# Chapter 7 <br> Northern Rock Sole 

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## EXECUTIVE SUMMARY

The following changes have been made to this assessment relative to the November 2005 SAFE:

## Changes to the input data

1) 2005 fishery age composition.
2) 2005 survey age composition.
3) 2006 trawl survey biomass point estimate and standard error.
4) Estimate of catch ( $t$ ) and discards through 6, September 2006.
5) Estimate of retained and discarded portions of the 2005 catch.

## Assessment results

1) The projected age $2+$ biomass for 2007 is $1,674,000 \mathrm{t}$.
2) The projected female spawning biomass for 2007 is $392,000 \mathrm{t}$.
3) The recommended 2007 ABC is $121,100 t$ based on an $\mathrm{F}_{40 \%}(0.144)$ harvest level.
4) The 2007 overfishing level is $144,000 \mathrm{t}$ based on an $\mathrm{F}_{35 \%}(0.174)$ harvest level.

|  | 2006 <br> Assessment Recommendations <br> for the 2007 harvest | 2005 Assessment Recommendations <br> for the 2006 harvest |
| :--- | :---: | :---: |
| Total biomass | $1,674,000 \mathrm{t}$ | $1,489,600 \mathrm{t}$ |
| ABC | $121,100 \mathrm{t}$ | $125,500 \mathrm{t}$ |
| Overfishing | $144,000 \mathrm{t}$ | $149,600 \mathrm{t}$ |
| $\mathrm{F}_{\text {ABC }}$ | $\mathrm{F}_{0.40}=0.144$ | $\mathrm{~F}_{0.40}=0.15$ |
| $\mathrm{~F}_{\text {overfishing }}$ | $\mathrm{F}_{0.35}=0.174$ | $\mathrm{~F}_{0.35}=0.18$ |
| $\mathrm{~B}_{40 \%}$ | $222,000 \mathrm{t}$ | $228,400 \mathrm{t}$ |
| $\mathrm{B}_{35 \%}$ | $194,300 \mathrm{t}$ | $199,800 \mathrm{t}$ |

## SSC comments from December 2005

The SSC looks forward to seeing the results of the management strategy evaluation to explore the consequences of a non-stationary spawner-recruit relationship.

See Tier 1 considerations section
The age-structured assessment model uses combined sex data but the size composition data shown in the figures suggests sexual dimorphism in growth and sex ratios that differ from 50:50. If there is sexual dimorphism in growth, then size-based selection in the fisheries will generate time-variations in sex ratios that can have important consequences to the stock's productivity. The SSC requests that the authors evaluate whether sex ratios differ from 50:50 and if there have been trends in sex ratio.

Northern rock sole exhibit sexually explicit differences in growth. Instead of implementing a split sex stock assessment model, the weight at age for males and females combined is calculated as the average of their sex-specific weight for each age. Male and female northern rock sole have the same weight-at-age from the juvenile stage until they become sexually mature (age of $50 \%$ maturity $=9$ years, see figure below). After maturation, when the weights at age diverge, the average is appropriate to calculate population biomass because males and females are found in nearly equal numbers in the shelf trawl surveys (see table below). However, a split sex model is a consideration to improve modeling the population dynamics of males and females at ages older than the age at maturation.

Northern Rock Sole


Average weight at age of northern rock sole, by sex, in the population from 2003-2005.

Proportion of male northern rock sole in the population estimated from the past 10 shelf surveys.

| year | Proportion <br> male |
| :---: | :---: |
| 1997 | 0.50 |
| 1998 | 0.51 |
| 1999 | 0.54 |
| 2000 | 0.47 |
| 2001 | 0.50 |
| 2002 | 0.47 |
| 2003 | 0.51 |
| 2004 | 0.52 |
| 2005 | 0.54 |
| 2006 | 0.52 |

## INTRODUCTION

Northern rock sole (Lepidopsetta polyxystra n. sp.) are distributed primarily on the eastern Bering Sea continental shelf and in much lesser amounts in the Aleutian Islands region. Two species of rock sole are known to occur in the North Pacific ocean, a northern rock sole (L. polyxystra) and a southern rock sole (L. bilineata) (Orr and Matarese 2000). These species have an overlapping distribution in the Gulf of Alaska, but the northern species comprise the majority of the Bering Sea and Aleutian Islands populations where they are managed as a single stock.

Centers of abundance occur off the Kamchatka Peninsula (Shubnikov and Lisovenko 1964), British Columbia (Forrester and Thompson 1969), the central Gulf of Alaska, and in the southeastern Bering Sea (Alton and Sample 1975). Adults exhibit a benthic lifestyle and occupy separate winter (spawning) and summertime feeding distributions on the southeastern Bering Sea continental shelf. Northern rock sole spawn during the winter-early spring period of December-March.

## CATCH HISTORY

Rock sole catches increased from an average of 7,000 $t$ annually from 1963-69 to 30,000 t between 1970 1975. Catches ( t ) since implementation of the MFCMA in 1977 are shown in Table 7.1, with catch data for 1980-88 separated into catches by non-U.S. fisheries; joint venture operations and DAP catches (where available). Prior to 1987, the classification of rock sole in the "other flatfish" management category prevented reliable estimates of DAP catch. Catches from 1989-2005 (domestic only) have averaged $48,175 \mathrm{t}$ annually. The size composition of the 2006 catch from observer sampling, by sex and management area, are shown in Figure 7.1 and the locations of the 2006 catch are presented for each month in the Appendix.

Rock sole are important as the target of a high value roe fishery occurring in February and March which accounts for the majority of the annual catch ( $62 \%$ in 2006). About $58 \%$ of the 2006 catch came from management areas 509 and 513 with the rest from areas 513,517 and 521. The 2006 catch of $35,907 \mathrm{t}$ comprised $29 \%$ of the ABC of $126,000 \mathrm{t}$ ( $89 \%$ of the TAC). Thus, rock sole remain lightly harvested in the Bering Sea and Aleutian Islands.

During the 2006 fishing season rock sole harvesting was temporarily closed in the Bering Sea and Aleutian Islands due to halibut bycatch restrictions on February 21 and April 12 (first and second seasonal apportionments were obtained). On August 7 directed rock sole harvesting was closed due to the attainment of the annual halibut bycatch allowance, after which the species could only be retained as bycatch.

Although female rock sole are highly desirable when in spawning condition, large amounts of rock sole are discarded overboard in the various Bering Sea trawl target fisheries. Estimates of retained and discarded catch from at-sea sampling for 1987-2005 are shown in Table 7.2. From 1987 to 2000 rock sole were discarded in greater amounts than they were retained, however the past five years there has been increased utilization of the catch. Fisheries with the highest discard amounts include the rock sole roe fishery, the yellowfin sole fishery and the Pacific cod fisheries (shown for 2004 and 2005 in Table 7.3).

## DATA

The data used in this assessment include estimates of total catch, trawl fishery catch-at-age, trawl survey age composition, trawl survey biomass estimates and sampling error, maturity observations from observer sampling and mean weight-at-age.

## Fishery Catch and Catch-at-Age

Available information include fishery total catch data from 1975-September 6, 2006 (Table 7.1) and fishery catch-at-age numbers from 1980-2005 (Table 7.4).

## Survey CPUE

Since rock sole are lightly exploited and are often taken incidentally in target fisheries for other species, CPUE from commercial fisheries are considered an unreliable method for detecting trends in abundance. It is therefore necessary to use research vessel survey data to assess the condition of these stocks.

Abundance estimates from the 1982 AFSC survey were substantially higher than from the 1981 survey data for a number of bottom-tending species such as flatfishes. This is coincident with the change in research trawl to the $83 / 112$ with better bottom tending characteristics. The increase in survey CPUE was particularly large for rock sole ( 6.5 to $12.3 \mathrm{~kg} / \mathrm{ha}$, Figure 7.2 ). Allowing the stock assessment model to fit these early survey estimates would most likely underestimate the true pre-1982 biomass, thus exaggerating the degree to which biomass increased during that period. Consequently, CPUE and biomass from the 1975-81 surveys are not used in the assessment model.

The CPUE trend indicates a significantly increasing population from 1982-92 when the mean CPUE more than tripled. The population leveled-off from 1994-98 when CPUE values indicated a high level of abundance. The 1999 value of $36.5 \mathrm{~kg} / \mathrm{ha}$ was the lowest observed since 1992 , possibly due to extremely low water temperatures. Since that time the value has been stable with a 2006 value of $47.8 \mathrm{~kg} / \mathrm{ha}$..

## Absolute Abundance

Estimates of rock sole biomass are also estimated from the AFSC surveys using stratified area-swept expansion of the CPUE data (Table 7.5). It should be recognized that these biomass estimates are point estimates from an "area-swept" bottom trawl survey. As a result they are uncertain. It is assumed that the sampling plan covers the distribution of the fish and that all fish in the path of the footrope of the trawl are captured. That is, there are no losses due to escape or gains due to gear herding effects. Due to sampling variability alone, the $95 \%$ confidence interval for the 2006 point estimate of the Bering Sea surveyed area is $1,918,800 \mathrm{t}-2,512,600 \mathrm{t}$.

Rock sole biomass was relatively stable through 1979 , but then increased substantially in the following years to $799,300 \mathrm{t}$ in 1984. In 1985 the estimate declined to $700,000 \mathrm{t}$ but increased again in 1986 to over 1 million $t$ and continued this trend through 1988. The 1989 and 1990 estimates were at a high and stable level (slightly less than the 1988 estimate) and continued to increase to the highest levels estimated by the trawl survey at 2.9 million metric tons in 1994 and 2.7 million $t$ in 1997. With the exception of the cold year in 1999 when all flatfish biomass estimates declined, the biomass estimates from the trawl survey have exhibited a stable trend since 1997.

The 2006 Aleutian Islands biomass estimate of $77,751 \mathrm{t}$ is $3 \%$ of the combined BSAI total. Since it is such a low proportion of the total biomass for this area, the Aleutian Islands biomass is not used in this assessment.

## Weight-at-age and Maturity-at-age

In conjunction with the large and steady increase in the rock sole stock size in the early 1980s, it was found that there was also a corresponding decrease in size-at-age for both sexes (Figure 7.3). This also caused a resultant decrease in weight-at-age as the population increased and expanded westward toward the shelf edge (Walters and Wilderbuer 2000). These updated values of weight-at-age (Table 7.6) were also applied to the populations in 2001-2006 to model the population dynamics of the rock sole population.

The length-weight relationship did not change significantly over this time period as discerned from an analysis of observations made in 1975, 1976 and 1988. The following parameters have been calculated for the length ( cm )-weight ( g ) relationship:

$$
\mathrm{W}=\mathrm{a} * \mathrm{~L}^{\mathrm{b}}
$$

No significant differences were found between sexes so that these parameters are for both sexes combined.

$$
\begin{array}{cc}
\underline{\mathrm{a}} & \underline{\mathrm{~b}} \\
0.007610 & 3.11976
\end{array}
$$

Maturity information available from anatomical scans collected by fishery observers during the 1993 and 1994 Bering Sea rock sole roe fishery are used in this assessment (Table 7.7). These data indicate that the age of $50 \%$ maturity occurs at $9-10$ years for female rock sole.

