# Chapter 2: Assessment of the Pacific Cod Stock in the Eastern Bering Sea and Aleutian Islands Area 

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## Summary of Major Changes

Relative to the November edition of last year's BSAI SAFE report, the following substantive changes have been made in the Pacific cod stock assessment.

Changes in the Input Data

1) Catch data for 2005 were updated, and preliminary catch data for 2006 were incorporated.
2) Commercial fishery size composition data were recompiled for all years.
3) Size composition data from the 1982-2005 EBS shelf bottom trawl surveys were recompiled.
4) Size composition data from the 2006 EBS shelf bottom trawl survey were incorporated.
5) The biomass estimate from the 2006 EBS shelf bottom trawl survey was incorporated (the 2006 estimate of $517,698 \mathrm{t}$ was down about $14 \%$ from the 2005 estimate).
6) The biomass estimate from the 2006 AI bottom trawl survey was incorporated (the 2006 estimate of $92,526 t$ was down about $19 \%$ from the 2004 estimate).
7) Age composition data from the 1994 and 2004-2005 EBS shelf bottom trawl surveys were incorporated.
8) Length-at-age and weight-at-length data from the 1994 and 2004-2005 EBS shelf bottom trawl surveys were incorporated.
9) Relative abundance indices and size composition data from the Japanese longline survey (annual from 1982 through 1994) and the U.S. longline survey (biennial from 1997 through 2005) were incorporated into some models, but not others.

## Changes in the Assessment Model

The model selected last year by the Plan Team and SSC is presented again, basically unchanged except for updated estimates of parameters governing life history schedules that can be reliably estimated outside of the stock assessment model (e.g., length-at-age parameters, weight-at-length parameters). In addition, eight alternative models are presented. Unlike the base model, in which the catchability coefficient for the EBS shelf bottom trawl survey is fixed at a value of 1.0 , the eight alternative models all attempt to estimate this parameter. The eight alternative models are distinguished from one another via a factorial design based on the following three questions:

1) Should data from the longline surveys be excluded or included?
2) Should the selectivity function be of the "double logistic" or "double normal" form?
3) Should the prior distributions receive full (1.0) or partial (0.5) weight in the objective function?

The model recommended by the authors is Model B1, in which the data from the longline surveys are excluded, the selectivity function is of the "double normal" form, and the prior distributions receive full weight in the objective function (Model B2, which is the same as Model B1 except with down-weighted priors, gives very similar results).

## Changes in Assessment Results

Free estimation of shelf trawl survey catchability by all of the alternative models (except those incorporating data from the longline surveys) tended to result in estimates of biomass somewhat higher than the estimates from last year's assessment (using the model selected by the Plan Team and SSC).

1) Based on Model B1, the projected 2007 female spawning biomass for the BSAI stock is 307,000 t, up about $10 \%$ from last year's estimate for 2006 and up about $25 \%$ from last year's $F_{A B C}$ projection for 2007.
2) Based on Model B1, the projected 2007 total age 3+ biomass for the BSAI stock is $960,000 \mathrm{t}$, up about 4\% from last year's estimate for 2006.
3) Based on Model B1, the recommended 2007 ABC for the BSAI stock is $176,000 \mathrm{t}$, down about $9 \%$ from the actual 2006 ABC and up about $19 \%$ from last year's $F_{A B C}$ projection for 2007.
4) Based on Model B1, the estimated 2007 OFL for the BSAI stock is $207,000 \mathrm{t}$, down about $10 \%$ from the actual 2006 OFL and up about $17 \%$ from last year's $F_{A B C}$ projection for 2007.

## Responses to Comments from the SSC and Plan Teams

## SSC Comments Specific to the Pacific Cod Assessments

From the December, 2005 minutes:"The Bering Sea model in particular suggests very high uncertainty about the true values of $M$ and $Q$, and the SSC suggests that the authors try to estimate only one of these parameters at a time, while leaving the other parameter fixed." The present assessment includes eight alternative models in which EBS shelf bottom trawl survey catchability $(\mathrm{Q})$ is estimated. All of the models leave the natural mortality rate $(\mathrm{M})$ fixed at its traditional value of 0.37 .

From the December, 2005 minutes: "The SSC requests a brief update on stock structure of Pacific cod when new genetic data become available. Although the assessments for the Bering Sea and Gulf of Alaska have "converged" on the same model in this year's assessment, there is little a priori reason to emphasize the use of the same model or the same parameter values across regions." A presentation to the SSC is planned for the coming year, perhaps as early as February (SSC minutes, October, 2005).

From the December, 2005 minutes: "We endorse the Plan Team's recommendation to continue work on size-at-maturity. To reiterate, although we concur that sufficient justification was provided for adopting the new maturity schedule, there is some concern over the timing (GOA) and location (BSAI) of the samples that were used for histological examination. For example, maturity data for the BSAI were obtained only on the spawning grounds and may lead to an underestimation of length-at-maturity if small mature fish have a higher probability of entering the spawning grounds than immature fish of the same size." A three-year study of Pacific cod maturity is currently underway. Results will be reported as soon as they become available.

From the December, 2005 minutes: "The SSC encourages the authors to explore the use of longer time series of CPUE in the GOA using ADF\&G and IPHC trawl survey data, similar to the GLM approach used in the GOA pollock assessment." A preliminary investigation into the possible use of ADF\&G survey data was presented in the 2004 GOA Pacific cod assessment. For this year's assessment, priority
for inclusion of additional survey time series was given to the Japanese longline survey and U.S. longline survey, per Plan Team request (see below).

From the December, 2005 minutes:"In next year's assessment, the SSC would like to see a summary table of the overall likelihood of the models that were fit and the contribution to this likelihood of the various components, similar to tables provided in other assessments." The table of likelihood component values (Table 2.17 in last year's assessment) has been restructured (Table 2.16 in the present assessment) so as to be more similar to its counterparts in some other assessments.

From the September, 2006 minutes: "The Plan Teams and SSC received a paper on estimating Pacific cod off-bottom distance from archival tag data that was collected for different purposes. The SSC encourages continued work along those lines, recognizing that such estimates could prove extremely valuable for improving survey estimates of abundance and stock assessments." Work on alternative methods of estimating survey catchability and selectivity, including the use of archival tag data, will continue. However, as suggested at the September Plan Team meeting (see Plan Team minutes), it was not possible to complete the studies based on archival tag data in time for use in the present assessment.

## SSC Comments on Assessments in General

From the December, 2005 minutes: "The SSC appreciates the inclusion of phase-plane diagrams of relative harvest rate versus biomass, but we recommend standardization of units along the axes in all chapters to facilitate comparisons across species. The SSC suggests considering a quad plot based on $F / F 35 \%$ versus $B / B 35 \%$." Figure 2.10 has been revised per the SSC's suggestion.

From the December, 2005 minutes:"The SAFEs have been improved overall by expanded sections on ecosystem considerations to include discussion of predator-prey interactions. To this end, tables and figures have been added from ECOPATH models. One problem that has arisen is that there is some confusion about whether the information presented is stomach contents data, output from a single-species model, or output from an ECOPATH model. Figures and tables should more explicitly describe the source of the information presented. To avoid confusion between statistically-driven single species models and manually-adjusted ECOPATH models, the word "estimate" should be reserved for output from single-species models. In the absence of a statistical fitting procedure, outputs from ECOPATH/ECOSIM models should be referred to as adjusted parameters or just outputs. When ECOPATH/ECOSIM parameters are assumed to take on particular values, such assumptions should be stated explicitly. Care should be taken to avoid mixing results from different model structures." The present assessment includes an attachment describing recent results from ecosystem models. Special attention was paid to use of appropriate terminology so as to avoid confusion regarding the sources of the information presented.

## Plan Team Comments

From the September, 2005 minutes:"The Teams suggested using the longline survey data in the model." The present assessment includes four alternative models that use data from the Japanese longline survey and the U.S. longline survey.

From the November, 2005 minutes: "For future assessments, the Teams recommend that the authors present a model where $Q$ is estimated (and/or prior is provided) and $M$ is fixed." This recommendation is similar to one made by the SSC (see above). The present assessment includes eight alternative models in which EBS shelf bottom trawl survey catchability $(Q)$ is estimated. All of the models leave the natural mortality rate $(M)$ fixed at its traditional value of 0.37 . A prior distribution for $Q$ is specified for all eight alternative models, but the prior distribution is relaxed in four of those models.

From the November, 2005 minutes:"The Teams recommend exploring estimation of natural mortality from existing mark-recapture data." Given the Teams' suggestion to leave $M$ fixed for the time being (along with a similar SSC recommendation), this suggestion was not addressed in the present assessment.

From the November, 2005 minutes:"In September, the Plan Teams recommended that stock assessment authors continue to work on incorporating ecosystem assessment information into their chapters as much as possible, and that the ecosystem modelers also try to work with specific stock assessments each year to better incorporate the information to the assessments. ... The Teams agreed and noted that the following priorities for next year might be useful: GOA arrowtooth, AI Pollock, AI Pacific cod." An attachment to the present assessment summarizes results from ecosystem models on the role of Pacific cod in the Eastern Bering Sea and Aleutian Islands ecosystems.

## INTRODUCTION

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $63^{\circ} \mathrm{N}$ latitude. Pacific cod is distributed widely over the eastern Bering Sea (EBS) as well as in the Aleutian Islands (AI) area. The resource in these two areas (BSAI) is managed as a single unit. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and Gulf of Alaska (GOA). Although at least one previous genetic study (Grant et al. 1987) failed to show significant evidence of stock structure within these areas, current genetic research underway at the Alaska Fisheries Science Center may soon shed additional light on the issue of stock structure of Pacific cod within the BSAI (M. Canino, AFSC, pers. commun.). Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the EBS or AI areas.

## FISHERY

Catches of Pacific cod taken in the EBS, AI, and BSAI for the periods 1964-1980 and 1981-2006 are shown in Tables 2.1a and 2.1b, 2.2a and 2.2b, and 2.3a and 2.3b, respectively. The catches in Tables 2.1a, 2.2a, and 2.3a are broken down by year and fleet sector (foreign, joint venture, domestic annual processing), while the catches in Tables 2.1b, 2.2 b , and 2.3 b are broken down by gear type as well. During the early 1960s, a Japanese longline fishery harvested BSAI Pacific cod for the frozen fish market. Beginning in 1964, the Japanese trawl fishery for walleye pollock (Theragra chalcogramma) expanded and cod became an important bycatch species and an occasional target species when high concentrations were detected during pollock operations. By the time that the Magnuson Fishery Conservation and Management Act went into effect in 1977, foreign catches of Pacific cod had consistently been in the $30,000-70,000 \mathrm{t}$ range for a full decade. In 1981, a U.S. domestic trawl fishery and several joint venture fisheries began operations in the BSAI. The foreign and joint venture sectors dominated catches through 1988, but by 1989 the domestic sector was dominant and by 1991 the foreign and joint venture sectors had been displaced entirely. Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components. Figure 2.1 shows areas in which sampled hauls or sets for each of the three main gear types (trawl, longline, and pot) were concentrated during 2005. To create these figures, the EEZ off Alaska was divided into $20 \mathrm{~km} \times 20 \mathrm{~km}$ squares. For each gear type, a square is shaded if more than two hauls/sets containing Pacific cod were sampled in it during 2005.

The history of acceptable biological catch (ABC) and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate (i.e., all-gear, combined area) commercial catches in Table 2.4. From 1980 through 2006, TAC averaged about $78 \%$ of ABC, and aggregate commercial catch averaged about $88 \%$ of TAC. In 10 of these 27 years ( $37 \%$ ), TAC equaled ABC exactly, and in 5 of these 27 years ( $19 \%$ ), catch exceeded TAC (by an average of $4 \%$ ). Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3 ) changes in the stock assessment model. For example, in the assessments for fishery years 1980 through 2005, seven different assessment models were used (Table 2.4). All assessments from 1993 through 2004 used the Stock Synthesis 1 modeling software with primarily length-based data, albeit with some changes in model structure from time to time. The assessment was migrated to Stock Synthesis 2
last year (Thompson and Dorn 2005). Historically, the great majority of the BSAI catch has come from the EBS area. During the most recent complete five-year period (2001-2005), the EBS accounted for an average of about $85 \%$ of the BSAI catch.

Current regulations specify that the BSAI Pacific cod TAC will be allocated initially according to gear type as follows: the trawl fishery will be allocated $47 \%$, the fixed gear (longline and pot) fishery will be allocated $51 \%$, and the jig fishery will be allocated $2 \%$; of the fixed gear allocation, the longline fishery will be allocated $80.3 \%$ (not counting catcher vessels less than 60 ft LOA ), the pot fishery will be allocated $18.3 \%$ (not counting catcher vessels less than 60 ft . LOA), and fixed-gear catcher vessels less than 60 ft . LOA will be allocated $1.4 \%$. Typically, as the harvest year progresses, it becomes apparent that one or more gear types will be unable to harvest their full allotment(s) by the end of the year. This is addressed by reallocating TAC between gear types in September of each year. Most often, such reallocations shift TAC from the trawl, jig, and sometimes pot components of the fishery to the longline catcher/processors. The longline catcher-processors typically receive $15,000-20,000 \mathrm{t}$ per year through such transfers.

The catches shown in Tables 2.1b, 2.2b, 2.3b, and 2.4 include estimated discards. Discard rates of Pacific cod in the various EBS and AI target fisheries are shown for each year 1991-2002 in Table 2.5a and for each year 2003-2004 in Table 2.5b.

