | 130 | 58 | 27 | 10 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 32 | 122 | 196 | 194 | 186 | 799 | 1329 | 1487 | 1739 | 1760 | 1393 | 946 | 590 | 315 | 190 | 111 | 62 | 26 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 52 | 120 | 162 | 147 | 140 | 461 | 756 | 1118 | 1584 | 1958 | 1796 | 1139 | 728 | 404 | 232 | 108 | 44 | 16 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 28 | 91 | 147 | 161 | 176 | 582 | 882 | 992 | 1186 | 1473 | 1570 | 1299 | 952 | 578 | 290 | 133 | 39 | 25 |
| 2007 | 0 | 1 | 2 | 8 | 13 | 28 | 56 | 151 | 231 | 283 | 384 | 440 | 1196 | 1843 | 2275 | 3112 | 3013 | 2579 | 2083 | 1633 | 1073 | 655 | 358 | 132 | 55 |


Table 2.7c-Length frequencies for the September-December trawl fishery by length bin.

Table 2.8a-Length frequencies for the January-May longline fishery by length bin.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 23 | 124 | 623 | 812 | 435 | 269 | 216 | 160 | 110 | 58 | 36 | 7 | 7 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 83 | 377 | 683 | 434 | 337 | 1135 | 2126 | 2432 | 1356 | 465 | 233 | 128 | 56 | 27 | 3 | 6 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 15 | 66 | 212 | 591 | 604 | 320 | 182 | 199 | 244 | 111 | 36 | 11 | 4 | 0 | 0 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 5 | 18 | 7 | 7 | 10 | 0 | 18 | 48 | 285 | 496 | 448 | 335 | 197 | 153 | 89 | 70 | 36 | 9 | 4 | 0 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 | 13 | 18 | 131 | 184 | 266 | 334 | 314 | 211 | 101 | 61 | 44 | 31 | 10 | 1 | 1 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 16 | 48 | 170 | 1116 | 1525 | 2035 | 2732 | 3421 | 3065 | 1838 | 792 | 334 | 163 | 88 | 36 | 7 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 6 | 19 | 40 | 41 | 46 | 416 | 800 | 1323 | 2414 | 3163 | 3015 | 2012 | 1015 | 437 | 155 | 70 | 24 | 6 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 12 | 34 | 186 | 550 | 1367 | 958 | 1828 | 3877 | 7018 | 8009 | 5977 | 3362 | 1591 | 537 | 175 | 44 | 7 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 30 | 81 | 121 | 385 | 1765 | 3055 | 3578 | 3014 | 3739 | 5900 | 5622 | 3348 | 1554 | 654 | 237 | 63 | 13 |
| 1987 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 5 | 18 | 88 | 425 | 1362 | 4950 | 5219 | 8337 | 14661 | 16709 | 12862 | 11421 | 9132 | 4689 | 1828 | 519 | 180 | 31 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 20 | 100 | 221 | 377 | 480 | 420 | 342 | 230 | 174 | 107 | 67 | 31 | 11 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 12 | 32 | 109 | 249 | 502 | 912 | 1150 | 978 | 700 | 406 | 248 | 137 | 84 | 33 | 14 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 7 | 24 | 96 | 256 | 869 | 1282 | 1329 | 1601 | 1886 | 1639 | 1158 | 837 | 528 | 307 | 165 | 75 | 25 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 9 | 21 | 49 | 167 | 369 | 1080 | 2056 | 2763 | 2413 | 1688 | 1280 | 946 | 692 | 425 | 229 | 98 | 49 | 12 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 9 | 20 | 54 | 144 | 733 | 1845 | 3065 | 3958 | 3309 | 1817 | 850 | 485 | 294 | 186 | 91 | 46 | 15 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 8 | 27 | 159 | 456 | 1231 | 2085 | 3529 | 4520 | 4244 | 2616 | 1121 | 443 | 214 | 111 | 54 | 28 | 13 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 6 | 18 | 62 | 181 | 996 | 2224 | 2991 | 3323 | 3131 | 2280 | 1335 | 655 | 297 | 146 | 82 | 36 | 16 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 10 | 24 | 62 | 190 | 970 | 2005 | 3472 | 4601 | 4086 | 2399 | 1243 | 658 | 376 | 172 | 71 | 25 | 10 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 16 | 34 | 88 | 253 | 843 | 1395 | 2048 | 2946 | 3141 | 2307 | 1196 | 518 | 233 | 125 | 52 | 18 | 8 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 12 | 45 | 261 | 775 | 1824 | 1807 | 2111 | 2533 | 2483 | 2089 | 1308 | 679 | 291 | 139 | 59 | 29 | 18 |
| 2000 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 12 | 45 | 154 | 364 | 1510 | 2476 | 2555 | 2115 | 1736 | 1219 | 708 | 361 | 176 | 72 | 36 | 14 | 4 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 6 | 23 | 67 | 98 | 203 | 909 | 1761 | 2404 | 2672 | 2095 | 1110 | 545 | 274 | 136 | 73 | 33 | 16 | 7 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 2 | 5 | 25 | 57 | 89 | 255 | 641 | 1465 | 1704 | 2443 | 3386 | 3031 | 1836 | 722 | 300 | 133 | 77 | 55 | 12 | 6 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 46 | 107 | 290 | 704 | 2109 | 3046 | 3153 | 2909 | 2552 | 1841 | 937 | 414 | 150 | 61 | 24 | 10 | 3 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 15 | 23 | 45 | 84 | 233 | 1128 | 2541 | 3874 | 4240 | 2951 | 1562 | 801 | 422 | 183 | 76 | 34 | 13 | 4 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 21 | 49 | 128 | 274 | 931 | 1516 | 2204 | 3184 | 3467 | 2395 | 947 | 380 | 182 | 73 | 28 | 9 | 2 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 10 | 31 | 70 | 146 | 750 | 1880 | 2391 | 2453 | 2317 | 1988 | 1335 | 650 | 248 | 101 | 38 | 11 | 4 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 6 | 16 | 32 | 138 | 372 | 1824 | 3744 | 6435 | 8854 | 7325 | 5413 | 3711 | 2466 | 1323 | 519 | 195 | 79 | 34 |

Table 2.8b—Length frequencies for the June-August longline fishery by length bin.

Table 2.8c-Length frequencies for the September-December longline fishery by length bin.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 54 | 344 | 719 | 770 | 275 | 94 | 49 | 32 | 16 | 7 | 2 | 0 | 0 |
| 1979 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 11 | 51 | 252 | 263 | 195 | 401 | 705 | 605 | 220 | 44 | 11 | 9 | 2 | 0 | 0 |
| 1980 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 18 | 235 | 558 | 679 | 652 | 350 | 194 | 138 | 76 | 25 | 5 | 0 | 1 | 0 |
| 1981 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 86 | 230 | 318 | 300 | 220 | 89 | 29 | 15 | 2 | 0 | 1 | 0 |
| 1982 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 14 | 33 | 92 | 235 | 460 | 773 | 1149 | 1066 | 614 | 235 | 77 | 27 | 6 | 2 | 2 |
| 1983 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 4 | 28 | 129 | 459 | 1162 | 1260 | 1544 | 1776 | 1561 | 991 | 476 | 148 | 37 | 9 | 6 |
| 1984 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 14 | 55 | 293 | 764 | 1721 | 2467 | 6595 | 12255 | 15779 | 15982 | 12816 | 8397 | 4192 | 1528 | 407 | 91 | 24 |
| 1985 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 4 | 23 | 116 | 605 | 5449 | 16095 | 14240 | 10594 | 17780 | 24998 | 19637 | 11586 | 6071 | 2786 | 920 | 215 | 51 |
| 1986 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 18 | 158 | 616 | 2233 | 5154 | 14368 | 23612 | 20725 | 10897 | 10483 | 9006 | 4991 | 2308 | 881 | 326 | 85 |
| 1987 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 3 | 9 | 30 | 147 | 593 | 4503 | 18418 | 29582 | 24338 | 25914 | 28336 | 20972 | 10694 | 6630 | 3800 | 1532 | 414 | 134 |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 19 | 87 | 270 | 489 | 604 | 569 | 456 | 306 | 182 | 108 | 67 | 31 | 14 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 4 | 7 | 11 | 24 | 94 | 231 | 406 | 623 | 759 | 807 | 766 | 615 | 421 | 239 | 138 | 62 | 25 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 10 | 49 | 147 | 160 | 152 | 173 | 152 | 120 | 98 | 78 | 55 | 35 | 18 | 6 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 6 | 16 | 27 | 84 | 295 | 603 | 870 | 1016 | 766 | 446 | 238 | 140 | 87 | 53 | 31 | 15 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 5 | 21 | 51 | 280 | 678 | 784 | 907 | 1101 | 1008 | 739 | 443 | 246 | 136 | 79 | 38 | 12 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 14 | 32 | 123 | 423 | 909 | 1132 | 871 | 640 | 560 | 453 | 321 | 178 | 82 | 33 | 10 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 6 | 11 | 20 | 42 | 114 | 329 | 639 | 1037 | 1424 | 1841 | 1709 | 1182 | 614 | 415 | 251 | 135 | 58 | 23 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 3 | 13 | 69 | 146 | 339 | 539 | 812 | 988 | 1140 | 1031 | 767 | 474 | 242 | 127 | 76 | 32 | 13 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 8 | 15 | 47 | 87 | 463 | 990 | 859 | 780 | 837 | 761 | 523 | 362 | 208 | 109 | 48 | 24 | 12 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 5 | 44 | 145 | 485 | 1156 | 1903 | 2459 | 1908 | 1086 | 665 | 411 | 240 | 131 | 62 | 23 | 9 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 6 | 26 | 126 | 268 | 767 | 1331 | 2128 | 2431 | 2309 | 1726 | 794 | 352 | 164 | 92 | 54 | 22 | 9 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 9 | 29 | 67 | 144 | 300 | 982 | 1601 | 1807 | 1785 | 1676 | 1314 | 763 | 375 | 163 | 71 | 29 | 13 | 6 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 10 | 31 | 136 | 734 | 1577 | 2155 | 2151 | 1869 | 1381 | 889 | 489 | 217 | 85 | 33 | 13 | 4 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 6 | 10 | 22 | 49 | 119 | 491 | 1008 | 1604 | 2136 | 2099 | 1574 | 883 | 500 | 268 | 118 | 45 | 14 | 4 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 8 | 19 | 49 | 125 | 556 | 976 | 1408 | 1504 | 1500 | 1457 | 1227 | 831 | 447 | 244 | 95 | 31 | 7 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 28 | 84 | 388 | 783 | 1259 | 1326 | 1171 | 1095 | 906 | 874 | 639 | 408 | 213 | 91 | 31 |


| Table 2.9a—Length frequencies for the January-May pot fishery by length bin. |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | ---: | ---: |
| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 12 | 33 |
| 1993 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 14 | 33 |
| 1994 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 38 | 136 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 19 | 96 | 224 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 8 | 20 | 145 | 430 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 5 | 10 | 70 | 204 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 60 | 126 |
| 1999 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 15 | 126 | 198 |
| 2000 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 27 | 226 | 526 |
| 2001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 4 | 6 | 69 | 218 |
| 2002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 5 | 45 | 179 |
| 2003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 3 | 4 | 23 | 115 | 292 |
| 2004 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 5 | 6 | 4 | 5 | 9 | 113 | 368 |
| 2005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 4 | 8 | 51 | 194 |
| 2006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 66 | 274 |
| 2007 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 16 | 180 | 489 |

Table 2.9b—Length frequencies for the June-August pot fishery by length bin.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 9 | 17 | 28 | 36 | 27 | 28 | 17 | 10 | 5 | 0 | 0 |
| 1991 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 18 | 32 | 44 | 45 | 34 | 23 | 11 | 7 | 2 | 1 | 0 |
| 1992 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 20 | 113 | 201 | 194 | 215 | 211 | 147 | 90 | 62 | 38 | 21 | 11 | 8 | 3 |
| 1995 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 31 | 55 | 73 | 125 | 140 | 113 | 89 | 59 | 33 | 27 | 20 | 8 | 5 |
| 1996 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 28 | 98 | 202 | 283 | 252 | 179 | 126 | 107 | 75 | 53 | 30 | 16 | 11 |
| 1997 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 15 | 51 | 110 | 171 | 235 | 152 | 74 | 37 | 23 | 13 | 9 | 7 | 3 |
| 1998 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 23 | 55 | 99 | 136 | 141 | 84 | 42 | 23 | 16 | 5 | 2 | 1 |


Table 2.10a-Length frequencies for the 1979-1981 EBS shelf bottom trawl survey by length bin.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1979 | 0 | 5 | 44 | 186 | 374 | 457 | 694 | 1764 | 2393 | 1884 | 1171 | 618 | 202 | 70 | 44 | 51 | 29 | 8 | 0 | 3 | 1 | 1 | 0 | 0 | 0 |
| 1980 | 0 | 6 | 85 | 241 | 82 | 42 | 224 | 687 | 929 | 1320 | 1542 | 2062 | 1364 | 893 | 333 | 100 | 33 | 31 | 19 | 6 | 2 | 0 | 0 | 0 | 0 |
| 1981 | 0 | 20 | 156 | 330 | 278 | 32 | 100 | 330 | 653 | 724 | 511 | 1063 | 1396 | 1746 | 1215 | 812 | 398 | 156 | 39 | 27 | 13 | 1 | 0 | 0 | 0 |

Table 2.10b-Length frequencies for the 1982-2007 EBS shelf bottom trawl survey by length bin.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19 | 17 | 97 | 234 | 148 | 37 | 28 | 132 | 403 | 766 | 750 | 416 | 0 | 15 | 1326 | 1288 | 11 | 874 | 474 | 210 | 90 | 29 | 9 | 4 | 0 | 0 |
| 1983 | 393 | 1396 | 1289 | 622 | 147 | 32 | 135 | 370 | 551 | 380 | 209 | 393 | 1367 | 1289 | 1341 | 1128 | 921 | 650 | 325 | 151 | 31 | 19 | 4 |  | 0 |
| 1984 | 70 | 129 | 82 | 142 | 282 | 920 | 1653 | 1712 | 1041 | 485 | 249 | 261 | 536 | 579 | 864 | 96 | 880 | 590 | 381 | 173 | 94 | 38 | 9 |  | 0 |
| 198 | 162 | 540 | 964 | 1537 | 1761 | 664 | 8 | 95 | 880 | 942 | 1154 | 1528 | 1879 | 67 | 80 | 543 | 687 | 674 | 496 | 253 | 111 | 38 | 17 | 5 | 0 |
| 1986 | 154 | 465 | 501 | 154 | 114 | 692 | 1775 | 1908 | 1585 | 1083 | 553 | 425 | 1069 | 1338 | 1203 | 628 | 416 | 453 | 370 | 264 | 119 | 74 | 21 | 13 | 0 |
| 7 | 18 | 69 | 250 | 398 | 267 | 85 | 440 | 99 | 779 | 606 | 617 | 57 | 1478 | 827 | 598 | 65 | 632 | 413 | 21 | 16 | 71 | 49 | 16 | 7 | 0 |
| 1988 | 8 | 49 | 76 | 88 | 109 | 233 | 279 | 84 | 641 | 625 | 491 | 660 | 1418 | 1306 | 1114 | 849 | 570 | 420 | 293 | 244 | 74 | 32 | 25 | 7 | 4 |
| 1989 | 24 | 154 | 298 | 205 | 70 | 34 | 82 | 87 | 39 | 348 | 339 | 366 | 871 | 1193 | 1294 | 1143 | 94 | 85 | 66 | 338 | 247 | 145 | 90 | 62 | 0 |
| 1990 | 201 | 88 | 699 | 355 | 133 | 122 | 249 | 292 | 321 | 276 | 175 | 123 | 194 | 223 | 346 | 419 | 283 | 26 | 182 | 128 | 82 | 33 | 26 | 11 | 3 |
| 1991 | 131 | 389 | 432 | 369 | 229 | 272 | 620 | 898 | 932 | 631 | 346 | 193 | 301 | 312 | 250 | 215 | 207 | 178 | 110 | 112 | 49 | 20 | 22 | 7 | 2 |
| 1992 | 18 | 6 | 517 | 698 | 556 | 435 | 854 | 1075 | 856 | 542 | 451 | 622 | 15 | 546 | 242 | 222 | 176 | 103 | 97 | 86 | 51 | 37 | 28 | 15 | 3 |
| 1993 | 114 | 924 | 1088 | 981 | 677 | 213 | 247 | 614 | 847 | 666 | 489 | 615 | 1071 | 665 | 399 | 267 | 230 | 85 | 62 | 48 | 37 | 20 | 23 | 14 | 6 |
| 1994 | 19 | 145 | 291 | 363 | 326 | 445 | 956 | 1922 | 2081 | 1121 | 444 | 522 | 1216 | 961 | 1059 | 920 | 565 | 288 | 92 | 46 | 34 | 60 | 16 | 22 | 9 |
| 1995 | 30 | 73 | 135 | 208 | 77 | 173 | 460 | 691 | 579 | 705 | 1064 | 12 | 1360 | 616 | 434 | 484 | 326 | 253 | 132 | 84 | 40 | 27 | 19 | 9 | 3 |
| 1996 | 14 | 65 | 164 | 198 | 110 | 103 | 357 | 99 | 677 | 526 | 499 | 744 | 1477 | 1404 | 908 | 499 | 288 | 237 | 148 | 109 | 71 | 25 | 16 | 7 | 3 |
| 1997 | 91 | 473 | 601 | 728 | 507 | 140 | 215 | 481 | 628 | 51 | 407 | 399 | 919 | 809 | 842 | 583 | 436 | 215 | 105 | 60 | 40 | 26 | 10 | 4 | 1 |
| 19 | 30 | 26 | 334 | 74 | 46 | 311 | 1151 | 1837 | 1396 | 655 | 379 | 367 | 659 | 458 | 378 | 39 | 333 | 244 | 132 | 64 | 33 | 29 | 9 | 10 | 1 |
| 1999 | 71 | 334 | 286 | 113 | 141 | 415 | 760 | 874 | 667 | 718 | 1169 | 1648 | 1854 | 768 | 493 | 447 | 337 | 252 | 132 | 89 | 62 | 37 | 24 | 7 | 2 |
| 20 | 174 | 917 | 1308 | 505 | 54 | 141 | 487 | 784 | 604 | 563 | 748 | 57 | 1718 | 1417 | 93 | 536 | 266 | 18 | 99 | 79 | 57 | 33 | 19 | 3 | 0 |
| 2001 | 95 | 646 | 1828 | 2113 | 1010 | 408 | 903 | 1990 | 2543 | 1613 | 705 | 486 | 1192 | 1276 | 1077 | 818 | 513 | 257 | 123 | 71 | 34 | 22 | 14 | 4 | 5 |
| 2002 | 31 | 190 | 374 | 352 | 105 | 209 | 664 | 1459 | 1449 | 1005 | 792 | 1216 | 1578 | 878 | 609 | 545 | 367 | 208 | 103 | 49 | 19 | 16 | 15 | 3 | 2 |
| 2003 | 19 | 283 | 633 | 774 | 682 | 489 | 182 | 252 | 682 | 837 | 974 | 1192 | 1974 | 1218 | 770 | 516 | 340 | 261 | 142 | 86 | 35 | 14 | 2 | 1 | 0 |
| 2004 | 24 | 275 | 483 | 562 | 318 | 218 | 484 | 729 | 930 | 979 | 711 | 578 | 806 | 925 | 844 | 714 | 474 | 283 | 211 | 111 | 82 | 34 | 15 | 5 | 4 |
| 2005 | 5 | 153 | 590 | 892 | 1018 | 1053 | 484 | 415 | 575 | 726 | 647 | 625 | 855 | 702 | 520 | 527 | 495 | 360 | 292 | 182 | 104 | 46 | 21 | 7 | 0 |
| 2006 | 478 | 1286 | 1075 | 884 | 317 | 165 | 266 | 604 | 753 | 866 | 706 | 533 | 728 | 855 | 643 | 494 | 395 | 320 | 259 | 238 | 144 | 76 | 35 | 14 | 1 |
| 2007 | 488 | 3110 | 2018 | 966 | 369 | 118 | 255 | 325 | 301 | 205 | 165 | 176 | 297 | 289 | 203 | 210 | 161 | 115 | 65 | 50 | 53 | 33 | 13 | 13 | 4 |

Table 2.11—Age compositions observed by the EBS shelf bottom trawl survey, 1994-2006.

| Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1994 | 0.0536 | 0.4015 | 0.1844 | 0.1259 | 0.1241 | 0.0837 | 0.0195 | 0.0050 | 0.0020 | 0.0001 | 0.0002 | 0.0000 |
| 1995 | 0.0276 | 0.2705 | 0.4401 | 0.1074 | 0.0803 | 0.0536 | 0.0106 | 0.0041 | 0.0042 | 0.0003 | 0.0008 | 0.0006 |
| 1996 | 0.0032 | 0.2306 | 0.2469 | 0.3568 | 0.0941 | 0.0541 | 0.0144 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1997 | 0.2355 | 0.1841 | 0.1737 | 0.1610 | 0.1225 | 0.0898 | 0.0227 | 0.0081 | 0.0009 | 0.0010 | 0.0006 | 0.0000 |
| 1998 | 0.0664 | 0.4546 | 0.2020 | 0.1137 | 0.0589 | 0.0596 | 0.0284 | 0.0140 | 0.0022 | 0.0000 | 0.0002 | 0.0000 |
| 1999 | 0.0715 | 0.1992 | 0.3090 | 0.2409 | 0.0806 | 0.0575 | 0.0266 | 0.0103 | 0.0036 | 0.0000 | 0.0007 | 0.0000 |
| 2000 | 0.2240 | 0.1162 | 0.1675 | 0.2476 | 0.1563 | 0.0595 | 0.0108 | 0.0120 | 0.0028 | 0.0026 | 0.0007 | 0.0000 |
| 2001 | 0.2598 | 0.2469 | 0.2052 | 0.0941 | 0.0915 | 0.0703 | 0.0236 | 0.0056 | 0.0014 | 0.0009 | 0.0006 | 0.0001 |
| 2002 | 0.0799 | 0.1868 | 0.3104 | 0.2443 | 0.0733 | 0.0575 | 0.0390 | 0.0065 | 0.0018 | 0.0005 | 0.0000 | 0.0001 |
| 2003 | 0.1487 | 0.1633 | 0.2546 | 0.2212 | 0.1220 | 0.0412 | 0.0291 | 0.0151 | 0.0033 | 0.0003 | 0.0003 | 0.0007 |
| 2004 | 0.1421 | 0.1622 | 0.2805 | 0.1301 | 0.1333 | 0.0908 | 0.0346 | 0.0177 | 0.0062 | 0.0011 | 0.0014 | 0.0000 |
| 2005 | 0.1836 | 0.2560 | 0.1868 | 0.1383 | 0.0621 | 0.0843 | 0.0485 | 0.0243 | 0.0106 | 0.0016 | 0.0040 | 0.0000 |
| 2006 | 0.3198 | 0.1443 | 0.1703 | 0.1180 | 0.0946 | 0.0632 | 0.0473 | 0.0290 | 0.0097 | 0.0029 | 0.0009 | 0.0002 |