## Survey and Fishery Age composition

Rock sole otoliths have routinely been collected during the trawl surveys since 1979 to provide estimates of the population age composition (Fig. 7.4, Table 7.8). Fishery size composition data from 1980-97 (prior to 1980 observer coverage was sparse and did not reflect the catch size composition) were applied to age-length keys from these surveys to provide a time-series of catch-at-age assuming that the mean length at age from the trawl survey was the same as the fishery in a given year. Estimation of the fishery age composition since 1997 use age-length keys derived from age structures collected annually from the fishery.

## ANALYTIC APPROACH

## Model Structure

The abundance, mortality, recruitment and selectivity of rock sole were assessed with a stock assessment model using the AD Model builder software. The conceptual model is a separable catch-age analysis that uses survey estimates of biomass and age composition as auxiliary information (Fournier and Archibald 1982). The model simulates the dynamics of the population and compares the expected values of the population characteristics to the characteristics observed from surveys and fishery sampling programs. This is accomplished by the simultaneous estimation of the parameters in the model using the maximum
likelihood estimation procedure. The fit of the simulated values to the observable characteristics is optimized by maximizing a $\log$ (likelihood) function given some distributional assumptions about the data.

The parameters estimated in the stock assessment model are classified by three likelihood components:

## $\underline{\text { Data Component }}$

Trawl fishery catch-at-age
Trawl survey population age composition
Trawl survey biomass estimates and S.E.

## Distribution assumption

Multinomial
Multinomial
Log normal

The total log likelihood is the sum of the likelihoods for each data component (Table 7-9). The likelihood components may be weighted by an emphasis factor, however, equal emphasis was placed on fitting each likelihood component in the rock sole assessment except for the catch weight. The AD Model Builder software fits the data components using automatic differentiation (Griewank and Corliss 1991) software developed as a set of libraries (AUTODIFF C++ library). Table $7-9$ presents the key equations used to model the rock sole population dynamics in the Bering Sea and Table 7-10 provides a description of the variables used in Table 7-9. The model of rock sole population dynamics was evaluated with respect to the observations of the time-series of survey and fishery age compositions and the survey biomass trend since 1982, and estimates of natural mortality and catchability.

## Parameters Estimated Independently

Rock sole maturity schedules were estimated independently as discussed in a previous section (Table 7.7) as were length at age and length-weight relationships.

## Parameters Estimated Conditionally

The parameters estimated by the model are presented below:

| Fishing <br> mortality | Selectivity | Year class <br> strength | Spawner- <br> recruit | catchability | M | Total |
| :---: | :---: | :--- | :--- | :---: | :---: | :---: |
| 32 | 4 | 51 | 2 | 1 | 1 | 91 |

The increase in the number of parameters estimated in this assessment compared to last year can be accounted for by the input of another year of fishery data and the entry of another year class into the observed population.

## Year class strengths

The population simulation specifies the numbers-at-age in the beginning year of the simulation, the number of recruits in each subsequent year, and the survival rate for each cohort as it progresses through the population using the population dynamics equations given in Table 7-9.

## Selectivity

Fishery and survey selectivity were modeled in this assessment using the logistic function, as shown in Table 7-9. The model was configured with the selectivity curve constrained to provide an asymptotic fit for the older fish in the fishery and survey, but still was allowed to estimate the shape of the logistic curve for young fish. The oldest year classes in the surveys and fisheries were truncated at 20 and allowed to accumulate into the age category $20+$ years.

## Fishing Mortality

The fishing mortality rates ( F ) for each age and year are calculated to approximate the catch weight by solving for F while still allowing for observation error in catch measurement. A large emphasis (300) was placed on the catch likelihood component.

## Natural Mortality

Assessments for rock sole in other areas assume $\mathrm{M}=0.20$ for rock sole on the basis of the longevity of the species. In a past BSAI assessment, the stock synthesis model was used to entertain a range of M values to evaluate the fit of the observable population characteristics over a range of natural mortality values (Wilderbuer and Walters 1992). The best fit occurred at $\mathrm{M}=0.18$ with the survey catchability coefficient (q) set equal to 1.0 . Since that time twelve more years of fishery and survey age composition data have become available as well as experimental estimates of catchability. This allows for natural mortality to be estimated as a conditional parameter in this assessment.

## Survey Catchability

Unusually low estimates of flatfish biomass were obtained for Bering Sea shelf flatfish species during the very cold year of 1999. These results suggest a relationship between bottom water temperature and trawl survey catchabililty, which has been documented for yellowfin sole and arrowtooth flounder in a recent BSAI SAFE document. To better understand how water temperature may affect the catchability of rock sole to the survey trawl, we estimated catchability in a linear model for each year within the stock assessment model as:

$$
q=\alpha+\beta T
$$

where q is catchability, T is the average annual bottom water temperature at survey stations less than 100 m , and $\alpha$ and $\beta$ are parameters estimated by the model. The model estimated values of $\alpha$ and $\beta$ at 1.77 and 0.021 , respectively. The small value for $\beta$ indicates that temperature has very little effect on trawl catchability of rock sole and the value of 1.77 obtained for $\alpha$ suggests that survey catchability $(\mathrm{q})$ is greater than 1.0, the value used in earlier assessments.

Experiments conducted in recent years on the standard research trawl used in the annual trawl surveys indicate that rock sole are herded by the bridles (in contact with the seafloor) from the area outside the net mouth into the trawl path (Somerton and Munro 2001). Rock sole survey trawl catchability was estimated at 1.4 from these experiments which indicate that the standard area-swept biomass estimate from the survey is an overestimate of the rock sole population biomass.

These experimental results, in combination with the results of the bottom temperature analysis above, provided a compelling reason to consider an alternative model where survey catchability is estimated. As in past assessments we use the value of $q$ from the herding experiment to constrain survey catchabliity and then estimate survey catchabililty as follows:

$$
\text { qlike }=0.5\left[\frac{q_{\exp }-q_{\bmod }}{\sigma_{\exp }}\right]^{2}
$$

where qlike is the survey catchability likelihood component, $\mathrm{q}_{\text {mod }}$ is the survey catchability parameter estimated by the model, $\mathrm{q}_{\exp }$ is the estimate of area-swept q from the herding experiment, and $\sigma$ is the standard error of the experimental estimate of $q$.

## Natural Mortality

With catchability constrained as described above, natural mortality was estimated as a free parameter. The best fit to the total $\log$ likelihood occurred at $M=0.156(q=1.52)$, slightly lower than the value of 0.16 estimated last year. To gain a better understanding of how changes in M affect the fits to the observed population characteristics (likelihood components), M was fixed at values ranging from 0.1 to 0.2. The log likelihood of the data components and the total log likelihood from these runs are shown below and the posterior probability distributions for M and q (from the model run with the best fit) are shown in Figure 7.5

|  | $\mathrm{M}=0.2$ | $\mathrm{M}=0.18$ | $\mathrm{M}=0.156$ | $\mathrm{M}=0.14$ | $\mathrm{M}=0.12$ | $\mathrm{M}=0.1$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Survey biomass likelihood | 67.043 | 52.846 | 42.875 | 42.397 | 51.028 | 69.59 |
| Catch likelihood | . 00108 | 0.000931 | 0.0014 | 0.0024 | 0.00512 | 0.0096 |
| Catch age comp likelihood | 682.141 | 674.723 | 669.257 | 667.717 | 667.717 | 669.342 |
| Survey age comp likelihood | 392.868 | 386.388 | 386.915 | 391.447 | 401.735 | 414.209 |
| Recruitment likelihood | 79.457 | 77.831 | 75.923 | 74.776 | 73.342 | 72.178 |
| q likelihood | 0.629 | 0.0798 | 2.933 | 7.234 | 16.68 | 29.352 |
| q estimate | 1.34 | 1.42 | 1.52 | 1.61 | 1.723 | 1.83 |
| Ending biomass | 1733.67 | 1647.29 | 1582.63 | 1561.21 | 1559.73 | 1559.73 |
| Total likelihood | 1222.14 | 1191.869 | 1177.904 | 1183.573 | 1210.508 | 1254.683 |

## Model Evaluation

The probability of M or q being different than the estimated values from the model run with $\mathrm{M}=0.156$ and $\mathrm{q}=0.152$ declines sharply within plus or minus 0.5 of the estimated value (Figure 7.5) indicating that these values are fairly well estimated, given the data. These estimates also provide the best model fit to the observable data. Therefore, a natural mortality rate of $0.156(q=0.152)$ is used in this assessment.

## MODEL RESULTS

## Fishing Mortality and Selectivity

The assessment model estimates of the annual fishing mortality on fully selected ages and the estimated annual exploitation rates (catch/total biomass) are given Table 7.11. The exploitation rate has averaged $3.6 \%$ from 1975-2005, indicating a lightly exploited stock. Age-specific selectivity estimated by the model (Table 7.12, Fig. 7.6) indicate that rock sole are $50 \%$ selected by the fishery at age of 8 and are nearly fully selected by age 13 (sexes combined).

## Abundance Trend

The stock assessment model indicates that rock sole total biomass was at low levels during the mid 1970s through 1982 (160,000-330,000 t, Fig. 7.6 and Table 7.13). From 1985-95, a period characterized by sustained above-average recruitment (1980-88 year classes, Fig. 7.6) and light exploitation, the estimated total biomass rapidly increased at a high rate to over 1.8 million $t$ by 1995 . Since then, the model indicates the population biomass declined $25 \%$ to 1.41 million $t$ in 2003 before increasing the past three years to 1.58 million $t$. The decline from 1995-2003 was attributable to the below average recruitment to the adult portion of the population during the 1990 s . The increase the past three years is the result of increased recruitment in 2000-2003. The female spawning biomass is estimated to be at a high, but slowly declining level of $416,000 \mathrm{t}$ in 2006 (Table 7.13 ). The model provides good fits to most of the strong year classes observed in the fishery and surveys during the time-series. These are shown in the Appendix with the model estimates of population numbers at age.

The model estimates of survey biomass (using trawl survey age-specific selectivity and the estimate of q applied to the total biomass, Fig. 7.6) correspond fairly well with the trawl survey biomass trend with the exception of the cold year of 1999. The model fits the 2000,2002 and 2006 survey estimates but does not match the higher estimates from the 2001and 2003-2005 surveys. Although 2006 was a relatively cold year in the eastern Bering Sea, the rock sole biomass estimate increased indicating the lack of a relationship between survey catchability and bottom temperatures, as shown for other flatfish species. Both the trawl survey and the model indicate the same increasing biomass trend from the late 1970s to the mid 1990s but the survey does not indicate the declining trend after the mid 1990s that the model estimates.

## Total Biomass

The stock assessment model estimates of total biomass (begin year population numbers multiplied by mid-year weight at age) is used to recommend the ABC for 2007. Including the 2006 catch of $35,907 \mathrm{t}$ through 6 September (including discards), the model projects the total biomass for 2007 at $\mathbf{1 , 6 7 4 , 0 0 0} \mathbf{t}$.

Recruitment Trends

Increases in abundance for rock sole during the 1980s can be attributed to the recruitment of a series of strong year classes (Figs. 7.4 and 7.6, Table 7.14). Rock sole ages have now been read for samples obtained in 2005 and show that the 1990 year class, which are 15 year old fish in 2005, comprise a significant part ( $11 \%$ ) of the survey and fishery age composition numbers. The 1987 year class is the largest estimated during the recruitment time-series and still comprise $7 \%$ of the estimated 2005 survey age composition numbers as eighteen year old fish. Recruitment during the 1990s, with the exception of the 1990 year class, was below the 27 year average but has recently improved as the 2001-2003 year classes appear much stronger as discerned from the 2004 and 2005 survey age samples.