## DATA

This section describes data used in the current stock assessment models. It does not attempt to summarize all available data pertaining to Pacific cod in the BSAI.

## Commercial Catch Data

## Catch Biomass

Catches (which may not include discards) taken in the EBS for the period 1964-1980 are shown in Table 2.6a and catches (including estimated discards) taken in the EBS for the period 1981-2005 are shown in Table 2.6b. Catches in these tables are broken down by the three main gear types and intra-annual periods consisting of the months January-May, June-August, and September-December. This particular division, which was suggested by participants in the EBS fishery, is intended to reflect actual intra-annual differences in fleet operation (e.g., fishing operations during the spawning period may be different than at other times of year). In years for which estimates of the distribution by gear or period were not available, proxies based on other years' distributions were used.

## Catch Size Composition

Fishery size compositions are presently available, by gear, for at least one gear type in every year from 1974 through the first part of 2006, with the exception of 1976. For ease of representation and analysis, length frequency data for Pacific cod can usefully be grouped according to the following set of 25 intervals or "bins," with the upper and lower boundaries shown in cm :

| BinNumber: | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| LowerBound: | 9 | 12 | 15 | 18 | 21 | 24 | 27 | 30 | 33 | 36 | 39 | 42 | 45 | 50 | 55 | 60 | 65 | 70 | 75 | 80 | 85 | 90 | 95 | 100 | 105 |
| UpperBound: | 11 | 14 | 17 | 20 | 23 | 26 | 29 | 32 | 35 | 38 | 41 | 44 | 49 | 54 | 59 | 64 | 69 | 74 | 79 | 84 | 89 | 94 | 99 | 104 | 110 |

The collections of relative length frequencies are shown by year, period, and size bin for the trawl fishery in Tables 2.7a, 2.7b, and 2.7c; the longline fishery in Tables 2.8a, 2.8b, and 2.8c; and the pot fishery in Tables 2.9a and 2.9b. Input sample sizes ( N ) for the multinomial distribution used in the stock assessment model are also shown. These are set equal to the square root of the total sample size.

## Survey Data

## EBS Shelf Bottom Trawl Survey

The relative size compositions from bottom trawl surveys of the EBS shelf conducted by the Alaska Fisheries Science Center since 1979 are shown in Tables 2.10a for the years 1979-1981 and 2.10b for the years 1982-2006, using the same length bins defined above for the commercial catch size compositions. The survey is shown as two separate time series because of a gear change that was instituted in 1982. Input sample sizes (N) for the multinomial distribution used in the stock assessment model are also shown. These are set equal to the square root of the total sample size in years 1982-1987 and 1990-2006. For other years, N was set equal to 100 , approximating the square root of the average average of the 10 known true sample sizes from the years 1986-1997.
Following a decade-long hiatus in production ageing of Pacific cod, the Age and Growth Unit of the Alaska Fisheries Science Center began ageing samples of Pacific cod from the EBS shelf bottom trawl surveys a few years ago (Roberson 2001, Roberson et al. 2005). To date, the otolith collections from the 1994 and 1996-2005 surveys have been read. The relative age compositions from these surveys are shown in Table 2.11. The number of fish aged for each of these years is shown below:

| Year: | 1994 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| N: | 715 | 252 | 719 | 635 | 860 | 864 | 950 | 947 | 1360 | 1040 | 609 |

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl surveys are shown in Table 2.12a (1979-1981) and 2.12b (1982-2006), together with the standard errors and upper and lower $95 \%$ confidence intervals (CI) for the biomass estimates. Survey results indicate that biomass increased steadily from 1978 through 1983, then remained relatively constant from 1983 through 1988. The highest biomass ever observed by the survey was the 1994 estimate of $1,368,120 \mathrm{t}$. Following the high observation in 1994, the survey biomass estimate declined steadily through 1998. The survey biomass estimates have remained in the 510,000-620,000 t range from 1997 through the present, except for 2001 , when the estimate was $833,626 \mathrm{t}$. The biomass estimate from 2001 appears likely to be an overestimate, given the magnitude of the implied increases relative to the 2000 survey ( $57 \%$ ) and the fact that the 2002-2006 estimates were much closer to the preceding estimates. The 2006 estimate was $517,698 \mathrm{t}$, a $14 \%$ drop from the 2005 value and the second lowest estimate in the post-1981 time series.

## EBS Slope Bottom Trawl Survey

The Alaska Fisheries Science Center conducted bottom trawl surveys of the EBS slope in 2002 and 2004. The relative size compositions from these surveys are shown in Table 2.13, using the same length bins defined above for the commercial catch size compositions. Input sample sizes (N) for the multinomial distribution used in the stock assessment model are also shown. These are set equal to the square root of the total sample size. A total of 468 fish were measured in the 2002 survey and a total of 531 fish were measured in the 2004 survey (note that these sample sizes are only about one-twentieth of the average sample size from the shelf survey). The biomass estimates and standard errors from the 2002 and 2004 surveys are shown below (all figures are in $\mathfrak{t}$ ):

| Year | Biomass | Standard Error |
| ---: | ---: | ---: |
| 2002 | 7511 | 1944 |
| 2004 | 5756 | 968 |

## Japanese and U.S. Longline Surveys

The Japanese longline survey was conducted annually from 1982-1994, and the U.S. longline survey has been conducted in the EBS biennially starting in 1997. These surveys are designed primarily to assess the abundance of sablefish (Anoplopoma fimbria), but Pacific cod are also captured in these surveys. Pacific cod size compositions from the Japanese and U.S. longline surveys are shown in Tables 2.14a and 2.14b.

Input sample sizes $(\mathrm{N})$ for the multinomial distribution used in the stock assessment model are also shown. These are set equal to the square root of the total sample size.

A problem arises in use of the longline survey catch rates as an index of abundance, however, in that most of the Pacific cod catches take place in the shallowest depth strata, where few sablefish are caught. Because few sablefish are caught in these strata, appropriate area expansion factors have not been computed, so the only index of abundance available for Pacific cod is a simple average catch per station. The time series of average Pacific cod catch (number of fish caught per station) and associated coefficients of variation are shown for the two surveys in Table 2.15. To make the abundance indices as meaningful as possible, the averages were computed only for those stations that were successfully sampled in every year. The numbers of stations that qualify under this criterion are not large. For the Japanese survey, 32 stations were successfully sampled every year, but only 11 stations were successfully sampled every year in the U.S. survey.

It should be emphasized that the abundance indices in Table 2.15 are relative indices at best. The Japanese survey in particular shows an enormous degree of year-to-year variability. Of the 12 year-toyear changes present in the Japanese time series, there were two one-year increases of well over $200 \%$ (i.e., the index more than tripled) and two other annual changes showed decreases of more than $50 \%$.

## Aleutian Bottom Trawl Survey

Biomass estimates for the Aleutian Islands region were derived from U.S.-Japan cooperative bottom trawl surveys conducted during the summers of 1980,1983 , and 1986, and by U.S. bottom trawl surveys of the same area in 1991, 1994, 1997, 2000, 2002, 2004, and 2006. These surveys covered both the Aleutian management area ( 170 degrees east to 170 degrees west) and a portion of the Bering Sea management area ("Southern Bering Sea") not covered by the EBS shelf bottom trawl surveys. The time series of biomass estimates from the overall Aleutian survey area are shown together with their sum below (all figures are in t ):

| Year | Survey Type | Aleutian Survey Area |
| :---: | :---: | ---: |
| 1980 | U.S.-Japan | 148,272 |
| 1983 | U.S.-Japan | 215,755 |
| 1986 | U.S.-Japan | 255,072 |
| 1991 | U.S. | 191,049 |
| 1994 | U.S. | 184,068 |
| 1997 | U.S. | 83,416 |
| 2000 | U.S. | 136,028 |
| 2002 | U.S. | 82,970 |
| 2004 | U.S. | 114,161 |
| 2006 | U.S. | 92,526 |

For many years, the assessments of Pacific cod in the BSAI used a weighted average formed from EBS and Aleutian survey biomass estimates to provide a conversion factor which was used to translate model projections of EBS catch and biomass into BSAI equivalents. Prior to the 2004 assessment, the weighted average was based on the sums of the biomass estimates from the EBS shelf and AI survey biomass time series. However, in December of 2003 the SSC requested that alternative methods of estimating relative biomass between the EBS and AI be explored. Following a presentation of some possible alternatives, the SSC recommended that an approach based on a simple Kalman filter be used (SSC Minutes, October, 2004). Applying the Kalman filter approach to the updated (through 2006) time series indicates that the best estimate of the current biomass distribution is $84 \% \mathrm{EBS}$ and $16 \% \mathrm{AI}$ (the previous proportions were $85 \%$ and $15 \%$, respectively). Because the $83-112$ net (with no roller gear) used in the EBS survey generally tends the bottom better than the polyethylene Noreastern net (with roller gear) used in the AI
survey, this ratio should tend to err on the conservative side (that is, the AI survey would be expected to miss more fish than the EBS survey, so the true portion in the AI should be higher than the ratio of the AI to AI+EBS survey estimates).

## ANALYTIC APPROACH

## Model Structure

## History of Model Structures Developed Under Stock Synthesis 1 and 2

Beginning with the 1993 SAFE report (Thompson and Methot 1993) and continuing through the 2004 SAFE report (Thompson and Dorn 2004), a model using the Stock Synthesis 1 (SS1) assessment program (Methot 1986, 1990, 1998, 2000) and based largely on length-structured data formed the primary analytical tool used to assess the EBS Pacific cod stock. It should be emphasized that the model has always been intended to assess only the EBS portion of the BSAI stock. Conversion of model estimates of EBS biomass and catch to BSAI equivalents has traditionally been accomplished by application of an expansion factor based on the relative survey biomasses between the EBS and AI.

SS1 is a program that used the parameters of a set of equations governing the assumed dynamics of the stock (the "model parameters") as surrogates for the parameters of statistical distributions from which the data are assumed to be drawn (the "distribution parameters"), and varies the model parameters systematically in the direction of increasing likelihood until a maximum is reached. The overall likelihood is the product of the likelihoods for each of the model components. In part because the overall likelihood can be a very small number, SS1 uses the logarithm of the likelihood as the objective function. Each likelihood component is associated with a set of data assumed to be drawn from statistical distributions of the same general form (e.g., multinomial, lognormal, etc.). Typically, likelihood components are associated with data sets such as catch size (or age) composition, survey size (or age) composition, and survey biomass (either relative or absolute).

SS1 permits each data time series to be divided into multiple segments, resulting in a separate set of parameter estimates for each segment. The EBS Pacific cod assessments, for example, have usually divided the shelf bottom trawl survey size composition time series into pre-1982 and post-1981 segments to account for the effects of a change in the trawl survey gear instituted in 1982. Also, to account for possible differences in selectivity between the mostly foreign (also joint venture) and mostly domestic fisheries, the fishery size composition time series have traditionally been split into pre-1989 and post1988 segments.

In the EBS Pacific cod model, each year has traditionally been partitioned into three seasons: JanuaryMay, June-August, and September-December (these seasonal boundaries were suggested by industry participants). Four fisheries have traditionally been defined: The January-May trawl fishery, the JuneDecember trawl fishery, the longline fishery, and the pot fishery.

Following a series of modifications from 1993 through 1997, the base model for EBS Pacific cod remained completely unchanged from 1997 through 2001. During the late 1990s, a number of attempts were made to estimate the natural mortality rate $M$ and the shelf bottom trawl survey catchability coefficient $Q$, but these were not particularly successful and the Plan Team and SSC always opted to retain the base model in which $M$ and $Q$ were fixed at their traditional values of 0.37 and 1.0, respectively.

A minor modification of the base model was suggested by the SSC in 2001, namely, that consideration be given to dividing the domestic era into pre-2000 and post-1999 segments. This modification was tested in the 2002 assessment (Thompson and Dorn 2002), where it was found to result in a statistically significant improvement in the model's ability to fit the data. In the 2004 assessment (Thompson and Dorn 2004),
further modifications were made to the base model. The 2004 model included a set of selectivity parameters for the EBS slope bottom trawl survey and added new likelihood components for the age compositions and length-at-age data from the 1998-2003 EBS shelf bottom trawl surveys and the size composition and biomass data from the 2002 and 2004 EBS slope bottom trawl surveys. Incorporation of age data and slope survey data had been suggested by the SSC (SSC minutes, December 2003).

A major change took place in the 2005 assessment (Thompson and Dorn 2005), as the model was migrated to the newly developed Stock Synthesis 2 (SS2) program, which makes use of the ADMB modeling architecture (Fournier 2005) currently used in most age-structured assessments of BSAI and GOA groundfish. The move to SS2 facilitated improved estimation of model parameters as well as statistical characterization of the uncertainty associated with parameter estimates and derived quantities such as spawning biomass. Three alternative models were presented in the 2005 assessment. Model 1 was identical to the SS1-based model used in the 2004 assessment. Model 2 was very similar to Model 1, but was explicitly Bayesian (i.e., prior distributions were specified for all model parameters) and it was configured under SS2 rather than SS1. Model 3 was similar to Model 2, except that values of the shelf bottom trawl survey catchability coefficient $Q$ and the natural mortality rate $M$ were estimated rather than fixed at the traditional values of 1.0 and 0.37 , respectively. The Plan Team and SSC both chose Model 2, feeling that moving from fixed values of $Q$ and $M$ to estimated values for both those parameters at the same time was too big a step. (It should be noted that fixing $Q$ is not the same as fixing the entire selectivity schedule, as selectivity parameters are still typically estimated even when $Q$ is fixed. However, fixing $Q$ at a particular value will usually influence the values of the estimated selectivity parameters.)