Table 2.12a—Abundance measured in units of biomass and numbers, with standard errors, as estimated by EBS shelf bottom trawl surveys, 1979-1981. For biomass, $95 \%$ confidence intervals (CI) are also shown. All biomass figures are expressed in metric tons. Population numbers are expressed in terms of individual fish. The actual standard errors for abundance measured in numbers during these years are unknown; the standard errors shown here are estimates obtained by assuming that the coefficient of variation was the same as for the biomass estimate.

|  | Abundance (biomass) |  |  |  | Abundance (numbers) |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Estimate | Standard Error | Lower 95\% CI | Upper 95\% CI | Estimate | Standard Error |
| 1979 | 754,314 | 97,844 | 562,539 | 946,089 | $1,530,429,650$ | $198,515,948$ |
| 1980 | 905,344 | 87,898 | 733,063 | $1,077,624$ | $1,084,147,540$ | $105,257,671$ |
| 1981 | $1,034,629$ | 123,849 | 791,885 | $1,277,373$ | $794,619,624$ | $95,118,971$ |

Table 2.12b- Abundance measured in units of biomass and numbers, with standard errors, as estimated by EBS shelf bottom trawl surveys, 1982-2007. For biomass, $95 \%$ confidence intervals (CI) are also shown. All biomass figures are expressed in metric tons. Population numbers are expressed in terms of individual fish.

|  | Abundance (biomass) |  |  |  | Abundance (numbers) |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Estimate | Standard Error | Lower 95\% CI | Upper 95\% CI | Estimate | Standard Error |
| 1982 | $1,012,856$ | 73,588 | 867,151 | $1,158,562$ | $583,715,842$ | $38,040,768$ |
| 1983 | $1,185,419$ | 120,868 | 941,146 | $1,429,692$ | $751,066,723$ | $80,440,661$ |
| 1984 | $1,048,595$ | 63,643 | 922,583 | $1,174,608$ | $680,914,697$ | $49,913,926$ |
| 1985 | $1,001,108$ | 55,845 | 890,536 | $1,111,681$ | $841,108,075$ | $112,271,991$ |
| 1986 | $1,117,774$ | 69,604 | 979,957 | $1,255,590$ | $838,123,105$ | $83,854,636$ |
| 1987 | $1,106,621$ | 68,682 | 970,630 | $1,242,612$ | $728,956,963$ | $48,520,099$ |
| 1988 | 959,000 | 76,265 | 807,996 | $1,110,004$ | $508,065,276$ | $35,526,047$ |
| 1989 | 836,177 | 62,981 | 711,475 | 960,878 | $292,210,905$ | $19,939,408$ |
| 1990 | 691,255 | 51,455 | 589,375 | 793,136 | $423,835,267$ | $36,466,423$ |
| 1991 | 517,209 | 38,158 | 441,657 | 592,761 | $488,861,768$ | $50,972,542$ |
| 1992 | 551,369 | 45,780 | 460,725 | 642,013 | $601,795,262$ | $70,551,400$ |
| 1993 | 690,535 | 54,380 | 582,862 | 798,208 | $851,863,422$ | $106,911,178$ |
| 1994 | $1,368,120$ | 250,044 | 868,032 | $1,868,209$ | $1,237,758,281$ | $153,120,867$ |
| 1995 | $1,003,096$ | 91,739 | 821,453 | $1,184,740$ | $757,657,482$ | $75,485,760$ |
| 1996 | 890,793 | 87,552 | 717,439 | $1,064,146$ | $609,304,214$ | $88,330,629$ |
| 1997 | 604,881 | 69,250 | 466,382 | 743,380 | $487,429,700$ | $72,155,388$ |
| 1998 | 558,419 | 45,182 | 468,960 | 647,879 | $537,278,347$ | $48,263,858$ |
| 1999 | 583,891 | 50,621 | 483,662 | 684,120 | $500,915,139$ | $46,536,008$ |
| 2000 | 528,466 | 43,037 | 443,253 | 613,679 | $481,358,109$ | $44,098,753$ |
| 2001 | 833,626 | 76,247 | 681,133 | 986,119 | $985,568,802$ | $94,981,577$ |
| 2002 | 618,680 | 69,082 | 480,516 | 756,845 | $566,471,072$ | $57,675,818$ |
| 2003 | 595,826 | 62,099 | 471,628 | 720,024 | $499,925,561$ | $62,237,449$ |
| 2004 | 596,464 | 35,191 | 526,787 | 666,142 | $424,075,921$ | $36,061,059$ |
| 2005 | 603,788 | 43,150 | 517,488 | 690,089 | $452,075,840$ | $63,294,550$ |
| 2006 | 517,698 | 28,341 | 461,583 | 573,813 | $393,993,981$ | $23,784,449$ |
| 2007 | 423,703 | 34,811 | 354,080 | 493,326 | $733,374,144$ | $195,954,076$ |

Table 2.13a—Actual length sample sizes from at-sea observers by gear and season, 1990-2006.

| Year | Twl(1) | Twl(2) | Twl(3) | Lgl(1) | Lgl(2) | Lgl(3) | Pot(1) | Pot(2) | Pot(3) |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 70,213 | 326 | 4,138 | 22,030 | 69,835 | 63,550 |  | 1,526 | 5,107 |
| 1991 | 124,193 | 1,701 |  | 54,119 | 58,576 | 85,581 |  | 6,388 | 7,351 |
| 1992 | 70,164 | 170 | 476 | 141,207 | 75,623 | 20,926 | 13,714 | 15,318 | 3,282 |
| 1993 | 73,155 |  | 354 | 121,206 |  |  | 7,665 |  |  |
| 1994 | 101,934 |  | 547 | 152,377 |  | 41,671 | 21,544 |  | 4,947 |
| 1995 | 65,390 |  | 667 | 138,341 |  | 66,596 | 31,515 | 5,090 | 6,036 |
| 1996 | 100,144 |  | 3,554 | 148,856 |  | 73,687 | 48,338 | 9,057 | 9,429 |
| 1997 | 96,103 | 499 |  | 173,973 |  | 144,358 | 28,305 | 5,180 | 8,992 |
| 1998 | 80,880 | 1,692 | 1,303 | 126,890 | 64 | 173,242 | 21,469 | 4,498 | 3,000 |
| 1999 | 36,374 | 144 | 761 | 77,556 | 7,244 | 47,216 | 13,456 |  | 2,613 |
| 2000 | 34,153 | 139 | 173 | 52,859 | 6,285 | 86,683 | 12,003 |  | 226 |
| 2001 | 18,125 | 1,922 | 1,214 | 64,859 | 16,300 | 87,251 | 9,432 |  | 5,113 |
| 2002 | 24,192 | 3,114 | 2,437 | 68,503 | 30,893 | 84,876 | 6,680 |  | 5,132 |
| 2003 | 25,989 | 6,165 | 1,975 | 95,501 | 36,712 | 102,683 | 8,931 |  | 5,872 |
| 2004 | 21,784 | 3,359 | 1,600 | 79,913 | 32,196 | 89,005 | 6,386 |  | 4,257 |
| 2005 | 23,500 | 1,249 | 184 | 67,901 | 37,570 | 85,617 | 5,173 |  | 4,516 |
| 2006 | 22,761 | 804 |  | 54,983 | 30,529 | 51,831 | 6,423 |  | 5,219 |

Table 2.13b—Harmonic mean sample sizes from a bootstrap simulation by gear and season, 1990-2006.

| Year | $\operatorname{Twl}(1)$ | $\operatorname{Twl}(2)$ | $\operatorname{Twl}(3)$ | $\operatorname{Lgl}(1)$ | $\operatorname{Lgl}(2)$ | $\operatorname{Lgl}(3)$ | $\operatorname{Pot}(1)$ | $\operatorname{Pot}(2)$ | $\operatorname{Pot}(3)$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 10,644 | 96 | 877 | 3,637 | 10,170 | 9,922 |  | 272 | 346 |
| 1991 | 12,787 | 163 |  | 8,022 | 10,080 | 15,353 |  | 1,517 | 1,375 |
| 1992 | 7,868 | 3 | 17 | 21,882 | 12,284 | 3,759 | 1,848 | 3,114 | 574 |
| 1993 | 8,742 |  | 27 | 15,863 |  |  | 1,374 |  |  |
| 1994 | 15,585 |  | 21 | 20,708 |  | 7,739 | 3,577 |  | 429 |
| 1995 | 8,777 |  | 173 | 23,094 |  | 9,465 | 5,046 | 979 | 708 |
| 1996 | 16,971 |  | 497 | 20,728 |  | 10,605 | 7,757 | 1,361 | 1,105 |
| 1997 | 17,763 | 86 |  | 24,208 |  | 21,020 | 4,621 | 922 | 1,494 |
| 1998 | 14,414 | 302 | 346 | 22,378 | 10 | 29,137 | 3,259 | 798 | 546 |
| 1999 | 11,416 | 38 | 245 | 26,617 | 2,629 | 16,268 | 4,351 | 826 |  |
| 2000 | 11,555 | 28 | 24 | 17,284 | 2,171 | 28,133 | 3,750 | 87 |  |
| 2001 | 5,947 | 597 | 377 | 21,330 | 5,591 | 29,835 | 3,034 | 87 |  |
| 2002 | 8,136 | 1,135 | 855 | 23,052 | 10,850 | 29,647 | 2,309 | 1,464 |  |
| 2003 | 8,267 | 2,057 | 754 | 31,835 | 12,208 | 32,752 | 3,018 | 1,553 |  |
| 2004 | 7,142 | 1,166 | 553 | 27,633 | 11,600 | 30,546 | 2,206 | 1,829 |  |
| 2005 | 7,945 | 379 | 60 | 25,101 | 13,207 | 29,536 | 1,813 | 1,414 |  |
| 2006 | 8,045 | 248 |  | 20,065 | 11,217 | 18,418 | 2,144 | 1,568 |  |
|  |  |  |  |  |  |  |  | 1,848 |  |

Table 2.13c—Harmonic means from bootstrap rescaled proportionally to exhibit an average of 300 .

| Year | Twl(1) | $\operatorname{Twl}(2)$ | $\operatorname{Twl}(3)$ | Lgl(1) | Lgl(2) | Lgl(3) | $\operatorname{Pot}(1)$ | $\operatorname{Pot}(2)$ | $\operatorname{Pot}(3)$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 387 | 3 | 32 | 132 | 370 | 361 |  | 10 | 13 |
| 1991 | 465 | 6 |  | 292 | 367 | 559 |  | 55 | 50 |
| 1992 | 286 | 0 | 1 | 796 | 447 | 137 | 67 | 113 | 21 |
| 1993 | 318 |  | 1 | 577 |  |  | 50 |  |  |
| 1994 | 567 |  | 1 | 754 |  | 282 | 130 |  | 16 |
| 1995 | 319 |  | 6 | 840 |  | 344 | 184 | 36 | 26 |
| 1996 | 618 |  | 18 | 754 |  | 386 | 282 | 50 | 40 |
| 1997 | 646 | 3 |  | 881 |  | 765 | 168 | 34 | 54 |
| 1998 | 525 | 11 | 13 | 814 | 0 | 1,060 | 119 | 29 | 20 |
| 1999 | 415 | 1 | 9 | 969 | 96 | 592 | 158 |  | 30 |
| 2000 | 420 | 1 | 1 | 629 | 79 | 1,024 | 136 |  | 3 |
| 2001 | 216 | 22 | 14 | 776 | 203 | 1,086 | 110 |  | 53 |
| 2002 | 296 | 41 | 31 | 839 | 395 | 1,079 | 84 |  | 57 |
| 2003 | 301 | 75 | 27 | 1,159 | 444 | 1,192 | 110 |  | 67 |
| 2004 | 260 | 42 | 20 | 1,006 | 422 | 1,112 | 80 |  | 51 |
| 2005 | 289 | 14 | 2 | 913 | 481 | 1,075 | 66 |  | 57 |
| 2006 | 293 | 9 |  | 730 | 408 | 670 | 78 | 67 |  |

Table 2.13d—Square roots of actual sample sizes.

| Year | Twl(1) | Twl(2) | Twl(3) | Lgl(1) | Lgl(2) | Lgl(3) | Pot(1) | $\operatorname{Pot}(2)$ | $\operatorname{Pot}(3)$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 265 | 18 | 64 | 148 | 264 | 252 |  | 39 | 71 |
| 1991 | 352 | 41 |  | 233 | 242 | 293 |  | 80 | 86 |
| 1992 | 265 | 13 | 22 | 376 | 275 | 145 | 117 | 124 | 57 |
| 1993 | 270 |  | 19 | 348 |  |  | 88 |  |  |
| 1994 | 319 |  | 23 | 390 |  | 204 | 147 |  | 70 |
| 1995 | 256 |  | 26 | 372 |  | 258 | 178 | 71 | 78 |
| 1996 | 316 |  | 60 | 386 |  | 271 | 220 | 95 | 97 |
| 1997 | 310 | 22 |  | 417 |  | 380 | 168 | 72 | 95 |
| 1998 | 284 | 41 | 36 | 356 | 8 | 416 | 147 | 67 | 55 |
| 1999 | 191 | 12 | 28 | 278 | 85 | 217 | 116 |  | 51 |
| 2000 | 185 | 12 | 13 | 230 | 79 | 294 | 110 | 15 |  |
| 2001 | 135 | 44 | 35 | 255 | 128 | 295 | 97 | 72 |  |
| 2002 | 156 | 56 | 49 | 262 | 176 | 291 | 82 |  | 72 |
| 2003 | 161 | 79 | 44 | 309 | 192 | 320 | 95 | 77 |  |
| 2004 | 148 | 58 | 40 | 283 | 179 | 298 | 80 |  | 65 |
| 2005 | 153 | 35 | 14 | 261 | 194 | 293 | 72 |  | 67 |
| 2006 | 151 | 28 |  | 234 | 175 | 228 | 80 | 72 |  |

Table 2.13e—Ratio of rescaled bootstrap harmonic means to square roots of actual sample sizes.

| Year | Twl(1) | Twl(2) | $\operatorname{Twl}(3)$ | $\operatorname{Lgl}(1)$ | $\operatorname{Lgl}(2)$ | $\operatorname{Lgl}(3)$ | $\operatorname{Pot}(1)$ | $\operatorname{Pot}(2)$ | $\operatorname{Pot}(3)$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 1.46 | 0.19 | 0.50 | 0.89 | 1.40 | 1.43 |  | 0.25 | 0.18 |
| 1991 | 1.32 | 0.14 |  | 1.25 | 1.52 | 1.91 |  | 0.69 | 0.58 |
| 1992 | 1.08 | 0.01 | 0.03 | 2.12 | 1.63 | 0.95 | 0.57 | 0.92 | 0.36 |
| 1993 | 1.18 |  | 0.05 | 1.66 |  |  | 0.57 |  |  |
| 1994 | 1.78 |  | 0.03 | 1.93 |  | 1.38 | 0.89 |  | 0.22 |
| 1995 | 1.25 |  | 0.24 | 2.26 |  | 1.33 | 1.03 | 0.50 | 0.33 |
| 1996 | 1.95 |  | 0.30 | 1.96 |  | 1.42 | 1.28 | 0.52 | 0.41 |
| 1997 | 2.09 | 0.14 |  | 2.11 |  | 2.01 | 1.00 | 0.47 | 0.57 |
| 1998 | 1.84 | 0.27 | 0.35 | 2.29 | 0.05 | 2.55 | 0.81 | 0.43 | 0.36 |
| 1999 | 2.18 | 0.11 | 0.32 | 3.48 | 1.12 | 2.72 | 1.37 |  | 0.59 |
| 2000 | 2.28 | 0.09 | 0.07 | 2.74 | 1.00 | 3.48 | 1.25 |  | 0.21 |
| 2001 | 1.61 | 0.50 | 0.39 | 3.05 | 1.59 | 3.68 | 1.14 |  | 0.75 |
| 2002 | 1.90 | 0.74 | 0.63 | 3.21 | 2.25 | 3.70 | 1.03 |  | 0.79 |
| 2003 | 1.87 | 0.95 | 0.62 | 3.75 | 2.32 | 3.72 | 1.16 |  | 0.87 |
| 2004 | 1.76 | 0.73 | 0.50 | 3.56 | 2.35 | 3.73 | 1.00 |  | 0.79 |
| 2005 | 1.89 | 0.39 | 0.16 | 3.51 | 2.48 | 3.67 | 0.92 | 0.85 |  |
| 2006 | 1.94 | 0.32 |  | 3.11 | 2.34 | 2.94 | 0.97 | 0.93 |  |

Table 2.13f—Ratio of rescaled bootstrap harmonic means to actual sample sizes.

| Year | $\operatorname{Twl}(1)$ | $\operatorname{Twl}(2)$ | $\operatorname{Twl}(3)$ | $\operatorname{Lgl}(1)$ | $\operatorname{Lgl}(2)$ | $\operatorname{Lgl}(3)$ | $\operatorname{Pot}(1)$ | $\operatorname{Pot}(2)$ | $\operatorname{Pot}(3)$ |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1990 | 0.15 | 0.29 | 0.21 | 0.17 | 0.15 | 0.16 |  | 0.18 | 0.07 |
| 1991 | 0.10 | 0.10 |  | 0.15 | 0.17 | 0.18 |  | 0.24 | 0.19 |
| 1992 | 0.11 | 0.02 | 0.03 | 0.15 | 0.16 | 0.18 | 0.13 | 0.20 | 0.17 |
| 1993 | 0.12 |  | 0.08 | 0.13 |  |  | 0.18 |  |  |
| 1994 | 0.15 |  | 0.04 | 0.14 |  | 0.19 | 0.17 |  | 0.09 |
| 1995 | 0.13 |  | 0.26 | 0.17 |  | 0.14 | 0.16 | 0.19 | 0.12 |
| 1996 | 0.17 |  | 0.14 | 0.14 |  | 0.14 | 0.16 | 0.15 | 0.12 |
| 1997 | 0.18 | 0.17 |  | 0.14 |  | 0.15 | 0.16 | 0.18 | 0.17 |
| 1998 | 0.18 | 0.18 | 0.27 | 0.18 | 0.16 | 0.17 | 0.15 | 0.18 | 0.18 |
| 1999 | 0.31 | 0.26 | 0.32 | 0.34 | 0.36 | 0.34 | 0.32 |  | 0.32 |
| 2000 | 0.34 | 0.20 | 0.14 | 0.33 | 0.35 | 0.32 | 0.31 |  | 0.38 |
| 2001 | 0.33 | 0.31 | 0.31 | 0.33 | 0.34 | 0.34 | 0.32 |  | 0.29 |
| 2002 | 0.34 | 0.36 | 0.35 | 0.34 | 0.35 | 0.35 | 0.35 |  | 0.30 |
| 2003 | 0.32 | 0.33 | 0.38 | 0.33 | 0.33 | 0.32 | 0.34 |  | 0.31 |
| 2004 | 0.33 | 0.35 | 0.35 | 0.35 | 0.36 | 0.34 | 0.35 |  | 0.33 |
| 2005 | 0.34 | 0.30 | 0.32 | 0.37 | 0.35 | 0.34 | 0.35 |  | 0.35 |
| 2006 | 0.35 | 0.31 |  | 0.36 | 0.37 | 0.36 | 0.33 |  | 0.35 |

Table 2.14—Multinomial sample sizes for length compositions.

| Year | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  | Trawl survey |  | IPHC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 |  |  |  |
| 1977 |  | 13 |  |  |  |  |  |  |  |  |  |  |
| 1978 | 4 |  | 23 | 21 | 25 | 18 |  |  |  |  |  |  |
| 1979 | 11 |  | 6 | 73 | 17 | 21 |  |  |  | 74 |  |  |
| 1980 | 61 |  |  | 19 | 10 | 22 |  |  |  | 74 |  |  |
| 1981 |  | 35 | 11 | 17 | 9 | 10 |  |  |  | 74 |  |  |
| 1982 | 29 |  | 12 | 13 | 8 | 35 |  |  |  |  | 166 |  |
| 1983 | 75 | 10 | 107 | 129 | 27 | 71 |  |  |  |  | 207 |  |
| 1984 | 121 | 70 | 31 | 111 | 74 | 617 |  |  |  |  | 191 |  |
| 1985 | 221 | 9 | 12 | 263 | 39 | 971 |  |  |  |  | 266 |  |
| 1986 | 211 | 5 | 11 | 246 | 10 | 784 |  |  |  |  | 242 |  |
| 1987 | 302 | 40 | 148 | 685 |  | 1304 |  |  |  |  | 167 |  |
| 1988 | 710 |  | 21 |  |  |  |  |  |  |  | 157 |  |
| 1989 | 431 |  |  |  |  |  |  |  |  |  | 157 |  |
| 1990 | 493 |  | 41 | 168 | 471 | 459 |  | 13 | 16 |  | 89 |  |
| 1991 | 592 | 8 |  | 371 | 467 | 711 |  | 70 | 64 |  | 114 |  |
| 1992 | 364 |  | 1 | 1013 | 569 | 174 | 86 | 144 | 27 |  | 151 |  |
| 1993 | 405 |  |  | 734 |  |  | 64 |  |  |  | 164 |  |
| 1994 | 721 |  | 1 | 958 |  | 358 | 166 |  | 20 |  | -219 |  |
| 1995 | 406 |  | 8 | 1069 |  | 438 | 234 | 45 | 33 |  | -145 |  |
| 1996 | 785 |  | 23 | 959 |  | 491 | 359 | 63 | 51 |  | -147 |  |
| 1997 | 822 | 4 |  | 1120 |  | 973 | 214 | 43 | 69 |  | -144 |  |
| 1998 | 667 | 14 | 16 | 1036 |  | 1349 | 151 | 37 | 25 |  | -151 |  |
| 1999 | 528 |  | 11 | 1232 | 122 | 753 | 201 |  | 38 |  | -184 |  |
| 2000 | 535 |  |  | 800 | 100 | 1302 | 174 |  |  |  | -197 |  |
| 2001 | 275 | 28 | 17 | 987 | 259 | 1381 | 140 |  | 68 |  | -311 |  |
| 2002 | 377 | 53 | 40 | 1067 | 502 | 1372 | 107 |  | 72 |  | -193 |  |
| 2003 | 383 | 95 | 35 | 1473 | 565 | 1516 | 140 |  | 85 |  | -194 |  |
| 2004 | 331 | 54 | 26 | 1279 | 537 | 1414 | 102 |  | 65 |  | -170 |  |
| 2005 | 368 | 18 |  | 1162 | 611 | 1367 | 84 |  | 73 |  | -178 |  |
| 2006 | 372 | 11 |  | 929 | 519 | 852 | 99 |  | 86 |  | -191 |  |
| 2007 | 340 | 36 |  | 669 | 116 |  | 92 |  |  |  | 157 | 44 |

Table 2.15a—Summary of statistics pertaining to four models.