## Tier 1 Considerations

The SSC has requested that flatfish assessments which have a lengthy time-series of stock and recruitment estimates explore management under a Tier 1 harvest policy. In the case of rock sole, the time series of recruitment estimates from this assessment is 28 years. MSY is an equilibrium concept and it's calculated value is dependent on both the spawner-recruit data, which we assume represents the equilibrium stock size-recruitment relationship, and the model used to fit the data. In the stock assessment model used here, a Ricker form of the stock-recruit relationship was fit to these data and estimates of $\mathrm{F}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$ (female spawning biomass) were calculated, assuming that the fit to the stock-recruitment data points represent the long-term productivity of the stock. However, very different estimates of $\mathrm{F}_{\text {MSY }}$ and $\mathrm{B}_{\text {MSY }}$ were obtained, depending on which years of stock-recruitment data points were included in the fitting procedure. Fitting the full time series since the regime shift in 1977 (19782000) gives values of $\mathbf{F}_{\mathrm{MSY}}=\mathbf{0 . 3 9}, \mathrm{B}_{\mathrm{MSY}}=\mathbf{1 3 9 , 3 5 0} \mathbf{t}$, and $\mathbf{M S Y}=\mathbf{1 2 8 , 7 0 0} \mathbf{t}$.

A recent analysis of flatfish recruitment give compelling evidence that temporal trends in winter spawning flatfish production in the Eastern Bering Sea are consistent with the hypothesis that decadal scale climate variability influences marine survival during the early life history period (Wilderbuer et al. 2002). Periods of estimated cross-shelf advection of flatfish larvae was found to coincide with synchronous above-average recruitment (1980s) whereas periods of weak advection or advection to the west were associated with poor recruitment (1990s) (Fig 7.7 in the 2003 SAFE). This trend has continued in recent years where strong-cross shelf advection in 2001-2003 again coincided with strong recruitment (see ecosystem consideration chapter in this SAFE report). These changes in stock productivity were found to coincide with a decadal scale shift in atmospheric forcing. When the spawner-recruit information from the 1978-89 (productive) period was fit, estimates were obtained as follows: $\mathbf{F}_{\text {MSY }}=\mathbf{0 . 2 8}$, $\mathbf{B}_{\text {MSY }}=3.72 \mathrm{E}+\mathbf{1 2} \mathbf{t}$, and MSY $=5.85 \mathrm{E}+\mathbf{1 2} \mathbf{t}$. These estimates are clearly unrealistic and unreliable and only result from 12 observations (estimates).

This exercise of fitting spawner-recruit observations calls into concern whether a single fit of stock recruitment time-series data is able to reliably capture the long-term reproductive potential of the rock sole stock, particularly given the length of the time-series and the stock dynamics which have occurred since 1975. The aforementioned analysis was performed for rock sole, arrowtooth flounder and flathead sole, species which spawn in the winter in offshore areas and are seemingly reliant upon advection to nursery areas 3-4 months later. The atmospheric forcing responsible for the advection properties during this time period appears to be the location of the springtime signature of the Aleutian Low Pressure field. Anomalous sea level pressure implies that westerly to south-westerly surface winds (on-shelf) predominated during 1977-1988, whereas during 1989-96 easterly (off-shelf) winds were predominate. These shifts in recruitment production may be a cause of concern if we assume that the long term productivity is closely related to only spawning stock size while ignoring mechanisms governing the variability in production which may correspond to decadal (or longer) shifts in environmental conditions.

Given these concerns, a management strategy simulation study was performed to determine how robust the tier 1 harvest strategy calculations are when fitting the full time series of spawner recruit estimates for a fish stock experiencing temporal changes in reproductive potential due to changing ocean conditions. The simulation study was set up with an operating model which simulated 60 future years of stock and recruitment where a new productivity regime occurred every 15 years alternating between high and low productivity as described above and shown in Figure 7.7 (yellowfin sole $\mathrm{s} / \mathrm{r}$ from two productivity regimes were used in this simulation). A simulated survey value was produced for each year which incorporated the variability from the changing recruitment productivity schedule. Similarly, survey and fishery age composition "observations" were input into the model for each year. The stock assessment model was then run for each year inside the operating model simulation and re-estimated the spawner recruit time-series (adding a new point each year), fit the Ricker form of the stock recruitment curve to the entire time-series, and calculated MSY and the harmonic mean of $\mathrm{F}_{\text {MSY }}$ (tier 1 calculations) to set the harvest for the next year. One thousand replicates were made for each year and the results were averaged to compare the "known" population, biomass and recruitment values with those estimated by the stock assessment model. Results indicate a consistent underestimate of the "true" recruitment and spawning biomass by the stock assessment model throughout the 60 year simulation, regardless of the productivity state (Figure 7.8). Thus the Tier 1 harvest control strategy, although it does not explicitly consider environmental change, appears to be robust to underlying changes in stock productivity.

Results from the previous Tier 1 calculations for rock sole indicate that the harmonic mean of the $\mathrm{F}_{\text {MSY }}$ estimate is very close to the geometric mean value of the $\mathrm{F}_{\text {MSY }}$ estimate due to the low variability in the parameter estimates. This indicates that the previous analysis was performed with very little uncertainty. To better understand how uncertainty in specific parameter estimates affects the Tier 1 harvest policy calculations for rock sole, the following analysis was undertaken. Selectivity, catchability and $M$ were selected as important parameters whose uncertainty may directly affect the pdf of the estimate of $\mathrm{F}_{\text {MSY }}$. Eleven different model configurations were chosen to illustrate the effect of a range of uncertainly in these parameter estimates (varying from small to large ( $0.03,0.4$ and 0.8 ) ) and how they affect the estimate of the harmonic mean of $\mathrm{F}_{\mathrm{MSY}}$.

The analysis provided the following results (Table 7.15). The values of $\mathrm{F}_{\text {MSY }}, \mathrm{B}_{\text {MSY }}$ and MSY are dependent on the years of stock size and recruitment selected to be fit by the model (Models 1-3). Using the full time-series (1978-2001, Model 1, Fig. 7.9) to fit the spawner-recruit curve indicates that the rock sole stock is most productive at a smaller stock size with the result that the $\mathrm{F}_{\text {MSY }}$ value is more than twice as high as the $\mathrm{F}_{40 \%}$. value (recall that $\mathrm{F}_{40 \%}=0.156$ ). When the 1989-2001 years are fit (Model 2), the $\mathrm{F}_{\text {MSY }}$ value is less than twice the $\mathrm{F}_{40 \%}$ value but only 12 data points are used to discern this relationship. Model 3 fit the 1978-88 data and resulted in an unrealistic high value for $\mathrm{B}_{\mathrm{MSY}}(4.24 \mathrm{E}+13)$. Using the estimates of recruitment and stock size from 1978-2001 as the basis for the spawner-recruit relationship (Model 1), uncertainty was introduced for the estimates of selectivity (Models 4and 5), catchability (Models 6 and 7) and natural mortality (Models 8 and 9). Model 10 incorporated high uncertainty in estimating q, M and selectivity. Adding uncertainty to selectivity resulted in the largest difference between the geometric mean and the harmonic mean of the estimate of $\mathrm{F}_{\text {MSY }}$ for these Model runs, but the introduced uncertainty only resulted in a $5 \%$ reduction. Similarly, the addition of uncertainty in estimating catchability and natural mortality resulted in a $1-2 \%$ reduction for the estimate of the harmonic mean (Models 6-9). Thus $\mathrm{F}_{\text {MSY }}$ appears to be well estimated by the model. The posterior distributions of $\mathrm{F}_{\text {MSY }}$ from the 10 model runs are shown in the Appendix.

ACCEPTABLE BIOLOGICAL CATCH

The reference fishing mortality rate for rock sole is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Equilibrium female spawning biomass for Tier 3 is calculated by applying the female spawning biomass per recruit resulting from a constant $\mathrm{F}_{0.40}$ harvest to an estimate of average equilibrium recruitment. For this assessment, year classes spawned in 1977 through 2001 are used to calculate the average equilibrium recruitment. This results in an estimate of $\mathbf{B}_{\mathbf{0 . 4 0}}=\mathbf{2 2 2 , 0 0 0} \mathbf{t}$. The stock assessment model estimates the 2007 level of female spawning biomass at $\mathbf{3 9 2 , 0 0 0} \mathbf{t}(\mathrm{B})$. Since reliable estimates of $\mathrm{B}, \mathrm{B}_{0.40}, \mathrm{~F}_{0.40}$, and $\mathrm{F}_{0.30}$ exist and $\mathrm{B}>\mathrm{B}_{0.40}(392,000>222,000$, fig. 7.6), rock sole reference fishing mortality can be defined in tier 3a. For the 2007 harvest: $\mathrm{F}_{\mathrm{ABC}} \bullet \mathrm{F}_{0.40}=0.144$ and $\mathrm{F}_{\text {overfishing }}=\mathrm{F}_{0.35}$ $=0.174$ (full selection F values).

The Tier 3 acceptable biological catch is estimated for 2007 by applying the $\mathrm{F}_{0.40}$ fishing mortality rate and age-specific fishery selectivities to the 2007 estimate of age-specific total biomass as follows:

$$
A B C=\sum_{a=a_{r}}^{a_{\text {nages }}} \bar{w}_{a} n_{a}\left(1-e^{-M-F s_{a}}\right) \frac{F s_{a}}{M+F s_{a}}
$$

where $S_{a}$ is the selectivity at age, $M$ is natural mortality, $W_{a}$ is the mean weight at age determined from recent surveys, and $n_{a}$ is the beginning of the year numbers at age. This results in a Tier 3a 2007 ABC of $\mathbf{1 2 1 , 1 0 0} t$ for the eastern Bering Sea portion of the stock.

The stock assessment analysis must also consider harvest limits, usually described as "overfishing" fishing mortality levels with corresponding yield amounts. Amendment 56 to the BS/AI FMP now sets the harvest limit at the $\mathrm{F}_{0.35}$ fishing mortality value. The overfishing fishing mortality value, ABC fishing mortality value and their corresponding yields are given as follows:

| Harvest level |  | F value |  |
| :--- | :--- | :--- | :--- |
|  |  |  |  |
|  |  |  |  |
| $\mathrm{F}_{0.35}$ | 0.174 |  | $144,000 \mathrm{t}$ Yield |
| $\mathrm{F}_{0.40}$ |  | 0.144 | $121,100 \mathrm{t}$ |

The time series of rock sole fishing mortality rates and female spawning biomass relative to $\mathrm{B}_{40 \%}$ and $\mathrm{F}_{40 \%}$ are shown in figure 7.8.

Alternatively, ABC can be calculated using Tier 1 methodology depending on whether the SSC determines that rock sole are in Tier 1 or Tier 3. It is critical for the Tier 1 calculations to identify which subset of the stock recruitment data is used. Using the full time series to fit the spawner recruit curve estimates that the stock is most productive at a small stock size. Thus MSY and $\mathrm{F}_{\text {MSY }}$ are high values and $\mathrm{B}_{\mathrm{MSY}}$ is a low value. If the stock was productive in the past at a small stock size because of non density dependent factors (environment), reducing the stock size to low levels could be detrimental to the longterm sustainability of the stock if the environment has changed from the earlier period. Since observations of rock sole recruitment at low stock sizes are not available from multiple time periods, it is uncertain if future recruitment events at low stock conditions would be as productive as during the late 1970s when the stock is estimated to have been at very low levels. The alternative data set to consider would be the model which uses 1989-2001, a period of high productivity, but it results in values of MSY, $\mathrm{F}_{\mathrm{MSY}}$ and $\mathrm{B}_{\mathrm{MSY}}$ that are similar to the full data set, but fits only 12 data points.