## Current Issues in Model Structure

## Estimation of EBS Shelf Bottom Trawl Survey Catchability

The SSC has requested that the 2006 assessment focus on estimating either $Q$ or $M$ (not both) while leaving the other parameter fixed at its traditional value of 1.0 or 0.37 , respectively (SSC minutes, December, 2005). The Plan Team was more explicit in its recommendation, suggesting that the 2006 assessment focus on estimating $Q$ while leaving $M$ fixed at its traditional value (Plan Team minutes, November, 2005).

Estimates of the selectivity schedule for the EBS shelf bottom trawl survey obtained in previous BSAI Pacific cod assessments have often tended to show a pronounced "kink," with survey selectivity increasing rapidly from a low value for the smallest fish up to a peak at some intermediate length, then decreasing rapidly as length increased further. It has been conjectured that this behavior was a result of fixing $Q$ at an artificially high level, thereby forcing a sharp kink in the selectivity curve so that, overall, the product of catchability and selectivity is approximately correct.

Although direct experimental evidence (as opposed to the types of indirect evidence coming from length compositions, age compositions, and abundance indices used in stock assessments) pertaining to the value of $Q$ for Pacific cod in the EBS shelf bottom trawl survey is becoming more available, it is still insufficient to enable estimation of this parameter outside the context of a full stock assessment model. Two types of direct experimental evidence are available: results of studies pertaining specifically to Pacific cod, and results of studies pertaining to closely related species.

Available experimental evidence regarding the value of $Q$ for Pacific cod in the EBS shelf bottom trawl survey includes the following: Munro and Somerton (2002) and Weinberg et al. (2002) showed that Pacific cod within the path of the net do not tend to escape under the footrope. Somerton (2004) showed that Pacific cod neither tend to escape around the sides of the net nor tend to be herded into the net by the doors. Von Szalay and Somerton (2005) showed that catch efficiency of Pacific cod decreased with increases in net spread and presumed decreases in net height, leaving open the possibility that some fish
occur in the water column above the headrope or are initially within the path of the net but escape over the headrope. Recently, Nichol et al. (unpubl. manuscr.) and Thompson and Nichol (unpubl. manuscr.) proposed methods for estimating the vertical distribution of Pacific cod relative to the bottom based on archival tag data. However, neither of these studies has been completed.

Available experimental evidence from closely related species includes the following: Winger et al. (2000) showed that catchability of Atlantic cod (Gadus morhua) may be highly sensitive to changes in towing speed. If the same holds true for Pacific cod, it is possible that some fish may be out-swimming the trawl survey net. Handegard et al. (2003) and Handegard and Tjostheim (2005) showed that some other gadids, including Atlantic cod, may tend to dive as a behavioral response to an approaching vessel or net, meaning that even a highly accurate estimate of "typical" vertical distribution may provide a biased picture of catchability.

Not only is the available experimental evidence regarding survey catchability of Pacific cod less than conclusive in some respects, the results of existing studies tend to obscure the distinction between age- or size-specific selectivity (a measure of how relative susceptibility to capture differs with age or size) and overall catchability (a measure of absolute susceptibility to capture for the most-selected age or size).

In summary, considering the indirect evidence from past stock assessments along with the available direct evidence from field experiments, it seems that enough uncertainty about the true value of $Q$ exists to warrant exploration of the possibility that $Q$ does not equal the traditional value of 1.0.

## Use of Longline Survey Data

For many years, data from the Japanese longline survey and U.S. longline survey have been a primary input to the BSAI and GOA assessments of sablefish. In 2005, the Plan Teams suggested using data from the longline surveys in the Pacific cod assessments as well (Plan Team minutes, September, 2005). There are some issues involved with use of the Pacific cod data from these surveys, as discussed under "Data" above. Nevertheless, relative abundance estimates (though not expanded by area) and size composition data are available annually from the Japanese longline survey from 1982 through 1994 and biennially from the U.S. longline survey from 1997 through 2005.

## Functional Form of the Selectivity Curve

Several options are included in SS2 for specifying the functional form of the selectivity curve. The most flexible and commonly used of these is the "double logistic" function, which the BSAI Pacific cod assessments have used ever since the first length-based SS1 version of the assessment in 1993 (Thompson and Methot 1993). This function has grown increasingly complicated over the years, starting from a fourparameter form in its original incarnation in SS1 and evolving to an eight-parameter form as currently implemented in SS2. The double logistic function consists of a pair of scaled logistic curves joined by a horizontal linear segment. The first (ascending) logistic curve begins at the minimum length specified in the data file ( 9 cm in the case of the EBS Pacific cod model), where the selectivity is less than 1.0, and ends at some intermediate length, where selectivity is exactly 1.0. A horizontal linear segment extends from the right-hand end of the first logistic to the left-hand end of the second logistic. Selectivity equals 1.0 throughout this linear segment. The second (descending) logistic curve begins at the end of the horizontal linear segment, where selectivity is still exactly 1.0 , and ends at the maximum length specified in the data file ( 110 cm in the case of the EBS Pacific cod model), where the selectivity is less than 1.0 . Eight parameters are used to define the double logistic selectivity function: the size at which selectivity first reaches a value of 1.0 (peak location), the selectivity at the minimum length represented in the data (S(Lmin)), the logit transform of the size corresponding to the inflection of the ascending logistic curve (logit(infl1)), the relative slope of the ascending logistic curve (slope1), the logit transform of the size corresponding to the inflection of the descending logistic curve (logit(infl2)), the relative slope of the descending logistic curve (slope2), the logit transform of the selectivity at the maximum length
represented in the data $(\operatorname{logit}(S(\operatorname{Lmax})))$, and the width of the length range at which selectivity equals 1.0 (peak width).

Another option provided by SS2 for the functional form of the selectivity curve is the "double normal" function, which involves a pair of curves reminiscent of the left and right halves of a pair of normal probability density functions joined by a horizontal linear segment. Like the double logistic function, the double normal function involves an ascending curve that reaches a maximum value of 1.0 at some point (peak location), a horizontal linear segment extending for some distance (peak width), and a descending curve that begins at the end of the horizontal linear segment. Contrasted with the double logistic function, the double normal function is simpler but less flexible, in that a single parameter defines the shape of the ascending curve and a single parameter defines the shape of the descending curve (as opposed to three parameters apiece in the double logistic). The parameters governing the shapes of the ascending and descending curves in the double normal are the log variances (Invar1 and Invar2, respectively) of the associated normal curves. Using the ascending curve as an example, selectivity at length len is given by:

$$
S(\text { len })=\exp \left(-\frac{(\text { len }- \text { peak location })^{2}}{\exp (\operatorname{lnvar} 1)}\right)
$$

Estimating or otherwise specifying eight parameters for each selectivity function (there are either 14 or 16 selectivity functions in the Pacific cod model, depending on whether the longline survey data are excluded or included), as required by use of the double logistic function, is a challenging undertaking, and it is worth exploring the possibility that a simpler functional form may not change the point estimates of the most important model outputs appreciably but may make those estimates less uncertain.

## Prior Distributions

Because SS2 is explicitly cast in a Bayesian framework, specification of a prior distribution is required for each parameter. Of course, a noninformative prior can be chosen for any or all parameters if so desired. However, use of informative priors is probably appropriate for at least some of the parameters in the EBS Pacific cod model, because both the Plan Team and the SSC have indicated in the past that certain values, or ranges of values, for various parameters are either relatively likely or unlikely. For example, the Plan Team has expressed concern that the estimates of large-fish selectivity in the EBS shelf bottom trawl survey obtained in many previous assessments may be too low (Plan Team minutes, November 2004). By utilizing a Bayesian framework, SS2 provides a logical means of integrating perspectives such as these into the stock assessment model. Use of informative priors can also help to stabilize parameter estimates.

Last year's assessment contained a thorough description of the prior distributions used, but the sensitivity of the results to those distributions was not made explicit in the SAFE report. One way to make such sensitivity more explicit would be to include model runs in which the contribution of the prior distributions to the overall objective function is downweighted.

## Model Structures Considered in This Year's Assessment

This year's BSAI Pacific cod assessment includes nine alternative models for the EBS portion of the stock. Model 0 , the base model, is the same as the model selected last year by the Plan Teams and SSC. In addition to the base model, eight other models are presented as possible alternatives. All models, including the base model and the eight alternatives, use the latest estimates of parameters governing the length-at-age and weight-at-length relationships, as well as the latest estimates of parameters governing variability in length at age and variability in estimated age (ageing error). Parameters governing the maturity-at-length schedule have not changed since last year. The eight alternative models differ from the base model in various respects, but two of these differences are consistent across all of the alternative models:

1) In all of the alternative models, catchability of the EBS shelf bottom trawl survey is estimated, rather than assumed to equal 1.0 as in the base model. Separate catchability coefficients are estimated for the pre-1982 and post-1981 portions of the time series because of a change in the survey gear instituted in 1982.
2) In all of the alternative models, all selectivity parameters are estimated, except that $S(L \min )$ in models using the double logistic selectivity function is set equal to 0.001 for all gear types other than the EBS shelf bottom trawl surveys ( $S(L \min$ ) is estimated for the EBS shelf bottom trawl surveys). Last year, it was not possible to estimate all remaining selectivity parameters statistically in the model chosen by the Plan Team and SSC, so the value of each peak location parameter in that model was chosen by other methods. The same (fixed) peak location values are used in this year's base model, but not the alternative models.

Although the eight alternative models share the above pair of features in common, they are distinguished from one another via a factorial design based on the following three questions:

1) Should data from the longline surveys be excluded or included?
2) Should the selectivity function be of the "double logistic" or "double normal" form?

3 ) Should the prior distributions receive full (1.0) or partial (0.5) weight in the objective function?
The eight alternative models address all possible combinations of answers to the above as follow:

| Model | Longline survey data | Selectivity function | Prior weight |
| :--- | :--- | :--- | :--- |
| Model A1 | Exclude | Double logistic | 1.0 |
| Model A2 | Exclude | Double logistic | 0.5 |
| Model B1 | Exclude | Double normal | 1.0 |
| Model B2 | Exclude | Double normal | 0.5 |
| Model C1 | Include | Double logistic | 1.0 |
| Model C2 | Include | Double logistic | 0.5 |
| Model D1 | Include | Double normal | 1.0 |
| Model D2 | Include | Double normal | 0.5 |

## Parameters Estimated Independently

## Natural Mortality

In the 1993 BSAI Pacific cod assessment (Thompson and Methot 1993), the natural mortality rate $M$ was estimated using SS1 at a value of 0.37. Although attempts have been made to re-estimate $M$ in some years (during the late 1990s and, most recently, in the 2005 assessment (Thompson and Dorn 2005)), all models of the BSAI Pacific cod stock accepted by the Plan Team and SSC since 1993 have ultimately retained a value of 0.37 for $M$, as have all subsequent assessments of the GOA Pacific cod stock (with one exception, in 1995). Other published estimates of $M$ for Pacific cod are shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

All models in the present assessment fix $M$ at the traditional value of 0.37 .

## Trawl Survey Catchability

In Model 0 , catchability for the EBS shelf bottom trawl survey is fixed at a value of 1.0 for both the pre1982 and post-1981 portions of the time series. In all other models, these parameters are estimated freely and separately for both portions of the time series.

## Length at Age

Parameters of the Brody growth equation, as formulated in SS2, were re-estimated this year based on all available data. The curve described by the updated parameter values is close to last year's curve. The new parameter values are: length at 1 year $=11.1 \mathrm{~cm}$, length at 12 years $=93.3 \mathrm{~cm}$, and Brody's growth coefficient $K=0.113$.

## Variability in Length at Age

The method for estimating variability in length at age was substantially improved this year by developing a formal statistical model based on SS2's required assumption that the coefficient of variation in length at age is a linear function of mean length at age. A lognormal distribution of lengths at age was assumed.
The new parameter estimates are: CV at age $1=0.16$, CV at age $13=0.065$.

## Variability in Estimated Age

Variability in estimated age in SS2 is based on the standard deviation of estimated age. Weighted least squares regression was used in the 2005 assessment (Thompson and Dorn 2005) to estimate a proportional relationship between standard deviation and age. The regression was re-run this year based on all available data. The new relationship is close to last year's. The new estimated proportionality is 0.103 (i.e, the standard deviation of estimated age was modeled as $0.103 \times$ age).

## Weight at Length

Parameters governing the allometric relationship between weight $(\mathrm{kg})$ and length $(\mathrm{cm})$ were re-estimated this year by log-log regression from the same data used to estimate the parameters of the length-at-age relationship. The curve described by the updated parameter values is close to last year's curve. The new parameter values are: multiplicative constant $=3.86 \times 10^{-6}$, and exponent $=3.266$.

## Maturity at Length

A detailed history and evaluation of parameter values used to describe maturity at length for BSAI Pacific cod was presented in the 2005 assessment (Thompson and Dorn 2005). The parameters used in last year's assessment, based on a study by Stark (2005), were as follows: length at $50 \%$ maturity $=58 \mathrm{~cm}$ and slope of linearized logistic equation $=-0.132$. The same parameter values are used for all models in this year's assessment.