## Parameter counts

No. fixed mortality/growth parameters
No. fixed selectivity parameters
Total no. fixed parameters
No. free mortality/growth parameters
No. stock-recruitment parameters
No. recruitment deviations
No. initial fishing mortalities
No. catchabilities
No. free selectivity parameters
No. selectivity deviations
Total no. free parameters
Total no. fixed and free parameters
Aggregate likelihood components
Abundance indices
Length compositions
Age compositions
Initial catch
Recruitment
Annual deviations
Total
Relative abundance likelihoods
Jan-May trawl fishery CPUE Jun-Aug trawl fishery CPUE Sep-Dec trawl fishery CPUE Jan-May longline fishery CPUE Jun-Aug longline fishery CPUE Sep-Dec longline fishery CPUE Jan-May pot fishery CPUE Jun-Aug pot fishery CPUE Sep-Dec pot fishery CPUE
Pre-1982 trawl survey abundance Post-1982 trawl survey abundance IPHC longline survey CPUE

## Size composition likelihoods

 Jan-May trawl fishery sizecomp Jun-Aug trawl fishery sizecomp Sep-Dec trawl fishery sizecomp Jan-May longline fishery sizecomp Jun-Aug longline fishery sizecomp Sep-Dec longline fishery sizecomp Jan-May pot fishery sizecomp Jun-Aug pot fishery sizecomp Sep-Dec pot fishery sizecomp Pre-1982 trawl survey sizecomp Post-1982 trawl survey sizecomp IPHC longline survey sizecomp| Model 1 | Model 2 | Model 3 | Model 4 |
| :---: | :---: | :---: | :---: |
|  |  |  |  |
| 1 | 1 | 0 | 0 |
| 30 | 32 | 31 | 38 |
| 31 | 33 | 31 | 38 |
| 5 | 5 | 6 | 6 |
| 2 | 2 | 2 | 2 |
| 33 | 33 | 33 | 28 |
| 1 | 1 | 1 | 1 |
| 2 | 2 | 2 | 1 |
| 42 | 40 | 41 | 28 |
| 29 | 29 | 29 | 52 |
| 114 | 112 | 114 | 118 |
| 145 | 145 | 145 | 156 |
| 71.60 | 71.09 | 52.57 | 60.50 |
| 1748.40 | 1775.08 | 1748.74 | 1781.17 |
| 206.66 | 176.57 | 179.87 | n/a |
| 0.01 | 0.01 | 0.16 | n/a |
| 29.66 | 31.03 | 45.72 | 17.17 |
| 9.46 | 20.74 | 10.81 | 18.73 |
| 2065.79 | 2074.52 | 2037.86 | 1877.57 |
| 48.41 | 50.83 | 44.16 | 48.65 |
| 29.46 | 31.47 | 26.52 | 31.21 |
| 38.99 | 40.88 | 34.55 | 40.38 |
| 9.54 | 10.65 | 10.09 | 12.17 |
| 6.37 | 7.92 | 5.95 | 10.11 |
| 10.14 | 13.27 | 7.38 | 17.13 |
| 3.83 | 3.10 | 6.17 | 3.07 |
| 0.48 | 0.48 | 0.44 | 0.49 |
| 11.58 | 11.37 | 11.88 | 10.81 |
| 4.16 | 5.57 | 5.47 | n/a |
| 67.44 | 65.52 | 47.10 | 60.50 |
| 31.72 | 33.13 | 37.01 | 36.68 |
| 480.73 | 490.26 | 494.53 | 504.15 |
| 103.25 | 102.82 | 103.00 | 77.84 |
| 69.40 | 68.66 | 66.24 | 60.27 |
| 331.07 | 331.62 | 323.65 | 296.77 |
| 122.72 | 120.39 | 120.55 | 120.86 |
| 377.69 | 374.26 | 370.99 | 364.71 |
| 37.70 | 38.98 | 37.04 | 43.75 |
| 22.40 | 22.29 | 22.36 | 24.19 |
| 41.57 | 41.73 | 41.06 | 46.25 |
| 24.22 | 25.54 | 23.41 | n/a |
| 136.87 | 157.78 | 145.16 | 241.15 |
| 0.80 | 0.76 | 0.75 | 1.23 |

Table 2.15b—Summary of key parameters from four models.

## Life history and recruitment

Natural mortality rate M
Mean Jan. population length at age 1
Mean Jan. population length at age 20
Brody growth rate K
Std. dev. of length at age 1
Std. dev. of length at age 20
Mean July survey length at age 1
Mean July survey length at age 2
Mean July survey length at age 3
Post-1976 log median recruits (1000s)
Pre-1977 log median recruits (1000s)
Standard deviation of log recruits
Catchability and selectivity
Pre-1982 trawl survey log catchability Post-1981 trawl survey log catchability Jan-May trawl fish. begin peak Jan-May trawl fish. sel. at max. size Jun-Aug trawl fish. begin peak Jun-Aug trawl fish. sel. at max. size Sep-Dec trawl fish. begin peak Sep-Dec trawl fish. sel. at max. size Jan-May longl. fish. begin peak Jan-May longl. fish. sel. at max. size Jun-Aug longl. fish. begin peak Jun-Aug longl. fish. sel. at max. size Sep-Dec longl. fish. begin peak
Sep-Dec longl. fish. sel. at max. size Jan-May pot fish. begin peak Jan-May pot fish. sel. at max. size Jun-Aug pot fish. begin peak Jun-Aug pot fish. sel. at max. size Sep-Dec pot fish. begin peak
Sep-Dec pot fish. sel. at max. size Pre-1982 trawl surv. begin peak Pre-1982 trawl surv. sel. at max. size Post-1981 trawl surv. begin peak Post-1981 trawl surv. sel. at max. size IPHC longline surv. begin peak IPHC longline surv. sel. at max. size Ave. post-82 surv. sel. x Q (60-80 cm)

| Model 1 | Model 2 | Model 3 | Model 4 |
| ---: | ---: | ---: | ---: |
|  |  |  |  |
| 0.34 | 0.37 | 0.22 | 0.46 |
| 6.99 | 7.14 | 6.83 | 6.33 |
| 94.03 | 96.89 | 93.47 | 102.55 |
| 0.22 | 0.21 | 0.23 | 0.20 |
| 3.11 | 3.23 | 3.05 | 4.12 |
| 9.26 | 9.28 | 9.45 | 13.76 |
| 17.16 | 17.15 | 17.10 | 16.85 |
| 32.83 | 32.69 | 32.89 | 34.23 |
| 45.48 | 45.39 | 45.58 | 46.21 |
| 13.59 | 13.77 | 12.40 | 13.84 |
| 11.91 | 11.83 | 10.85 | 14.59 |
| 0.78 | 0.78 | 0.78 | 0.60 |
|  |  |  |  |
| 0.11 | -0.05 | 0.98 | $\mathrm{n} / \mathrm{a}$ |
| 0.33 | -0.17 | 0.42 | -0.38 |
| 74.99 | 76.93 | 75.06 | 81.56 |
| 0.71 | 0.72 | 0.70 | 0.57 |
| 69.60 | 73.08 | 70.24 | 86.31 |
| 1.00 | 1.00 | 1.00 | 1.00 |
| 86.66 | 89.62 | 88.68 | 99.75 |
| 1.00 | 1.00 | 1.00 | 1.00 |
| 65.20 | 66.04 | 65.05 | 67.66 |
| 0.33 | 0.35 | 0.32 | 0.45 |
| 65.35 | 66.98 | 65.32 | 70.76 |
| 1.00 | 1.00 | 1.00 | 1.00 |
| 66.04 | 67.24 | 65.96 | 70.01 |
| 0.70 | 0.72 | 0.68 | 1.00 |
| 67.97 | 68.55 | 67.97 | 69.83 |
| 0.46 | 0.50 | 0.45 | 0.67 |
| 67.02 | 67.79 | 66.95 | 67.99 |
| 0.56 | 0.61 | 0.54 | 1.00 |
| 64.16 | 65.50 | 64.07 | 68.28 |
| 1.00 | 1.00 | 1.00 | 1.00 |
| 1.22 | 1.25 | 1.22 | $\mathrm{n} / \mathrm{a}$ |
| 0.13 | 0.19 | 0.14 | $\mathrm{n} / \mathrm{a}$ |
| 1.11 | 3.82 | 1.09 | 34.15 |
| 0.10 | 0.11 | 0.04 | 1.00 |
| 68.24 | 68.39 | 68.28 | 66.75 |
| 0.50 | 0.54 | 0.51 | 1.00 |
| 0.56 | 0.64 | 1.08 | 0.68 |
|  |  |  |  |

Table 2.15c—Summary of key management reference points from four models.

|  | Model 1 | Model 2 | Model 3 | Model 4 |
| :--- | ---: | ---: | ---: | ---: |
| Examples of historical biomass (BS) |  |  |  | n/a |
| BS total biomass 1977 | 528 | 367 | 126 | 1875 |
| BS total biomass 1985 | 2501 | 2179 | 1165 | 1018 |
| BS total biomass 2007 | 1051 | 1074 | 510 | 116 |
| BS age 3+ biomass 1977 | 487 | 326 | 1153 | 1818 |
| BS age 3+ biomass 1985 | 2469 | 2147 | 1153 |  |
| BS age 3+ biomass 2007 | 1000 | 991 | 489 | 917 |
| BS female spawning biomass 1977 | 128 | 76 | 19 | n/a |
| BS female spawning biomass 1985 | 932 | 773 | 440 | 613 |
| BS female spawning biomass 2007 | 370 | 353 | 181 | 310 |
| Projected biomass (BSAI) |  |  |  |  |
| BSAI total biomass 2008 | 1266 | 1474 | 584 | 1465 |
| BSAI total biomass 2009 | 1513 | 2006 | 830 | 1956 |
| BSAI age 3+ biomass 2008 | 1080 | 1137 | 504 | 1054 |
| BSAI age 3+ biomass 2009 | 1423 | 1904 | 796 | 1861 |
| BSAI female spawning biomass 2008 | 398 | 396 | 192 | 341 |
| BSAI female spawning biomass 2009 | 395 | 422 | 237 | 357 |
| Spawning biomass reference points |  |  |  |  |
| B100\% | 1350 | 1290 | 1420 | 748 |
| B40\% | 540 | 516 | 568 | 299 |
| B35\% | 473 | 452 | 497 | 262 |
| Proportion of B100\% in 2008 | 0.29 | 0.31 | 0.14 | 0.46 |
| Proportion of B100\% in 2009 | 0.29 | 0.33 | 0.17 | 0.48 |
| ABC reference points |  |  |  |  |
| F40\% | 0.31 | 0.32 | 0.21 | 0.45 |
| maxFABC 2008 | 0.22 | 0.24 | 0.063 | 0.45 |
| maxFABC 2009 | 0.22 | 0.26 | 0.080 | 0.45 |
| BSAI ABC 2007 (Council adopted) | 176 | 176 | 176 | 176 |
| BSAI maxABC 2008 (from model) | 150 | 182 | 22.4 | 281 |
| BSAI maxABC 2009 (from model) | 162 | 240 | 38.1 | 327 |
| rel. change in ABC (2007 to 2008) | -0.15 | 0.03 | -0.87 | 0.60 |
| rel. change in ABC (2007 to 2009) | -0.08 | 0.36 | -0.78 | 0.86 |
| OFL reference points |  |  |  |  |
| F35\% | 0.37 | 0.39 | 0.25 | 0.54 |
| FOFL 2008 | 0.26 | 0.29 | 0.075 | 0.54 |
| FOFL 2009 (Scenario 6) | 0.26 | 0.30 | 0.094 | 0.54 |
| BSAI OFL 2007 (Council adopted) | 207 | 207 | 207 | 207 |
| BSAI OFL 2008 (from model) | 176 | 214 | 26.5 | 332 |
| BSAI OFL 2009 (from model) | 190 | 283 | 45.0 | 389 |
| rel. change in OFL (2007 to 2008) | -0.15 | 0.03 | -0.87 | 0.60 |
| rel. change in OFL (2007 to 2009) | -0.08 | 0.37 | -0.78 | 0.88 |

Table 2.16a-Estimates and standard deviations of parameters (except annual devs) from four models.

| Parameter | Model 1 |  | Model 2 |  | Model 3 |  | Model 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Est. | S.D. | Est. | S.D. | Est. | S.D. | Est. | S.D. |
| Natural mortality |  |  |  |  | 0.216 | 0.016 | 0.464 | 0.018 |
| Mean length at age 1 | 6.989 | 0.288 | 7.138 | 0.295 | 6.826 | 0.290 | 6.333 | 0.294 |
| Mean length at age 20 | 94.025 | 0.851 | 96.888 | 1.086 | 93.466 | 0.835 | 102.550 | 1.079 |
| Brody growth rate | 0.225 | 0.004 | 0.213 | 0.005 | 0.228 | 0.004 | 0.204 | 0.004 |
| Std. dev. of length at age 1 | 3.106 | 0.161 | 3.227 | 0.168 | 3.046 | 0.164 | 4.121 | 0.124 |
| Std. dev. of length at age 20 (offset) | 1.093 | 0.071 | 1.057 | 0.078 | 1.133 | 0.073 | 1.206 | 0.086 |
| Post-1976 log median recruits | 13.593 | 0.068 | 13.769 | 0.068 | 12.403 | 0.141 | 13.837 | 0.159 |
| Pre-1977 log median recruits (offset) | -1.680 | 0.150 | -1.936 | 0.123 | -1.549 | 0.113 | 0.749 | 0.103 |
| Initial fishing mortality | 0.162 | 0.051 | 0.271 | 0.094 | 0.850 | 0.226 | 0.728 | 0.078 |
| Pre-1982 survey catchability (log) | -0.113 | 0.130 | -0.049 | 0.135 | 0.984 | 0.141 | n/a | n/a |
| Post-1981 survey catchability (log) | -0.331 | 0.071 | -0.173 | 0.081 | 0.419 | 0.081 | -0.382 | 0.113 |
| Jan-May trawl beg. peak | 74.991 | 1.001 | 76.929 | 0.956 | 75.055 | 0.974 | 81.563 | 0.893 |
| Jan-May trawl asc. width | 6.420 | 0.041 | 6.435 | 0.037 | 6.495 | 0.044 | 6.507 | 0.035 |
| Jan-May trawl des. width | 4.185 | 0.635 | 4.228 | 0.804 | 4.254 | 0.623 |  |  |
| Jan-May trawl final sel. | 0.907 | 0.278 | 0.951 | 0.372 | 0.827 | 0.278 | 0.299 | 0.471 |
| Jun-Aug trawl beg. peak | 69.601 | 3.987 | 73.084 | 4.308 | 70.242 | 4.379 | 86.314 | 5.714 |
| Jun-Aug trawl asc. width | 6.334 | 0.222 | 6.406 | 0.213 | 6.434 | 0.242 | 6.733 | 0.203 |
| Sep-Dec trawl beg. peak | 86.656 | 4.563 | 89.621 | 4.238 | 88.682 | 4.765 | 99.754 | 4.663 |
| Sep-Dec trawl asc. width | 6.650 | 0.180 | 6.675 | 0.156 | 6.762 | 0.182 | 6.843 | 0.140 |
| Jan-May longline beg. peak | 65.197 | 0.389 | 66.043 | 0.387 | 65.046 | 0.372 | 67.656 | 0.319 |
| Jan-May longline asc. width | 5.292 | 0.031 | 5.317 | 0.029 | 5.319 | 0.031 | 5.365 | 0.025 |
| Jan-May longline des. width | 5.142 | 0.135 | 5.189 | 0.160 | 5.209 | 0.135 | 5.330 | 0.265 |
| Jan-May longline final sel. | -0.714 | 0.132 | -0.615 | 0.152 | -0.768 | 0.135 | -0.195 | 0.204 |
| Jun-Aug longline beg. peak | 65.347 | 0.816 | 66.979 | 0.775 | 65.324 | 0.791 | 70.762 | 0.757 |
| Jun-Aug longline asc. width | 5.150 | 0.074 | 5.235 | 0.065 | 5.180 | 0.074 | 5.429 | 0.057 |
| Sep-Dec longline beg. peak | 66.042 | 0.495 | 67.241 | 0.488 | 65.958 | 0.480 | 70.006 | 0.496 |
| Sep-Dec longline peak width | -1.912 | 0.619 | -1.711 | 0.711 | -1.948 | 0.671 |  |  |
| Sep-Dec longline asc. width | 5.232 | 0.039 | 5.284 | 0.035 | 5.263 | 0.039 | 5.414 | 0.034 |
| Sep-Dec longline des. width | 4.799 | 0.754 | 5.054 | 1.345 | 4.968 | 0.795 |  |  |
| Sep-Dec longline final sel. | 0.852 | 0.289 | 0.940 | 0.550 | 0.776 | 0.314 |  |  |
| Jan-May pot beg. peak | 67.973 | 0.684 | 68.555 | 0.712 | 67.966 | 0.688 | 69.835 | 0.793 |
| Jan-May pot asc. width | 5.068 | 0.073 | 5.081 | 0.072 | 5.094 | 0.074 | 5.122 | 0.074 |
| Jan-May pot des. width | 4.436 | 0.413 | 4.434 | 0.497 | 4.503 | 0.409 | 3.993 | 0.915 |
| Jan-May pot final sel. | -0.160 | 0.225 | 0.004 | 0.254 | -0.217 | 0.228 | 0.716 | 0.326 |
| Jun-Aug pot beg. peak | 67.022 | 1.758 | 67.788 | 1.674 | 66.954 | 1.778 | 67.992 | 1.545 |
| Jun-Aug pot asc. width | 5.072 | 0.184 | 5.102 | 0.170 | 5.093 | 0.188 | 5.084 | 0.162 |
| Jun-Aug pot des. width | 4.247 | 1.200 | 4.237 | 1.424 | 4.374 | 1.117 |  |  |
| Jun-Aug pot final sel. | 0.259 | 0.435 | 0.429 | 0.505 | 0.152 | 0.429 |  |  |
| Sep-Dec pot beg. peak | 64.163 | 1.414 | 65.500 | 1.387 | 64.070 | 1.436 | 68.278 | 1.502 |
| Sep-Dec pot asc. width | 4.754 | 0.182 | 4.851 | 0.166 | 4.773 | 0.187 | 5.032 | 0.157 |
| Pre-1982 survey beg. peak | 1.222 | 0.293 | 1.247 | 0.418 | 1.216 | 0.338 | n/a | n/a |
| Pre-1982 survey peak width | -2.159 | 0.165 | -2.162 | 0.236 | -2.173 | 0.188 | n/a | /a |
| Pre-1982 survey asc. width | -3.968 | 3.202 | -3.834 | 4.002 | -3.992 | 3.775 | n/a | n/a |
| Pre-1982 survey final sel. | -1.927 | 1.016 | -1.478 | 1.074 | -1.853 | 0.900 | n/a | $\mathrm{n} / \mathrm{a}$ |
| Post-1981 survey beg. peak | 1.115 | 0.022 | 3.823 | 0.184 | 1.095 | 0.018 | 34.151 | 0.699 |
| Post-1981 survey peak width | -1.462 | 0.216 |  |  | -1.993 | 0.340 |  |  |
| Post-1981 survey asc. width | -4.432 | 0.490 | 2.438 | 0.236 | -4.629 | 0.489 | 4.534 | 0.422 |
| Post-1981 survey des. width | 1.893 | 0.616 | 2.368 | 0.295 | 2.740 | 0.448 |  |  |
| Post-1981 survey initial sel. |  |  |  |  |  |  | -1.492 | 0.179 |
| Post-1981 survey final sel. | -2.157 | 0.617 | -2.120 | 0.714 | -3.106 | 1.334 |  |  |
| IPHC survey beg. peak | 68.243 | 5.603 | 68.389 | 0.382 | 68.284 | 0.176 | 66.749 | 3.781 |
| IPHC survey peak width | -3.315 | 4.581 |  |  |  |  |  |  |
| IPHC survey asc. width | 4.510 | 0.777 | 4.489 | 0.290 | 4.510 | 0.292 | 4.295 | 0.657 |
| IPHC survey final sel. | -0.004 | 0.915 | 0.144 | 0.826 | 0.040 | 0.779 |  |  |

Table 2.16b—Estimates and standard deviations of annual recruitment devs estimated by four models.

|  | Model 1 |  | Model 2 |  | Model 3 |  | Model 4 |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Year | Est. | S.D. | Est. | S.D. | Est. | S.D. | Est. |
| S.D. |  |  |  |  |  |  |  |  |
| 1974 | 1.873 | 0.208 | 1.775 | 0.213 | 1.502 | 0.175 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 1975 | 0.183 | 0.693 | 0.389 | 0.598 | -0.411 | 0.745 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 1976 | 1.599 | 0.249 | 1.580 | 0.254 | 1.122 | 0.269 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 1977 | 1.456 | 0.096 | 1.274 | 0.098 | 1.199 | 0.092 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 1978 | 0.287 | 0.178 | 0.208 | 0.166 | 0.032 | 0.184 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ |
| 1979 | 0.568 | 0.109 | 0.483 | 0.105 | 0.358 | 0.114 | 0.052 | 0.128 |
| 1980 | -0.361 | 0.152 | -0.229 | 0.133 | -0.419 | 0.148 | -0.505 | 0.114 |
| 1981 | -1.025 | 0.186 | -1.158 | 0.181 | -1.220 | 0.200 | -1.519 | 0.162 |
| 1982 | 0.830 | 0.052 | 0.778 | 0.050 | 0.804 | 0.051 | 0.930 | 0.051 |
| 1983 | -0.865 | 0.156 | -1.029 | 0.157 | -0.919 | 0.166 | -0.424 | 0.109 |
| 1984 | 0.571 | 0.058 | 0.572 | 0.053 | 0.612 | 0.058 | 0.758 | 0.051 |
| 1985 | -0.065 | 0.084 | -0.204 | 0.083 | -0.035 | 0.087 | 0.184 | 0.068 |
| 1986 | -0.730 | 0.115 | -0.713 | 0.106 | -0.639 | 0.118 | -0.357 | 0.083 |
| 1987 | -1.315 | 0.155 | -1.204 | 0.133 | -1.208 | 0.162 | -0.766 | 0.091 |
| 1988 | -0.408 | 0.077 | -0.526 | 0.076 | -0.269 | 0.080 | -0.416 | 0.067 |
| 1989 | 0.306 | 0.054 | 0.239 | 0.052 | 0.463 | 0.058 | 0.335 | 0.047 |
| 1990 | 0.215 | 0.057 | 0.137 | 0.055 | 0.387 | 0.061 | 0.405 | 0.047 |
| 1991 | -0.432 | 0.075 | -0.482 | 0.071 | -0.254 | 0.079 | -0.268 | 0.063 |
| 1992 | 0.403 | 0.044 | 0.353 | 0.041 | 0.592 | 0.051 | 0.470 | 0.041 |
| 1993 | -0.594 | 0.074 | -0.637 | 0.070 | -0.431 | 0.079 | -0.251 | 0.057 |
| 1994 | -0.603 | 0.063 | -0.683 | 0.061 | -0.469 | 0.067 | -0.420 | 0.051 |
| 1995 | -0.498 | 0.060 | -0.555 | 0.057 | -0.368 | 0.064 | -0.578 | 0.056 |
| 1996 | 0.413 | 0.043 | 0.321 | 0.041 | 0.527 | 0.047 | 0.356 | 0.041 |
| 1997 | -0.272 | 0.059 | -0.280 | 0.055 | -0.162 | 0.063 | 0.102 | 0.042 |
| 1998 | -0.170 | 0.055 | -0.181 | 0.051 | -0.056 | 0.058 | -0.057 | 0.047 |
| 1999 | 0.427 | 0.045 | 0.372 | 0.042 | 0.530 | 0.048 | 0.508 | 0.040 |
| 2000 | 0.055 | 0.050 | 0.109 | 0.047 | 0.153 | 0.052 | 0.538 | 0.040 |
| 2001 | -0.594 | 0.064 | -0.579 | 0.061 | -0.525 | 0.065 | -0.440 | 0.068 |
| 2002 | -0.393 | 0.062 | -0.338 | 0.058 | -0.353 | 0.063 | 0.071 | 0.055 |
| 2003 | -0.523 | 0.074 | -0.501 | 0.070 | -0.521 | 0.077 | -0.240 | 0.075 |
| 2004 | -1.065 | 0.098 | -0.904 | 0.096 | -0.991 | 0.099 | -0.403 | 0.099 |
| 2005 | -0.405 | 0.131 | -0.038 | 0.112 | -0.290 | 0.115 | 0.074 | 0.126 |
| 2006 | 1.134 | 0.207 | 1.651 | 0.176 | 1.258 | 0.192 | 1.861 | 0.180 |
|  |  |  |  |  |  |  |  |  |