Therefore it seems acceptable to select the 1978-2001 data set for the Tier 1 harvest recommendation (Model 1 in Table 7.15) where $\mathrm{F}_{\text {harmonic mean }}=0.383$ which gives a Tier 1 ABC harvest recommendation of 197,600 $\mathbf{t}$ and OFL value of $199,800 \mathrm{t}$

Depending on which stock recruitment subset is used for the Tier 1 calculations, significantly different stock recruitment relationships are found. These results illustrate the non-stationarity of stockrecruitment relationships for Bering Sea rock sole and bring into question whether a single stock recruit curve can adequately define the dynamics of the stock. Therefore, this assessment recommends retaining rock sole in Tier 3.

## BIOMASS PROJECTIONS

As in past years, a standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2006 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2007 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2006. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2007, are as follows ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC , so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2007 recommended in the assessment to the $\max F_{A B C}$ for 2006. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $75 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2002-2006 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{\text {TAC }}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above $1 / 2$ of its MSY level in 2007 and above its MSY level in 2017 under this scenario, then the stock is not overfished.)

Scenario 7: In 2007 and 2008, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2019 under this scenario, then the stock is not approaching an overfished condition.)

Simulation results shown in Table 7.16 indicate that rock sole are currently not overfished and are not approaching an overfished condition. If harvested at the average F from 2002-2006, rock sole female spawning biomass is projected to decline through 2007 due to the reduced recruitment observed during the 1990s (fig. 7.9), but slowly increase thereafter.

## Scenario Projections and Two-Year Ahead Overfishing Level

In addition to the seven standard harvest scenarios, Amendments $48 / 48$ to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2007, it does not provide the best estimate of OFL for 2008, because the mean 2008 catch under Scenario 6 is predicated on the 2007 catch being equal to the 2007 OFL, whereas the actual 2007 catch will likely be less than the 2007 ABC. Therefore, the projection model was re-run with the 2007 catch fixed equal to the 2006 catch and the 2008 fishing mortality rate fixed at $\mathrm{F}_{\mathrm{ABC}}$.

| Tier 3a |  |  |  |
| :---: | :---: | :---: | :---: |
| Year | Catch | ABC | OFL |
| 2007 | 35,907 | 121,100 | 144,400 |
| 2008 | 35,907 | 127,600 | 152,000 |

## Tier 1

| Year | Catch | ABC | OFL |
| :---: | :---: | :---: | :---: |
| 2007 | 35,900 | 197,600 | 199,800 |
| 2008 | 35,900 | 268,400 | 271,400 |

## ECOSYSTEM CONSIDERATIONS

## Ecosystem Effects on the stock

1) Prey availability/abundance trends

Rock sole diet by life stage varies as follows: Larvae consume plankton and algae, early juveniles consume zooplankton, late juvenile stage and adults prey includes bivalves, polychaetes, amphipods, mollusks and miscellaneous crustaceans. Information is not available to assess the abundance trends of
the benthic infauna of the Bering Sea shelf. The original description of infaunal distribution and abundance by Haflinger (1981) resulted from sampling conducted in 1975 and 1976 and has not be resampled since. The large populations of flatfish which have occupied the middle shelf of the Bering Sea over the past twenty years for summertime feeding do not appear food-limited. These populations have fluctuated due to the variability in recruitment success which suggests that the primary infaunal food source has been at an adequate level to sustain the rock sole resource.

## 2) Predator population trends

As juveniles, it is well-documented from studies in other parts of the world that flatfish are prey for shrimp species in near shore areas. This has not been reported for Bering Sea rock sole due to a lack of juvenile sampling and collections in near shore areas, but is thought to occur. As late juveniles they are found in stomachs of pollock, Pacific cod, yellowfin sole, skates and Pacific halibut; mostly on small rock sole ranging from 5 to 15 cm standard length..

Past, present and projected future population trends of these predator species can be found in their respective SAFE chapters in this volume. Encounters between rock sole and their predators may be limited as their distributions do not completely overlap in space and time.

## 3) Changes in habitat quality

Changes in the physical environment which may affect rock sole distribution patterns, recruitment success, migration timing and patterns are catalogued in the Ecosystem Considerations Appendix of this SAFE report. Habitat quality may be enhanced during years of favorable cross-shelf advection (juvenile survival) and warmer bottom water temperatures with reduced ice cover (higher metabolism with more active feeding).

## Fishery Effects on the ecosystem

1) The rock sole target fishery contribution to the total bycatch of other non-prohibited species is shown for 1991-2005 in Table 7.17. The rock sole target fishery contribution to the total bycatch of prohibited species is shown for 2003 and 2004 in Table 13 of the Economic SAFE (Appendix C) and is summarized for 2004 as follows:

2) Relative to the predator needs in space and time, the rock sole target fishery is not very selective for fish between $5-15 \mathrm{~cm}$ and therefore has minimal overlap with removals from predation.
3) The target fishery is not perceived to have an effect on the amount of large size target fish in the population due to the history of very light exploitation (3\%) over the past 28 years.
4) Rock sole fishery discards are presented in the Catch History section.
5) It is unknown what effect the fishery has had on rock sole maturity-at-age and fecundity.
6) Analysis of the benthic disturbance from the rock sole fishery is available in the Preliminary draft of the Essential Fish Habitat environmental Impact Statement.

## Ecosystem effects on rock sole

| Indicator | Observation | Interpretation | Evaluation |
| :--- | :---: | :---: | :---: |
| Prey availability or abundance trends <br> Benthic infauna |  |  |  |
|  | Stomach contents | Stable, data limited | Unknown |

## Predator population trends

| Fish (Pollock, Pacific cod, <br> halibut, yellowfin sole, skates) | Stable | Possible increases to rock <br> sole mortality |  |
| :--- | :--- | :--- | :--- |
| Changes in habitat quality <br> Temperature regime | Cold years rock sole <br> catchability and herding may <br> decrease | Likely to affect surveyed <br> stock | No concern (dealt <br> with in model) |
| Winter-spring environmental <br> conditions | Affects pre-recruit survival | Probably a number of <br> factors | Causes natural <br> variability |
| Pock sole effects on ecosystem |  | Observation | Interpretation |

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Table 7.1--Rock sole catch ( t ) from 1977 - September 3, 2006.

| Year | Foreign | Joint-Venture | Domestic | Total |
| :---: | :---: | :---: | :---: | :---: |
| 1977 | 5,319 |  |  | 5,319 |
| 1978 | 7,038 |  |  | 7,038 |
| 1979 | 5,874 |  |  | 5,874 |
| 1980 | 6,329 | 2,469 |  | 8,798 |
| 1981 | 3,480 | 5,541 |  | 9,021 |
| 1982 | 3,169 | 8,674 |  | 11,843 |
| 1983 | 4,479 | 9,140 |  | 13,619 |
| 1984 | 10,156 | 27,523 |  | 37,679 |
| 1985 | 6,671 | 12,079 |  | 18,750 |
| 1986 | 3,394 | 16,217 |  | 19,611 |
| 1987 | 776 | 11,136 | 28,910 | 40,822 |
| 1988 |  | 40,844 | 45,522 | 86,366 |
| 1989 |  | 21,010 | 47,902 | 68,912 |
| 1990 |  | 10,492 | 24,761 | 35,253 |
| 1991 |  |  | 60,587 | 60,587 |
| 1992 |  |  | 56,998 | 56,998 |
| 1993 |  |  | 63,953 | 63,953 |
| 1994 |  |  | 59,606 | 59,606 |
| 1995 |  |  | 58,870 | 58,870 |
| 1996 |  |  | 46,928 | 46,928 |
| 1997 |  |  | 67,564 | 67,564 |
| 1998 |  |  | 33,642 | 33,642 |
| 1999 |  |  | 40,510 | 40,510 |
| 2000 |  |  | 49,264 | 49,264 |
| 2001 |  |  | 29,255 | 29,255 |
| 2002 |  |  | 41,331 | 41,331 |
| 2003 |  |  | 35,395 | 35,395 |
| 2004 |  |  | 47,637 | 47,637 |
| 2005 |  |  | 35,546 | 35,456 |
| 2006 |  |  | 35,907 | 35,907 |

Table 7.2 Retained and discarded catch ( t ) in Bering Sea fisheries, 1987-2005.

| Year | Retained (t) | Discarded (t) | \% Retained |
| :---: | :---: | :---: | :---: |
|  |  |  |  |
| 1987 | 14,209 | 14,701 | 49 |
| 1988 | 22,374 | 23,148 | 49 |
| 1989 | 23,544 | 24,358 | 49 |
| 1990 | 12,170 | 12,591 | 49 |
| 1991 | 25,406 | 35,181 | 42 |
| 1992 | 21,317 | 35,681 | 37 |
| 1993 | 22,589 | 45,669 | 33 |
| 1994 | 20,951 | 39,945 | 34 |
| 1995 | 21,761 | 33,108 | 40 |
| 1996 | 19,770 | 27,158 | 42 |
| 1997 | 27,743 | 39,821 | 41 |
| 1998 | 12,645 | 20,999 | 38 |
| 1999 | 15,224 | 25,286 | 38 |
| 2000 | 22,151 | 27,113 | 45 |
| 2001 | 19,299 | 9,956 | 66 |
| 2002 | 23,607 | 17,724 | 57 |
| 2003 | 19,492 | 15,903 | 55 |
| 2004 | 26,600 | 21,037 | 56 |
| 2005 | 23,172 | 12,376 | 65 |

Table 7.3--Discarded and retained rock sole catch ( t ), by target fishery, in 2004 and 2005.

| 2004 |  |  |  |
| :---: | :---: | :---: | :---: |
| target fishery | Retained | Discarded | total |
| Atka mackerel | 36 | 113 | 149 |
| Bottom pollock | 209 | 38 | 248 |
| Pacific cod | 2,601 | 6,563 | 9,163 |
| Mid-water pollock | 1,330 | 924 | 2,254 |
| Sablefish | 1 | 0 | 1 |
| Rockfish | 4 | 3 | 7 |
| Arrowtooth flounder | 16 | 31 | 48 |
| Flathead sole | 999 | 982 | 1,981 |
| Rock sole | 15,655 | 8,138 | 23,793 |
| Yellowfin sole | 5,696 | 4,167 | 9,863 |
| Greenland turbot | 0 | 1 | 1 |
| Other flatfish | 51 | 39 | 91 |
| Other species | 0 | 37 | 38 |
| Total catch |  |  | 47,637 |
| 2005 |  |  |  |
|  | Retained | Discarded | Total |
| Atka mackerel | 81 | 69 | 151 |
| Bottom pollock | 52 | 28 | 80 |
| Pacific cod | 2,778 | 4,787 | 7,565 |
| Mid-water pollock | 491 | 499 | 990 |
| Sablefish | 1 | 0 | 1 |
| Rockfish | 0 | 2 | 2 |
| Arrowtooth flounder | 101 | 36 | 136 |
| Flathead sole | 570 | 545 | 1,114 |
| Rock sole | 13,300 | 2,559 | 15,858 |
| Yellowfin sole | 5,779 | 3,817 | 9,596 |
| Greenland turbot | 0 | 0 | 0 |
| Other flatfish | 18 | 32 | 51 |
| Other species | 0 | 0 | 0 |
| halibut | 0 | 2 | 2 |
| Total catch |  |  | 35,546 |