## Parameters Estimated Conditionally

Parameters estimated conditionally (i.e., within individual SS2 runs, based on the data and the parameters estimated independently) by all nine models consist of the following:

1) log-scale mean recruitment for the post-1976 environmental regime
2) annual $\log$-scale recruitment deviations
3) EBS slope bottom trawl survey catchability
4) initial fishing mortality rates (the population is assumed to be in equilibrium in 1964)

Estimation of catchability coefficients for surveys other than the EBS slope bottom trawl survey varies by survey as follows:

1) Pre-1982 EBS shelf bottom trawl survey catchability: all models except Model 0
2) Post-1981 EBS shelf bottom trawl survey catchability: all models except Model 0
3) Japanese longline survey catchability: Models C1, C2, D1, and D2 only
4) U.S. longline survey catchability: Models C1, C2, D1, and D2 only

Recall that all models consider three bottom trawl surveys (pre-1982 shelf trawl survey, post-1981 shelf trawl survey, slope trawl survey) and a total of 11 gear- and era-specific fisheries (four gears, consisting of the January-May trawl fishery, June-December trawl fishery, longline fishery, and pot fishery; and three eras, consisting of the years 1964-1988, 1989-1999, and 2000-2006, except that there was no significant pot fishery during the 1964-1988 era). In addition, Models C1, C2, D1, and D2 consider two longline surveys (Japan and U.S.). The total number of selectivity parameters estimated conditionally therefore varies by model as follows:

1) Model 0 uses the 8 -parameter double logistic function to describe selectivity for 3 surveys and 11 fisheries, which would total 112 selectivity parameters, except that $S(L \min )$ is fixed at a value of 0.001 for the slope trawl survey and all fisheries and peak location is fixed at various values for all surveys and fisheries, bringing the total of estimated selectivity parameters down to 86 .
2) Models A1 and A2 use the 8-parameter double logistic function to describe selectivity for 3 surveys and 11 fisheries, which would total 112 selectivity parameters, except that that $S(\operatorname{Lmin})$ is fixed at a value of 0.001 for the slope trawl survey and all fisheries, bringing the total of estimated selectivity parameters down to 100 .
3) Models B1 and B2 use the 4-parameter double normal function to describe selectivity for 3 surveys and 11 fisheries, with no parameters fixed, giving a total of 56 estimated selectivity parameters.
4) Models C 1 and C 2 use the 8-parameter double logistic function to describe selectivity for 5 surveys and 11 fisheries, which would total 128 selectivity parameters, except that that $S(\operatorname{Lmin})$ is fixed at a value of 0.001 for the slope trawl survey, the Japanese and U.S. longline surveys, and all fisheries, bringing the total of estimated selectivity parameters down to 114 .
5) Models D1 and D2 use the 4-parameter double normal function to describe selectivity for 5 surveys and 11 fisheries, with no parameters fixed, giving a total of 64 estimated selectivity parameters.

For all parameters estimated within individual SS2 runs, the estimator used is the mode of the logarithm of the joint posterior distribution, which is in turn calculated as the sum of the logarithms of the parameter-specific prior distributions (see below) and the logarithm of the likelihood function.

In addition to the above, there are two other sets of parameters that are estimated conditionally, but not in the same sense as the above parameters. The first of these is the full set of year-, season-, and gearspecific fishing mortality rates. The fishing mortality rates are determined exactly rather than estimated
statistically because SS 2 assumes that the input total catch data are true values rather than estimates, so the fishing mortality rates can be computed algebraically given the other parameter values and the input catch data.

The second set of parameters that is estimated conditionally, but in a manner different from the other parameters, consists of two parameters that help to describe the distribution of individual recruitments. These are estimated iteratively (i.e., between SS2 runs rather than within an individual SS2 run). In SS2, log-scale recruitment is modeled in terms of a mean, a standard deviation $\left(\sigma_{R}\right)$, and annual deviations from the mean. The parameters are automatically scaled so that the average annual deviation from the mean is zero. A problem arises, however, in attempting to model the effects of the major environmental regime shift that occurred in 1977 (e.g., Hare and Mantua 2000), because the available information indicates strongly that year classes of Pacific cod were much smaller (in magnitude) during the pre-1977 regime than during the post-1976 regime. Establishing different pre-1977 and post-1976 log-scale means is easily accomplished in SS2 by creating a regime shift "dummy variable" for each year in the time series and estimating a link between mean log-scale recruitment and the dummy variable. However, $\sigma_{R}$ cannot be linked to the dummy variable in SS2. This implies that the mean recruitment deviation for each portion of the time series (pre-1977 and post-1976) will not necessarily equal zero, even though SS2 forces the mean recruitment deviation for the overall time series to equal zero. This, in turn, implies that the estimates of the pre- and post-regime shift means will be confounded with the estimate of $\sigma_{R}$.

To resolve the problem of confounding between the estimates of the pre-1977 and post-1976 recruitment $\log$-scale means with the estimate of $\sigma_{R}$, the following iterative algorithm was adopted in last year's assessment (Thompson and Dorn 2005) and retained this year to implement the 1977 environmental regime shift in SS2:

1) Candidate values for the pre-1977 log-scale mean and $\sigma_{R}$ were chosen.
2) SS2 was allowed to estimate the post-1976 log-scale mean and the recruitment deviations for the entire time series (deviations are expressed as the difference between the logarithm of annual recruitment at age 0 and the log-scale mean for the respective environmental regime), conditional on the candidate values for the pre-1977 log-scale mean and $\sigma_{R}$.
3) The mean of the estimated pre-1977 recruitment deviations and the standard deviation of the entire time series of recruitment deviations were computed.
4) If the absolute value of the mean computed in Step 3 was less than 0.005 and the standard deviation computed in Step 3 was equal to $\sigma_{R}$ within three significant digits, the candidate values were determined to be the final estimates. If either of these conditions did not hold, the candidate value for the pre-1977 log-scale mean was set equal to the old value plus the mean computed in Step 3, the candidate value for $\sigma_{R}$ was set equal to the standard deviation computed in Step 3, and the process returned to Step 2. (Occasionally, the change in candidate values between iterations deviated slightly from this algorithm if the prescribed changes seemed to small or too large.)
The above algorithm was tested many times under different initial candidate values and consistently returned the same final estimates, so long as the initial candidate values were feasible. It should also be noted that the path to convergence was not always smooth or rapid.

## Prior Distributions

If an informative prior distribution was placed on a parameter, it is described in the following paragraphs (all distributions are normal). If a particular parameter is not listed, it is because a noninformative prior (i.e., a normal distribution with a very large variance) was used. Except for the prior distribution for shelf bottom trawl survey catchability, all priors are identical to those used in last year's assessment
(Thompson and Dorn 2005).

## Parameters with priors based on a specified coefficient of variation (CV)

Log shelf bottom trawl survey catchability $\ln (Q)$ : A mean of zero and a standard deviation of 0.294 were specified, corresponding to a lognormal prior distribution on $Q$ with a mean of 1.0 and a CV of $30 \%$, corresponding to the mean and CV used to specify a prior distribution for $Q$ during the late 1990s.

Initial fishing mortality: The mean was set at 0.1 , reflecting the conventional wisdom that the stock was lightly exploited during the 1960 s . The standard deviation was set at 0.03 , corresponding to a CV of $30 \%$.

Double logistic selectivity parameter $S($ Lmin ): For the EBS slope bottom trawl survey, the Japanese and U.S. longline surveys, and all commercial fisheries, this was not an estimated parameter, but was set at a fixed value of 0.001 . This choice was based on the fact that almost no fish in the sub- 18 cm range are taken by these gears and because preliminary model runs invariably resulted in this parameter being bound at whatever minimum value was specified. For the EBS shelf bottom trawl surveys, the prior distribution was assigned a mean of 0.2 and a standard deviation of 0.06 , corresponding to a $30 \% \mathrm{CV}$. In contrast to the commercial fisheries, $12 \%$ of the average shelf bottom trawl survey size composition since 2000 has consisted of fish smaller than 18 cm .

Double logistic selectivity parameters slope1 and slope2: These two parameters had identical priors, with the mean set at 0.2 and the standard deviation set at 0.06 , corresponding to a $30 \% \mathrm{CV}$. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of these parameters.

Double logistic selectivity parameter peak width: The mean was set at 10 and the standard deviation was set at 3 , corresponding to a $30 \% \mathrm{CV}$. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of this parameter, in addition to results from preliminary model runs which, for the double logistic form at least, indicated that values much higher than 10 tended to cause the model to get "stuck." Although the peak width parameter is also used in the double normal functional form, an informative prior was not specified when the parameter was used in that context.

## Parameters with priors based on one or both endpoints of the $98 \%$ confidence interval

Double logistic selectivity parameters logit(infl1) and logit(infl2): These two parameters had identical priors, with the mean set at 0 and the standard deviation set at 0.944 . The mean corresponds to an inflection point located midway between Lmin and peak location, in the case of infl1, or between peak location + peak width and Lmax, in the case of infl2. The mean and standard deviation together imply a $98 \%$ confidence interval extending from $10 \%$ to $90 \%$ of the difference between Lmin and peak location, in the case of infl1, or between peak location + peak width and Lmax, in the case of infl2. The choice of mean was based on a subjective examination of the shape of the selectivity curve under different values of these parameters.

Double logistic selectivity parameter logit(S(Lmax)): The mean was set at 2.197 and the standard deviation was set at 0.944 . The mean corresponds to a selectivity of 0.9 at Lmax. The mean and standard deviation together imply a $1 \%$ chance of selectivity at Lmax being less than 0.5 . These parameter values were chosen in part to reflect the Plan Team's belief that selectivity of large fish in the bottom trawl survey should be fairly high.

## Parameters with priors based on the data

Selectivity parameter peak location (used in both the double logistic and double normal functional forms): The mean and standard deviation were set individually for each selectivity curve by identifying the length associated with the maximum frequency in each length frequency record, then computing the mean and standard deviation (weighted by the square root of sample size) for each respective gear type and portion
of the time series. This was done in order to give the model a reasonable starting value and place reasonable constraints on peak location, a parameter which is typically very difficult to estimate. Extensive testing during the 2005 assessment (Thompson and Dorn 2005) indicated that the value of this parameter can be quite important in determining model results and that free estimation (with a reasonably strong prior) was much more likely to find an optimal value than profiling manually over the range of possible integer values, especially considering the practical difficulty of manually tuning 14-16 such parameters (one peak location for each selectivity curve) at the same time. The resulting means (cm) and standard deviations ( cm ) for peak location in each of the potential 16 selectivity curves were as follow:

| Fishery/Survey | Years | Mean | Std. Dev. |
| :--- | :---: | ---: | ---: |
| Jan-May Trawl Fishery | $1964-1988$ | 60.7 | 9.4 |
| Jan-May Trawl Fishery | $1989-1999$ | 58.9 | 10.6 |
| Jan-May Trawl Fishery | $2000-2006$ | 64.1 | 26.8 |
| Jul-Dec Trawl Fishery | $1964-1988$ | 61.5 | 9.2 |
| Jul-Dec Trawl Fishery | $1989-1999$ | 62.7 | 12.7 |
| Jul-Dec Trawl Fishery | $2000-2006$ | 60.6 | 10.2 |
| Longline Fishery | $1964-1988$ | 63.4 | 6.4 |
| Longline Fishery | $1989-1999$ | 62.6 | 4.6 |
| Longline Fishery | $2000-2006$ | 59.2 | 3.2 |
| Pot Fishery | $1989-1999$ | 63.9 | 4.3 |
| Pot Fishery | $2000-2006$ | 61.2 | 3.2 |
| Shelf Bottom Trawl Survey | $1979-1981$ | 41.7 | 6.9 |
| Shelf Bottom Trawl Survey | $1982-2006$ | 35.4 | 11.8 |
| Slope Bottom Trawl Survey | $2002-2004$ | 55.1 | 5.0 |
| Japanese Longline Survey | $1982-1994$ | 64.2 | 4.3 |
| U.S. Longline Survey | $1997-2005$ | 62.9 | 2.5 |

## Likelihood Components

Likelihood components included in all nine models were of five types: size composition, age composition, survey abundance, mean size at age, and recruitment deviations. All nine models included at least seven size composition components in the likelihood: one each for the January-May trawl fishery, the June-December trawl fishery, the longline fishery, the pot fishery, the pre-1982 shelf trawl survey, the post-1981 shelf trawl survey, the slope trawl survey. In addition, Models C1, C2, D1, and D2 included size composition components for the Japanese longline survey and the U.S. longline survey. Only one age composition component and one size-at-age component appear in the likelihood, because all age data currently come from the post- 1982 shelf trawl survey. All nine models included at least three survey abundance components in the likelihood: one each for the pre-1982 shelf trawl survey, the post-1981 shelf trawl survey, and the slope trawl survey. In addition, Models C1, C2, D1, and D2 included survey abundance components for the Japanese longline survey and the U.S. longline survey.
In SS2, emphasis factors are specified to determine which likelihood components receive the greatest attention during the parameter estimation process. The prior distributions are also assigned an emphasis. As in previous assessments, each likelihood component in each model was given an emphasis of 1.0 in the present assessment. The prior distributions were given an emphasis of 1.0 in Models $0, \mathrm{~A} 1, \mathrm{~B} 1, \mathrm{C} 1$, and D1 and an emphasis of 0.5 in Models A2, B2, C2, and D2.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery, and time period within the year. In the parameter estimation process, SS2 weights a given size composition observation (i.e., the size frequency distribution observed in a given year, gear/fishery, and period) according to the emphasis associated with the respective likelihood component
and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which SS1 was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to compensate for contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. As in previous assessments, the present assessment uses a multinomial sample size equal to the square root of the true length sample size, rather than the true length sample size itself. Given the true length sample sizes observed in the EBS Pacific cod data, this procedure tends to give values somewhat below 400 while still providing SS2 with usable information regarding the appropriate effort to devote to fitting individual length samples. Multinomial length sample sizes derived by this procedure for the commercial fishery size compositions are shown in Tables 2.7-2.9, for the shelf bottom trawl surveys in Tables 2.10a and 2.10b, for the slope bottom trawl survey in Table 2.13, for the Japanese longline survey in Table 2.14a, and for the U.S. longline survey in Table 2.14b.