Table 2.16c—Estimates and standard deviations of selectivity parameter devs estimated by four models.

| Survey | Parameter | Year | Model 1 |  | Model 2 |  | Model 3 |  | Model 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | Est. | S.D. | Est. | S.D. | Est. | S.D. | Est. | S.D. |
| Pre-1982 survey | asc. width | 1979 | 0.033 | 0.144 | 0.026 | 0.144 | 0.023 | 0.146 | n/a | n/a |
| Pre-1982 survey | asc. width | 1980 | 0.051 | 0.137 | 0.038 | 0.136 | 0.047 | 0.137 | n/a | n/a |
| Pre-1982 survey | asc. width | 1981 | -0.083 | 0.139 | -0.064 | 0.137 | -0.071 | 0.139 | n/a | n/a |
| Post-1981 survey | asc. width | 1982 | 0.072 | 0.090 | -0.249 | 0.150 | 0.013 | 0.116 | -0.668 | 0.236 |
| Post-1981 survey | asc. width | 1983 | 0.001 | 0.058 | 0.063 | 0.122 | -0.004 | 0.064 | -0.380 | 0.173 |
| Post-1981 survey | asc. width | 1984 | 0.097 | 0.090 | 0.014 | 0.142 | 0.078 | 0.104 | 0.126 | 0.111 |
| Post-1981 survey | asc. width | 1985 | -0.120 | 0.073 | 0.299 | 0.134 | -0.138 | 0.087 | 0.320 | 0.123 |
| Post-1981 survey | asc. width | 1986 | 0.072 | 0.065 | 0.168 | 0.140 | 0.077 | 0.072 | 0.021 | 0.109 |
| Post-1981 survey | asc. width | 1987 | -0.069 | 0.102 | 0.182 | 0.149 | -0.073 | 0.115 | 0.206 | 0.135 |
| Post-1981 survey | asc. width | 1988 | 0.123 | 0.097 | -0.050 | 0.159 | 0.142 | 0.103 | -0.014 | 0.126 |
| Post-1981 survey | asc. width | 1989 | 0.234 | 0.076 | -0.556 | 0.108 | 0.269 | 0.081 | -0.824 | 0.281 |
| Post-1981 survey | asc. width | 1990 | 0.003 | 0.064 | 0.031 | 0.129 | 0.034 | 0.067 | -0.169 | 0.185 |
| Post-1981 survey | asc. width | 1991 | 0.049 | 0.065 | 0.016 | 0.134 | 0.079 | 0.068 | -0.014 | 0.134 |
| Post-1981 survey | asc. width | 1992 | -0.246 | 0.116 | 0.317 | 0.142 | -0.236 | 0.123 | 0.506 | 0.197 |
| Post-1981 survey | asc. width | 1993 | -0.232 | 0.099 | 0.376 | 0.134 | -0.236 | 0.110 | 0.511 | 0.197 |
| Post-1981 survey | asc. width | 1994 | 0.090 | 0.070 | 0.117 | 0.145 | 0.104 | 0.076 | 0.176 | 0.122 |
| Post-1981 survey | asc. width | 1995 | 0.221 | 0.084 | 0.036 | 0.150 | 0.241 | 0.087 | 0.094 | 0.131 |
| Post-1981 survey | asc. width | 1996 | 0.272 | 0.093 | -0.205 | 0.161 | 0.283 | 0.092 | -0.010 | 0.146 |
| Post-1981 survey | asc. width | 1997 | 0.092 | 0.051 | -0.206 | 0.098 | 0.101 | 0.054 | 0.097 | 0.129 |
| Post-1981 survey | asc. width | 1998 | 0.172 | 0.065 | -0.199 | 0.110 | 0.198 | 0.070 | -0.029 | 0.108 |
| Post-1981 survey | asc. width | 1999 | 0.191 | 0.060 | -0.484 | 0.092 | 0.217 | 0.065 | -0.069 | 0.110 |
| Post-1981 survey | asc. width | 2000 | 0.072 | 0.051 | -0.317 | 0.086 | 0.083 | 0.053 | -0.311 | 0.138 |
| Post-1981 survey | asc. width | 2001 | -0.289 | 0.100 | 0.327 | 0.132 | -0.328 | 0.113 | 0.510 | 0.182 |
| Post-1981 survey | asc. width | 2002 | 0.052 | 0.062 | -0.351 | 0.105 | 0.052 | 0.069 | -0.170 | 0.158 |
| Post-1981 survey | asc. width | 2003 | -0.031 | 0.059 | 0.061 | 0.119 | -0.067 | 0.072 | 0.279 | 0.130 |
| Post-1981 survey | asc. width | 2004 | -0.010 | 0.061 | -0.075 | 0.112 | -0.041 | 0.073 | 0.125 | 0.146 |
| Post-1981 survey | asc. width | 2005 | -0.288 | 0.117 | 0.312 | 0.141 | -0.295 | 0.123 | 0.529 | 0.188 |
| Post-1981 survey | asc. width | 2006 | -0.273 | 0.117 | 0.211 | 0.143 | -0.300 | 0.123 | -0.439 | 0.217 |
| Post-1981 survey | asc. width | 2007 | -0.256 | 0.134 | 0.161 | 0.164 | -0.252 | 0.136 | -0.400 | 0.181 |
| Post-1981 survey | initial sel. | 1982 | n/a | n/a | n/a | n/a | n/a | n/a | 0.037 | 0.230 |
| Post-1981 survey | initial sel. | 1983 | n/a | n/a | n/a | n/a | n/a | n/a | -0.451 | 0.202 |
| Post-1981 survey | initial sel. | 1984 | n/a | n/a | n/a | n/a | n/a | n/a | 0.210 | 0.294 |
| Post-1981 survey | initial sel. | 1985 | n/a | n/a | n/a | n/a | n/a | n/a | 0.353 | 0.241 |
| Post-1981 survey | initial sel. | 1986 | n/a | n/a | n/a | n/a | n/a | n/a | -0.083 | 0.237 |
| Post-1981 survey | initial sel. | 1987 | n/a | n/a | n/a | n/a | n/a | n/a | 0.219 | 0.352 |
| Post-1981 survey | initial sel. | 1988 | n/a | n/a | n/a | n /a | n/a | n/a | 0.324 | 0.293 |
| Post-1981 survey | initial sel. | 1989 | n/a | n/a | n/a | n/a | n/a | n/a | 0.300 | 0.176 |
| Post-1981 survey | initial sel. | 1990 | n/a | n/a | n/a | n/a | n/a | n/a | -0.526 | 0.239 |
| Post-1981 survey | initial sel. | 1991 | n/a | n/a | n/a | n/a | n/a | n/a | -0.126 | 0.251 |
| Post-1981 survey | initial sel. | 1992 | n/a | n/a | n/a | n/a | n/a | n/a | -0.113 | 0.348 |
| Post-1981 survey | initial sel. | 1993 | n/a | n/a | n/a | n/a | n/a | n/a | -0.261 | 0.311 |
| Post-1981 survey | initial sel. | 1994 | n/a | n/a | n/a | n/a | n/a | n/a | 0.278 | 0.324 |
| Post-1981 survey | initial sel. | 1995 | n/a | n/a | n/a | n/a | n/a | n/a | 0.157 | 0.329 |
| Post-1981 survey | initial sel. | 1996 | n/a | n/a | n/a | n/a | n/a | n/a | 0.224 | 0.301 |
| Post-1981 survey | initial sel. | 1997 | n/a | n/a | n/a | n/a | n/a | n/a | 0.062 | 0.294 |
| Post-1981 survey | initial sel. | 1998 | n/a | n/a | n/a | n/a | n/a | n/a | 0.207 | 0.215 |
| Post-1981 survey | initial sel. | 1999 | n/a | n/a | n/a | n/a | n/a | n/a | 0.161 | 0.206 |
| Post-1981 survey | initial sel. | 2000 | n/a | n/a | n/a | n/a | n/a | n/a | -0.250 | 0.186 |
| Post-1981 survey | initial sel. | 2001 | n/a | n/a | n/a | n/a | n/a | n/a | 0.283 | 0.272 |
| Post-1981 survey | initial sel. | 2002 | n/a | n/a | n/a | n/a | n/a | n/a | -0.234 | 0.249 |
| Post-1981 survey | initial sel. | 2003 | n/a | n/a | n/a | n/a | n/a | n/a | 0.402 | 0.310 |
| Post-1981 survey | initial sel. | 2004 | n/a | n/a | n/a | n/a | n/a | n/a | 0.094 | 0.366 |
| Post-1981 survey | initial sel. | 2005 | n/a | n/a | n/a | n/a | n/a | n/a | 0.272 | 0.357 |
| Post-1981 survey | initial sel. | 2006 | n/a | n/a | n/a | n/a | n/a | n/a | -0.808 | 0.254 |
| Post-1981 survey | initial sel. | 2007 | n/a | n/a | n/a | n/a | n/a | n/a | -0.730 | 0.278 |

Table 2.17a—Root mean squared errors for fishery CPUE and survey relative abundance time series.

## Root Mean Squared Error

| Fishery/Survey | 1 | 2 | 3 | 4 |
| :--- | ---: | ---: | ---: | ---: |
| Jan-May trawl fishery | 0.23 | 0.24 | 0.22 | 0.23 |
| Jun-Aug trawl fishery | 0.43 | 0.44 | 0.41 | 0.43 |
| Sep-Dec trawl fishery | 0.59 | 0.61 | 0.56 | 0.61 |
| Jan-May longline fishery | 0.19 | 0.20 | 0.20 | 0.22 |
| Jun-Aug longline fishery | 0.23 | 0.25 | 0.25 | 0.28 |
| Sep-Dec longline fishery | 0.21 | 0.24 | 0.18 | 0.27 |
| Jan-May pot fishery | 0.11 | 0.10 | 0.13 | 0.10 |
| Jun-Aug pot fishery | 0.10 | 0.10 | 0.09 | 0.10 |
| Sep-Dec pot fishery | 0.30 | 0.29 | 0.32 | 0.28 |
| Pre-1982 trawl survey | 0.21 | 0.24 | 0.24 | n/a |
| Post-1982 trawl survey | 0.23 | 0.24 | 0.20 | 0.24 |
| IPHC longline survey | 0.29 | 0.30 | 0.30 | 0.32 |

Table 2.17b-Correlations between observed data and model estimates for fishery CPUE and survey relative abundance time series.

## Correlation (observed:estimated)

| Fishery/Survey | 1 | 2 | 3 | 4 |
| :--- | ---: | ---: | ---: | ---: |
| Jan-May trawl fishery | 0.62 | 0.57 | 0.66 | 0.58 |
| Jun-Aug trawl fishery | 0.43 | 0.31 | 0.48 | 0.29 |
| Sep-Dec trawl fishery | 0.13 | 0.04 | 0.31 | 0.06 |
| Jan-May longline fishery | 0.18 | 0.08 | 0.34 | 0.14 |
| Jun-Aug longline fishery | 0.25 | 0.10 | 0.31 | 0.04 |
| Sep-Dec longline fishery | 0.17 | -0.10 | 0.53 | -0.15 |
| Jan-May pot fishery | 0.73 | 0.74 | 0.63 | 0.76 |
| Jun-Aug pot fishery | 0.87 | 0.87 | 0.88 | 0.86 |
| Sep-Dec pot fishery | 0.55 | 0.55 | 0.48 | 0.53 |
| Pre-1982 trawl survey | 0.98 | 1.00 | 0.77 | $\mathrm{n} / \mathrm{a}$ |
| Post-1982 trawl survey | 0.61 | 0.59 | 0.74 | 0.61 |
| IPHC longline survey | -0.45 | -0.46 | -0.38 | -0.38 |

Table 2.18a—Median effective multinomial sample size from the data ("input") and estimated for each fishery and survey size composition time series.

Median Effective $\mathbf{N}$

|  | Models using base data file |  |  |  | Model 4 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Fishery/Survey | Input | 1 | 2 | 3 | Input | Output |
| Jan-May trawl fishery | 372 | 215 | 223 | 205 | 377 | 208 |
| Jun-Aug trawl fishery | 18 | 56 | 52 | 55 | 18 | 64 |
| Sep-Dec trawl fishery | 17 | 62 | 63 | 61 | 17 | 62 |
| Jan-May longline fishery | 734 | 279 | 261 | 281 | 929 | 469 |
| Jun-Aug longline fishery | 116 | 231 | 221 | 239 | 259 | 301 |
| Sep-Dec longline fishery | 711 | 240 | 220 | 238 | 784 | 301 |
| Jan-May pot fishery | 140 | 223 | 245 | 224 | 140 | 218 |
| Jun-Aug pot fishery | 45 | 116 | 129 | 111 | 45 | 141 |
| Sep-Dec pot fishery | 64 | 196 | 188 | 191 | 64 | 151 |
| Pre-1982 trawl survey | 74 | 62 | 55 | 66 | n/a | n/a |
| Post-1982 trawl survey (years without ages) | 164 | 212 | 141 | 164 | 164 | 216 |
| Post-1982 trawl survey (years with ages) | -184 | 141 | 105 | 133 | 184 | 154 |
| IPHC longline survey | 44 | 925 | 967 | 968 | 44 | 178 |

Table 2.18b—Mean effective multinomial sample size from the data ("input") and estimated for each fishery and survey size composition time series.

## Mean Effective $\mathbf{N}$

|  | Models using base data file |  |  | Model 4 |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Fishery/Survey | Input | 1 | 2 | 3 | Input |

Table 2.19—Estimates of Pacific cod fishing mortality rates, expressed on an annual time scale (Model 1). Empty cells indicate that recorded catch was negligible or that no catch was recorded.

|  |  | Trawl |  | Longline |  |  | Pot |  |  |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
| Year | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 |
| 1977 | 0.050 | 0.016 | 0.026 | 0.007 | 0.001 | 0.010 |  |  |  |
| 1978 | 0.051 | 0.016 | 0.027 | 0.007 | 0.001 | 0.009 |  |  |  |
| 1979 | 0.032 | 0.010 | 0.016 | 0.004 | 0.001 | 0.006 |  |  |  |
| 1980 | 0.025 | 0.007 | 0.012 | 0.004 |  | 0.004 |  |  |  |
| 1981 | 0.013 | 0.009 | 0.020 | 0.001 |  | 0.003 |  |  |  |
| 1982 | 0.013 | 0.009 | 0.012 |  |  | 0.001 |  |  |  |
| 1983 | 0.021 | 0.011 | 0.014 | 0.002 |  | 0.001 |  |  |  |
| 1984 | 0.024 | 0.011 | 0.015 | 0.003 | 0.001 | 0.010 |  |  |  |
| 1985 | 0.029 | 0.014 | 0.013 | 0.009 | 0.001 | 0.012 |  |  |  |
| 1986 | 0.033 | 0.013 | 0.014 | 0.006 |  | 0.010 |  |  |  |
| 1987 | 0.037 | 0.008 | 0.014 | 0.015 |  | 0.016 |  |  |  |
| 1988 | 0.076 | 0.014 | 0.033 |  |  | 0.001 |  |  |  |
| 1989 | 0.082 | 0.010 | 0.015 | 0.003 | 0.003 | 0.004 |  |  |  |
| 1990 | 0.072 | 0.008 | 0.009 | 0.012 | 0.011 | 0.014 |  | 0.001 |  |
| 1991 | 0.090 | 0.013 | 0.006 | 0.028 | 0.018 | 0.030 |  | 0.001 | 0.002 |
| 1992 | 0.060 | 0.011 | 0.007 | 0.067 | 0.024 | 0.007 | 0.004 | 0.006 |  |
| 1993 | 0.072 | 0.005 | 0.012 | 0.066 |  |  | 0.003 |  |  |
| 1994 | 0.062 | 0.005 | 0.022 | 0.069 |  | 0.021 | 0.006 |  | 0.003 |
| 1995 | 0.090 | 0.008 | 0.015 | 0.079 |  | 0.031 | 0.015 | 0.004 | 0.003 |
| 1996 | 0.080 | 0.003 | 0.014 | 0.073 |  | 0.029 | 0.023 | 0.008 | 0.003 |
| 1997 | 0.088 | 0.004 | 0.012 | 0.087 |  | 0.055 | 0.019 | 0.005 | 0.004 |
| 1998 | 0.056 | 0.007 | 0.016 | 0.079 |  | 0.040 | 0.014 | 0.004 | 0.002 |
| 1999 | 0.059 | 0.004 | 0.006 | 0.088 | 0.003 | 0.033 | 0.016 | 0.002 | 0.003 |
| 2000 | 0.057 | 0.005 | 0.007 | 0.059 | 0.002 | 0.049 | 0.026 |  |  |
| 2001 | 0.027 | 0.007 | 0.008 | 0.051 | 0.008 | 0.052 | 0.017 | 0.001 | 0.005 |
| 2002 | 0.042 | 0.009 | 0.006 | 0.064 | 0.013 | 0.042 | 0.015 | 0.001 | 0.004 |
| 2003 | 0.037 | 0.009 | 0.004 | 0.063 | 0.011 | 0.044 | 0.020 |  | 0.006 |
| 2004 | 0.043 | 0.011 | 0.004 | 0.063 | 0.010 | 0.044 | 0.015 | 0.001 | 0.004 |
| 2005 | 0.047 | 0.007 | 0.001 | 0.064 | 0.013 | 0.052 | 0.015 |  | 0.005 |
| 2006 | 0.053 | 0.007 | 0.001 | 0.069 | 0.017 | 0.037 | 0.020 |  | 0.005 |
| 2007 | 0.046 | 0.011 | 0.002 | 0.068 | 0.016 | 0.042 | 0.019 |  | 0.006 |

Table 2.20—Schedules of Pacific cod selectivities at length in the commercial fisheries as defined by final parameter estimates (Model 1). Lengths (cm) correspond to mid-points of size bins. Note that trawl survey selectivities are age-based rather than length-based.

|  | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Len. | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | IPHC |
| 10.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 13.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 16.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 19.5 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 22.5 | 0.01 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 25.5 | 0.02 | 0.03 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 28.5 | 0.03 | 0.05 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 31.5 | 0.04 | 0.07 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 34.5 | 0.07 | 0.11 | 0.03 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 37.5 | 0.10 | 0.16 | 0.04 | 0.02 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| 40.5 | 0.14 | 0.22 | 0.06 | 0.05 | 0.03 | 0.03 | 0.01 | 0.01 | 0.01 | 0.00 |
| 43.5 | 0.20 | 0.30 | 0.09 | 0.09 | 0.06 | 0.07 | 0.02 | 0.03 | 0.03 | 0.00 |
| 47.5 | 0.29 | 0.42 | 0.14 | 0.21 | 0.16 | 0.16 | 0.07 | 0.09 | 0.09 | 0.01 |
| 52.5 | 0.44 | 0.59 | 0.22 | 0.44 | 0.38 | 0.38 | 0.22 | 0.27 | 0.31 | 0.07 |
| 57.5 | 0.61 | 0.77 | 0.33 | 0.74 | 0.70 | 0.68 | 0.50 | 0.57 | 0.68 | 0.28 |
| 62.5 | 0.78 | 0.91 | 0.47 | 0.96 | 0.95 | 0.94 | 0.83 | 0.88 | 0.98 | 0.70 |
| 67.5 | 0.91 | 0.99 | 0.62 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.99 |
| 72.5 | 0.99 | 1.00 | 0.77 | 0.93 | 1.00 | 1.00 | 0.99 | 0.96 | 1.00 | 0.69 |
| 77.5 | 1.00 | 1.00 | 0.90 | 0.73 | 1.00 | 0.97 | 0.79 | 0.76 | 1.00 | 0.50 |
| 82.5 | 0.92 | 1.00 | 0.98 | 0.53 | 1.00 | 0.86 | 0.57 | 0.61 | 1.00 | 0.50 |
| 87.5 | 0.79 | 1.00 | 1.00 | 0.40 | 1.00 | 0.77 | 0.48 | 0.57 | 1.00 | 0.50 |
| 92.5 | 0.72 | 1.00 | 1.00 | 0.35 | 1.00 | 0.72 | 0.46 | 0.56 | 1.00 | 0.50 |
| 97.5 | 0.71 | 1.00 | 1.00 | 0.33 | 1.00 | 0.70 | 0.46 | 0.56 | 1.00 | 0.50 |
| 102.5 | 0.71 | 1.00 | 1.00 | 0.33 | 1.00 | 0.70 | 0.46 | 0.56 | 1.00 | 0.50 |
| 107.5 | 0.71 | 1.00 | 1.00 | 0.33 | 1.00 | 0.70 | 0.46 | 0.56 | 1.00 | 0.50 |


|  | Trawl survey |  |
| ---: | ---: | ---: |
| Age | Pre82 | Post81 |
| 0 | 0.00 | 0.00 |
| 1 | 0.10 | 0.40 |
| 2 | 1.00 | 1.00 |
| 3 | 1.00 | 1.00 |
| 4 | 0.73 | 1.00 |
| 5 | 0.13 | 1.00 |
| 6 | 0.13 | 0.96 |
| 7 | 0.13 | 0.73 |
| 8 | 0.13 | 0.44 |
| 9 | 0.13 | 0.24 |
| 10 | 0.13 | 0.14 |
| 11 | 0.13 | 0.11 |
| 12 | 0.13 | 0.11 |
| 13 | 0.13 | 0.10 |
| 14 | 0.13 | 0.10 |
| 15 | 0.13 | 0.10 |
| 16 | 0.13 | 0.10 |
| 17 | 0.13 | 0.10 |
| 18 | 0.13 | 0.10 |
| 19 | 0.13 | 0.10 |
| 20 | 0.13 | 0.10 |