Table 7.5 Bottom trawl survey biomass estimates ( t$)$ from the Eastern Bering Sea shelf and the Aleutian Islands for northern rock sole.

| year | Bering Sea | Aleutians |
| :---: | :---: | :---: |
| 1975 | 175,500 |  |
| 1979 | 194,700 |  |
| 1980 | 283,800 | 28,500 |
| 1981 | 302,400 |  |
| 1982 | 578,800 |  |
| 1983 | 713,000 | 23,300 |
| 1984 | 799,300 |  |
| 1985 | 700,100 |  |
| 1986 | $1,031,400$ | 26,900 |
| 1987 | $1,269,700$ |  |
| 1988 | $1,480,100$ |  |
| 1989 | $1,138,600$ |  |
| 1990 | $1,381,300$ |  |
| 1991 | $1,588,300$ | 37,325 |
| 1992 | $1,543,900$ |  |
| 1993 | $2,123,500$ |  |
| 1994 | $2,894,200$ | 54,785 |
| 1995 | $2,175,040$ |  |
| 1996 | $2,183,000$ |  |
| 1997 | $2,710,900$ | 56,154 |
| 1998 | $2,168,700$ |  |
| 1999 | $1,689,100$ |  |
| 2000 | $2,127,700$ | 45,949 |
| 2001 | $2,135,400$ |  |
| 2002 | $1,921,400$ | 57,700 |
| 2003 | $2,424,800$ |  |
| 2004 | $2,182,100$ | 63,900 |
| 2005 | $2,119,100$ |  |
| 2006 | $2,215,670$ | 77,751 |

Table 7－6－－Rock sole weight－at－age（grams）by age and year determined from 1980－2000 from length－at－age and length－weight relationships from

| － | $\underset{\sim}{N}$ | N | $\stackrel{\text { N }}{\sim}$ | $\stackrel{\sim}{\infty}$ | N | $\stackrel{\text { ¢ }}{ }$ | $\stackrel{\sim}{2}$ | \％ | $\stackrel{7}{6}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | ${ }_{\sim}^{\infty}$ | $\stackrel{\text { ® }}{ }$ | ल | G | $\stackrel{2}{2}$ | N | \％ | 亿 | 2 | 亿 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\bigcirc$ | $\stackrel{ \pm}{\text { ¢ }}$ | $\stackrel{\text { J }}{\sim}$ | $\stackrel{\text { N }}{\sim}$ | $\stackrel{\sim}{0}$ | N | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{2}$ | \％ | $\underset{\sim}{7}$ | ƠO | $\stackrel{\sim}{\sim}$ | N | $\stackrel{\text { ® }}{ }$ | ल | G | N | $\stackrel{\bigcirc}{\wedge}$ | \％ | 亿 | 亿 | \％ |
| $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{2}$ | $\stackrel{\text { ¹ }}{\sim}$ | ¢ | N | \％ | $\stackrel{\sim}{2}$ | $\stackrel{\downarrow}{\sim}$ | กิ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | \％ | $\stackrel{\text { ® }}{ }$ | $\overline{0}$ | $\bigcirc$ | N | $\hat{0}$ | \％ | 亿 | 2 | 亿 |
| ， | N | N | $\stackrel{\text { N }}{\sim}$ | $\cdots$ | N | in | $\cdots$ | $\stackrel{\rightharpoonup}{\sim}$ | ？ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{n}$ | $\cdots$ | $\stackrel{\text { ® }}{ }$ | $\overline{6}$ | ন | $\underset{\infty}{\infty}$ | $\stackrel{\sim}{\sim}$ | ¢ | $\stackrel{\square}{4}$ | \％ | $\stackrel{+}{6}$ |
| $\stackrel{\sim}{\square}$ | $\stackrel{\sim}{\gtrless}$ | $\stackrel{\sim}{\gtrless}$ | $\stackrel{\bigcirc}{\sim}$ | n | $\stackrel{\infty}{\infty}$ | in | \％ | ® | $\stackrel{\imath}{6}$ | bib | $\cdots$ | $\cdots$ | $\begin{aligned} & \circ \\ & \text { in } \end{aligned}$ | $\overline{0}$ | ì | $\underset{\infty}{\infty}$ | $\frac{0}{6}$ | へ̀ | त̀ | in | ì |
| $\cdots$ | $\stackrel{\sim}{6}$ | ה | N | $\stackrel{\infty}{\infty}$ | $\stackrel{\circ}{7}$ | \％ | $\frac{0}{6}$ | $\stackrel{\infty}{+}$ | $\bar{\square}$ | $\stackrel{\square}{n}$ | ¢ | \％ | $\cdots$ | ot | $\cdots$ | N | $\stackrel{ \pm}{n}$ | in | N | $\stackrel{N}{\text { in }}$ | in |
| $\pm$ | ¢ | \％ | $\stackrel{ \pm}{4}$ | n | $\stackrel{\infty}{\text { O}}$ | $\stackrel{\text {－}}{*}$ | No | $\frac{\mathrm{N}}{6}$ | $\frac{n}{0}$ | n | oి | 亏 | n | i | i | $\stackrel{\circ}{\gtrless}$ | $\cdots$ | $\frac{m}{n}$ | $\cdots$ | $\cdots$ | $m$ |
| $\cdots$ | $\frac{m}{6}$ | $\frac{m}{6}$ | N | $\cdots$ | $\stackrel{\circ}{\text { ¢ }}$ | in | 年 | $\stackrel{\text { m }}{\sim}$ | $\stackrel{\infty}{\circ}$ | $\cdots$ | $\cdots$ | ¢ | N | $\cdots$ | $\frac{\pi}{n}$ | 읏 | $\stackrel{+}{\square}$ | \％ | $\stackrel{\sim}{7}$ | $\stackrel{\text { ？}}{7}$ | $\underset{F}{F}$ |
| N | N | N | $\frac{n}{7}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{\circ}{7}$ | \％ | $\stackrel{\infty}{+}$ | N | $\begin{aligned} & \text { n } \\ & i \end{aligned}$ | $\frac{n}{n}$ | in | $\stackrel{7}{\sim}$ | $\underset{\downarrow}{\infty}$ | $\underset{\sim}{\text { - }}$ | $\underset{+}{+}$ | $8$ | $\stackrel{+}{4}$ | $\underset{\sim}{\infty}$ | $\underset{\sim}{\infty}$ | N | $\underset{\sim}{\infty}$ |
| こ | $\stackrel{\sim}{n}$ | $\stackrel{\sim}{n}$ | $\stackrel{\circ}{+}$ | ¢ | $\stackrel{\odot}{+}$ | ¢ | $\frac{n}{\tau}$ | $\stackrel{\infty}{\text { ¢ }}$ | $\stackrel{ \pm}{\sim}$ | へ | $\stackrel{\sim}{\sim}$ | \％ | \％ | $\stackrel{\circ}{7}$ | ¢ | $\underset{F}{\mathcal{F}}$ | $\stackrel{\infty}{\ni}$ | $\cdots$ | en | n | en |
|  | $\underset{N}{N}$ | $\underset{i}{N}$ | $\frac{\pi}{7}$ | ¢ | $\stackrel{8}{7}$ | \％ | $\underset{+}{\infty}$ | N | $\stackrel{\imath}{n}$ | $\frac{n}{n}$ | $\stackrel{\sim}{n}$ | ¢ | $\underset{+}{\infty}$ | 氐 | $\stackrel{+}{+}$ | $8$ | $\stackrel{ \pm}{4}$ | $\underset{\sim}{\infty}$ | $\underset{\sim}{\infty}$ | $\underset{\sim}{\text { ® }}$ | $\underset{\sim}{\text { ® }}$ |
|  | N | N | $\stackrel{\circ}{+}$ | $\stackrel{8}{+}$ | $\stackrel{\odot}{+}$ | $\stackrel{\infty}{+}$ | $\stackrel{\sim}{\gamma}$ | $\stackrel{\infty}{\text { ¢ }}$ | $\stackrel{ \pm}{\text { N }}$ | へ | $\stackrel{\sim}{\sim}$ | ๙ | \％ | $\stackrel{\circ}{\circ}$ | $\hat{O}$ | $\underset{F}{\mathcal{F}}$ | $\frac{\infty}{\ni}$ | $\cdots$ | en | ¢ | $\cdots$ |
|  | ＊ | ন | $\stackrel{\sim}{\sim}$ | $\stackrel{\hat{p}}{n}$ | $\stackrel{ \pm}{7}$ | \％ | ๆ | フ | in | $\stackrel{\infty}{n}$ | $\stackrel{\Re}{f}$ | $\underset{\sim}{\sim}$ | $\underset{\sim}{\text { o }}$ | － | $\stackrel{\infty}{\stackrel{\infty}{n}}$ | $\frac{\infty}{\ni}$ | へ | $\frac{\infty}{m}$ | $\frac{\infty}{m}$ | $\stackrel{\infty}{m}$ | $\stackrel{\infty}{m}$ |
|  | oे | ） | $\underset{\sim}{\infty}$ | n | nిల | $\cdots$ | $\underset{\sim}{\infty}$ | $\stackrel{\circ}{\square}$ | $\frac{n}{7}$ |  | ） | $\underset{\sim}{\aleph}$ | － | $\stackrel{ \pm}{m}$ | İ | $\underset{\sim}{\underset{\sim}{2}}$ | N |  | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\circ}{\sim}$ |
|  | ＋ | m | 入 | 눈 | ৪ | N | $\stackrel{\circ}{4}$ | $\stackrel{\text { F }}{ }$ | 8 | $\underset{m}{n}$ | \％ | $\stackrel{\circ}{\text { N}}$ | ले | 入 | N | Ǹ | $\stackrel{\sim}{\circ}$ | N | Nิ | N | N |
|  | ત | ત | $\stackrel{\sim}{n}$ | $\stackrel{\rightharpoonup}{\sim}$ | $\stackrel{\infty}{\sim}$ | U | $\hat{N}$ | m | No | $\stackrel{\downarrow}{\text { ¢ }}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{-}$ | $\stackrel{\infty}{\infty}$ | $\stackrel{ \pm}{-}$ | $\stackrel{\bigcirc}{\stackrel{-}{-}}$ | N | $\stackrel{\otimes}{\sim}$ | $\stackrel{\infty}{\perp}$ | $\stackrel{\infty}{\perp}$ | $\stackrel{\infty}{\sim}$ | $\stackrel{\infty}{\sim}$ |
|  | Nิ | ત̀ | $\pm$ | $\stackrel{\sim}{2}$ | $\stackrel{\infty}{\infty}$ | ） | $\stackrel{\infty}{\circ}$ | $\stackrel{\sim}{\sim}$ | ત̀ | $\stackrel{ \pm}{-}$ | － | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{0}$ | $\pm$ | 三 | $\underset{\sim}{2}$ | $\cdots$ | 은 | 이 | 으 | 인 |
|  | $\stackrel{\sim}{n}$ | $\cdots$ | $\bigcirc$ | $\bigcirc$ | Ј | $\stackrel{\text { n }}{=}$ | $\bigcirc$ | 윽 | N | $\stackrel{\circ}{\circ}$ | $\infty$ | N | U | $\infty$ | ล | 2 | $\infty$ | $\infty$ | $\infty$ | $\infty$ | $\infty$ |
|  | $\stackrel{\square}{\sim}$ | $\stackrel{\bigcirc}{\sim}$ | － | 2 | G | $\stackrel{\sim}{\sim}$ | $\stackrel{\circ}{7}$ | $\underset{\sim}{~}$ | n | $\stackrel{\infty}{\sim}$ | F | m | m | $\cdots$ | i | $\stackrel{\text { \％}}{ }$ | n | ช | 9 | ¢ | 子 |
|  | m | m | $\stackrel{\sim}{\circ}$ | $\cdots$ | － | N | N | e | ते | へ | － | － | $\cdots$ | $\stackrel{\rightharpoonup}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{ \pm}{\sim}$ | $\cdots$ | $\cdots$ | $\cdots$ | $\cdots$ |
|  | $\bigcirc$ | $\bigcirc$ | $\stackrel{\sim}{\sim}$ | N | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\sim$ | － | $\checkmark$ | $\bigcirc$ | $\stackrel{\sim}{2}$ | $\bigcirc$ | $a$ | N | $\sim$ | $\infty$ | $\infty$ | $\infty$ | $\infty$ | $\infty$ |
|  | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ |
|  | $\stackrel{\circ}{\circ}$ | $\stackrel{-}{\circ}$ | $\stackrel{\sim}{\circ}$ | $\stackrel{\infty}{\circ}$ | $\stackrel{+}{\circ}$ | $\stackrel{\infty}{\infty}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{N}{\infty}$ | $\begin{aligned} & \infty \\ & \stackrel{\infty}{\circ} \end{aligned}$ | $\stackrel{\otimes}{\circ}$ | $8$ | Ј | $\underset{\sigma}{\mathrm{N}}$ | $\grave{\sigma}$ | す | $\stackrel{\sim}{2}$ | $\stackrel{\circ}{2}$ | 人 | $\stackrel{\infty}{\circ}$ | ล2 | 8 |