## Use of Age Composition Data in Parameter Estimation

Like the size composition data, the age composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear/fishery (in this case, the EBS shelf bottom trawl survey), and time period within the year (in this case, the June-August period). However, selection of an appropriate input sample size is more complicated for age composition data than for length composition data, because age composition data are generated not only from the set of otolith readings but from the estimated size composition as well. Therefore, even if a square root transformation is appropriate for size composition data, taking the square root of the number of otoliths read may underestimate the weight that should be given to the age composition data. The 2004 assessment (Thompson and Dorn 2004) introduced a method for setting an input sample size appropriate to age composition, a method which has been retained since. The steps are as follow:

1) The proportions of age at length are assumed to be approximately multivariate normally distributed, with a variance-covariance matrix determined by the matrix of proportions and the number of otoliths actually read at each length. A set of 10,000 random age-length keys was then simulated.
2) Survey numbers at each length are assumed to be approximately lognormally distributed with a mean equal to the point estimate and for that length and a constant (across lengths) coefficient of variation (CV) equal to the amount that sets the sum of the variances in numbers at length equal to the variance of the survey estimate of population size. A set 10,000 of random numbers-atlength distributions was then simulated.
3) For each combination of randomly simulated age-key and numbers-at-length distribution, an effective sample size was computed.
4) The input sample size was set equal to the harmonic mean of the distribution of randomly simulated effective sample sizes, based on the asymptotic equivalence of these two quantities. The following table was thereby obtained for the age composition data (the last row shows the values used as input sample sizes):

| Year | 1994 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Number of fish aged: | 715 | 252 | 719 | 635 | 860 | 864 | 950 | 947 | 1360 | 1040 | 609 |
| Sqrt. of no, fish aged: | 27 | 16 | 27 | 25 | 29 | 29 | 31 | 31 | 37 | 32 | 25 |
| CV of nos. at length: | 0.78 | 0.93 | 1.12 | 0.51 | 0.60 | 0.63 | 0.63 | 0.65 | 0.87 | 0.64 | 1.06 |
| Harmonic mean: | 67 | 43 | 47 | 107 | 131 | 136 | 111 | 108 | 77 | 157 | 53 |

Note that this procedure gives an input sample size larger than would be achieved simply by taking the square root of the number of fish aged (third row in the above table). This reflects the added precision achieved by use of both age-at-length and numbers-at-length data in constructing a numbers-at-age
estimate. To avoid double counting of the same data, all nine models ignore length composition data from the EBS shelf bottom trawl surveys in years where age data are available.

It may be noted that all but one of the harmonic mean effective sample sizes computed above is smaller than the sample sizes obtained for the corresponding length compositions using the square root method in the preceding subsection, suggesting that the two methods of computing sample sizes are not entirely consistent. This is not surprising, given that the square root method was adopted only as a simple approximation in the first place, but it does suggest a need for further work in this area.

## Use of Size-at-Age Data in Parameter Estimation

Each size at age datum is assumed to be drawn from a normal distribution specific for that age and year. The model's estimate of mean size at age serves as the mean for that year's distribution, and the standard deviation is inversely proportional to the sample size (Methot 2000, Methot 2005a).

## Use of Survey Abundance Data in Parameter Estimation

Each year's survey abundance datum is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey abundance in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey abundance datum's standard error to the survey biomass datum itself serves as the distribution's coefficient of variation.

## Use of Recruitment Deviation "Data" in Parameter Estimation

The recruitment deviations likelihood component is different from traditional likelihoods because it does not involve "data" in the same sense that traditional likelihoods do. Instead, the log-scale recruitment deviation plays the role of the datum and the log-scale recruitment mean and $\sigma_{R}$ play the role of the parameters in a normal distribution, but, of course, all of these are treated as parameters by SS2.

## MODEL EVALUATION

As described in the preceding section, nine models are evaluated in the present assessment. Model 0 is very similar to the model selected last year by the Plan Team and SSC, except for use of updated values for those parameters that are estimated independently (i.e., outside of the SS2 model). Model 0 fixes the catchability coefficient for the EBS shelf bottom trawl survey at the traditional value of 1.0. The eight alternative models attempt to estimate catchability for all surveys, and differ from one another with respect to use or exclusion of longline survey data, choice of functional form for selectivity, and the weight assigned to prior distributions in the objective function. All models appeared to converge successfully and the Hessian matrices from all models were positive definite. However, it should be noted that it was typically more difficult to achieve convergence for the models associated with deemphasized prior distributions (Models A2, B2, C2, and D2). To achieve convergence, those models were initialized with the parameter estimates from their respective "full prior" counterparts. Even then, convergence was sometimes achieved only after considerable trial and error, particularly in the case of models utilizing the double logistic selectivity function (Models A2 and C2). Also, models using the longline survey data (Models C1, C2, D1, and D2) had a difficult time converging unless estimation of the longline survey selectivity parameters was moved to the last phases in the estimation routine.

## Overall Conclusions Common to All Models

Before choosing a preferred model, it is important to note that, in many respects, the descriptions of the stock provided by all of the models are, qualitatively at least, very similar. For example, Figure 2.2 compares numbers of age 0 fish for the years 1977-2005 as estimated by all the models. All the models are in basic agreement as to which year classes appear to be strong and which appear to be weak (of course, there is estimation error associated with all of the points shown in Figure 2.2, but to keep the figure legible, only the point estimates are shown). In particular, all the models agree that the 2000-2004
year classes currently appear to be weak. Figure 2.3 compares female spawning biomass for the years 1977-2006 as estimated by all the models. The overall shapes of all the estimated time series are again qualitatively similar, with the main difference being one of scale. From about 1993 to the present, all models indicate that female spawning biomass has been fairly stable, although the trend over the last couple of years is downward in all models. As far as the prognosis for the future is concerned, again the models are in qualitative agreement, with all models projecting continued declines for the next 2-3 years, as shown in Figure 2.4 (note that the spawning biomasses in Figure 2.3 are from the assessment model, which is configured for the EBS portion of the stock only, whereas the spawning biomasses in Figure 2.4 are from the projection model, which is configured for the overall BSAI stock, so the endpoints of the two time series do not match). It should be emphasized that the projections shown in Figure 2.4 represent the average of a large number of stochastic projections. The averages rather than the ranges are plotted because of the large number of models being compared.

## Comparing and Contrasting the Models

Table 2.16 presents a summary of some key results from last year's assessment (based on the model chosen by the Plan Team and SSC) and compares them with the corresponding results from Model 0 and the eight alternative models. The table is structured as follows:

Row 1: Model names.
Rows 2-4: Factors that distinguish the eight alternative models from each other.
Rows 5-7: Parameters governing the distribution of recruitments. Row 5 shows the standard deviation of the distribution of log-scale recruitment deviations, row 6 shows the median logscale recruitment for the post-1976 environmental regime, and row 7 shows the $\log$ of the ratio of median $\log$-scale recruitments between the pre-1977 and post-1976 environmental regimes (i.e., a negative value in row 7 means that median recruitment was lower in the pre-1977 regime than in the post-1976 regime).

Rows 8-10: Parameters or function values characterizing shelf trawl survey catchability and selectivity. Row 8 shows the catchability for the pre-1982 portion of the time series, row 9 shows the catchability for the post-1981 portion of the time series, and row 10 shows the estimated post1981 shelf trawl survey selectivity for fish 90 cm in length. The full selectivity schedules for the post-1981 shelf trawl survey are compared in Figure 2.5.

Rows 11-15: Log likelihood values related to survey abundance indices (by convention, all log likelihood, $\log$ prior, and $\log$ objective function values are multiplied by -1 ). These rows show the values of the log likelihoods pertaining to the abundance data from the pre-1982 shelf trawl survey, post-1981 shelf trawl survey, slope trawl survey, Japanese longline survey, and U.S. longline survey, respectively.

Rows 16-24: Log likelihood values related to size composition. These rows show the values of the log likelihoods pertaining to the size composition data from the January-May trawl fishery, June-December trawl fishery, longline fishery, pot fishery, pre-1982 shelf trawl survey, post-1981 shelf trawl survey, slope trawl survey, Japanese longline survey, and U.S. longline survey, respectively.

Rows 25-27: Other log likelihoods. Row 25 shows the log likelihood pertaining to the post-1981 shelf trawl survey age composition data, row 26 shows the log likelihood pertaining to the post1981 shelf trawl survey size-at-age data, and row 27 shows the log likelihood pertaining to recruitment deviations.

Row 28: Log prior distributions.

Row 29: Log posterior distribution (the objective function). This row shows the sum of the previous 18 rows, except that the log prior distribution is weighted by a factor of 0.5 in Models A2, B2, C2, and D2.

Table 2.17 continues the comparison by presenting results for several management-related quantities. Values obtained from the SS2 model are shown in normal font and values obtained from the projection model are shown in bold font. All values pertain to the overall BSAI stock, not just the EBS portion of the stock assessed by the SS2 model. The table is structured as follows:

Rows 1-4: Same as Table 2.16.
Rows 5-6: BSAI total biomass for 2005 and 2006.
Rows 7-10: BSAI female spawning biomass for 2005-2008. Note that there is a mismatch between values obtained from SS2 and those obtained from the projection model, because SS2 computes spawning biomass at the start of the year whereas the projection model computes spawning biomass at the month of peak spawning.

Rows 11-14: BSAI female spawning biomass for 2005-2008 expressed as a proportion of equilibrium unfished spawning biomass (again, there is a slight mismatch between the SS2 and projection model estimates of equilibrium unfished spawning biomass).

Rows 15-19: Current (2006) BSAI ABC and projected maximum permissible ABC for 20072008, with the proportional year-to-year changes implied by those ABCs.

Rows 20-24: Similar to rows 15-19, but for OFL instead of ABC.
For the length composition and age composition components of the likelihood, past assessments have included a comparison of input sample sizes and "effective" output sample sizes. The rationale is as follows: Once maximum likelihood estimates of the model parameters have been obtained, SS2 computes an "effective" sample size for the length or age composition data specific to a particular year, gear, and season within the year. Roughly, the effective sample size can be interpreted as the multinomial sample size that would typically be required in order to produce the given fit. More precisely, it is the sample size that sets the sum of the marginal variances of the proportions implied by the multinomial distribution equal to the sum of the squared differences between the sample proportions and the estimated proportions (McAllister and Ianelli 1997). As a function of a multinomial random variable, the effective sample size has its own distribution. The harmonic mean of the distribution is asymptotically equal to the true sample size in the multinomial distribution. Thus, if the effective sample size is less than the true sample size in the multinomial distribution, it is reasonable to conclude that the fit is not as good as expected. The following table shows the average of the input sample sizes (Input N ) for each length or age composition component and the ratio between the average effective sample size and the average input sample size under each model (a higher ratio implies a better fit):

|  |  |  |  | Model |  |  |  |  |  |  |  |
| :--- | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Gear | Type | Input N | 0 | A1 | A2 | B1 | B2 | C1 | C2 | D1 |
| D2 |  |  |  |  |  |  |  |  |  |  |  |
| Jan-May trawl fish. | Length | 169 | 1.55 | 1.74 | 1.86 | 1.52 | 1.52 | 1.91 | 2.00 | 1.50 | 1.51 |
| Jun-Dec trawl fish. | Length | 42 | 1.96 | 2.07 | 2.12 | 1.99 | 1.94 | 1.97 | 1.97 | 1.97 | 1.94 |
| longline fishery | Length | 191 | 1.58 | 1.57 | 1.61 | 1.79 | 1.80 | 1.54 | 1.56 | 1.74 | 1.76 |
| pot fishery | Length | 100 | 2.33 | 2.29 | 2.40 | 2.44 | 2.45 | 2.31 | 2.45 | 2.55 | 2.57 |
| pre-82 shelf survey | Length | 100 | 0.72 | 0.70 | 0.69 | 0.64 | 0.64 | 0.57 | 0.56 | 0.54 | 0.54 |
| post-81 shelf survey | Length | 104 | 1.11 | 1.12 | 1.11 | 0.93 | 0.93 | 1.08 | 1.07 | 0.93 | 0.93 |
| slope survey | Length | 23 | 5.00 | 4.68 | 5.85 | 10.27 | 9.91 | 5.56 | 6.98 | 11.21 | 10.65 |
| Japan LL survey | Length | 140 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 1.22 | 1.19 | 1.23 | 1.23 |
| U.S. LL survey | Length | 88 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 2.17 | 2.21 | 1.95 | 2.12 |
| post-81 shelf survey | Age | 94 | 0.63 | 0.62 | 0.61 | 0.60 | 0.60 | 0.61 | 0.61 | 0.62 | 0.61 |

Two points should be noted regarding the shelf survey length composition components: 1) The true input sample sizes for the pre-1982 portion of the time series are unknown, so the assumed value of 100 is only a guess. 2) To avoid double-counting, results for the post-1981 shelf survey length composition component do not include years for which age data are available.

## Evaluation Criteria

The values of the various components of the objective function are often fairly close across models, or involve tradeoffs that make it difficult to choose one model over another. The same conclusion holds for the effective sample sizes associated with the length and age composition data. Because all of the models seem to perform reasonably well in terms of fitting the data, the following criteria are therefore proposed:

1) The model should describe a plausible selectivity schedule for the post-1981 shelf trawl survey.
2) The model should not depend on data that require further validation before they can be considered ready for use in the stock assessment.
3) The model should converge well (e.g., not be too dependent on initial parameter estimates).
4) The model should not depend too strongly on the prior distributions.