Table 2.21—Schedules of Pacific cod length (cm) by season and age as estimated by Model 1. Sea1 = Jan-Jun, Sea2 = Jul-Aug, Sea3 = Sep-Dec.

|  | Population |  |  | Trawl fishery |  |  | ongline fishery |  |  | Pot fishery |  |  | rawl survey |  | IPHC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | a1 | Sea2 | Sea3 | eal | Sea2 | S | Sea1 | Sea2 | Sea3 | 1 | Sea | Sea3 | Pre82 | Post81 |  |
|  | 5 | 7. | 7.3 | 13.84 | 13.7 | 12 | 0 | 10 | 10. | 10 | 10.82 | 10.72 | 10.82 | 10.82 | 82 |
|  | 11.03 | 17 | 22 | 15.67 | 20 | 25.25 | 12.28 | 18.44 | 27 | 12 | 17.48 | 23 | 17.16 | 17.16 | , |
|  | 27.97 | 32.83 | 36.79 | 20 | 35.93 | 40.32 | 34.60 | 40.31 | 43. | 35.60 | 41.16 | 45. | 2.83 |  |  |
|  |  |  |  |  |  |  |  | 51.24 |  | 49.23 |  |  |  |  |  |
|  | 52.33 | 55 | 5 | 55.06 | 57 | 61.08 | 56.16 | 58 | 60.85 | 5 | 59.65 | 61.21 | 55.45 | 55.45 |  |
|  | 60.97 | 63.44 | 65.4 | 62.94 | 64.55 | 68.09 | 62.69 | 65.07 | 66.6 | 63.89 | 65.30 | 66 | 63.44 | 63.44 | 66.04 |
|  | 67.87 | 69 | 71 |  | 70.38 | 73.5 |  | 70.50 | 71 | 68.66 | 69.91 | 71 | 69.84 |  |  |
|  | 73.38 |  | 76.2 |  | 7.21 |  | 71.9 | 75.2 | 75.80 | 72.6 | . |  | 74.96 |  |  |
|  | 77.78 | 79.0 | 80 | 77.52 | 79.1 | 81 | 75 | 79.1 | 79.27 | 76.05 | 77.55 |  | 4 |  |  |
|  | 81.30 | 82.31 | 83.1 | 80.63 | 82.37 | 84.03 | 46 | 82.36 | 82.15 | 79. | 0.74 | 83 | . 30 | 82.30 |  |
|  | 84 | 84.91 | 85.5 | 83.2 | 84.94 | 86.26 | 81.06 | 84.93 | 84.52 | 81.8 | 83. | 85 | 90 | 84.90 |  |
|  | 86.35 | 87.0 | 87 | 85.35 | 86.99 | 88. |  | 86.99 | 6. | 84.20 | 85.6 | 87. | 86.97 | 86.97 |  |
|  | 88.15 | 88.6 |  |  | 88.62 | 89.4 |  | 88.62 |  | 86.13 | 87.46 | 89.0 | 88.61 | 88.61 |  |
|  | 89 | 99 | 90 |  | .9 | 90.6 | 86.67 | 89.9 |  | 87.69 | 8.8 | 90.2 | 89.91 | 89.91 |  |
|  | 90.72 | 91.0 | 91.32 | 89 | 90.9 | 91 |  | 90.94 | 90.2 | 88.9 | 90.01 | 91. | 90.93 | 90 |  |
|  | 91.63 | 91.90 | 92.1 | 90.55 | 91.75 |  | 88 | 91.75 | 91. | 89.94 | 90.90 |  | 91.75 |  |  |
|  | 92 |  |  |  | 2.3 |  |  | 92.3 | 1.71 | 0.73 |  |  |  |  |  |
|  | 92.95 | 93.1 | 93 |  | 92.9 | , | 9 | 92.90 | 92.2 | 91.3 | 92.1 | 93. | 22.89 | 92. |  |
| 18 | 93.4 | 93.55 | - |  | 93.30 |  | 90 | 93.30 |  | 91.85 | 2.61 | 93.4 | 93. |  |  |
| 19 | 93.78 | 93.89 | 93.9 | 仡 68 | 93.62 | 93.9 | 91.26 | 93.62 | , | 92.25 | 2.95 | 93.70 | 3.62 | , | 93.38 |
| 20 | 94 | 94.1 | 94.24 | 92.97 | 93.88 | 94.16 | 91.58 | .8 | 93.18 | 2.56 | 3.23 | 93.9 | 3.87 | 93. |  |

Table 2.22—Schedules of Pacific cod weight (kg) by season and age as estimated by Model 1. Sea1 = Jan-Jun, Sea2 = Jul-Aug, Sea3 = Sep-Dec.

|  | Population |  |  | Trawl fishery |  |  | Longline fishery |  |  | Pot fishery |  |  | Trawl survey |  | IPHC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Sea1 | Sea2 | Sea3 | Sea1 | Sea2 | Sea3 | Sea1 | Sea2 | Sea3 | Sea1 | Sea2 | Sea3 | Pre82 | Post81 |  |
| 0 | 0.01 | 0.01 | 0.01 | 0.02 | 0.02 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 |
| 1 | 0.02 | 0.05 | 0.11 | 0.03 | 0.08 | 0.16 | 0.02 | 0.07 | 0.22 | 0.02 | 0.05 | 0.14 | 0.05 | 0.05 | 0.05 |
| 2 | 0.23 | 0.38 | 0.54 | 0.32 | 0.50 | 0.72 | 0.43 | 0.71 | 0.92 | 0.48 | 0.76 | 1.07 | 0.38 | 0.38 | 0.88 |
| 3 | 0.80 | 1.07 | 1.32 | 1.01 | 1.27 | 1.63 | 1.19 | 1.53 | 1.78 | 1.34 | 1.61 | 1.90 | 1.07 | 1.07 | 1.96 |
| 4 | 1.68 | 2.02 | 2.33 | 1.95 | 2.23 | 2.73 | 2.07 | 2.41 | 2.68 | 2.25 | 2.50 | 2.72 | 2.02 | 2.02 | 2.78 |
| 5 | 2.74 | 3.12 | 3.45 | 3.01 | 3.27 | 3.89 | 2.96 | 3.34 | 3.62 | 3.13 | 3.37 | 3.64 | 3.12 | 3.12 | 3.47 |
| 6 | 3.88 | 4.25 | 4.58 | 4.06 | 4.34 | 4.99 | 3.82 | 4.36 | 4.59 | 3.97 | 4.22 | 4.66 | 4.25 | 4.25 | 4.19 |
| 7 | 4.99 | 5.34 | 5.65 | 5.04 | 5.39 | 6.00 | 4.65 | 5.39 | 5.52 | 4.77 | 5.08 | 5.68 | 5.34 | 5.34 | 5.06 |
| 8 | 6.02 | 6.35 | 6.62 | 5.93 | 6.37 | 6.91 | 5.43 | 6.37 | 6.39 | 5.57 | 5.96 | 6.63 | 6.35 | 6.35 | 6.01 |
| 9 | 6.95 | 7.23 | 7.47 | 6.75 | 7.25 | 7.70 | 6.19 | 7.24 | 7.19 | 6.37 | 6.81 | 7.48 | 7.23 | 7.23 | 6.94 |
| 10 | 7.76 | 8.00 | 8.20 | 7.48 | 8.01 | 8.39 | 6.90 | 8.00 | 7.89 | 7.13 | 7.59 | 8.20 | 8.00 | 8.00 | 7.76 |
| 11 | 8.44 | 8.64 | 8.81 | 8.13 | 8.65 | 8.97 | 7.55 | 8.65 | 8.49 | 7.83 | 8.27 | 8.81 | 8.64 | 8.64 | 8.45 |
| 12 | 9.01 | 9.18 | 9.32 | 8.69 | 9.18 | 9.45 | 8.12 | 9.18 | 9.00 | 8.43 | 8.84 | 9.32 | 9.18 | 9.18 | 9.02 |
| 13 | 9.48 | 9.62 | 9.73 | 9.16 | 9.62 | 9.84 | 8.60 | 9.62 | 9.42 | 8.93 | 9.31 | 9.73 | 9.62 | 9.62 | 9.49 |
| 14 | 9.86 | 9.97 | 10.06 | 9.54 | 9.98 | 10.17 | 9.01 | 9.97 | 9.77 | 9.35 | 9.69 | 10.06 | 9.97 | 9.97 | 9.86 |
| 15 | 10.17 | 10.26 | 10.33 | 9.86 | 10.26 | 10.42 | 9.34 | 10.26 | 10.04 | 9.68 | 10.00 | 10.33 | 10.26 | 10.26 | 10.16 |
| 16 | 10.42 | 10.49 | 10.55 | 10.11 | 10.49 | 10.63 | 9.62 | 10.49 | 10.27 | 9.96 | 10.25 | 10.55 | 10.49 | 10.49 | 10.40 |
| 17 | 10.62 | 10.67 | 10.72 | 10.32 | 10.67 | 10.80 | 9.84 | 10.67 | 10.45 | 10.18 | 10.45 | 10.72 | 10.67 | 10.67 | 10.59 |
| 18 | 10.78 | 10.82 | 10.86 | 10.48 | 10.82 | 10.93 | 10.01 | 10.82 | 10.59 | 10.35 | 10.60 | 10.86 | 10.82 | 10.82 | 10.75 |
| 19 | 10.90 | 10.94 | 10.97 | 10.62 | 10.94 | 11.04 | 10.16 | 10.94 | 10.71 | 10.49 | 10.73 | 10.97 | 10.94 | 10.94 | 10.87 |
| 20 | 11.00 | 11.03 | 11.05 | 10.72 | 11.03 | 11.12 | 10.27 | 11.03 | 10.80 | 10.61 | 10.83 | 11.06 | 11.03 | 11.03 | 10.97 |

Table 2.23—Time series of EBS (not expanded to BSAI) Pacific cod female spawning biomass (t) for the years 1977-2006 as estimated last year under the Plan Team's and SSC's preferred model and this year under Model 1, 1977-2007 The columns labeled "L95\%CI" and "U95\%CI" represent the lower and upper bounds of the 95\% confidence interval.

|  | Last Year's Values |  |  | This Year's Values |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Sp. Bio. | L95\%CI | U95\%CI | Sp. Bio. | L95\%CI | U95\%CI |
| 1977 | 56,590 | 39,103 | 74,077 | 127,575 | 47,260 | 207,890 |
| 1978 | 78,325 | 57,381 | 99,269 | 155,625 | 68,818 | 242,432 |
| 1979 | 114,795 | 87,281 | 142,309 | 202,290 | 101,556 | 303,024 |
| 1980 | 181,760 | 144,358 | 219,162 | 283,920 | 162,116 | 405,724 |
| 1981 | 290,795 | 239,529 | 342,061 | 425,240 | 274,389 | 576,091 |
| 1982 | 424,045 | 357,133 | 490,957 | 625,900 | 438,524 | 813,276 |
| 1983 | 544,850 | 465,244 | 624,456 | 819,800 | 603,485 | $1,036,115$ |
| 1984 | 613,850 | 527,493 | 700,207 | 924,700 | 701,054 | $1,148,346$ |
| 1985 | 630,500 | 542,936 | 718,064 | 932,450 | 720,084 | $1,144,816$ |
| 1986 | 622,950 | 537,388 | 708,512 | 888,200 | 694,836 | $1,081,564$ |
| 1987 | 619,300 | 537,036 | 701,564 | 849,750 | 674,585 | $1,024,915$ |
| 1988 | 607,300 | 529,182 | 685,418 | 815,200 | 655,774 | 974,626 |
| 1989 | 564,850 | 491,797 | 637,903 | 760,400 | 615,438 | 905,362 |
| 1990 | 516,550 | 449,321 | 583,779 | 702,300 | 571,901 | 832,699 |
| 1991 | 454,815 | 394,277 | 515,353 | 620,050 | 505,419 | 734,681 |
| 1992 | 378,065 | 324,193 | 431,937 | 512,550 | 413,021 | 612,079 |
| 1993 | 344,165 | 295,331 | 392,999 | 451,425 | 362,648 | 540,202 |
| 1994 | 351,985 | 306,049 | 397,921 | 445,295 | 361,897 | 528,693 |
| 1995 | 360,540 | 315,910 | 405,170 | 448,745 | 367,448 | 530,042 |
| 1996 | 350,860 | 306,281 | 395,439 | 432,445 | 352,105 | 512,785 |
| 1997 | 343,040 | 297,689 | 388,391 | 416,680 | 337,350 | 496,010 |
| 1998 | 314,645 | 268,605 | 360,685 | 375,345 | 298,165 | 452,525 |
| 1999 | 308,685 | 261,600 | 355,770 | 351,455 | 276,238 | 426,672 |
| 2000 | 319,535 | 270,639 | 368,431 | 345,010 | 269,609 | 420,411 |
| 2001 | 342,440 | 291,318 | 393,562 | 358,985 | 280,568 | 437,402 |
| 2002 | 366,965 | 314,358 | 419,572 | 384,300 | 301,471 | 467,129 |
| 2003 | 376,425 | 323,431 | 429,419 | 404,030 | 316,403 | 491,657 |
| 2004 | 376,585 | 323,995 | 429,175 | 424,690 | 331,946 | 517,434 |
| 2005 | 360,260 | 308,790 | 411,730 | 429,370 | 332,850 | 525,890 |
| 2006 | 326,400 | 276,697 | 376,103 | 407,020 | 309,816 | 504,224 |
| 2007 | $n / a$ | $n / a$ | $n / a$ | 369,640 | 274,504 | 464,776 |

Table 2.24—Time series of EBS (not expanded to BSAI) Pacific cod age 0 recruitment (1000s of fish) as estimated last year under the Plan Team's and SSC's preferred model and this year under Model 1, 19772006. The columns labeled "L95\%CI" and "U95\%CI" represent the lower and upper bounds of the 95\% confidence interval for each cohort.

|  | Last Year's Values |  |  | This Year's Values |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Recruits | L95\%CI | U95\%CI | Recruits | L95\%CI | U95\%CI |
| 1977 | $1,611,960$ | $1,292,760$ | $2,009,960$ | $2,533,200$ | $1,869,348$ | $3,197,052$ |
| 1978 | $1,014,290$ | 755,490 | $1,361,690$ | 787,020 | 486,160 | $1,087,880$ |
| 1979 | 947,821 | 723,421 | $1,241,821$ | $1,042,300$ | 796,849 | $1,287,751$ |
| 1980 | 453,442 | 302,942 | 678,742 | 411,490 | 283,294 | 539,686 |
| 1981 | 886,610 | 704,310 | $1,116,110$ | 211,920 | 128,575 | 295,265 |
| 1982 | $1,508,730$ | $1,280,230$ | $1,778,030$ | $1,354,400$ | $1,149,286$ | $1,559,514$ |
| 1983 | 383,242 | 263,542 | 557,342 | 248,630 | 167,343 | 329,917 |
| 1984 | $1,210,830$ | $1,030,230$ | $1,423,130$ | $1,045,200$ | 892,093 | $1,198,307$ |
| 1985 | 418,040 | 315,040 | 554,740 | 553,370 | 447,936 | 658,804 |
| 1986 | 392,177 | 299,587 | 513,377 | 284,700 | 218,781 | 350,619 |
| 1987 | 313,653 | 227,433 | 432,553 | 158,470 | 109,815 | 207,125 |
| 1988 | 906,898 | 766,798 | $1,072,598$ | 392,570 | 320,744 | 464,396 |
| 1989 | $1,139,520$ | 975,220 | $1,331,520$ | 801,750 | 682,919 | 920,581 |
| 1990 | 658,085 | 534,385 | 810,485 | 731,990 | 631,722 | 832,258 |
| 1991 | 926,882 | 787,582 | $1,090,882$ | 383,500 | 317,436 | 449,564 |
| 1992 | $1,000,980$ | 858,580 | $1,166,980$ | 883,170 | 770,819 | 995,521 |
| 1993 | 373,064 | 285,674 | 487,164 | 325,880 | 272,607 | 379,153 |
| 1994 | 565,069 | 460,369 | 693,569 | 323,260 | 271,453 | 375,067 |
| 1995 | 985,921 | 844,021 | $1,151,721$ | 359,010 | 295,639 | 422,381 |
| 1996 | $1,106,130$ | 960,530 | $1,273,830$ | 892,480 | 749,316 | $1,03,644$ |
| 1997 | 600,909 | 500,609 | 721,309 | 449,700 | 378,282 | 521,118 |
| 1998 | 830,782 | 710,482 | 971,382 | 497,090 | 416,483 | 579,497 |
| 1999 | $1,023,880$ | 890,480 | $1,177,280$ | 904,670 | 761,386 | $1,047,954$ |
| 2000 | 528,671 | 442,611 | 631,471 | 623,330 | 528,149 | 719,311 |
| 2001 | 462,633 | 381,223 | 561,433 | 325,960 | 266,200 | 385,720 |
| 2002 | 429,282 | 342,522 | 537,982 | 398,510 | 325,433 | 471,587 |
| 2003 | 394,653 | 298,673 | 521,453 | 350,100 | 277,070 | 423,130 |
| 2004 | 303,430 | 193,130 | 476,630 | 203,560 | 154,074 | 253,046 |
| 2005 | 675,083 | 448,783 | $1,015,383$ | 393,800 | 273,891 | 513,709 |
| 2006 | $n / a$ | $n / a$ | $n / a$ | $1,835,100$ | $1,013,252$ | $2,656,948$ |

Table 2.25-Time series of EBS Pacific cod catch divided by age 3+ biomass as estimated last year under the Plan Team's and SSC’s preferred model and this year under Model 1, 1977-2007. The last entry in each column is based on partial catches for the respective year, because the year was/is still in progress at the time of the assessment.

| Year | Last Year’s Values | This Year’s Values |
| :--- | ---: | ---: |
| 1977 | 0.11 | 0.07 |
| 1978 | 0.12 | 0.08 |
| 1979 | 0.05 | 0.05 |
| 1980 | 0.04 | 0.03 |
| 1981 | 0.04 | 0.03 |
| 1982 | 0.04 | 0.03 |
| 1983 | 0.05 | 0.04 |
| 1984 | 0.07 | 0.05 |
| 1985 | 0.08 | 0.06 |
| 1986 | 0.07 | 0.06 |
| 1987 | 0.08 | 0.06 |
| 1988 | 0.12 | 0.09 |
| 1989 | 0.11 | 0.09 |
| 1990 | 0.13 | 0.10 |
| 1991 | 0.17 | 0.14 |
| 1992 | 0.14 | 0.12 |
| 1993 | 0.11 | 0.10 |
| 1994 | 0.14 | 0.13 |
| 1995 | 0.18 | 0.16 |
| 1996 | 0.18 | 0.16 |
| 1997 | 0.21 | 0.19 |
| 1998 | 0.15 | 0.15 |
| 1999 | 0.13 | 0.13 |
| 2000 | 0.13 | 0.14 |
| 2001 | 0.12 | 0.12 |
| 2002 | 0.13 | 0.13 |
| 2003 | 0.14 | 0.13 |
| 2004 | 0.16 | 0.14 |
| 2005 | 0.17 | 0.15 |
| 2006 | 0.19 | 0.15 |
| 2007 | n/a | 0.15 |

Table 2.26—Projections for BSAI Pacific cod catch (t), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in 2008-2020 (Scenarios 1 and 2), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 150,000 | 150,000 | 150,000 | 150,000 | 0 |
| 2009 | 161,000 | 161,000 | 162,000 | 162,000 | 111 |
| 2010 | 257,000 | 258,000 | 259,000 | 264,000 | 2,792 |
| 2011 | 350,000 | 367,000 | 369,000 | 397,000 | 16,197 |
| 2012 | 325,000 | 365,000 | 377,000 | 468,000 | 50,648 |
| 2013 | 255,000 | 347,000 | 362,000 | 507,000 | 83,911 |
| 2014 | 210,000 | 334,000 | 346,000 | 525,000 | 102,413 |
| 2015 | 182,000 | 327,000 | 336,000 | 525,000 | 107,342 |
| 2016 | 169,000 | 321,000 | 328,000 | 514,000 | 106,534 |
| 2017 | 161,000 | 320,000 | 324,000 | 507,000 | 105,837 |
| 2018 | 166,000 | 321,000 | 323,000 | 512,000 | 106,547 |
| 2019 | 167,000 | 316,000 | 324,000 | 516,000 | 107,519 |
| 2020 | 167,000 | 319,000 | 325,000 | 522,000 | 110,055 |
| Biomass projections: |  |  |  |  |  |


| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 398,000 | 398,000 | 398,000 | 398,000 | 0 |
| 2009 | 394,000 | 395,000 | 395,000 | 395,000 | 263 |
| 2010 | 452,000 | 454,000 | 455,000 | 462,000 | 3,589 |
| 2011 | 529,000 | 542,000 | 547,000 | 579,000 | 17,567 |
| 2012 | 544,000 | 584,000 | 596,000 | 687,000 | 50,533 |
| 2013 | 501,000 | 583,000 | 603,000 | 770,000 | 93,272 |
| 2014 | 457,000 | 563,000 | 594,000 | 820,000 | 123,433 |
| 2015 | 427,000 | 555,000 | 585,000 | 834,000 | 134,722 |
| 2016 | 406,000 | 548,000 | 577,000 | 827,000 | 134,494 |
| 2017 | 397,000 | 542,000 | 570,000 | 830,000 | 131,220 |
| 2018 | 396,000 | 542,000 | 567,000 | 816,000 | 130,130 |
| 2019 | 400,000 | 541,000 | 567,000 | 816,000 | 131,630 |
| 2020 | 403,000 | 538,000 | 568,000 | 823,000 | 134,414 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.22 | 0.22 | 0.22 | 0.22 | 0.00 |
| 2009 | 0.22 | 0.22 | 0.22 | 0.22 | 0.00 |
| 2010 | 0.25 | 0.26 | 0.26 | 0.26 | 0.00 |
| 2011 | 0.30 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2012 | 0.31 | 0.31 | 0.31 | 0.31 | 0.00 |
| 2013 | 0.28 | 0.31 | 0.30 | 0.31 | 0.01 |
| 2014 | 0.26 | 0.31 | 0.30 | 0.31 | 0.02 |
| 2015 | 0.24 | 0.31 | 0.29 | 0.31 | 0.02 |
| 2016 | 0.23 | 0.31 | 0.29 | 0.31 | 0.03 |
| 2017 | 0.22 | 0.31 | 0.29 | 0.31 | 0.03 |
| 2018 | 0.22 | 0.31 | 0.29 | 0.31 | 0.03 |
| 2019 | 0.22 | 0.31 | 0.29 | 0.31 | 0.03 |
| 2020 | 0.22 | 0.31 | 0.29 | 0.31 | 0.03 |