Table 7-7.--Mean length-at-age (cm) and proportion mature for female Bering Sea rock sole from observer anatomical scans during the 1993-94

| Age | Length-at-age | Proportion mature |
| :---: | :---: | :---: |
| 1 | 6.7 | 0 |
| 2 | 10.8 | 0.006 |
| 3 | 15.4 | 0.003 |
| 4 | 23.6 | 0.012 |
| 5 | 27.1 | 0.039 |
| 6 | 30.1 | 0.098 |
| 7 | 32.6 | 0.198 |
| 8 | 34.6 | 0.330 |
| 9 | 36.4 | 0.470 |
| 10 | 37.8 | 0.590 |
| 11 | 39.0 | 0.680 |
| 12 | 40.0 | 0.746 |
| 13 | 40.8 | 0.795 |
| 14 | 41.5 | 0.830 |
| 15 | 42.1 | 0.856 |
| 16 | 41.6 | 0.875 |
| 17 | 43.0 | 0.889 |
| 18 | 43.4 | 0.900 |
| 19 | 43.7 | 0.908 |
| 20 | 44.0 | 0.915 |
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Table 7.9--Key equations used in the population dynamics model.

$$
\begin{array}{ll}
N_{t, 1}=R_{t}=R_{0} e^{\tau_{t}}, \tau_{t} \sim N\left(0, \delta^{2}\right) & \text { Recruitment 1956-75 } \\
N_{t, 1}=R_{t}=R_{\gamma} e^{\tau_{t}}, \tau_{t} \sim N\left(0, \delta^{2}{ }_{R}\right) & \text { Recruitment 1976-96 } \\
C_{t, a}=\frac{F_{t, a}}{Z_{t, a}}\left(1-e^{-z_{t, a}}\right) N_{t, a} & \text { Catch in year } t \text { for age } a \text { fish } \\
N_{t+1, a+1}=N_{t, a} e^{-z_{t, a}} & \text { Numbers of fish in year } t+1 \text { at age } a \\
N_{t+1, A}=N_{t, A-1} e^{-z_{t, A-1}}+N_{t, A} e^{-z_{t, A}} & \text { Numbers of fish in the "plus group" } \\
S_{t}=\sum N_{t, a} W_{t, a} \phi_{a} & \text { Spawning biomass } \\
Z_{t, a}=F_{t, a}+M & \text { Total mortality in year } t \text { at age } a \\
F_{t, a}=s_{a} \mu^{F} e^{e x p} \varepsilon^{\varepsilon_{t}} & \varepsilon^{F}{ }_{t} \sim N\left(o, \sigma^{2}\right) \\
S_{a}=\frac{\text { Fishing mortality }}{1+\left(e^{-\alpha+\beta a}\right)} & \\
C_{t}=\sum C_{t, a} & \text { Age-specific fishing selectivity } \\
P_{t, a}=C_{t, a} / C_{t} & \text { Total catch in numbers } \\
& \text { Proportion at age in catch }
\end{array}
$$

$$
\operatorname{SurB}_{t}=q \sum N_{t, a} W_{t, a} v_{a}
$$

Survey biomass

$$
L=\sum_{t, a} m_{t} p_{t, a} \ln \frac{\hat{p_{t, a}}}{p_{t, a}}+(-0.5) \sum_{t}\left[\left(\ln \frac{\operatorname{surB}_{t}}{\hat{\operatorname{sur} B_{t}}} 1 / \sigma_{t}\right)^{2}-\ln \sigma_{t}\right]
$$

Total log likelihood

Table 7.10--Variables used in the population dynamics model.

| Variables |  |
| :---: | :--- |
| $R_{t}$ | Age 1 recruitment in year $t$ |
| $R_{0}$ | Geometric mean value of age 1 recruitment, 1956-75 |
| $R_{\gamma}$ | Geometric mean value of age 1 recruitment, 1976-96 |
| $\tau_{t}$ | Recruitment deviation in year $t$ |
| $N_{t, a}$ | Number of fish in year $t$ at age $a$ |
| $C_{t, a}$ | Catch numbers of fish in year $t$ at age $a$ |
| $P_{t, a}$ | Proportion of the numbers of fish age $a$ in year $t$ |
| $C_{t}$ | Total catch numbers in year $t$ |
| $W_{t, a}$ | Mean body weight (kg) of fish age $a$ in year $t$ |
| $\phi_{a}$ | Proportion of mature females at age $a$ |
| $F_{t, a}$ | Instantaneous annual fishing mortality of age $a$ fish in year $t$ |
| M | Instantaneous natural mortality, assumed constant over all ages and years |
| $Z_{t, a}$ | Instantaneous total mortality for age $a$ fish in year $t$ |
| $s_{a}$ | Age-specific fishing gear selectivity |
| $\mu^{F}$ | Median year-effect of fishing mortality |
| $\varepsilon_{t}^{F}$ | The residual year-effect of fishing mortality |
| $v_{a}$ | Age-specific survey selectivity |
| $\alpha$ | Slope parameter in the logistic selectivity equation |
| $\beta$ | Age at 50\% selectivity parameter in the logistic selectivity equation |
| $\sigma_{t}$ | Standard error of the survey biomass in year $t$ |

Table 7.11--Model estimates of rock sole fishing mortality and exploitation rate (catch/total biomass).

| year | Full selection <br> $\mathbf{F}$ | Exploitation <br> rate |
| :---: | :---: | :---: |
| 1975 | 0.183 | 0.075 |
| 1976 | 0.137 | 0.059 |
| 1977 | 0.063 | 0.029 |
| 1978 | 0.072 | 0.035 |
| 1979 | 0.054 | 0.026 |
| 1980 | 0.074 | 0.034 |
| 1981 | 0.070 | 0.030 |
| 1982 | 0.098 | 0.036 |
| 1983 | 0.096 | 0.032 |
| 1984 | 0.244 | 0.078 |
| 1985 | 0.101 | 0.033 |
| 1986 | 0.085 | 0.028 |
| 1987 | 0.131 | 0.042 |
| 1988 | 0.236 | 0.078 |
| 1989 | 0.164 | 0.058 |
| 1990 | 0.070 | 0.029 |
| 1991 | 0.108 | 0.048 |
| 1992 | 0.096 | 0.044 |
| 1993 | 0.087 | 0.041 |
| 1994 | 0.072 | 0.036 |
| 1995 | 0.056 | 0.032 |
| 1996 | 0.044 | 0.027 |
| 1997 | 0.062 | 0.040 |
| 1998 | 0.029 | 0.020 |
| 1999 | 0.035 | 0.025 |
| 2000 | 0.041 | 0.031 |
| 2001 | 0.025 | 0.019 |
| 2002 | 0.036 | 0.028 |
| 2003 | 0.033 | 0.025 |
| 2004 | 0.047 | 0.033 |
| 2005 | 0.037 | 0.024 |
| 2006 |  | 0.023 |
|  |  |  |

Table 7.12 --Model estimates of rock sole age-specific fishery and survey selectivities.

| Age | Fishery (1980- <br> $\mathbf{2 0 0 5})$ | Survey (1982- <br> 2005) |
| :---: | :---: | :---: |
| 1 |  |  |
| 2 | 0.00 | 0.01 |
| 3 | 0.00 | 0.07 |
| 4 | 0.01 | 0.31 |
| 5 | 0.03 | 0.73 |
| 6 | 0.07 | 0.94 |
| 7 | 0.15 | 0.99 |
| 8 | 0.30 | 1.00 |
| 9 | 0.51 | 1.00 |
| 10 | 0.72 | 1.00 |
| 11 | 0.86 | 1.00 |
| 12 | 0.94 | 1.00 |
| 13 | 0.97 | 1.00 |
| 14 | 0.99 | 1.00 |
| 15 | 0.99 | 1.00 |
| 16 | 0.99 | 1.00 |
| 17 | 0.99 | 1.00 |
| 18 | 0.99 | 1.00 |
| 19 | 0.99 | 1.00 |
| 20 | 0.99 | 1.00 |

Table 7-13.--Model estimates of rock sole age $2+$ total biomass ( t ) and female spawning biomass $(\mathrm{t})$ from the 2005 and 2006 assessments.

|  | 2006 Assessment |  | 2005 Assessment |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Age 2+ Total biomass | Female Spawning biomass | Age 2+ Total biomass | Female Spawning biomass |
| 1975 | 160,817 | 27,848 | 166,861 | 28,832 |
| 1976 | 167,951 | 30,076 | 174,053 | 31,134 |
| 1977 | 178,839 | 33,742 | 184,969 | 34,870 |
| 1978 | 200,139 | 39,175 | 206,361 | 40,343 |
| 1979 | 225,700 | 43,817 | 232,018 | 44,974 |
| 1980 | 259,326 | 48,558 | 265,684 | 49,662 |
| 1981 | 297,975 | 52,852 | 304,269 | 53,879 |
| 1982 | 332,107 | 50,108 | 338,300 | 50,906 |
| 1983 | 430,971 | 58,123 | 437,949 | 58,848 |
| 1984 | 480,610 | 67,142 | 487,474 | 67,731 |
| 1985 | 562,836 | 75,108 | 570,795 | 75,550 |
| 1986 | 707,629 | 91,285 | 717,432 | 91,506 |
| 1987 | 970,799 | 124,181 | 980,323 | 122,955 |
| 1988 | 1,100,660 | 151,037 | 1,129,070 | 153,761 |
| 1989 | 1,193,690 | 168,745 | 1,250,680 | 178,239 |
| 1990 | 1,196,070 | 194,195 | 1,275,390 | 211,603 |
| 1991 | 1,274,350 | 223,071 | 1,324,550 | 233,235 |
| 1992 | 1,300,510 | 236,212 | 1,351,000 | 246,572 |
| 1993 | 1,544,380 | 296,538 | 1,590,060 | 305,326 |
| 1994 | 1,633,760 | 330,768 | 1,682,820 | 341,168 |
| 1995 | 1,828,790 | 427,806 | 1,882,750 | 441,283 |
| 1996 | 1,746,090 | 424,934 | 1,801,140 | 439,875 |
| 1997 | 1,674,820 | 447,120 | 1,699,090 | 453,401 |
| 1998 | 1,645,320 | 474,604 | 1,698,000 | 491,680 |
| 1999 | 1,597,300 | 494,626 | 1,632,550 | 507,024 |
| 2000 | 1,571,670 | 512,966 | 1,589,040 | 519,762 |
| 2001 | 1,521,480 | 517,297 | 1,550,680 | 529,498 |
| 2002 | 1,458,840 | 507,435 | 1,464,900 | 512,040 |
| 2003 | 1,406,200 | 480,927 | 1,412,190 | 482,592 |
| 2004 | 1,432,080 | 463,318 | 1,434,280 | 462,144 |
| 2005 | 1,502,560 | 443,759 | 1,489,600 | 439,752 |
| 2006 | 1,582,630 | 416,440 |  |  |