## Selection of Final Model

Criterion \#1 argues against choosing Model 0. As Figure 2.5 shows, Model 0 's selectivity schedule for the post-1981 shelf trawl survey shows a pronounced kink that is very difficult to justify on theoretical grounds. The eight alternative models all result in much more plausible selectivity schedules for this survey.

Criterion \#2 argues against choosing Models C1, C2, D1, and D2, which are the models that utilize data from the Japanese longline survey and the U.S. longline survey. While it may be possible to develop usable indices from these surveys in the future, the present indices seem too problematic, for the following reasons: 1) the available abundance indices for Pacific cod (unlike those for sablefish) do not include appropriate area expansion factors, 2) the interannual variability in the available abundance indices from the Japanese longline survey is extreme, and 3) the sample size in the U.S. longline survey is small (only 11 stations have been successfully sampled in every year).
Criterion \#3 argues against Models A2, B2, C2, and D2, which are the models with de-emphasized priors and that typically had to be initiated with the converged parameter estimates from their respective "full prior" counterparts in order to converge successfully. Also, the models that used the longline survey data (Models C1, C2, D1, and D2) had difficulty converging unless estimation of longline survey selectivity parameters was delayed until other parameters had been estimated (i.e., moved to a later phase).

Criterion \#4 argues against Model A1. Using relative change in estimated 2006 spawning biomass as an indicator of sensitivity, Model A1 is seen to be much more sensitive to the emphasis assigned to the prior distributions than any of the other "full prior" models (Models B1, C1, and D1). The relative change in 2006 spawning biomass between Model A1 (full prior) and Model A2 (de-emphasized prior) was $16 \%$, compared to $-3 \%, 3 \%$, and $-2 \%$ for the relative changes between Models B1 and B2, C1 and C2, and D1 and D 2 , respectively.

By process of elimination, then, Model B1 is therefore recommended as the preferred model. If Model A1, which has many more parameters, were to have given results substantially different from Model B1, it might be argued that Model B1 is under-parameterized. However, results from Models A1 and B1 are fairly similar, indicating that the more parsimonious parameterization used in Model B1 does not cause the model to overlook key details. It may also be noted that Model B2 gives results extremely similar to those from Model B1, suggesting that this model could also be a viable candidate, particularly in a future
assessment if further work confirms the stability of the model when less informative priors are specified. Another consideration pertaining to future assessment work is that Model B1 may have potential to overcome some of the past difficulties encountered in attempting to estimate $M$ and $Q$ for Pacific cod using models based on the double logistic selectivity function.

## Final Parameter Estimates and Associated Schedules

Final estimates of some key scalar parameters (i.e., parameters that do not define length-specific schedules) corresponding to Model B1 are shown in Table 2.16. Another scalar parameter estimated by SS2 is the equilibrium fishing mortality rate at the start of the time series, which had a value of 0.075 in Model B1.

Estimates of year-, season-, and gear-specific fishing mortality rates from Model B1 are shown in Table 2.18, estimates of regime-specific median recruitments and annual recruitment deviations from Model B1 are shown in Table 2.19, and estimates of selectivity parameters from Model B1 are shown in Table 2.20.

Schedules of selectivity at length from Model B1 are shown for the commercial fisheries in Table 2.21a and for the bottom trawl surveys in Table 2.21b. The schedules in Tables 2.21a and 2.21b are plotted in Figure 2.6.
Schedules of length at age, proportion mature at age, and weight at age from Model B1 are shown in Table 2.22.

## RESULTS

## Definitions

The biomass estimates presented here will be defined in three ways: 1) age $3+$ biomass, consisting of the biomass of all fish aged three years or greater in January of a given year; 2) spawning biomass, consisting of the biomass of all spawning females in a given year; and 3) survey biomass, consisting of the biomass of all fish that the model estimates should have been observed by the survey in July of a given year. The recruitment estimates presented here will be defined as numbers of age 0 fish in a given year. The fishing mortality rates presented here will be defined as full-selection, instantaneous fishing mortality rates expressed on a per annum scale. In all comparisons involving last year's results, it is important to note that table entries labeled "Last Year's Values" do not correspond to the values given in last year's SAFE report, because the values given in last year's SAFE report corresponded to the authors' preferred model, not the model chosen by the Plan Team and SSC. Instead, table entries labeled "Last Year's Values" correspond to the results given last year under the model chosen by the Plan Team and SSC.

## Biomass

Table 2.23 shows the time series of EBS (not expanded to BSAI) Pacific cod female spawning biomass for the years 1977-2006 as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1. Both estimated time series are accompanied by their respective $95 \%$ confidence intervals.

The estimated EBS female spawning biomass time series and confidence intervals from Model B1 are shown, together with the Model B1's estimated time series of EBS age 3+ biomass, in Figure 2.7. Figure 2.7 also compares the observed and model-estimated time series from the EBS shelf bottom trawl survey. All three biomass trends estimated by Model B1 are fairly flat from about 1992 through about 2004, but all three show a declining trend for at least the last couple of years.

## Recruitment

Table 2.24 shows the time series of EBS (not expanded to BSAI) Pacific cod age 0 recruitment (1000s of fish) for the years 1977-2005 as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1. Both estimated time series are accompanied by their respective $95 \%$ confidence intervals.

Model B1's recruitment estimates for the entire time series (1964-2005) are shown in Figure 2.8, along with their respective $95 \%$ confidence intervals and regime-specific averages. For the time series as a whole, the largest year classes appear to have been the 1976-1977 cohorts. Other large cohorts include the $1978,1982,1984,1989,1992,1996$, and 1999 year classes. Of the five classes spawned immediately after the strong 1999 year class, however, none have $95 \%$ confidence intervals that extend above the 1977-2005 average. One potential bright spot on the horizon is the 2005 year class, whose point estimate is just below the 1977-2005 average. However, its confidence interval is fairly large, since the only data currently available to estimate its strength is the size composition data from the 2006 shelf trawl survey.

To date, it has not been possible to estimate a reliable stock-recruitment relationship for this stock. With the move to SS2, prospects for future estimation of such a relationship should improve. In the interim, Figure 2.9 is provided to give some indication of the relationship between stock and recruitment. The Ricker (1954) curve shown in this figure (fit by maximum likelihood, ignoring process error) is intended to be illustrative only, and is not recommended for management purposes.

## Exploitation

Table 2.25 shows the time series of EBS Pacific cod catch divided by age 3+ biomass for the years 19772006 as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1.

The average value of this ratio over the entire time series is about 0.12 , slightly less than the average value of 0.13 obtained in the model chosen last year by the Plan Team and SSC. The estimated values exceed the average for every year after 1989 except 1993, whereas none of the estimated values exceed the average in any year prior to 1990. This finding is similar to that obtained in past assessments.

Figure 2.10 plots the trajectory of relative fishing mortality and relative female spawning biomass from 1977 through 2006 based on Model B1, overlaid with the current harvest control rules (fishing mortality rates in the figure are standardized relative to $F_{35 \%}$ and biomasses are standardized relative to $B_{35 \%}$, per SSC request). The entire trajectory lies underneath the $F_{\text {OFL }}$ control rule except for the years 1977-1979. For the period since 1980, the entire trajectory also fell below the $\operatorname{maxF}_{A B C}$ control rule, except for 1995 and 1997, when the fishing mortality rate appears to have exceeded the retroactively calculated $\max _{A B C}$. It should also be noted that the current harvest control rules did not go into effect until 1999.

## PROJECTIONS AND HARVEST ALTERNATIVES

## Amendment 56 Reference Points

Amendment 56 to the BSAI Groundfish Fishery Management Plan (FMP) defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL ( $F_{O F L}$ ), the maximum permissible ABC , and the fishing mortality rate used to set the maximum permissible ABC. The fishing mortality rate used to set ABC ( $F_{A B C}$ ) may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the BSAI are managed under Tier 3 of Amendment 56. Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the
fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

$$
\begin{aligned}
& \text { 3a)Stock status: } B / B_{40 \%}>1 \\
& F_{O F L}=F_{35 \%} \\
& F_{A B C} \leq F_{40 \%} \\
& \text { 3b)Stock status: } 0.05<B / B_{40 \%} \leq 1 \\
& F_{O F L}=F_{35 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& F_{A B C} \leq F_{40 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95
\end{aligned}
$$

3c)Stock status: $B / B_{40 \%} \leq 0.05$

$$
\begin{aligned}
& F_{O F L}=0 \\
& F_{A B C}=0
\end{aligned}
$$

Estimation of the $B_{40 \%}$ reference point used in the above formulae requires an assumption regarding the equilibrium level of recruitment. In this assessment, it is assumed that the equilibrium level of recruitment is equal to the post-1976 average (i.e., the arithmetic mean of all estimated recruitments from year classes spawned in 1977 or later). Other useful biomass reference points which can be calculated using this assumption are $B_{100 \%}$ and $B_{35 \%}$, defined analogously to $B_{40 \%}$. These reference points are estimated as follows, based on Model B1:

$$
\begin{array}{rccc}
\text { Reference point: } & B_{35 \%} & B_{40 \%} & B_{100 \%} \\
\text { BSAI: } & 280,000 \mathrm{t} & 320,000 \mathrm{t} & 800,000 \mathrm{t} \\
\text { EBS: } & 235,000 \mathrm{t} & 269,000 \mathrm{t} & 672,000 \mathrm{t}
\end{array}
$$

For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. For this assessment, the apportionment was based on Model B1's estimates of fishing mortality by gear for the three most recent complete years of data (2003-2005). The average fishing mortality rates for those years implied that total fishing mortality was divided among the three main gear types according to the following percentages: trawl $31.0 \%$, longline $58.8 \%$, and pot $10.2 \%$. This apportionment results in estimates of $F_{35 \%}$ and $F_{40 \%}$ equal to 0.42 and 0.34 , respectively.

## Specification of OFL and Maximum Permissible ABC

BSAI spawning biomass for 2007 is estimated by Model B1 at a value of $307,000 \mathrm{t}(\mathrm{EBS}$ value $=258,000$ t ). This is about $4 \%$ below the BSAI $B_{40 \%}$ value of $320,000 \mathrm{t}($ EBS value $=269,000 \mathrm{t}$ ), thereby placing Pacific cod in sub-tier "b" of Tier 3. Given this, Model B1 estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2007 as follows:

| Quantity | Overfishing Level | Maximum Permissible ABC |
| :--- | ---: | ---: |
| EBS catch: | $174,000 \mathrm{t}$ | $148,000 \mathrm{t}$ |
| BSAI catch: | $207,000 \mathrm{t}$ | $176,000 \mathrm{t}$ |
| Fishing mortality rate: | 0.39 | 0.33 |

The age 3+ biomass estimates for 2007 from Model B1 are $960,000 \mathrm{t}$ and $807,000 \mathrm{t}$ for the BSAI and EBS, respectively.

## ABC Recommendation

## Review of Past Approaches

BSAI Pacific cod ABCs for the years 1998-2002 were based on a harvest strategy that attempted to address some of the statistical uncertainty in the assessment model, namely the uncertainty surrounding parameters the natural mortality rate $M$ and survey catchability $Q$ (Thompson and Dorn 1997, 1998, 1999). For the 2001-2002 ABCs, the strategy was simplified by assuming that the ratio between the recommended $F_{A B C}$ and $F_{40 \%}$ estimate given in the 1999 assessment ( 0.87 ) was an appropriate factor by which to multiply the current maximum permissible $F_{A B C}$ to obtain a recommended $F_{A B C}$ (Thompson and Dorn 2001). For the 2003 and 2004 ABCs, concerns regarding the performance of the assessment model led to a decision that kept ABC constant at the 2002 level of $223,000 \mathrm{t}$, well below the maximum permissible level estimated in the respective assessments (Thompson and Dorn 2002, 2003). In the 2004 assessment (Thompson and Dorn 2004), the maximum permissible value for the 2005 ABC was estimated to be $227,000 \mathrm{t}$, only slightly higher than the 2003-2004 ABCs of 223,000 t. Because the 2003-2004 "constant catch" ABCs were intended to provide a precautionary alternative to the model's maximum permissible ABCs, it seemed appropriate in the 2004 assessment to consider another method for recommending ABC . This method was based on a consideration of the mean-variance tradeoff associated with future catches predicted by the standard projection model, and resulted in a 2005 ABC of 206,000 t. In the 2005 assessment, the Plan Team and SSC selected a model that resulted in a maximum permissible ABC of $194,000 \mathrm{t}$, which was adopted as the 2006 ABC.

## Recommendation for 2007

Based on Model B1, the maximum permissible ABC (Tier 3b) for 2007 is $176,000 \mathrm{t}$. To provide some context for this value, the time series of ABCs for the 16 years following 1990 shows that ABC has ranged from a low of $164,500 \mathrm{t}$ to a high of $328,000 \mathrm{t}$, with an average of about $221,000 \mathrm{t}$, (Table 2.4). A 2007 ABC of $176,000 \mathrm{t}$ would be the second lowest ABC since 1990, and the decrease from the 2006 ABC ( $14,000 \mathrm{t}$ ) would represent the seventh largest one-year decrease in the time series since 1990. Given the magnitude of this decrease and the fact that it follows immediately on the heels of two consecutive decreases of similar magnitude, there does not seem to be any compelling reason to recommend an ABC lower than the maximum permissible value for 2007. Therefore, $176,000 \mathrm{t}$ is the recommended ABC for 2007. It should be noted that all models considered in this year's assessment, including Model B1, project the maximum permissible ABC to continue declining for at least the next couple of years while the weak 2000-2004 year classes work their way through the age structure.