Table 2.27—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=1 / 2 \max F_{A B C}$ in 2007-2019 (Scenario 3), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 136,000 | 136,000 | 136,000 | 136,000 | 0 |
| 2009 | 149,000 | 149,000 | 149,000 | 149,000 | 2 |
| 2010 | 208,000 | 208,000 | 209,000 | 210,000 | 612 |
| 2011 | 251,000 | 258,000 | 260,000 | 278,000 | 9,997 |
| 2012 | 244,000 | 271,000 | 279,000 | 342,000 | 34,831 |
| 2013 | 217,000 | 270,000 | 282,000 | 385,000 | 57,507 |
| 2014 | 199,000 | 266,000 | 281,000 | 409,000 | 69,408 |
| 2015 | 185,000 | 266,000 | 279,000 | 415,000 | 72,935 |
| 2016 | 176,000 | 263,000 | 275,000 | 408,000 | 72,658 |
| 2017 | 169,000 | 263,000 | 273,000 | 404,000 | 72,158 |
| 2018 | 170,000 | 262,000 | 272,000 | 410,000 | 72,815 |
| 2019 | 173,000 | 261,000 | 272,000 | 408,000 | 73,828 |
| 2020 | 175,000 | 259,000 | 273,000 | 424,000 | 75,780 |
| Biomass projections: |  |  |  |  |  |


| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 399,000 | 399,000 | 399,000 | 399,000 | 0 |
| 2009 | 401,000 | 401,000 | 401,000 | 402,000 | 271 |
| 2010 | 464,000 | 467,000 | 468,000 | 474,000 | 3,738 |
| 2011 | 561,000 | 575,000 | 579,000 | 613,000 | 18,421 |
| 2012 | 615,000 | 658,000 | 671,000 | 765,000 | 52,993 |
| 2013 | 601,000 | 694,000 | 715,000 | 896,000 | 101,334 |
| 2014 | 562,000 | 699,000 | 729,000 | 983,000 | 142,263 |
| 2015 | 528,000 | 697,000 | 731,000 | $1,020,000$ | 165,097 |
| 2016 | 498,000 | 698,000 | 728,000 | $1,040,000$ | 173,335 |
| 2017 | 478,000 | 694,000 | 722,000 | $1,060,000$ | 174,606 |
| 2018 | 468,000 | 697,000 | 717,000 | $1,040,000$ | 175,177 |
| 2019 | 470,000 | 694,000 | 715,000 | $1,040,000$ | 177,340 |
| 2020 | 468,000 | 687,000 | 716,000 | $1,040,000$ | 180,757 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2009 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2010 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2011 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2012 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2013 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2014 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2015 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2016 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2017 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2018 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2019 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |
| 2020 | 0.20 | 0.20 | 0.20 | 0.20 | 0.00 |

Table 2.28—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that the upper bound on $F_{A B C}$ is set at $F_{60 \%}$ (Scenario 4), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 105,000 | 105,000 | 105,000 | 105,000 | 0 |
| 2009 | 118,000 | 118,000 | 118,000 | 118,000 | 2 |
| 2010 | 165,000 | 165,000 | 165,000 | 166,000 | 462 |
| 2011 | 202,000 | 207,000 | 209,000 | 223,000 | 7,583 |
| 2012 | 201,000 | 223,000 | 229,000 | 277,000 | 26,816 |
| 2013 | 183,000 | 225,000 | 234,000 | 314,000 | 45,214 |
| 2014 | 170,000 | 223,000 | 235,000 | 339,000 | 55,620 |
| 2015 | 160,000 | 225,000 | 235,000 | 343,000 | 59,278 |
| 2016 | 152,000 | 223,000 | 233,000 | 343,000 | 59,562 |
| 2017 | 147,000 | 223,000 | 231,000 | 340,000 | 59,318 |
| 2018 | 146,000 | 222,000 | 231,000 | 345,000 | 59,837 |
| 2019 | 150,000 | 222,000 | 231,000 | 345,000 | 60,660 |
| 2020 | 150,000 | 220,000 | 232,000 | 356,000 | 62,237 |
| Biomass projections: |  |  |  |  |  |


| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 401,000 | 401,000 | 401,000 | 401,000 | 0 |
| 2009 | 416,000 | 416,000 | 416,000 | 417,000 | 271 |
| 2010 | 490,000 | 492,000 | 493,000 | 500,000 | 3,739 |
| 2011 | 601,000 | 615,000 | 619,000 | 653,000 | 18,465 |
| 2012 | 673,000 | 717,000 | 729,000 | 825,000 | 53,657 |
| 2013 | 674,000 | 769,000 | 791,000 | 978,000 | 104,797 |
| 2014 | 642,000 | 786,000 | 818,000 | $1,090,000$ | 150,814 |
| 2015 | 609,000 | 791,000 | 828,000 | $1,150,000$ | 178,821 |
| 2016 | 579,000 | 797,000 | 829,000 | $1,170,000$ | 190,716 |
| 2017 | 557,000 | 793,000 | 826,000 | $1,190,000$ | 193,876 |
| 2018 | 546,000 | 798,000 | 822,000 | $1,180,000$ | 195,088 |
| 2019 | 548,000 | 797,000 | 821,000 | $1,190,000$ | 197,504 |
| 2020 | 548,000 | 793,000 | 822,000 | $1,180,000$ | 201,226 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2009 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2010 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2011 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2012 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2013 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2014 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2015 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2016 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2017 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2018 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2019 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |
| 2020 | 0.15 | 0.15 | 0.15 | 0.15 | 0.00 |

Table 2.29—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=0$ in 2007-2019 (Scenario 5), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 0 | 0 |
| 2020 | 0 | 0 | 0 | 0 | 0 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 409,000 | 409,000 | 409,000 | 409,000 | 0 |
| 2009 | 467,000 | 467,000 | 467,000 | 468,000 | 271 |
| 2010 | 582,000 | 585,000 | 586,000 | 592,000 | 3,741 |
| 2011 | 750,000 | 764,000 | 768,000 | 802,000 | 18,601 |
| 2012 | 900,000 | 945,000 | 959,000 | $1,060,000$ | 55,758 |
| 2013 | 975,000 | $1,080,000$ | $1,100,000$ | $1,310,000$ | 116,438 |
| 2014 | 988,000 | $1,160,000$ | $1,200,000$ | $1,520,000$ | 181,814 |
| 2015 | 984,000 | $1,210,000$ | $1,260,000$ | $1,670,000$ | 232,320 |
| 2016 | 966,000 | $1,250,000$ | $1,300,000$ | $1,760,000$ | 262,961 |
| 2017 | 938,000 | $1,280,000$ | $1,320,000$ | $1,840,000$ | 278,436 |
| 2018 | 924,000 | $1,290,000$ | $1,330,000$ | $1,870,000$ | 286,143 |
| 2019 | 934,000 | $1,310,000$ | $1,340,000$ | $1,870,000$ | 291,846 |
| 2020 | 943,000 | $1,320,000$ | $1,350,000$ | $1,890,000$ | 297,867 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2009 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2010 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2011 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2012 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2013 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2014 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2015 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2016 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2017 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2018 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2019 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 2020 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 2.30—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=F_{\text {OFL }}$ in 2007-2019 (Scenario 6), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 176,000 | 176,000 | 176,000 | 176,000 | 0 |
| 2009 | 181,000 | 181,000 | 181,000 | 181,000 | 127 |
| 2010 | 284,000 | 286,000 | 287,000 | 292,000 | 3,178 |
| 2011 | 378,000 | 399,000 | 405,000 | 452,000 | 24,872 |
| 2012 | 341,000 | 407,000 | 415,000 | 523,000 | 62,322 |
| 2013 | 257,000 | 373,000 | 386,000 | 560,000 | 101,898 |
| 2014 | 212,000 | 345,000 | 364,000 | 568,000 | 119,982 |
| 2015 | 184,000 | 333,000 | 352,000 | 571,000 | 123,047 |
| 2016 | 172,000 | 327,000 | 344,000 | 555,000 | 121,118 |
| 2017 | 165,000 | 329,000 | 340,000 | 553,000 | 120,121 |
| 2018 | 169,000 | 327,000 | 340,000 | 554,000 | 121,095 |
| 2019 | 172,000 | 323,000 | 341,000 | 560,000 | 122,449 |
| 2020 | 171,000 | 327,000 | 343,000 | 565,000 | 125,284 |

## Biomass projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 396,000 | 396,000 | 396,000 | 396,000 | 0 |
| 2009 | 383,000 | 383,000 | 383,000 | 383,000 | 262 |
| 2010 | 434,000 | 436,000 | 437,000 | 444,000 | 3,569 |
| 2011 | 503,000 | 516,000 | 520,000 | 551,000 | 17,067 |
| 2012 | 508,000 | 545,000 | 556,000 | 640,000 | 47,349 |
| 2013 | 459,000 | 530,000 | 551,000 | 705,000 | 85,769 |
| 2014 | 418,000 | 509,000 | 537,000 | 743,000 | 109,629 |
| 2015 | 390,000 | 501,000 | 527,000 | 745,000 | 116,282 |
| 2016 | 375,000 | 494,000 | 520,000 | 736,000 | 114,049 |
| 2017 | 367,000 | 493,000 | 514,000 | 734,000 | 110,752 |
| 2018 | 368,000 | 494,000 | 512,000 | 732,000 | 110,494 |
| 2019 | 370,000 | 490,000 | 513,000 | 726,000 | 112,356 |
| 2020 | 370,000 | 493,000 | 514,000 | 738,000 | 114,860 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.26 | 0.26 | 0.26 | 0.26 | 0.00 |
| 2009 | 0.25 | 0.26 | 0.26 | 0.26 | 0.00 |
| 2010 | 0.29 | 0.29 | 0.29 | 0.30 | 0.00 |
| 2011 | 0.34 | 0.35 | 0.35 | 0.37 | 0.01 |
| 2012 | 0.34 | 0.37 | 0.36 | 0.37 | 0.01 |
| 2013 | 0.31 | 0.36 | 0.35 | 0.37 | 0.02 |
| 2014 | 0.28 | 0.35 | 0.34 | 0.37 | 0.03 |
| 2015 | 0.26 | 0.34 | 0.33 | 0.37 | 0.04 |
| 2016 | 0.25 | 0.33 | 0.33 | 0.37 | 0.04 |
| 2017 | 0.24 | 0.33 | 0.33 | 0.37 | 0.04 |
| 2018 | 0.24 | 0.34 | 0.32 | 0.37 | 0.04 |
| 2019 | 0.25 | 0.33 | 0.32 | 0.37 | 0.04 |
| 2020 | 0.25 | 0.33 | 0.32 | 0.37 | 0.04 |

Table 2.31—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in each year 2007-2008 and $F=F_{\text {OFL }}$ thereafter (Scenario 7), with random variability in future recruitment.

## Catch projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 2008 | 150,000 | 150,000 | 150,000 | 150,000 | 0 |
| 2009 | 161,000 | 161,000 | 162,000 | 162,000 | 111 |
| 2010 | 301,000 | 304,000 | 304,000 | 310,000 | 3,258 |
| 2011 | 389,000 | 409,000 | 415,000 | 456,000 | 23,444 |
| 2012 | 344,000 | 408,000 | 417,000 | 526,000 | 61,953 |
| 2013 | 257,000 | 374,000 | 387,000 | 561,000 | 102,040 |
| 2014 | 212,000 | 345,000 | 365,000 | 568,000 | 120,158 |
| 2015 | 184,000 | 333,000 | 352,000 | 571,000 | 123,146 |
| 2016 | 172,000 | 327,000 | 344,000 | 555,000 | 121,163 |
| 2017 | 165,000 | 329,000 | 340,000 | 553,000 | 120,140 |
| 2018 | 169,000 | 327,000 | 340,000 | 554,000 | 121,103 |
| 2019 | 172,000 | 323,000 | 341,000 | 560,000 | 122,452 |
| 2020 | 171,000 | 327,000 | 343,000 | 565,000 | 125,285 |
| Biomass projections: |  |  |  |  |  |


| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 398,000 | 398,000 | 398,000 | 398,000 | 0 |
| 2009 | 394,000 | 395,000 | 395,000 | 395,000 | 263 |
| 2010 | 449,000 | 452,000 | 453,000 | 459,000 | 3,569 |
| 2011 | 511,000 | 524,000 | 528,000 | 559,000 | 17,140 |
| 2012 | 511,000 | 548,000 | 560,000 | 645,000 | 47,884 |
| 2013 | 460,000 | 531,000 | 552,000 | 708,000 | 86,375 |
| 2014 | 418,000 | 509,000 | 538,000 | 744,000 | 110,059 |
| 2015 | 389,000 | 501,000 | 527,000 | 746,000 | 116,519 |
| 2016 | 375,000 | 494,000 | 520,000 | 737,000 | 114,159 |
| 2017 | 367,000 | 493,000 | 514,000 | 734,000 | 110,798 |
| 2018 | 368,000 | 494,000 | 512,000 | 732,000 | 110,512 |
| 2019 | 370,000 | 490,000 | 513,000 | 726,000 | 112,362 |
| 2020 | 370,000 | 493,000 | 514,000 | 738,000 | 114,862 |

Fishing mortality projections:

| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 2008 | 0.22 | 0.22 | 0.22 | 0.22 | 0.00 |
| 2009 | 0.22 | 0.22 | 0.22 | 0.22 | 0.00 |
| 2010 | 0.30 | 0.30 | 0.31 | 0.31 | 0.00 |
| 2011 | 0.35 | 0.36 | 0.36 | 0.37 | 0.01 |
| 2012 | 0.35 | 0.37 | 0.36 | 0.37 | 0.01 |
| 2013 | 0.31 | 0.36 | 0.35 | 0.37 | 0.02 |
| 2014 | 0.28 | 0.35 | 0.34 | 0.37 | 0.03 |
| 2015 | 0.26 | 0.34 | 0.33 | 0.37 | 0.04 |
| 2016 | 0.25 | 0.33 | 0.33 | 0.37 | 0.04 |
| 2017 | 0.24 | 0.33 | 0.33 | 0.37 | 0.04 |
| 2018 | 0.24 | 0.34 | 0.32 | 0.37 | 0.04 |
| 2019 | 0.25 | 0.33 | 0.32 | 0.37 | 0.04 |
| 2020 | 0.25 | 0.33 | 0.32 | 0.37 | 0.04 |

Table 2.32a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod trawl fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS Pacific cod trawl fishery |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 1508 | 1365 | 893 | 1280 | 749 | 925 | 0.22 | 0.26 | 0.20 | 0.23 | 0.12 | 0.12 |
| Skates | 678 | 676 | 946 | 981 | 583 | 1303 | 0.04 | 0.04 | 0.07 | 0.06 | 0.03 | 0.05 |
| Shark | 0 | 0 | 0 | 9 | 2 | 3 | 0.00 | 0.00 | 0.00 | 0.15 | 0.09 | 0.08 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 | 0.08 |
| Sleepershk | 8 | 33 | 4 | 0 | 12 | 10 | 0.03 | 0.10 | 0.01 | 0.00 | 0.02 | 0.01 |
| Octopus | 29 | 19 | 17 | 68 | 17 | 30 | 0.14 | 0.13 | 0.13 | 0.19 | 0.09 | 0.08 |
| Squid | 7 | 1 | 0 | 2 | 4 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 1 | 0 | 1 | 0 | 0 | 0 | 0.03 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.00 | 0.00 | 0.00 | 0.71 | 0.00 |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.03 | 0.00 | 0.00 | 0.01 | 0.00 |
| Sandfish | 0 | 0 | 3 | 0 | 0 | 1 | 0.27 | 0.08 | 0.91 | 0.02 | 0.05 | 0.36 |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.90 | 0.01 |
| Grenadier | 1 | 6 | 0 | 3 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 231 | 232 | 195 | 302 | 220 | 157 | 0.16 | 0.21 | 0.20 | 0.24 | 0.18 | 0.14 |
| Crabs | 10 | 6 | 5 | 8 | 3 | 6 | 0.03 | 0.03 | 0.05 | 0.06 | 0.02 | 0.04 |
| Starfish | 133 | 63 | 83 | 109 | 57 | 98 | 0.02 | 0.02 | 0.03 | 0.03 | 0.01 | 0.02 |
| Jellyfish | 948 | 213 | 416 | 413 | 112 | 93 | 0.11 | 0.03 | 0.06 | 0.04 | 0.03 | 0.05 |
| Invertunid | 1 | 9 | 3 | 11 | 1 | 51 | 0.00 | 0.02 | 0.02 | 0.01 | 0.00 | 0.05 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.10 | 0.09 | 0.01 | 0.06 | 0.00 | 0.00 |
| Sponge | 73 | 34 | 39 | 28 | 9 | 13 | 0.23 | 0.09 | 0.22 | 0.30 | 0.05 | 0.08 |
| Anemone | 14 | 5 | 18 | 10 | 6 | 9 | 0.08 | 0.05 | 0.11 | 0.03 | 0.03 | 0.03 |
| Tunicate | 6 | 10 | 0 | 67 | 5 | 1 | 0.00 | 0.01 | 0.00 | 0.06 | 0.00 | 0.00 |
| Benthinv | 25 | 18 | 11 | 23 | 6 | 12 | 0.04 | 0.03 | 0.05 | 0.06 | 0.01 | 0.03 |
| Snails |  | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| echinoderm | 13 | 4 | 13 | 13 | 20 | 14 | 0.31 | 0.20 | 0.54 | 0.33 | 0.50 | 0.46 |
| Coral | 0 | 0 | 0 | 4 | 0 | 0 | 0.02 | 0.01 | 0.04 | 0.37 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.07 | 0.03 | 0.01 | 0.00 | 0.01 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |

Table 2.32b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod trawl fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 14 | 4 | 9 | 0.01 | 0.00 | 0.01 |
| Birds | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 |
| Bivalves | 1 | 10 | 0 | 0.05 | 0.52 | 0.03 |
| Brittle star unidentified | 1 | 1 | 0 | 0.02 | 0.03 | 0.00 |
| Capelin |  | 0 |  |  | 0.02 |  |
| Corals Bryozoans | 1 | 1 | 0 | 0.28 | 0.25 | 0.06 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 62 | 27 | 1 | 0.27 | 0.30 | 0.02 |
| Eulachon |  | 0 | 0 |  | 0.00 | 0.00 |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 4 | 2 | 1 | 0.43 | 0.40 | 0.23 |
| Grenadier | 14 | 9 | 0 | 0.01 | 0.00 | 0.00 |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 5 | 3 | 1 | 0.04 | 0.05 | 0.01 |
| Invertebrate unidentified | 5 | 4 | 0 | 0.01 | 0.01 | 0.00 |
| Lanternfishes (myctophidae) |  | 0 |  |  | 0.07 |  |
| Large Sculpins | 547 | 1422 | 897 | 0.39 | 0.32 | 0.22 |
| Misc crabs | 7 | 3 | 2 | 0.13 | 0.09 | 0.07 |
| Misc crustaceans | 0 | 0 | 0 | 0.24 | 0.20 | 0.07 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 174 | 152 | 149 | 0.35 | 0.30 | 0.31 |
| Misc inverts (worms etc) | 0 | 0 | 0 | 0.07 | 0.02 | 0.00 |
| Octopus | 14 | 44 | 12 | 0.10 | 0.12 | 0.05 |
| Other osmerids | 0 | 0 |  | 0.01 | 0.09 |  |
| Other Sculpins | 854 | 95 | 58 | 0.22 | 0.18 | 0.12 |
| Pacific Sand lance | 0 | 0 | 0 | 0.45 | 0.40 | 0.59 |
| Pandalid shrimp | 0 | 0 | 0 | 0.15 | 0.18 | 0.01 |
| Polychaete unidentified |  | 0 | 0 |  | 0.01 | 0.08 |
| Scypho jellies | 727 | 699 | 391 | 0.11 | 0.10 | 0.06 |
| Sea anemone unidentified | 14 | 16 | 12 | 0.10 | 0.09 | 0.12 |
| Sea pens whips | 0 | 1 | 0 | 0.01 | 0.05 | 0.01 |
| Sea star | 118 | 91 | 81 | 0.03 | 0.03 | 0.03 |
| Shark | 10 | 29 | 11 | 0.03 | 0.08 | 0.05 |
| Skate | 1010 | 1355 | 570 | 0.06 | 0.07 | 0.03 |
| Snails | 14 | 13 | 3 | 0.07 | 0.05 | 0.02 |
| Sponge unidentified | 3 | 7 | 3 | 0.01 | 0.08 | 0.04 |
| Squid | 5 | 4 | 1 | 0.00 | 0.00 | 0.00 |
| Stichaeidae | 0 | 0 | 0 | 0.12 | 0.07 | 0.14 |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 11 | 10 | 12 | 0.36 | 0.43 | 0.48 |

Table 2.33a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod longline fishery, 1997-2002. The first part of the table ("Bycatch in...") shows the amount (t) of each species group taken as bycatch in the EBS Pacific cod longline fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS P. cod longline fishery |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 706 | 931 | 821 | 801 | 1142 | 1383 | 0.11 | 0.18 | 0.18 | 0.14 | 0.19 | 0.18 |
| Skates | 12961 | 12808 | 9178 | 11578 | 11932 | 17507 | 0.77 | 0.70 | 0.69 | 0.68 | 0.66 | 0.66 |
| Shark | 27 | 48 | 18 | 47 | 17 | 22 | 0.50 | 0.40 | 0.11 | 0.78 | 0.70 | 0.48 |
| Salmonshk | 0 | 1 | 1 | 0 | 1 | 10 | 0.00 | 0.05 | 0.04 | 0.01 | 0.05 | 0.22 |
| Dogfish | 4 | 5 | 5 | 8 | 11 | 8 | 1.00 | 0.90 | 0.99 | 0.98 | 0.83 | 0.92 |
| Sleepershk | 67 | 114 | 99 | 114 | 240 | 250 | 0.24 | 0.34 | 0.35 | 0.33 | 0.37 | 0.30 |
| Octopus | 15 | 15 | 13 | 29 | 15 | 76 | 0.07 | 0.10 | 0.10 | 0.08 | 0.08 | 0.19 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.60 | 0.00 | 0.80 | 0.00 | 0.00 |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.56 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 437 | 604 | 356 | 364 | 162 | 336 | 0.15 | 0.12 | 0.08 | 0.09 | 0.07 | 0.06 |
| Otherfish | 43 | 27 | 38 | 38 | 71 | 122 | 0.03 | 0.03 | 0.04 | 0.03 | 0.06 | 0.11 |
| Crabs | 1 | 0 | 0 | 1 | 1 | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 |
| Starfish | 136 | 141 | 250 | 132 | 319 | 384 | 0.02 | 0.04 | 0.08 | 0.04 | 0.08 | 0.08 |
| Jellyfish | 5 | 7 | 24 | 2 | 2 | 5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 10 | 12 | 1 | 6 | 10 | 11 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 |
| seapen/whip | 2 | 2 | 4 | 3 | 6 | 41 | 0.83 | 0.79 | 0.87 | 0.63 | 0.79 | 0.95 |
| Sponge | 1 | 1 | 2 | 1 | 0 | 5 | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.03 |
| Anemone | 76 | 58 | 123 | 200 | 115 | 195 | 0.42 | 0.51 | 0.73 | 0.58 | 0.55 | 0.59 |
| Tunicate | 1 | 1 | 0 | 2 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 7 | 5 | 10 | 11 | 12 | 12 | 0.01 | 0.01 | 0.04 | 0.03 | 0.02 | 0.03 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1.00 | 0.00 |
| echinoderm | 1 | 0 | 3 | 0 | 0 | 0 | 0.02 | 0.00 | 0.11 | 0.00 | 0.00 | 0.01 |
| Coral | , | 0 |  |  | 1 | 2 | 0.07 | 0.02 | 0.04 | 0.30 | 0.01 | 0.03 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 26 | 33 | 17 | 24 | 13 | 13 | 0.98 | 0.86 | 0.81 | 0.97 | 0.88 | 0.96 |