Table 7.14--Estimated age 4 recruitment of rock sole (thousands of fish) from the 2005and 2006 assessments.

| Year <br> class | 2006 <br> Assessment | $\mathbf{2 0 0 5}$ <br> Assessment |
| :---: | :---: | :---: |
| 1971 | 98,808 | 103,394 |
| 1972 | 81,991 | 85,659 |
| 1973 | 110,582 | 115,024 |
| 1974 | 152,144 | 157,486 |
| 1975 | 402,202 | 414,711 |
| 1976 | 227,282 | 233,752 |
| 1977 | 344,873 | 353,351 |
| 1978 | 395,694 | 403,688 |
| 1979 | 504,458 | 513,323 |
| 1980 | 972,309 | 991,091 |
| 1981 | 977,751 | $1,004,120$ |
| 1982 | 849,637 | 880,221 |
| 1983 | $1,504,390$ | $1,567,640$ |
| 1984 | $1,213,520$ | $1,269,970$ |
| 1985 | $1,205,890$ | $1,265,990$ |
| 1986 | $1,947,190$ | $2,039,130$ |
| 1987 | $3,324,650$ | $3,476,670$ |
| 1988 | $1,229,230$ | $1,269,080$ |
| 1989 | 889,434 | 908,157 |
| 1990 | $1,980,720$ | $2,065,460$ |
| 1991 | 930,758 | 942,572 |
| 1992 | 482,322 | 492,312 |
| 1993 | 839,262 | 884,285 |
| 1994 | 408,186 | 414,131 |
| 1995 | 405,161 | 378,866 |
| 1996 | 560,748 | 576,643 |
| 1997 | 263,701 | 256,290 |
| 1998 | 438,617 | 388,800 |
| 1999 | 528,293 | 547,629 |
| 2000 | $1,216,100$ |  |
| 2001 | $2,607,450$ |  |
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Table 7.15. Results of the northern rock sole Tier 1 analysis from 10 models that use different levels of uncertainty in the estimates of fishery selectivity, natural mortality and catchability. Values that change between runs are highlighted.

|  | Years <br> used in <br> S/R fit | Selectivity <br> CV | q <br> sigma | M <br> sigma | F $_{\text {MSY }}$ | Harmonic <br> mean of <br> FMSY |
| :---: | :--- | :---: | :---: | :---: | :---: | :---: |
| Model 1 | $1978-2001$ | 0.03 | 0.056 | 0.05 | 0.389 | 0.383 |
| Model 2 | $1989-2001$ | 0.03 | 0.056 | 0.05 | 0.259 | 0.236 |
| Model 3 | $1978-1988$ | 0.03 | 0.056 | 0.05 | 0.296 | 0.292 |
| Model 4 | $1978-2001$ | 0.4 | 0.056 | 0.05 | 0.389 | 0.380 |
| Model 5 | $1978-2001$ | 0.8 | 0.056 | 0.05 | 0.389 | 0.371 |
| Model 6 | $1978-2001$ | 0.03 | 0.4 | 0.05 | 0.409 | 0.403 |
| Model 7 | $1978-2001$ | 0.03 | 0.8 | 0.05 | 0.410 | 0.404 |
| Model 8 | $1978-2001$ | 0.03 | 0.056 | 0.3 | 0.389 | 0.383 |
| Model 9 | $1978-2001$ | 0.03 | 0.056 | 0.8 | 0.389 | 0.383 |
| Model | $1978-2001$ | 0.8 | 0.8 | 0.8 | 0.410 | 0.392 |
| 10 |  |  |  |  |  |  |

Table 7.16--Projections of rock sole female spawning biomass $(1,000 \mathrm{~s} t)$, future catch $(1,000 \mathrm{~s} t)$ and full selection fishing mortality rates for seven future harvest scenarios. 2006 ABC is highlighted.

| Scenarios 1 and 2 <br> Maximum ABC harvest permissible <br> Female <br> Year |  |  |  |
| :--- | :---: | :---: | :---: |
| spwn bio | catch | F |  |
| 2006 | 409.799 | 35.900 | 0.040 |
| 2007 | 391.995 | 121.100 | 0.144 |
| 2008 | 370.523 | 117.853 | 0.144 |
| 2009 | 368.124 | 122.110 | 0.144 |
| 2010 | 378.314 | 129.935 | 0.144 |
| 2011 | 389.427 | 135.435 | 0.144 |
| 2012 | 391.953 | 135.495 | 0.144 |
| 2013 | 380.380 | 129.328 | 0.144 |
| 2014 | 361.696 | 120.888 | 0.144 |
| 2015 | 347.258 | 114.815 | 0.144 |
| 2016 | 334.407 | 110.020 | 0.144 |
| 2017 | 318.685 | 104.910 | 0.144 |
| 2018 | 301.192 | 99.756 | 0.144 |
| 2019 | 292.282 | 96.737 | 0.144 |

Scenario 4
Harvest at average F over the past 5 years

## Female

| Year | spwn bio | catch | F |
| :---: | ---: | :---: | ---: |
| 2006 | 409.799 | 35.900 | 0.040 |
| 2007 | 394.841 | 42.100 | 0.048 |
| 2008 | 402.762 | 62.796 | 0.069 |
| 2009 | 417.972 | 67.517 | 0.069 |
| 2010 | 444.325 | 74.019 | 0.069 |
| 2011 | 473.133 | 79.783 | 0.069 |
| 2012 | 493.855 | 82.891 | 0.069 |
| 2013 | 499.836 | 82.568 | 0.069 |
| 2014 | 496.153 | 80.540 | 0.069 |
| 2015 | 495.130 | 79.323 | 0.069 |
| 2016 | 492.678 | 78.260 | 0.069 |
| 2017 | 479.671 | 75.915 | 0.069 |
| 2018 | 460.520 | 72.872 | 0.069 |
| 2019 | 452.602 | 71.574 | 0.069 |

Scenario 3
1/2 Maximum ABC harvest permissible
Female

| Year | spwn bio | catch | F |
| :---: | ---: | :---: | :---: |
| 2006 | 409.799 | 35.900 | 0.040 |
| 2007 | 394.195 | 60.550 | 0.070 |
| 2008 | 396.874 | 24.445 | 0.027 |
| 2009 | 426.340 | 27.144 | 0.027 |
| 2010 | 466.779 | 30.591 | 0.027 |
| 2011 | 511.034 | 33.876 | 0.027 |
| 2012 | 548.241 | 36.169 | 0.027 |
| 2013 | 570.615 | 37.035 | 0.027 |
| 2014 | 582.281 | 37.096 | 0.027 |
| 2015 | 595.952 | 37.405 | 0.027 |
| 2016 | 606.284 | 37.647 | 0.027 |
| 2017 | 601.363 | 37.115 | 0.027 |
| 2018 | 586.366 | 36.093 | 0.027 |
| 2019 | 584.789 | 35.897 | 0.027 |

Scenario 5
No fishing
Female

| Year | spwn bio | catch | F |
| :---: | :---: | :---: | :---: |
| 2006 | 409.799 | 0 | 0 |
| 2007 | 396.276 | 0 | 0 |
| 2008 | 421.079 | 0 | 0 |
| 2009 | 458.148 | 0 | 0 |
| 2010 | 506.550 | 0 | 0 |
| 2011 | 560.312 | 0 | 0 |
| 2012 | 607.885 | 0 | 0 |
| 2013 | 641.272 | 0 | 0 |
| 2014 | 663.967 | 0 | 0 |
| 2015 | 688.732 | 0 | 0 |
| 2016 | 709.266 | 0 | 0 |
| 2017 | 711.232 | 0 | 0 |
| 2018 | 700.412 | 0 | 0 |
| 2019 | 705.191 | 0 | 0 |

Table 7.16-continued.

| Scenario 6 |  |  |  |
| :---: | :---: | :---: | :---: |
| Determination of whether rock sole are currently overfished $B 35=194.3$ |  |  |  |
|  | Female |  |  |
| Year | spwn bio | catch | F |
| 2006 | 409.799 | 35.900 | 0.040 |
| 2007 | 391.114 | 144.372 | 0.174 |
| 2008 | 360.949 | 137.469 | 0.174 |
| 2009 | 352.326 | 140.307 | 0.174 |
| 2010 | 357.249 | 147.533 | 0.174 |
| 2011 | 362.923 | 151.774 | 0.174 |
| 2012 | 360.225 | 149.668 | 0.174 |
| 2013 | 344.320 | 140.708 | 0.174 |
| 2014 | 322.569 | 129.701 | 0.174 |
| 2015 | 306.083 | 121.807 | 0.174 |
| 2016 | 292.176 | 115.948 | 0.174 |
| 2017 | 276.897 | 110.122 | 0.174 |
| 2018 | 261.066 | 104.472 | 0.174 |
| 2019 | 252.788 | 98.635 | 0.169 |


| Scenario 7 <br> Determination of whether rock sole are <br> approaching <br> an overfished <br> condition | Female | B35=194.3 |  |
| :--- | :---: | :---: | :---: |
| Year | spwn bio | catch | F |
| 2006 | 409.799 | 35.900 | 0.040 |
| 2007 | 391.994 | 121.110 | 0.144 |
| 2008 | 370.526 | 117.853 | 0.144 |
| 2009 | 367.392 | 145.788 | 0.174 |
| 2010 | 369.708 | 152.113 | 0.174 |
| 2011 | 373.432 | 155.670 | 0.174 |
| 2012 | 369.296 | 153.071 | 0.174 |
| 2013 | 352.970 | 144.024 | 0.174 |
| 2014 | 331.286 | 133.224 | 0.174 |
| 2015 | 315.065 | 125.662 | 0.174 |
| 2016 | 301.609 | 119.915 | 0.174 |
| 2017 | 285.503 | 112.460 | 0.171 |
| 2018 | 269.487 | 104.216 | 0.166 |
| 2019 | 261.707 | 100.051 | 0.163 |

Table 7.17-Catch and bycatch in the rock sole target fisheries, 1991-2004, from blend of regional office reported catch and observer sampling.