## Area Allocation of Harvests

At present, ABC of BSAI Pacific cod is not allocated by area. However, the Council is presently considering the possibility of specifying separate harvests in the EBS and AI.

## Standard Harvest and Recruitment Scenarios and Projection Methodology

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2006 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2007 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end)
catch for 2006. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2007, are as follow ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2007 recommended in the assessment to the max $F_{A B C}$ for 2007. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2002-2006 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{\text {TAC }}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be 1) above its MSY level in 2007 or 2) above $1 / 2$ of its MSY level in 2007 and above its MSY level in 2017 under this scenario, then the stock is not overfished.)

Scenario 7: In 2007 and 2008, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2019 under this scenario, then the stock is not approaching an overfished condition.)

## Projections and Status Determination

## Scenario Projections and Two-Year Ahead Overfishing Level

Projections corresponding to the standard scenarios are shown for Model B1 in Tables 2.26-2.31 (Table 2.26 combines scenarios 1 and 2 , which are redundant).

In addition to the seven standard harvest scenarios, Amendments 48/48 to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2007, it does not provide the best estimate of OFL for 2008, because the mean 2007 catch under Scenario 6 is predicated on the 2007 catch being equal to the 2007 OFL, whereas the actual 2007 catch will likely be less than the 2007 OFL. Table 2.17 contains the appropriate one- and two-year ahead projections for both ABC and OFL under any of the nine models considered in the present assessment.

## Status Determination

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock overfished? This depends on the stock's estimated spawning biomass in 2007:
a. If spawning biomass for 2007 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2007 is estimated to be above $B_{35 \%}$ the stock is above its MSST.
c. If spawning biomass for 2007 is estimated to be above $1 / 2 B_{35 \%}$ but below $B 35 \%$, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 2.30). If the mean spawning biomass for 2017 is below $B_{35 \%}$, the stock is below its MSST. Otherwise, the stock is above its MSST.
Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7 (Table 2.31):
a. If the mean spawning biomass for 2009 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2009 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c. If the mean spawning biomass for 2009 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2019. If the mean spawning biomass for 2019 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

In the case of BSAI Pacific cod, spawning biomass for 2007 is estimated to be above $B_{35 \%}$ under Model B1. Therefore, the stock is above its MSST and is not overfished. Mean spawning biomass for 2009 in Table 2.31 is above $1 / 2 B_{35 \%}$, but below $B_{35 \%}$, and mean spawning biomass for 2019 is above $B_{35 \%}$. Therefore, the stock is not approaching an overfished condition.

## ECOSYSTEM CONSIDERATIONS

Attachment 2.1 contains a summary of new results from ecosystem models on the role of Pacific Cod in the Eastern Bering Sea and Aleutian Islands ecosystems. The material in the present section is largely unchanged from last year's assessment.

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Boldt (ed.), 2005). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g.,

Hare and Mantua 2000). In the present assessment, an attempt was made to estimate the change in median recruitment of EBS Pacific cod associated with the 1977 regime shift. According to Model B1, pre-1977 median recruitment was only about $31 \%$ of post-1976 median recruitment. Establishing a link between environment and recruitment within a particular regime is more difficult. In the 2004 assessment (Thompson and Dorn 2004), for example, the correlations between age 1 recruits spawned since 1977 and monthly values of the Pacific Decadal Oscillation (Mantua et al. 1997) were computed and found to be very weak.

The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), Lang et al. (2003), Westrheim (1996), and Yang (2004). The composition of Pacific cod prey varies to some extent by time and area. In terms of percent occurrence, some of the most important items in the diet of Pacific cod in the BSAI and GOA have been polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, some of the most important dietary items have been euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, some of the most important dietary items have been walleye pollock, fishery offal, yellowfin sole, and crustaceans. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include Pacific cod, halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

## Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Bycatch of Nontarget and "Other" Species

Bycatch of nontarget species and members of the "other species" group are shown in the following set of tables (for the 2003-2005 tables, the "hook and line" gear type includes both longline and jig gear): Tables 2.32a and 2.32b show bycatch for the EBS Pacific cod trawl fishery in 1997-2002 and 2003-2005, respectively. Tables 2.33 a and 2.33 b show bycatch for the EBS Pacific cod longline fishery in 1997-2002 and the EBS Pacific cod hook and line fishery in 2003-2005, respectively. Tables 2.34a and 2.34b show bycatch for the EBS Pacific cod pot fishery in 1997-2002 and 2003-2005, respectively. Tables 2.35a and 2.35 b show bycatch for the AI Pacific cod trawl fishery in 1997-2002 and 2003-2005, respectively. Tables 2.36a and 2.36 b show bycatch for the AI Pacific cod longline fishery in 1997-2002 and the AI Pacific cod hook and line fishery in 2003-2005, respectively. Tables 2.37 shows bycatch for the AI Pacific cod pot fishery in 1997-2002 (no data exist for this fishery in 2003-2005).

It is not clear how much bycatch of a particular species constitutes "too much" in the context of ecosystem concerns. As a first step toward possible prioritization of future investigation into this question, it might be reasonable to focus on those species groups for which a Pacific cod fishery had a bycatch in excess of 100 t and accounted for more than $10 \%$ of the total bycatch in at least two of the three most recent years. This criterion results in the following list of impacted species groups (an "X" indicates that the criterion was met for that area/species/gear combination).

| Area | Species group | Trawl | Hook and Line |
| :---: | :---: | :---: | :---: |
| EBS | Grenadier |  | X |
| EBS | Large sculpins | X | X |
| EBS | Misc. fish | X |  |
| EBS | Other sculpins |  | X |
| EBS | Shark |  | X |
| EBS | Skate |  | X |
| AI | Skate |  | X |

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery operates to some extent in the same geographic areas used by Steller sea lion as foraging grounds (Livingston (ed.), 2002).

The Fisheries Interaction Team of the Alaska Fisheries Science Center has been engaged in research to determine the effectiveness of recent management measures designed to mitigate the impacts of the Pacific cod fisheries (among others) on Steller sea lions. Results from studies conducted in 2002-2003 were summarized by Conners et al. (2004). These studies included a tagging feasibility study, which may evolve into an ongoing research effort capable of providing information on the extent and rate to which Pacific cod move in and out of various portions of Steller sea lion critical habitat. Nearly 6,000 cod with spaghetti tags were released, of which approximately 1,000 had been returned as of September, 2003.

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the hook and line fishery for Pacific cod (Tables 2.33b and 2.36b). Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross (Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft . LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed sets was as follows:

| Gear | BS | AI | GOA |
| :--- | ---: | ---: | ---: |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513,517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort were dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.

Impacts of the Pacific cod fisheries on essential fish habitat were further analyzed in an environmental impact statement by NMFS (2005).

## Data Gaps and Research Priorities

Understanding of the above ecosystem considerations would be improved if future research were directed toward closing certain data gaps. Such research would have several foci, including the following: 1) ecology of the Pacific cod stock, including spatial dynamics, trophic and other interspecific relationships, and the relationship between climate and recruitment; 2) behavior of the Pacific cod fishery, including spatial dynamics; 3) determinants of trawl survey catchability and selectivity; 4) ecology of species taken as bycatch in the Pacific cod fisheries, including estimation of biomass, carrying capacity, and resilience; and 5) ecology of species that interact with Pacific cod, including estimation of biomass, carrying capacity, and resilience.

## SUMMARY

The major results of the Pacific cod stock assessment are summarized in Table 2.38.

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Table 2.1a-Summary of 1964-1980 catches ( t ) of Pacific cod in the Eastern Bering Sea by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

| Eastern Bering Sea only: |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: |
| Year | Foreign | Joint Venture | Domestic | Total |
| 1964 | 13408 | 0 | 0 | 13408 |
| 1965 | 14719 | 0 | 0 | 14719 |
| 1966 | 18200 | 0 | 0 | 18200 |
| 1967 | 32064 | 0 | 0 | 32064 |
| 1968 | 57902 | 0 | 0 | 57902 |
| 1969 | 50351 | 0 | 0 | 50351 |
| 1970 | 70094 | 0 | 0 | 70094 |
| 1971 | 43054 | 0 | 0 | 43054 |
| 1972 | 42905 | 0 | 0 | 42905 |
| 1973 | 53386 | 0 | 0 | 53386 |
| 1974 | 62462 | 0 | 0 | 62462 |
| 1975 | 51551 | 0 | 0 | 51551 |
| 1976 | 50481 | 0 | 0 | 50481 |
| 1977 | 33335 | 0 | 0 | 33335 |
| 1978 | 42512 | 0 | 31 | 42543 |
| 1979 | 32981 | 8370 | 780 | 33761 |
| 1980 | 35058 |  | 2433 | 45861 |

Table 2.1b—Summary of 1981-2005 catches ( t ) of Pacific cod in the Eastern Bering Sea by fleet sector and gear type. All catches include discards. LLine = longline, Subt. = sector subtotal. Catches for 2006 are through early October.

Eastern Bering Sea only:

|  | Foreign |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |
| 1981 | 30347 | 5851 | 36198 | 7410 | 7410 | 12884 | 1 | 0 | 14 | 12899 | 56507 |  |  |
| 1982 | 23037 | 3142 | 26179 | 9312 | 9312 | 23893 | 5 | 0 | 1715 | 25613 | 61104 |  |  |
| 1983 | 32790 | 6445 | 39235 | 9662 | 9662 | 45310 | 4 | 21 | 569 | 45904 | 94801 |  |  |
| 1984 | 30592 | 26642 | 57234 | 24382 | 24382 | 43274 | 8 | 0 | 205 | 43487 | 125103 |  |  |
| 1985 | 19596 | 36742 | 56338 | 35634 | 35634 | 51425 | 50 | 0 | 0 | 51475 | 143447 |  |  |
| 1986 | 13292 | 26563 | 39855 | 57827 | 57827 | 37646 | 48 | 62 | 167 | 37923 | 135605 |  |  |
| 1987 | 7718 | 47028 | 54746 | 47722 | 47722 | 46039 | 1395 | 1 | 0 | 47435 | 149903 |  |  |
| 1988 | 0 | 0 | 0 | 106592 | 106592 | 93706 | 2474 | 299 | 0 | 96479 | 203071 |  |  |
| 1989 | 0 | 0 | 0 | 44612 | 44612 | 119631 | 13935 | 145 | 0 | 133711 | 178323 |  |  |
| 1990 | 0 | 0 | 0 | 8078 | 8078 | 115493 | 47114 | 1382 | 0 | 163989 | 172067 |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 129392 | 76734 | 3343 | 0 | 209469 | 209469 |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 77259 | 80174 | 7512 | 33 | 164978 | 164978 |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 81790 | 49295 | 2098 | 2 | 133185 | 133185 |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 84931 | 78566 | 8037 | 730 | 172264 | 172264 |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 110956 | 97665 | 19275 | 599 | 228496 | 228496 |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 91910 | 88882 | 28006 | 267 | 209064 | 209064 |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 93924 | 117008 | 21493 | 173 | 232598 | 232598 |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 60780 | 84323 | 13232 | 192 | 158526 | 158526 |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 51902 | 81463 | 12399 | 100 | 145865 | 145865 |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 53815 | 81640 | 15849 | 68 | 151372 | 151372 |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 35655 | 90360 | 16385 | 52 | 142452 | 142452 |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 51065 | 100269 | 15051 | 166 | 166552 | 166552 |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 47580 | 106967 | 21957 | 155 | 176659 | 176659 |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 57784 | 109692 | 17238 | 231 | 184945 | 184945 |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 52604 | 112994 | 17104 | 104 | 182807 | 182807 |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 54844 | 88254 | 17578 | 78 | 158753 | 158753 |  |  |

Table 2.2a-Summary of 1964-1980 catches (t) of Pacific cod in the Aleutian Islands region by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

Aleutian Islands region only:

| Year | Foreign | Joint Venture | Domestic | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1964 | 241 | 0 | 0 | 241 |
| 1965 | 451 | 0 | 0 | 451 |
| 1966 | 154 | 0 | 0 | 154 |
| 1967 | 293 | 0 | 0 | 293 |
| 1968 | 289 | 0 | 0 | 289 |
| 1969 | 220 | 0 | 0 | 220 |
| 1970 | 283 | 0 | 0 | 283 |
| 1971 | 2078 | 0 | 0 | 2078 |
| 1972 | 435 | 0 | 0 | 435 |
| 1973 | 977 | 0 | 0 | 977 |
| 1974 | 1379 | 0 | 0 | 1379 |
| 1975 | 2838 | 0 | 0 | 2838 |
| 1976 | 4190 | 0 | 0 | 4190 |
| 1977 | 3262 | 0 | 0 | 3262 |
| 1978 | 3295 | 0 | 0 | 3295 |
| 1979 | 5593 | 0 | 0 | 5593 |
| 1980 | 5788 | 0 | 0 | 5788 |

Table 2.2b-Summary of 1981-2006 catches (t) of Pacific cod in the Aleutian Islands region by fleet sector and gear type. All catches include discards. LLine $=$ longline, Subt. $=$ sector subtotal. Catches for 2006 are through early October.