Table 2.33b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod hook-and-line (including jigs) fishery, 2003-2005. The first part of the table ("Bycatch") shows the amount (t) of each species group taken as bycatch in the EBS Pacific cod hook-and-line fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 19972002.

|  | Byatch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Birds | 6 | 6 | 2 | 0.93 | 0.93 | 0.44 |
| Bivalves | 4 | 6 | 5 | 0.36 | 0.33 | 0.68 |
| Brittle star unidentified | 0 | 0 | 0 | 0.01 | 0.00 | 0.01 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 1 | 1 | 1 | 0.23 | 0.23 | 0.30 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 4 | 8 | 16 | 0.02 | 0.09 | 0.25 |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier | 1 | 16 | 91 | 0.01 | 0.08 | 0.08 |
| Greenlings | 3 | 1 | 1 | 0.28 | 0.23 | 0.20 |
| Grenadier | 221 | 202 | 158 | 0.08 | 0.10 | 0.12 |
| Gunnels |  | 0 | 0 |  | 1.00 | 1.00 |
| Hermit crab unidentified | 1 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Invertebrate unidentified | 14 | 2 | 3 | 0.02 | 0.00 | 0.01 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 194 | 1087 | 865 | 0.14 | 0.24 | 0.21 |
| Misc crabs | 1 | 1 | 9 | 0.01 | 0.02 | 0.24 |
| Misc crustaceans | 0 | 0 | 0 | 0.02 | 0.00 | 0.43 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 44 | 58 | 26 | 0.09 | 0.12 | 0.05 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.00 | 0.01 |
| Octopus | 41 | 37 | 20 | 0.30 | 0.10 | 0.08 |
| Other osmerids |  |  | 0 |  |  | 0.00 |
| Other Sculpins | 993 | 234 | 163 | 0.25 | 0.44 | 0.33 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified | 0 | 0 | 0 | 0.13 | 0.01 | 0.64 |
| Scypho jellies | 16 | 4 | 1 | 0.00 | 0.00 | 0.00 |
| Sea anemone unidentified | 79 | 94 | 69 | 0.58 | 0.53 | 0.69 |
| Sea pens whips | 6 | 10 | 19 | 0.86 | 0.84 | 0.88 |
| Sea star | 288 | 288 | 202 | 0.07 | 0.10 | 0.08 |
| Shark | 140 | 146 | 128 | 0.50 | 0.42 | 0.55 |
| Skate | 13519 | 13863 | 13219 | 0.74 | 0.75 | 0.78 |
| Snails | 5 | 6 | 6 | 0.03 | 0.02 | 0.05 |
| Sponge unidentified | 3 | 1 | 2 | 0.01 | 0.01 | 0.02 |
| Squid | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Stichaeidae | 0 |  |  | 0.05 |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |

Table 2.34a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod pot fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod pot fishery, broken down by year. The second part of the table
("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS Pacific cod pot fishery |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 351 | 267 | 438 | 494 | 315 | 384 | 0.05 | 0.05 | 0.10 | 0.09 | 0.05 | 0.05 |
| Skates | 1 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Octopus | 79 | 95 | 80 | 199 | 140 | 254 | 0.38 | 0.65 | 0.64 | 0.56 | 0.75 | 0.65 |
| Squid | 0 | 0 | 0 | 0 | 1 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 27 | 44 | 32 | 12 | 48 | 23 | 0.02 | 0.04 | 0.03 | 0.01 | 0.04 | 0.02 |
| Crabs | 1 | 1 | 4 | 2 | 1 | 2 | 0.00 | 0.00 | 0.04 | 0.01 | 0.01 | 0.01 |
| Starfish | 64 | 14 | 15 | 35 | 31 | 11 | 0.01 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 |
| Jellyfish | 11 | 1 | 16 | 0 | 6 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sponge | 0 | 0 | 0 | 0 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Anemone | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 8 | 3 | 4 | 11 | 4 | 9 | 0.01 | 0.01 | 0.02 | 0.03 | 0.01 | 0.02 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| echinoderm | 1 | 0 | 0 | 2 | 1 | 0 | 0.02 | 0.02 | 0.02 | 0.04 | 0.02 | 0.01 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |

Table 2.34b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod pot fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Byatch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0.01 | 0.00 | 0.01 |
| Bivalves | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 |
| Brittle star unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 0 |  | 0 | 0.01 |  | 0.01 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 |  |  | 0.00 |  |  |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 0 | 0 | 0.06 | 0.07 | 0.14 |
| Grenadier |  |  |  |  |  |  |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Invertebrate unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 122 | 191 | 109 | 0.09 | 0.04 | 0.03 |
| Misc crabs | 0 | 1 | 1 | 0.01 | 0.02 | 0.04 |
| Misc crustaceans | 0 | 0 |  | 0.00 | 0.01 |  |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 30 | 13 | 14 | 0.06 | 0.03 | 0.03 |
| Misc inverts (worms etc) |  |  |  |  |  |  |
| Octopus | 49 | 57 | 187 | 0.35 | 0.15 | 0.76 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 133 | 13 | 2 | 0.03 | 0.03 | 0.00 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified |  |  |  |  |  |  |
| Scypho jellies | 2 | 1 | 3 | 0.00 | 0.00 | 0.00 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Sea pens whips | 0 |  |  | 0.00 |  |  |
| Sea star | 41 | 30 | 27 | 0.01 | 0.01 | 0.01 |
| Shark |  |  |  |  |  |  |
| Skate | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Snails | 7 | 1 | 2 | 0.04 | 0.00 | 0.02 |
| Sponge unidentified | 1 | 1 | 0 | 0.00 | 0.01 | 0.00 |
| Squid |  |  | 1 |  |  | 0.00 |
| Stichaeidae |  |  |  |  |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 1 | 1 | 0 | 0.04 | 0.06 | 0.01 |

Table 2.35a-Bycatch of nontarget and "other" species taken in the AI Pacific cod trawl fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

|  | Bycatch in AI Pacific cod trawl fishery |  |  |  |  |  | Proportion of total AI catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 107 | 146 | 131 | 257 | 102 | 131 | 0.14 | 0.14 | 0.14 | 0.18 | 0.06 | 0.12 |
| Skates | 37 | 95 | 38 | 72 | 49 | 97 | 0.04 | 0.08 | 0.05 | 0.04 | 0.02 | 0.14 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 4 | 0 | 0 | 0.00 | 0.00 | 0.00 | 1.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 |
| Octopus | 2 | 2 | 9 | 2 | 1 | 9 | 0.06 | 0.05 | 0.04 | 0.03 | 0.03 | 0.38 |
| Squid | 1 | 0 | 0 | 1 | 2 | 4 | 0.01 | 0.01 | 0.01 | 0.07 | 0.30 | 0.25 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.95 | 0.00 | 1.00 | 1.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | . 00 |  | 1.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 6 | 38 | 29 | 25 | 26 | 15 | 0.04 | 0.14 | 0.09 | 0.12 | 0.11 | 0.07 |
| Crabs | 1 | 1 | 0 | 0 | 1 | 2 | 0.13 | 0.44 | 0.27 | 0.22 | 0.42 | 0.88 |
| Starfish | 2 | 3 | 5 | 5 | 5 | 5 | 0.12 | 0.15 | 0.29 | 0.20 | 0.17 | 0.46 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.17 | 0.00 | 0.99 | 0.01 | 0.44 |
| Invertunid | 0 | 2 | 3 | 6 | 2 | 0 | 0.00 | 0.03 | 0.34 | 0.40 | 0.36 | 0.02 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.85 | 0.23 | 0.54 | 0.33 | 0.08 | 0.16 |
| Sponge | 4 | 52 | 15 | 15 | 13 | 28 | 0.02 | 0.47 | 0.10 | 0.21 | 0.18 | 0.16 |
| Anemone | 0 | 0 | 1 | 0 | 0 | 0 | 0.09 | 0.08 | 0.41 | 0.17 | 0.05 | 0.17 |
| Tunicate | 0 | 0 | 0 | 0 | 1 | 0 | 0.63 | 0.75 | 0.08 | 0.58 | 0.40 | 0.07 |
| Benthinv | 4 | 3 | 1 | 2 | 3 | 6 | 0.90 | 0.68 | 0.16 | 0.73 | 0.76 | 0.92 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 1 | 1 | 1 | 1 | 2 | 0.16 | 0.26 | 0.23 | 0.35 | 0.44 | 0.75 |
| Coral | 2 | 8 | 2 | 8 | 3 | 11 | 0.07 | 0.48 | 0.03 | 0.24 | 0.15 | 0.52 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.05 | 0.00 | 0.11 | 0.19 | 0.10 |
| Birds | 0 | 1 | 0 | 0 | 0 | 0 | 0.02 | 0.11 | 0.02 | 0.04 | 0.01 | 0.16 |

Table 2.35b-Bycatch of nontarget and "other" species taken in the AI Pacific cod trawl fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.05 | 0.16 | 0.37 |
| Birds | 0 | 0 | 0 | 0.21 | 0.01 | 0.38 |
| Bivalves | 15 | 1 | 0 | 0.99 | 0.92 | 0.81 |
| Brittle star unidentified |  | 0 | 0 |  | 0.05 | 0.01 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 24 | 11 | 12 | 0.40 | 0.35 | 0.24 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 | 1 | 0 | 0.08 | 0.51 | 0.00 |
| Eulachon |  |  | 0 |  |  | 0.68 |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 0 | 0 | 0.66 | 0.05 | 0.01 |
| Grenadier |  | 4 | 0 |  | 0.01 | 0.00 |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 0 | 0 | 0 | 0.80 | 0.98 | 0.09 |
| Invertebrate unidentified | 0 | 0 | 0 | 0.09 | 0.00 | 0.02 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 78 | 159 | 88 | 0.37 | 0.23 | 0.18 |
| Misc crabs | 1 | 1 | 0 | 0.73 | 0.59 | 0.52 |
| Misc crustaceans | 0 | 0 | 0 | 0.99 | 0.29 | 0.98 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 28 | 15 | 19 | 0.23 | 0.10 | 0.12 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.29 | 1.00 |
| Octopus | 6 | 5 | 3 | 0.36 | 0.28 | 0.40 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 122 | 1 | 3 | 0.31 | 0.01 | 0.04 |
| Pacific Sand lance | 0 |  | 0 | 1.00 |  | 1.00 |
| Pandalid shrimp | 0 | 0 | 0 | 0.06 | 0.01 | 0.03 |
| Polychaete unidentified |  | 0 | 0 |  | 0.13 | 0.97 |
| Scypho jellies | 0 | 0 | 1 | 0.17 | 0.49 | 0.44 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.61 | 0.31 | 0.32 |
| Sea pens whips | 0 | 0 | 0 | 0.34 | 0.91 | 0.42 |
| Sea star | 5 | 3 | 2 | 0.49 | 0.27 | 0.17 |
| Shark | 0 | 2 | 2 | 0.01 | 0.43 | 0.10 |
| Skate | 72 | 76 | 65 | 0.13 | 0.09 | 0.11 |
| Snails | 1 | 1 | 0 | 0.52 | 0.50 | 0.21 |
| Sponge unidentified | 24 | 18 | 22 | 0.30 | 0.13 | 0.28 |
| Squid | 3 | 2 | 1 | 0.10 | 0.11 | 0.07 |
| Stichaeidae |  |  | 0 |  |  | 0.00 |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 1 | 1 | 0 | 0.40 | 0.43 | 0.15 |

Table 2.36a-Bycatch of nontarget and "other" species taken in the AI Pacific cod longline fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod longline fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

|  | Bycatch in AI Pacific cod longline fishery |  |  |  |  |  | Proportion of total AI catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 334 | 597 | 356 | 662 | 1004 | 214 | 0.43 | 0.55 | 0.37 | 0.47 | 0.63 | 0.19 |
| Skates | 338 | 727 | 473 | 1397 | 2184 | 246 | 0.39 | 0.64 | 0.59 | 0.77 | 0.87 | 0.35 |
| Shark | 0 | 1 | 0 | 0 | 0 | 0 | 0.78 | 0.04 | 0.05 | 0.03 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 1 | 0 | 0.96 | 0.55 | 0.84 | 0.85 | 0.31 | 0.54 |
| Sleepershk | 0 | 0 | 1 | 0 | 1 | 2 | 0.00 | 0.00 | 0.02 | 0.00 | 0.03 | 0.49 |
| Octopus | 10 | 21 | 9 | 13 | 21 | 8 | 0.27 | 0.47 | 0.05 | 0.20 | 0.51 | 0.32 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0.00 |  | 0.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| Grenadier | 397 | 83 | 215 | 151 | 6 | 88 | 0.14 | 0.05 | 0.07 | 0.05 | 0.00 | 0.03 |
| Otherfish | 2 | 5 | 2 | 6 | 10 | 3 | 0.02 | 0.02 | 0.01 | 0.03 | 0.04 | 0.01 |
| Crabs | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.01 | 0.01 | 0.04 | 0.00 |
| Starfish | 3 | 7 | 4 | 13 | 16 | 3 | 0.22 | 0.41 | 0.28 | 0.51 | 0.59 | 0.25 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Invertunid | 0 | 1 | 0 | 1 | 0 | 0 | 0.00 | 0.01 | 0.02 | 0.06 | 0.08 | 0.02 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.21 | 0.44 | 0.54 | 0.92 | 0.56 |
| Sponge | 0 | 4 | 3 | 11 | 4 | 1 | 0.00 | 0.04 | 0.02 | 0.15 | 0.06 | 0.00 |
| Anemone | 0 | 0 | 1 | 1 | 0 | 1 | 0.34 | 0.57 | 0.32 | 0.59 | 0.47 | 0.69 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 | 0.24 | 0.00 | 0.00 |
| Benthinv | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.02 | 0.06 | 0.04 | 0.03 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 0 | 0 | 0 | 0 | 0 | 0.10 | 0.04 | 0.00 | 0.09 | 0.04 | 0.02 |
| Coral | 0 | 1 | 2 | 6 | 3 | 1 | 0.02 | 0.03 | 0.04 | 0.17 | 0.16 | 0.03 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.09 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Birds | 2 | 2 | 2 | 2 | 1 | 0 | 0.75 | 0.45 | 0.55 | 0.66 | 0.48 | 0.16 |

Table 2.36b-Bycatch of nontarget and "other" species taken in the AI Pacific cod hook-and-line (including jigs) fishery, 2003-2005. The first part of the table ("Bycatch") shows the amount (t) of each species group taken as bycatch in the AI Pacific cod hook-and-line fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.09 | 0.00 | 0.01 |
| Birds | 0 | 0 | 0 | 0.03 | 0.21 | 0.29 |
| Bivalves | 0 | 0 | 0 | 0.00 | 0.02 | 0.18 |
| Brittle star unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 1 | 1 | 0 | 0.01 | 0.05 | 0.01 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier | 0 | 0 | 0 | 0.30 | 0.00 | 0.00 |
| Greenlings | 0 | 0 | 0 | 0.08 | 0.16 | 0.02 |
| Grenadier | 46 | 8 | 0 | 0.01 | 0.01 | 0.00 |
| Gunnels |  |  | 0 |  |  | 0.00 |
| Hermit crab unidentified | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Invertebrate unidentified | 0 | 1 | 0 | 0.00 | 0.12 | 0.03 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 28 | 133 | 91 | 0.14 | 0.19 | 0.18 |
| Misc crabs | 0 | 0 | 0 | 0.00 | 0.01 | 0.01 |
| Misc crustaceans | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 1 | 3 | 1 | 0.01 | 0.02 | 0.00 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.00 | 0.00 |
| Octopus | 8 | 8 | 4 | 0.54 | 0.49 | 0.55 |
| Other osmerids |  |  | 0 |  |  | 0.00 |
| Other Sculpins | 31 | 63 | 1 | 0.08 | 0.41 | 0.01 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified | 0 | 0 | 0 | 1.00 | 0.00 | 0.03 |
| Scypho jellies | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.24 | 0.23 | 0.58 |
| Sea pens whips | 0 | 0 | 0 | 0.46 | 0.09 | 0.15 |
| Sea star | 1 | 6 | 3 | 0.10 | 0.47 | 0.25 |
| Shark | 0 | 0 | 0 | 0.01 | 0.08 | 0.02 |
| Skate | 105 | 402 | 245 | 0.20 | 0.48 | 0.43 |
| Snails | 0 | 0 | 0 | 0.01 | 0.03 | 0.05 |
| Sponge unidentified | 2 | 5 | 2 | 0.02 | 0.04 | 0.03 |
| Squid |  | 0 |  |  | 0.00 |  |
| Stichaeidae | 0 |  |  | 0.00 |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 0 | 0 | 0 | 0.02 | 0.11 | 0.01 |

Table 2.37-Bycatch of nontarget and "other" species taken in the AI Pacific cod pot fishery, 1997-2002. The first part of the table ("Bycatch in...") shows the amount (t) of each species group taken as bycatch in the AI Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

|  | Bycatch in AI Pacific cod pot fishery |  |  |  |  |  | Proportion of total AI catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 7 | 12 | 221 | 211 | 42 | 0 | 0.01 | 0.01 | 0.23 | 0.15 | 0.03 | 0.00 |
| Skates | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Octopus | 24 | 18 | 182 | 47 | 17 | 0 | 0.62 | 0.40 | 0.90 | 0.75 | 0.41 | 0.00 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0.00 |  | 0.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 0 | 0 | 7 | 1 | 4 | 0 | 0.00 | 0.00 | 0.02 | 0.01 | 0.02 | 0.00 |
| Crabs | 0 | 0 | 1 | 1 | 0 | 0 | 0.00 | 0.06 | 0.51 | 0.61 | 0.31 | 0.00 |
| Starfish | 0 | 0 | 1 | 1 | 0 | 0 | 0.00 | 0.00 | 0.05 | 0.05 | 0.00 | 0.00 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.07 | 0.00 | 0.00 |
| Sponge | 0 | 0 | 0 | 4 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.06 | 0.00 | 0.00 |
| Anemone | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 0 | 0 | 1 | 0 | 0 | 0 | 0.00 | 0.01 | 0.09 | 0.12 | 0.00 | 0.00 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 0 | 1 | 1 | 0 | 0 | 0.01 | 0.00 | 0.20 | 0.18 | 0.00 | 0.00 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |

Table 2.38—Summary of major results for the stock assessment of Pacific cod in the BSAI region.

| Tier | 3b |
| :--- | ---: |
| Reference mortality rates |  |
| $M$ | 0.34 |
| $F_{40 \%}$ | 0.31 |
| $F_{35 \%}$ | 0.37 |
| Equilibrium spawning biomass |  |
| $B_{35 \%}$ | $472,000 \mathrm{t}$ |
| $B_{40 \%}$ | $540,000 \mathrm{t}$ |
| $B_{100 \%}$ | $1,350,000 \mathrm{t}$ |
| Projected biomass for 2008 |  |
| Spawning (at max FABC) | $398,000 \mathrm{t}$ |
| Age 3+ | $1,080,000 \mathrm{t}$ |
| ABC for 2008 | 0.22 |
| $F_{\text {ABC }}$ (maximum permissible) | 0.22 |
| $F_{\text {ABC }}$ (recommended) | $150,000 \mathrm{t}$ |
| ABC (maximum permissible) | $150,000 \mathrm{t}$ |
| ABC (recommended) |  |
| Overfishing level for 2008 | 0.26 |
| Fishing Mortality |  |
| Catch |  |



Figure 2.1a-Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in January-May 2006, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.1b-Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in June-August 2006, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.1c-Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in September-December 2006, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.1d—Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in January-May 2007, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.1e-Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in June-August 2007, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.2-Comparison of log numbers (1000s) of age 0 EBS Pacific cod under four alternative models.


Figure 2.3-Comparison of female spawning biomass of EBS Pacific cod under four alternative models.


Figure 2.4a. Model 1 selectivity.


Figure 2.4b. Model 2 selectivity.


Figure 2.4c. Model 3 selectivity.


Figure 2.4d. Model 4 selectivity.


Figure 2.5a. Observed and estimated relative abundance (Model 1).


Figure 2.5b. Observed and estimated relative abundance (Model 2).


Figure 2.5c. Observed and estimated relative abundance (Model 3).


Figure 2.5d. Observed and estimated relative abundance (Model 4).


Figure 2.6-Biomass time trends (age 3+ biomass, female spawning biomass, survey biomass) of EBS Pacific cod as estimated by Model 1.


Figure 2.7-Time series of EBS Pacific cod recruitment at age 0 , with $95 \%$ confidence intervals, as estimated by Model 1. Red line = average.


Figure 2.8-Age 0 recruitment versus female spawning biomass for Pacific cod during the years 19772006 as estimated by Model 1, with Ricker stock-recruitment curve (for illustrative purposes only).


Figure 2.9-Trajectory of Pacific cod fishing mortality and female spawning biomass as estimated by Model 1, 1977-present. Because Pacific cod is a key prey of Steller sea lions, harvests of Pacific cod would be restricted to incidental catch in the event that spawning biomass fell below $B_{20 \%}$.

# Attachment 2.1: Results from Ecosystem Models on the Role of Pacific Cod In the Eastern Bering Sea and Aleutian Islands Ecosystems 

Sarah Gaichas and Kerim Aydin

Pacific cod are important predators in the Eastern Bering Sea (EBS) and Aleutian Islands (AI) ecosystems. While they are managed similarly in both ecosystems, food web modeling suggests key differences in cod's ecosystem role in the AI and EBS. The first key difference between ecosystems relates to cod's relative density in its continental shelf habitats in each system: because the AI has a much smaller area of shelf relative to the EBS (and the Gulf of Alaska, GOA), the smaller survey biomass estimate of cod in this area translates into a higher density in tons per square kilometer relative to the density in the EBS (Figure 1, left panel). Although the density of cod differs between systems, the relative effects of fishing and predation mortality as estimated within food web models constructed for each ecosystem (Aydin et al. in press) are similar between the AI, EBS, and GOA. Here, sources of mortality are compared against the total production of cod as estimated in the BSAI and GOA cod stock assessment models (see Annex A, "Production rates," for detailed methods). The "unknown" mortality in Figure 1 (left) represents the difference between the stock assessment estimated cod production and the known sources of fishing and predation mortality. While nearly half of cod production as estimated by the stock assessment appears to be "unused" in all three ecosystems, it is also clear that cod have relatively more fishing mortality than predation mortality in all three ecosystems (Figure 1, right panel). This suggests that changing fishing mortality is likely to affect cod population trajectories; therefore, we may ask what ecosystem effects changes in cod mortality might cause in each ecosystem.