| Species | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003* | 2004 | 2005 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Walleye Pollock | 9,711 | 9,825 | 18,583 | 15,784 | 7,766 | 7,698 | 9,123 | 3,955 | 5,207 | 5,481 | 4,577 | 9,942 | 4,643 | 8,937 | 7,240 |
| Arrowtooth Flounder | 254 | 473 | 1,143 | 1,782 | 507 | 1,341 | 411 | 300 | 69 | 216 | 835 | 314 | 419 | 346 | 599 |
| Pacific Cod | 4,262 | 4,651 | 8,160 | 6,358 | 9,796 | 6,965 | 8,947 | 3,529 | 3,316 | 4,219 | 3,391 | 4,366 | 3,195 | 5,648 | 5,192 |
| Groundfish, General | 1,693 | 3,000 | 3,091 | 3,266 | 1,605 | 1,581 | 1,381 | 909 | 537 | 1,186 | 1,198 | 692 | 978 | 801 |  |
| Rock Sole | 22,067 | 24,873 | 39,857 | 40,139 | 29,241 | 18,380 | 32,477 | 13,092 | 16,047 | 29,042 | 14,437 | 20,168 | 18,681 | 24,287 | 16,667 |
| Flathead Sole |  |  | 2,140 | 1,702 | 1,147 | 1,302 | 2,373 | 1,223 | 575 | 1,806 | 1,051 | 771 | 744 | 881 | 850 |
| Sablefish | 9 | 0 | 4 | 16 | 3 | 3 | 1 | 0 | 2 | 5 | 12 | 4 | 2 | 9 | 4 |
| Atka Mackerel | 3 | 10 | 15 | 0 |  | 0 | 0 | 9 | 0 | 38 | 3 | 0 | 1 | 16 | 48 |
| Pacific Ocean Perch | 37 | 10 | 15 | 62 | 4 | 2 |  | 1 | 0 | 0 | 0 | 0 |  |  |  |
| Rex Sole |  |  | 79 | 145 | 108 | 48 | 11 | 12 | 5 | 4 | 18 | 7 |  |  |  |
| Flounder, General | 2,610 | 4,550 | 2,221 | 2,756 | 1,636 | 1,591 | 1,498 | 342 | 362 | 1,184 | 726 | 307 | 783 | 820 | 937 |
| Squid |  | 0 | 0 | 0 |  |  |  |  |  |  | 0 |  |  |  |  |
| Dover Sole |  |  |  | 0 |  |  |  |  |  |  |  |  |  |  |  |
| Thornyhead |  |  |  | 8 |  |  |  |  |  |  |  |  |  |  |  |
| Shortraker/ Rougheye | 8 | 0 | 2 | 21 |  |  |  | 1 |  |  |  |  |  |  |  |
| Butter Sole |  |  | 38 | 11 | 1 | 5 | 79 | 53 | 38 | 156 | 72 | 94 |  |  |  |
| Unsp. pelagic rockfish |  |  |  | 5 |  |  |  |  |  |  |  |  |  |  |  |
| Rougheye Rockfish |  |  | 0 |  | 0 |  |  |  |  |  |  |  |  |  |  |
| Starry Flounder |  |  | 230 | 85 | 0 | 1 | 99 | 72 | 34 | 214 | 152 | 329 |  |  |  |
| Northern Rockfish |  |  |  | 29 |  |  |  |  | 2 |  |  | 1 |  |  |  |




Figure 7.1-Size composition of rock sole, by sex and area, in the 2006 catch as determined from observer sampling.

Rock sole (L. polyxystra + L. bilineata)
AFSC survey data: standard shelf area



Fig. 7.3. Mean lengths at age (mm) by year of survey for eastern Bering Sea northern rocksole ages 3-9 for each sex during 1975-1998. Growth curves are shown for the 1979 (79yc) and 1987 ( 87 yc ) year classes. Dotted lines indicate no data during the period.
(From Walters and Wilderbuer, 2000, p.20)


Figure 7.4—Age composition of northern rock sole from the AFSC annual trawl survey.


Figure 7.4--continued.


Figure 7.5-Posterior distributions of catchability (top panel) and natural mortality (bottom panel) from a thinned chain of 1,000 results from 1 million MCMC runs of the model with the best fit to all the observed data with natural mortality estimated as a free parameter and catchability estimated with a contraint from the results of a herding experiment.


Figure 7.6--Stock assessment model estimates of total 2+ biomass (top left panel), fit to trawl survey biomass (top right panel), age-specific fishery and survey selectivity (middle left panel) and average annual fishing mortality rate (middle right panel), female spawning biomass (bottom right panel) and estimated age 1 recruitment (bottom right panel).


Figure 7.7. Ricker curve fit to yellowfin sole female spawning biomass-age 2 recruitment numbers for two productivity regimes: 1954-99 (all years, red line and open circles) and 1978-99. These estimates provided the foundation for initial simulation trials for underlying "true" operational model.


Figure 7.8. Results of the MSE analysis used to evaluate the Tier 1 harvest policy using Bering Sea yellowfin sole population dynamics from two productivity regimes alternative every 15 over a 60 year time horizon


Figure 7.9--Ricker (1958) model fit to spawner-recruit estimates from 22 years.


Figure 7.8-Relationship of annual rock sole female spawning biomass and full selection fishing mortality to B40 and F40.


Figure 7.9-Projection of rock sole female spawning biomass when fishing in the future at the average F of the past five years.

## Appendix

1) Observed fishery trawl locations, by quarter, for the 2006 fishing season.
2) Figures showing the fit of the stock assessment model to the time-series of fishery and trawl survey age compositions (survey and fishery observations are the solid lines).
3) Table of the assessment model estimates of population numbers at age 1975-2005.
4) Table of total population removals of rock sole from Alaska Fisheries Science Center research activities, 1977-2006.
5) TAC and ABC of BSAI northern rock sole from 1989-2006.
6) Posterior distributions of some parameters of interest from the stock assessment model.
7) Posterior distributions





Fits to the survey age composition


## Survey







Fit to survey age composition (continued)



Fit to the fishery age composition age composition (continued)
Fishery

















Fishery



Total catch ( t ) of rock sole in Alaska Fisheries Science Center research catches in the Bering Sea and Aleutian Islands, 1977-2006.

| year | research catch $(\mathbf{t})$ |
| :---: | :---: |
| 1977 | 10 |
| 1978 | 14 |
| 1979 | 13 |
| 1980 | 20 |
| 1981 | 12 |
| 1982 | 26 |
| 1983 | 59 |
| 1984 | 63 |
| 1985 | 34 |
| 1986 | 53 |
| 1987 | 52 |
| 1988 | 82 |
| 1989 | 83 |
| 1990 | 88 |
| 1991 | 97 |
| 1992 | 46 |
| 1993 | 75 |
| 1994 | 113 |
| 1995 | 99 |
| 1996 | 72 |
| 1997 | 91 |
| 1998 | 79 |
| 1999 | 72 |
| 2000 | 72 |
| 2001 | 81 |
| 2002 | 69 |
| 2003 | 75 |
| 2004 | 84 |
| 2005 | 74 |


| TAC |  | ABC |
| ---: | ---: | ---: |
| $\mathbf{1 9 8 9}$ | 90,762 | 171,000 |
| $\mathbf{1 9 9 0}$ | 60,000 | 216,300 |
| $\mathbf{1 9 9 1}$ | 90,000 | 246,500 |
| $\mathbf{1 9 9 2}$ | 40,000 | 260,800 |
| $\mathbf{1 9 9 3}$ | 75,000 | 185,000 |
| $\mathbf{1 9 9 4}$ | 75,000 | 313,000 |
| $\mathbf{1 9 9 5}$ | 60,000 | 347,000 |
| $\mathbf{1 9 9 6}$ | 70,000 | 361,000 |
| $\mathbf{1 9 9 7}$ | 97,185 | 296,000 |
| $\mathbf{1 9 9 8}$ | 100,000 | 312,000 |
| $\mathbf{1 9 9 9}$ | 120,000 | 309,000 |
| $\mathbf{2 0 0 0}$ | 137,760 | 230,000 |
| $\mathbf{2 0 0 1}$ | 75,000 | 228,000 |
| $\mathbf{2 0 0 2}$ | 54,000 | 225,000 |
| $\mathbf{2 0 0 3}$ | 44,000 | 110,000 |
| $\mathbf{2 0 0 4}$ | 41,000 | 139,000 |
| $\mathbf{2 0 0 5}$ | 41,500 | 132,000 |
| $\mathbf{2 0 0 6}$ | 41,500 | 126,000 |



Posterior distributions of selected parameter estimates from the preferred stock assessment model run.

Selected parameter estimates and their standard deviations

|  | Name | value | standard dev |  | name | value | Standard dev |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: |
|  | q_survey | 1.536 | 0.054 | 1984 | total biomass | 480.610 | 20.907 |
|  | M | 0.156 | 0.005 | 1985 | total biomass | 562.840 | 25.022 |
|  | mean_log_rec | 0.289 | 0.146 | 1986 | total biomass | 707.630 | 30.709 |
|  | sel_slope_fsh | 0.899 | 0.023 | 1987 | total biomass | 970.800 | 41.192 |
|  | sel_slope_srv | 1.787 | 0.077 | 1988 | total biomass | 1100.700 | 46.719 |
|  | sel50_fsh | 7.954 | 0.114 | 1989 | total biomass | 1193.700 | 53.073 |
|  | sel50_srv | 3.449 | 0.064 | 1990 | total biomass | 1196.100 | 53.294 |
|  | F40 | 0.134 | 0.005 | 1991 | total biomass | 1274.400 | 55.946 |
|  | F35 | 0.161 | 0.006 | 1992 | total biomass | 1300.500 | 57.416 |
|  | F30 | 0.195 | 0.008 | 1993 | total biomass | 1544.400 | 68.152 |
|  | R_logalpha | -3.313 | 0.205 | 1994 | total biomass | 1633.800 | 71.998 |
|  | R_logbeta | -5.006 | 0.119 | 1995 | total biomass | 1828.800 | 79.468 |
|  | Fmsy | 0.327 | 0.055 | 1996 | total biomass | 1746.100 | 75.292 |
|  | logFmsy | -1.117 | 0.167 | 1997 | total biomass | 1674.800 | 71.020 |
|  | msy | 126.100 | 19.249 | 1998 | total biomass | 1645.300 | 69.899 |
|  | Bmsy | 136.330 | 12.527 | 1999 | total biomass | 1597.300 | 66.842 |
| 1975 | total biomass | 160.820 | 9.567 | 2000 | total biomass | 1571.700 | 65.546 |
| 1976 | total biomass | 167.950 | 10.100 | 2001 | total biomass | 1521.500 | 64.068 |
| 1977 | total biomass | 178.840 | 10.640 | 2002 | total biomass | 1458.800 | 61.645 |
| 1978 | total biomass | 200.140 | 11.335 | 2003 | total biomass | 1406.200 | 60.626 |
| 1979 | total biomass | 225.700 | 12.206 | 2004 | total biomass | 1432.100 | 63.608 |
| 1980 | total biomass | 259.330 | 13.276 | 2005 | total biomass | 1502.600 | 71.509 |
| 1981 | total biomass | 297.970 | 14.565 | 2006 | total biomass | 1582.600 | 83.401 |
| 1982 | total biomass | 332.110 | 15.397 |  |  |  |  |
| 1983 | total biomass | 430.970 | 19.353 |  |  |  |  |



Posterior distributions of Fmsy from 8 model runs used to analyze a Tier 1 harvest policy.