To determine the potential ecosystem effects of changing total cod mortality, we first examine the diet data collected for cod. Diet data are collected aboard NMFS bottom trawl surveys in both the EBS and AI ecosytems during the summer (May - August); this comparison uses diet data collected in the early 1990's in each ecosystem. In the EBS, 2436 cod stomachs were collected during the 1991 bottom trawl survey and used in this analysis. In the AI, a total of 1181 cod stomachs were collected between the 1991 and 1994 bottom trawl surveys ( $\mathrm{n}=659$ and 533, respectively) and used in this analysis. The diet compositions reported here reflect the size and spatial distribution of cod in each survey (see Annex A, "Diet calculations" for detailed methods). While the diet compositions reported here most accurately reflect early 1990's conditions in the BSAI, it is possible to update this information and examine changes in cod diets over time; that more extensive analysis is planned for a future assessment.

Food habits data show that Pacific cod have an extremely varied diet in both ecosystems (Figure 2). In the EBS, pollock are a major diet item for cod ( $26 \%$ of diet), but in the AI Atka mackerel and sculpins are the predominant fish prey for cod ( $15 \%$ of diet each), with pollock comprising less than $5 \%$ of the diet. In both ecosystems, Pandalid and non-Pandalid (NP) shrimp and various crabs are important prey, but other major prey items differ by ecosystem and seem to relate to the relative importance of benthic and pelagic pathways in each ecosystem as discussed in Aydin et al (in review). Commerically important crab species such as snow crab (C. opilio) and tanner crab (C. bairdi) make up 9\% of cod diets in the EBS, but less than $3 \%$ in the AI, reflecting the stronger benthic energy flow in the EBS. In contrast, squids make up over $6 \%$ of cod diets in the AI, but are very small proportions of diets in the EBS, reflecting the stronger pelagic energy flow in the AI. Myctophids are also found in cod diets only in the AI, reflecting the oceanic nature of the food web there. Cod are clearly opportunistic predators in both ecosystems, feeding on a variety of fish and invertebrates, and scavenging as well. Fishery offal makes up $5-7 \%$ of cod diets in
both systems, indicating that while fishing causes cod mortality, it also contributes to cod production (although much fishery offal comes from fisheries directed at pollock, not cod).

Using diet data for all predators of cod and consumption estimates for those predators, as well as fishery catch data, we next estimate the sources of cod mortality in the AI and EBS (see detailed methods in Annex A). As described above, sources of mortality are compared against the total production of cod as estimated in the BSAI cod stock assessment model. Mortality sources for cod are similar when comparing fisheries, but different when comparing predators between the EBS and AI. In both ecosystems, the trawl and longline fisheries for cod were the largest mortality sources for cod in the early 1990s (Figure 3). The next largest source of cod mortality is the pollock trawl fishery in the EBS and the directed Atka mackerel ("Other groundfish") fishery in the AI, which retains incidentally caught cod. In the EBS, pollock predation ranks next, and in the AI, adult and juvenile Steller sea lion predation represents the largest single source of predation mortality for cod. Cod cannibalism is a significant source of cod mortality only in the EBS, and flatfish trawl fisheries round out the large cod mortality sources in that ecosystem. Therefore, we see groundfish-dominated predation mortality sources for cod in the EBS, but sea-lion dominated predation mortality in the AI.

After comparing the different diet compositions and mortality sources of cod in each ecosystem, we shift focus slightly to view cod within the context of the larger EBS and AI food webs (Figure 4). Visually, it is apparent that cod's direct trophic relationships in each ecosystem include a majority of species groups; there are few boxes not connected to cod. However, comparing these food webs show further differences in cod trophic relationships between ecosystems. In the EBS, the significant predators of cod (blue boxes joined by blue lines) include the cod fisheries, the pollock fishery, and resident seals (upper panel of Figure 4). Significant prey of cod (green boxes joined by green lines) include the many species shown in Figure 2. Light blue boxes in the EBS food web represent species which are both predators and prey of cod at some stage of life, with the most significant predator/prey of cod being pollock. In contrast, there are no species groups in the AI which are both predator and prey to cod (Figure 4, lower panel).

We can investigate whether these differences in cod diet, mortality, and relationships between the EBS and AI might suggest different ecosystem roles for cod in these areas. We use the diet and mortality results integrated with information on uncertainty in the food web using the Sense routines (Aydin et al. in press) and a perturbation analysis with each model food web to explore the ecosystem relationships of cod further. Two questions are important in determining the ecosystem role of cod: which species groups are cod important to, and which species groups are important to cod? First, the importance of cod to other groups within the EBS and AI ecosystems was assessed using a model simulation analysis where cod survival was decreased (mortality was increased) by a small amount, $10 \%$, over 30 years to determine the potential effects on other living groups. This analysis also incorporated the uncertainty in model parameters using the Sense routines, resulting in ranges of possible outcomes which are portrayed as $50 \%$ confidence intervals (boxes in Figure 5) and 95\% confidence intervals (error bars in Figure 5). Species showing the largest median changes from baseline conditions are presented in descending order from left to right. Therefore, the largest change resulting from a $10 \%$ decrease in cod survival in both ecosystems is a decrease in adult cod biomass, as might have been expected from such a perturbation. However, the decrease in biomass resulting from the same perturbation is different between the EBS and AI: the 50\% intervals range from a $7-11 \%$ decrease in the AI, to a $7-17 \%$ decrease in the EBS (Figure 5).

The simulated decrease in cod survival affects the fisheries for cod similarly in the EBS and AI. After the decreased adult cod biomass, the next largest effect of the perturbation predicted by the models is a decrease in the "biomass" (catch) of the pot, longline, and trawl fisheries targeting adult cod in the EBS (Figure 5, top panel). In the AI ecosystem model, adult sablefish are predicted to have a larger change from the cod manipulation than the fisheries, although the predicted increase in sablefish biomass is much more uncertain than the predicted decrease in fishery catch in the AI (bottom panel, Figure 5). We discuss the sablefish result in detail below; for this discussion, we note that the cod fisheries in the AI are
behaving similarly to the cod fisheries in the EBS after the simulated decrease in cod survival. Since cod fisheries are extremely specialized predators of cod, it makes sense that they are most sensitive to changes in the survival of cod in each ecosystem. It is notable that none of the other predators of cod showed a significant sensitivity to a $10 \%$ decrease in cod survival. Pollock and sea lions ranked highest as nonfishery mortality sources of cod in the EBS and AI, respectively, but neither of these species were predicted to have significant changes in biomass in either ecosystem in this analysis: neither EBS pollock nor AI sea lions showed enough change from the baseline condition to be included in the plots. While these predators may cause significant cod mortality in each system, this analysis suggests that none of them are dependent on cod to the extent that small changes in cod survival affect their biomass in a predictable manner. It may be that these predator species would react more strongly to larger changes in cod survival; this could be further analyzed with different perturbation analyses.

In contrast with the predators of cod, a $10 \%$ decrease in cod survival is predicted to change the biomass of some cod prey, and even some species not directly connected to cod. In the EBS, greenling biomass is predicted to increase as a result of the perturbation, as are tanner crab and king crab biomass, albeit wth less certainty (Figure 5, top panel). In the AI, a larger set of species appear to react more strongly to increases in cod mortality than in the other two systems: sablefish, rex sole, arrowtooth flounder, and sleeper sharks are all predicted to increase in biomass in addition to greenlings and small sculpins (Figure 5). Of these, only rex sole, greenlings and other sculpins are direct cod prey; the change in adult sablefish and adult arrowtrooth biomass apparently arises from reduced cod predation mortality on the juveniles of each species in the AI ecosystem model: cod cause $80 \%$ of juvenile sablefish and juvenile arrowtooth mortality in the AI model. Sleeper sharks are neither predators nor prey of cod in the AI, suggesting that decreased cod survival has strong indirect effects in this ecosystem. Some of these differences in species sensitivity to cod mortality arise from the differences in cod diet in each system, but it seems likely that the higher sensitivity of multiple species to cod in the AI may also be due to cod's higher biomass per unit area there relative to the EBS. This in turn suggests that in the AI there may be stronger potential ecosystem effects of cod fishing than in the other two systems.

To determine which groups were most important to cod in each ecosystem, we conducted the inverse of the analysis presented above. In this simulation, each species group in the ecosystem had survival reduced by $10 \%$ and the system was allowed to adjust over 30 years. The strongest median effects on EBS and AI adult cod are presented in Figure 6. The largest effect on adult cod was the reduction in biomass resulting from the reduced survival of juvenile cod, followed by the expected direct effect, reduced biomass of adult cod in response to reduced survival of adult cod, in both ecosystems (Figure 6). Beyond these direct single species effects, cod appear most sensitive in all ecosystems to bottom up effects from both pelagic and benthic production pathways (small phytoplankton and benthic detritus). However, the bottom up effect is most pronounced in the AI, where the upper 95\% intervals for the percent change of cod indicate that cod biomass will almost certainly decrease as a result of decreased survival of small phytoplankton, benthic detritus, and large phytoplankton (Figure 6). In contrast, the EBS model prediction is that cod biomass is likely to decrease from decreased survival of small phytoplankton and benthic detritus, but the detritus $95 \%$ intervals cross the x axis indicating that no change is also a possible outcome.

While decreased survival of primary producers appears to hurt cod, there are few species groups in either ecosystem which appear to benefit cod through reduced survival. In other words, they have no obvious single competitor or predator supressing cod biomass in the AI or EBS. In general, reduced "survival" (lower catch) of fisheries means more cod in the EBS and AI. In the EBS, reduced survival of other sculpins may increase cod biomass to some extent (Figure 6), which may seem counterintuitive given that reduced cod survival appeared to increase other sculpin biomass in the AI (Figure 5). While adult cod eat other sculpins, other sculpins in turn eat juvenile cod in the EBS (Figure 7), likely accounting for the results shown in Figure 6.

The results of these perturbation analyses suggest that the regional level of management applied to Pacific cod should be modified to account for differences between ecosystems. The food web relationships of cod are demonstrably different between the EBS and AI ecosystems, where they are currently assessed and managed identically. The impacts of changing cod survival (and by extension, fishing mortality) differ by ecosystem as well, with the impacts felt most strongly and with highest certainty in the AI ecosystem according to this analysis. Therefore, it seems that the cod fishery in the AI should be managed separately from that in the EBS to ensure that any potential ecosystem effects of changing fishing mortality might be monitored at the appropriate scale.

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Figure 1. Comparative biomass density (left) and mortality sources (right) for Pacific cod in the AI, EBS, and GOA ecosystems. For the AI and GOA, biomass density (left) is the average biomass from early 1990s NMFS bottom trawl surveys divided by the total area surveyed. For the EBS, biomass density is the stock assessment estimated adult (age 3+) biomass for 1991 (Thompson and Dorn 2005) divided by the total area covered by the EBS bottom trawl survey. Total cod production (right) is derived from cod stock assessments for the early 1990's, and partitioned according to fishery catch data and predation mortality estimated from cod predator diet data (Aydin et al. in press). See Annex A for detailed methods.


Figure 2. Comparison of Pacific cod diet compositions for the EBS (top) and AI (bottom) ecosystems. Diets are estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI). See Annex A for detailed methods.


AI P. Cod mortality


Figure 3. Comparison of Pacific cod mortality sources for the EBS (top) and AI (bottom) ecosystems. Mortality sources reflect cod predator diets estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI), cod predator consumption rates estimated from stock assessments and other studies, and catch of cod by all fisheries in the same time periods (Aydin et al. in press). See Annex A for detailed methods.



Figure 4. Adult and juvenile cod in the EBS (top) and AI (bottom) food webs. Predators of cod are dark blue, prey of cod are green, and species that are both predators and prey of cod are light blue. Box size is proportional to biomass and lines between boxes represent the most significant energy flows.

BS P. Cod effects on other species


AI P. Cod effects on other species


Figure 5. Effect of changing cod survival on fishery catch (yellow) and biomass of other species (dark red): EBS (top) and AI (bottom), from a simulation analysis where cod survival was decreased by $10 \%$ and the rest of the ecosystem adjusted to this decrease for 30 years. Boxes show resulting percent change in the biomass of each species on the x axis after 30 years for $50 \%$ of feasible ecosystems, error bars show results for $95 \%$ of feasible ecosystems (see Aydin et al. in press for detailed Sense methods).


Figure 6. Effect of reducing fisheries catch (yellow) and other species survival (dark red) on cod biomass: EBS (top) and AI (bottom), from a simulation analysis where survival of each X axis species group was decreased by $10 \%$ and the rest of the ecosystem adjusted to this decrease for 30 years. Boxes show resulting percent change in the biomass of adult cod after 30 years for $50 \%$ of feasible ecosystems, error bars show results for $95 \%$ of feasible ecosystems (see Aydin et al. in press for detailed Sense methods).


Figure 7. Juvenile cod mortality sources: EBS (top) and AI (bottom). Mortality sources reflect juvenile cod predator diets estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI), cod predator consumption rates estimated from stock assessments and other studies, and catch of cod by all fisheries in the same time periods (Aydin et al. in press). See Annex A for detailed methods.

## Annex A

## Diet composition calculations

Notation:
DC = diet composition
$\mathrm{W}=$ weight in stomach
$\mathrm{n}=$ prey
$\mathrm{p}=$ predator
$\mathrm{s}=$ predator size class
$\mathrm{h}=$ survey haul
r = survey stratum
B = biomass estimate
$\mathrm{v}=$ survey
a = assessment
$\mathrm{R}=$ ration estimate

The diet composition for a species is calculated from stomach sampling beginning at the level of the individual survey haul (1), combining across hauls within a survey stratum (2), weighting stratum diet compositions by stratum biomass (3), and finally combining across predator size classes by weighting according to size-specific ration estimates and biomass from stock assessment estimated age structure (4). Ration calculations are described in detail below.
Diet composition (DC) of prey $n$ in predator $p$ of size $s$ in haul $h$ is the total weight of prey $n$ in all of the stomachs of predator $p$ of size $s$ in the haul divided by the sum over all prey in all of the stomachs for that predator size class in that haul:

$$
\begin{equation*}
D C_{n, p, s, h}=W_{n, p, s, h} / \sum_{n} W_{n, p, s, h} \tag{1}
\end{equation*}
$$

Diet composition of prey $n$ in predator $p$ of size $s$ in survey stratum $r$ is the average of the diet compositions across hauls within that stratum:

$$
\begin{equation*}
D C_{n, p, s, r}=\sum_{h} D C_{n, p, s, h} / h \tag{2}
\end{equation*}
$$

Diet composition of prey n in predator p of size s for the entire area t is the sum over all strata of the diet composition in stratum $r$ weighted by the survey biomass proportion of predator $p$ of size $s$ in stratum $r$ :

$$
\begin{equation*}
D C_{n, p, s, t}=\sum_{r} D C_{n, p, s, r} * B_{p, s, r}^{v} / \sum_{r} B_{p, s, r}^{v} \tag{3}
\end{equation*}
$$

Diet composition of prey $n$ in predator $p$ for the entire area $t$ is the sum over all predator sizes of the diet composition for predator $p$ of size $s$ as weighted by the relative stock assessment biomass of predator size $s$ times the ration of predator $p$ of size $s$ :

$$
\begin{equation*}
D C_{n, p, t}=\sum_{s} D C_{n, p, s, t} * B_{p, s}^{a} * R_{p, s} / \sum_{s} B_{p, s}^{a} * R_{p, s} \tag{4}
\end{equation*}
$$

## Ration Calculations

Size specific ration (consumption rate) for each predator was determined by the method of fitting the generalized Von Bertalanffy growth equations (Essington et al. 2001) to weight-at-age data collected aboard NMFS bottom trawl surveys.

The generalized Von Bertalanffy growth equation assumes that both consumption and respiration scale allometrically with body weight, and change in body weight over time (dW/dT) is calculated as follows (Paloheimo and Dickie 1965):

$$
\begin{equation*}
\frac{d W_{t}}{d t}=H \cdot W_{t}^{d}-k \cdot W_{t}^{n} \tag{5}
\end{equation*}
$$

Here, $W_{t}$ is body mass, $t$ is the age of the fish (in years), and $H, d, k$, and $n$ are allometric parameters. The term $H \cdot W_{t}^{d}$ is an allometric term for "useable" consumption over a year, in other words, the consumption (in wet weight) by the predator after indigestible portions of the prey have been removed and assuming constant caloric density between predator and prey. Total consumption $(Q)$ is calculated as $(1 / A) \cdot H \cdot W_{t}^{d}$, where $A$ is a scaling fraction between predator and prey wet weights that accounts for indigestible portions of the prey and differences in caloric density. The term $k \cdot W_{t}^{n}$ is an allometric term for the amount of biomass lost yearly as respiration.

Based on an analysis performed across a range of fish species, Essington et al. (2001) suggested that it is reasonable to assume that the respiration exponent $n$ is equal to 1 (respiration linearly proportional to body weight). In this case, the differential equation above can be integrated to give the following solution for weight-at-age:

$$
\begin{equation*}
W_{t}=W_{\infty} \cdot\left(1-e^{-k(1-d)\left(t-t_{0}\right)}\right) \frac{1}{1-d} \tag{6}
\end{equation*}
$$

Where $W_{\infty}$ (asymptotic body mass) is equal to $(H / k)^{\frac{1}{1-d}}$, and $t_{0}$ is the weight of the organism at time $=0$. If the consumption exponent $d$ is set equal to $2 / 3$, this equation simplifies into the "specialized" von Bertalanffy length-at-age equation most used in fisheries management, with the "traditional" von Bertalanffy K parameter being equal to the $k$ parameter from the above equations divided by 3 .

From measurements of body weight and age, equation 2 can be used to fit four parameters ( $W_{\infty}, d, k$, and $t_{0}$ ) and the relationship between $W_{\infty}$ and the $H, k$, and $d$ parameters can then be used to determine the consumption rate $H \cdot W_{t}^{d}$ for any given age class of fish. For these calculations, weight-at-age data available and specific to the modeled regions were fit by minimizing the difference between $\log$ (observed) and $\log$ (predicted) body weights as calculated by minimizing negative log likelihood:
observation error was assumed to be in weight but not aging. A process-error model was also examined but did not give significantly different results.

Initial fitting of 4-parameter models showed, in many cases, poor convergence to unique minima and shallow sum-of-squares surfaces: the fits suffered especially from lack of data at the younger age classes that would allow fitting to body weights near $\mathrm{t}=0$ or during juvenile, rapidly growing life stages. To counter this, the following multiple models were tested for goodness-of-fit:

1. All four parameters estimated by minimization;
2. $d$ fixed at $2 / 3$ (specialized von Bertalanffy assumption)
3. $d$ fixed at 0.8 (median value based on metaanalysis by Essington et al. 2001).
4. $t_{0}$ fixed at 0 .
5. $d$ fixed at $2 / 3$ with $t_{0}$ fixed at 0 , and $d$ fixed at 0.8 with $t_{0}$ fixed at 0 .

The multiple models were evaluated using Aikeike's Information Criterion, AIC (spreadsheet review). In general, the different methods resulted in a twofold range of consumption rate estimates; consistently, model \#3, $d$ fixed at 0.8 while the other three parameters were free, gave the most consistently good results using the AIC. In some cases model \#1 was marginally better, but in some cases, model \#1 failed to converge. The poorest fits were almost always obtained by assuming that d was fixed at $2 / 3$.

To obtain absolute consumption $(Q)$ for a given age class, the additional parameter $A$ is required to account for indigestible and otherwise unassimilated portions of prey. We noted that the range of indigestible percentage for a wide range of North Pacific zooplankton and fish summarized in Davis (2003) was between $5-30 \%$, with major zooplankton (copepods and euphasiids), as well as many forage fish, having a narrower range of indigestible percentages, generally between $10-20 \%$. Further, bioenergetics models, for example for walleye pollock (Buckley and Livingston 1994), indicate that nitrogenous waste (excretion) and egestion resulted in an additional 20-30\% loss of consumed biomass. As specific bioenergetics models were not available for most species, we made a uniform assumption of a total non-respirative loss of $40 \%$ (from a range of $25-60 \%$ ) for all fish species, with a corresponding $A$ value of 0.6 .

Finally, consumption for a given age class was scaled to population-level consumption using the available numbers-at-age data from stock assessments, or using mortality rates from stock assessments and the assumption of an equilibrium age structure in cases where numbers-at-age reconstructions were not available.

## Production rates

Production per unit biomass $(P / B)$ and consumption per unit biomass $(Q / B=R$, ration above) for a given population depend heavily on the age structure, and thus mortality rate of that population. For a population with an equilibrium age structure, assuming exponential mortality and Von Bertalanffy growth, $P / B$ is in fact equal to total mortality $Z$ (Allen 1971) and $Q / B$ is equal to $(Z+3 K) / A$, where $K$ is Von Bertalanffy's K , and A is a scaling factor for indigestible proportions of prey (Aydin 2004). If a population is not in equilibrium, $\mathrm{P} / \mathrm{B}$ may differ substantially from Z although it will still be a function of mortality.

For the Bering Sea, Aleutian Islands, and Gulf of Alaska ECOPATH models, P/B and Q/B values depend on available mortality rates, which were taken from estimates or literature values used in single-species models of the region. It is noted that the single-species model assumptions of constant natural mortality
are violated by definition in multispecies modeling; therefore, these estimates should be seen as "priors" to be input into the ECOPATH balancing procedures or other parameter-fitting (e.g. Bayesian) techniques.

Several methods were used to calculate P/B, depending on the level of data available. Proceeding from most data to least data, the following methods were used:

1. If a population is not in equilibrium, total production $P$ for a given age class over the course of a year can be approximated as ( $\mathrm{N}_{\mathrm{at}} \cdot \Delta \mathrm{W}_{\mathrm{at}}$ ), where $\mathrm{N}_{\mathrm{at}}$ is the number of fish of a given age class in a given year, exponentially averaged to account for mortality throughout the year, and $\Delta \mathrm{W}_{\mathrm{at}}$ is the change in body weight of that age class over that year. For a particular stock, if weight-at-age data existed for multiple years, and stock-assessment reconstructed numbers-at-age were also available, production was calculated by summing this equation over all assessed age classes. Walleye pollock P/B for both the EBS and GOA were calculated using this method: examining the components of this sum over the years showed that numbers-at-age variation was responsible for considerably more variability in overall $\mathrm{P} / \mathrm{B}$ than was weight-at-age variation.
2. If stock assessment numbers-at-age were available, but a time series of weight-at-age was not available and some weight-at-age data was available, the equation in (1), above, was used, however, the change in body weight over time was estimated using fits to the generalized Von Bertalanffy equations described in the consumption section, above.
3. If no stock assessment of numbers-at-age was available, the population was assumed to be in equilibrium, so that $\mathrm{P} / \mathrm{B}$ was taken to equal Z . In cases for many nontarget species, estimates of Z were not available so estimates of M were taken from conspecifics with little assumed fishing mortality for this particular calculation.

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