Aleutian Islands region only:

|  | Foreign |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |  |
| 1981 | 2680 | 235 | 2915 | 1749 | 1749 | 2744 | 26 | 0 | 0 | 2770 | 7434 |  |  |  |
| 1982 | 1520 | 476 | 1996 | 4280 | 4280 | 2121 | 0 | 0 | 0 | 2121 | 8397 |  |  |  |
| 1983 | 1869 | 402 | 2271 | 4700 | 4700 | 1459 | 0 | 0 | 0 | 1459 | 8430 |  |  |  |
| 1984 | 473 | 804 | 1277 | 6390 | 6390 | 314 | 0 | 0 | 0 | 314 | 7981 |  |  |  |
| 1985 | 10 | 829 | 839 | 5638 | 5638 | 460 | 0 | 0 | 0 | 460 | 6937 |  |  |  |
| 1986 | 5 | 0 | 5 | 6115 | 6115 | 784 | 1 | 1 | 0 | 786 | 6906 |  |  |  |
| 1987 | 0 | 0 | 0 | 10435 | 10435 | 2662 | 22 | 88 | 0 | 2772 | 13207 |  |  |  |
| 1988 | 0 | 0 | 0 | 3300 | 3300 | 1698 | 137 | 30 | 0 | 1865 | 5165 |  |  |  |
| 1989 | 0 | 0 | 0 | 6 | 6 | 4233 | 284 | 19 | 0 | 4536 | 4542 |  |  |  |
| 1990 | 0 | 0 | 0 | 0 | 0 | 6932 | 602 | 7 | 0 | 7541 | 7541 |  |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 3414 | 3203 | 3180 | 0 | 9797 | 9797 |  |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 14558 | 22108 | 6317 | 84 | 43068 | 43068 |  |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 17312 | 16860 | 0 | 33 | 34204 | 34204 |  |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 14382 | 7009 | 147 | 0 | 21539 | 21539 |  |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 10574 | 4935 | 1024 | 0 | 16534 | 16534 |  |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 21179 | 5819 | 4611 | 0 | 31609 | 31609 |  |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 17349 | 7151 | 575 | 89 | 25164 | 25164 |  |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 20531 | 13771 | 424 | 0 | 34726 | 34726 |  |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 16437 | 7874 | 3750 | 69 | 28130 | 28130 |  |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 20362 | 16183 | 3107 | 33 | 39684 | 39684 |  |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 15826 | 17817 | 544 | 19 | 34207 | 34207 |  |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 27929 | 2865 | 7 | 0 | 30801 | 30801 |  |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 31478 | 974 | 2 | 0 | 32455 | 32455 |  |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 25766 | 3099 | 0 | 0 | 28865 | 28865 |  |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 19613 | 3001 | 0 | 13 | 22627 | 22627 |  |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 19843 | 3214 | 189 | 6 | 23252 | 23252 |  |  |  |

Table 2.3a-Summary of 1964-1980 catches ( t ) of Pacific cod in the combined Eastern Bering Sea and Aleutian Islands region by fleet sector. Catches by gear are not available for these years. Catches may not always include discards.

Eastern Bering Sea and Aleutian Islands region combined:

| Year | Foreign | Joint Venture | Domestic | Total |
| ---: | ---: | ---: | ---: | ---: |
| 1964 | 13649 | 0 | 0 | 13649 |
| 1965 | 15170 | 0 | 0 | 15170 |
| 1966 | 18354 | 0 | 0 | 18354 |
| 1967 | 32357 | 0 | 0 | 32357 |
| 1968 | 58191 | 0 | 0 | 58191 |
| 1969 | 50571 | 0 | 0 | 50571 |
| 1970 | 70377 | 0 | 0 | 70377 |
| 1971 | 45132 | 0 | 0 | 45132 |
| 1972 | 43340 | 0 | 0 | 43340 |
| 1973 | 54363 | 0 | 0 | 54363 |
| 1974 | 63841 | 0 | 0 | 63841 |
| 1975 | 54389 | 0 | 0 | 54389 |
| 1976 | 54671 | 0 | 0 | 54671 |
| 1977 | 36597 | 0 | 0 | 36597 |
| 1978 | 45807 | 0 | 31 | 45838 |
| 1979 | 38574 | 0 | 780 | 39354 |
| 1980 | 40846 | 830 | 2433 | 51649 |

Table 2.3b-Summary of 1981-2006 catches ( t ) of Pacific cod in the combined Eastern Bering Sea and Aleutian Islands region by fleet sector and gear type. All catches include discards. LLine = longline, Subt. $=$ sector subtotal. Catches for 2006 are through early October.

Eastern Bering Sea and Aleutian Islands region combined:

|  | Foreign |  |  | Joint Venture |  |  |  |  | Domestic Annual Processing |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: |
| Year | Trawl | LLine | Subt. | Trawl | Subt. | Trawl | LLine | Pot | Other | Subt. | Total |  |  |  |
| 1981 | 33027 | 6086 | 39113 | 9159 | 9159 | 15628 | 27 | 0 | 14 | 15669 | 63941 |  |  |  |
| 1982 | 24557 | 3618 | 28175 | 13592 | 13592 | 26014 | 5 | 0 | 1715 | 27734 | 69501 |  |  |  |
| 1983 | 34659 | 6847 | 41506 | 14362 | 14362 | 46769 | 4 | 21 | 569 | 47363 | 103231 |  |  |  |
| 1984 | 31065 | 27446 | 58511 | 30772 | 30772 | 43588 | 8 | 0 | 205 | 43801 | 133084 |  |  |  |
| 1985 | 19606 | 37571 | 57177 | 41272 | 41272 | 51885 | 50 | 0 | 0 | 51935 | 150384 |  |  |  |
| 1986 | 13297 | 26563 | 39860 | 63942 | 63942 | 38430 | 49 | 63 | 167 | 38709 | 142511 |  |  |  |
| 1987 | 7718 | 47028 | 54746 | 58157 | 58157 | 48701 | 1417 | 89 | 0 | 50207 | 163110 |  |  |  |
| 1988 | 0 | 0 | 0 | 109892 | 109892 | 95404 | 2611 | 329 | 0 | 98344 | 208236 |  |  |  |
| 1989 | 0 | 0 | 0 | 44618 | 44618 | 123864 | 14219 | 164 | 0 | 138247 | 182865 |  |  |  |
| 1990 | 0 | 0 | 0 | 8078 | 8078 | 122425 | 47716 | 1389 | 0 | 171530 | 179608 |  |  |  |
| 1991 | 0 | 0 | 0 | 0 | 0 | 132806 | 79937 | 6523 | 0 | 219266 | 219266 |  |  |  |
| 1992 | 0 | 0 | 0 | 0 | 0 | 91818 | 102282 | 13829 | 117 | 208046 | 208046 |  |  |  |
| 1993 | 0 | 0 | 0 | 0 | 0 | 99102 | 66155 | 2098 | 35 | 167389 | 167389 |  |  |  |
| 1994 | 0 | 0 | 0 | 0 | 0 | 99313 | 85575 | 8184 | 730 | 193802 | 193802 |  |  |  |
| 1995 | 0 | 0 | 0 | 0 | 0 | 121530 | 102600 | 20299 | 599 | 245029 | 245029 |  |  |  |
| 1996 | 0 | 0 | 0 | 0 | 0 | 113089 | 94701 | 32617 | 267 | 240673 | 240673 |  |  |  |
| 1997 | 0 | 0 | 0 | 0 | 0 | 111273 | 124159 | 22068 | 262 | 257762 | 257762 |  |  |  |
| 1998 | 0 | 0 | 0 | 0 | 0 | 81310 | 98094 | 13657 | 192 | 193253 | 193253 |  |  |  |
| 1999 | 0 | 0 | 0 | 0 | 0 | 68339 | 89337 | 16150 | 169 | 173995 | 173995 |  |  |  |
| 2000 | 0 | 0 | 0 | 0 | 0 | 74177 | 97823 | 18956 | 101 | 191056 | 191056 |  |  |  |
| 2001 | 0 | 0 | 0 | 0 | 0 | 51482 | 108177 | 16929 | 71 | 176659 | 176659 |  |  |  |
| 2002 | 0 | 0 | 0 | 0 | 0 | 78994 | 103134 | 15058 | 166 | 197352 | 197352 |  |  |  |
| 2003 | 0 | 0 | 0 | 0 | 0 | 79059 | 107941 | 21959 | 156 | 209114 | 209114 |  |  |  |
| 2004 | 0 | 0 | 0 | 0 | 0 | 83550 | 112790 | 17239 | 231 | 213810 | 213810 |  |  |  |
| 2005 | 0 | 0 | 0 | 0 | 0 | 72217 | 115995 | 17104 | 117 | 205434 | 205434 |  |  |  |
| 2006 | 0 | 0 | 0 | 0 | 0 | 74687 | 91468 | 17767 | 84 | 182005 | 182005 |  |  |  |

Table 2.4-History of Pacific cod ABC, TAC, total BSAI catch, and type of stock assessment model used to recommend ABC. Catch for 2006 is current through early October. "SS1" refers to Stock Synthesis 1 and "SS2" refers to Stock Synthesis 2. Each cell in the "Stock Assessment Model" column lists the type of model used to recommend the ABC in the corresponding row, meaning that the model was produced in the year previous to the one listed in the corresponding row.

| Year | ABC | TAC | Catch | Stock assessment model (from previous year) |
| ---: | ---: | ---: | ---: | ---: |
| 1980 | 148,000 | 70,700 | 45,947 | projection of 1979 survey numbers at age |
| 1981 | 160,000 | 78,700 | 63,941 | projection of 1979 survey numbers at age |
| 1982 | 168,000 | 78,700 | 69,501 | projection of 1979 survey numbers at age |
| 1983 | 298,200 | 120,000 | 103,231 | projection of 1979 survey numbers at age |
| 1984 | 291,300 | 210,000 | 133,084 | projection of 1979 survey numbers at age |
| 1985 | 347,400 | 220,000 | 150,384 | projection of 1979-1985 survey numbers at age |
| 1986 | 249,300 | 229,000 | 142,511 | separable age-structured model |
| 1987 | 400,000 | 280,000 | 163,110 | separable age-structured model |
| 1988 | 385,300 | 200,000 | 208,236 | separable age-structured model |
| 1989 | 370,600 | 230,681 | 182,865 | separable age-structured model |
| 1990 | 417,000 | 227,000 | 179,608 | separable age-structured model |
| 1991 | 229,000 | 229,000 | 219,266 | separable age-structured model |
| 1992 | 182,000 | 182,000 | 208,046 | SS1 model (age-based data) |
| 1993 | 164,500 | 164,500 | 167,389 | SS1 model (length-based data) |
| 1994 | 191,000 | 191,000 | 193,802 | SS1 model (length-based data) |
| 1995 | 328,000 | 250,000 | 245,029 | SS1 model (length-based data) |
| 1996 | 305,000 | 270,000 | 240,673 | SS1 model (length-based data) |
| 1997 | 306,000 | 270,000 | 257,762 | SS1 model (length-based data) |
| 1998 | 210,000 | 210,000 | 193,253 | SS1 model (length-based data) |
| 1999 | 177,000 | 177,000 | 173,995 | SS1 model (length-based data) |
| 2000 | 193,000 | 193,000 | 191,056 | SS1 model (length-based data) |
| 2001 | 188,000 | 188,000 | 176,659 | SS1 model (length-based data) |
| 2002 | 223,000 | 200,000 | 197,352 | SS1 model (length-based data) |
| 2003 | 223,000 | 207,500 | 209,114 | SS1 model (length-based data) |
| 2004 | 223,000 | 215,500 | 213,810 | SS1 model (length-based data) |
| 2005 | 206,000 | 206,000 | 164,404 | SS1 model (length- and age-based data) |
| 2006 | 194,000 | 194,000 | 182,005 | SS2 model (length- and age-based data) |

Table 2.5a-Pacific cod discard rates by area, target species/group, and year for the period 1991-2002 (see Table 2.5b for the period 2003-2004). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

| Eastern Bering Sea <br> Target species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Arrowtooth flounder | 0.61 | 0.00 | 0.94 |  | 0.66 | 0.08 | 0.07 | 1.00 | 1.00 | 0.99 | 1.00 | 0.22 |
| Atka mackerel | 1.00 |  | 0.70 | 1.00 |  | 0.23 |  | 0.51 | 0.00 | 0.00 | 1.00 |  |
| Flathead sole |  |  |  |  | 0.39 | 0.58 | 0.10 | 0.75 | 0.87 | 0.75 | 0.00 | 1.00 |
| Greenland turbot | 0.01 | 0.00 | 0.12 | 0.04 | 0.35 | 0.09 | 0.03 | 0.04 | 0.13 | 0.10 | 0.01 | 0.18 |
| Other flatfish | 0.63 | 0.31 | 0.47 | 0.88 | 0.22 | 0.28 | 0.91 | 0.28 | 0.33 | 0.32 | 0.00 | 0.00 |
| Other species | 0.04 | 0.99 | 0.38 |  | 1.00 | 1.00 | 0.01 | 0.95 | 0.07 | 0.92 | 0.08 | 0.00 |
| Pacific cod | 0.03 | 0.04 | 0.08 | 0.06 | 0.07 | 0.04 | 0.03 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 |
| Pollock | 0.70 | 0.85 | 0.73 | 0.68 | 0.21 | 0.41 | 0.24 | 0.42 | 0.49 | 0.68 | 0.84 | 0.52 |
| Rock sole | 1.00 | 0.00 | 0.08 | 0.87 | 0.25 | 0.90 |  | 1.00 | 0.02 | 0.16 | 1.00 | 1.00 |
| Rockfish | 1.00 | 0.00 | 0.89 | 0.01 | 0.84 | 0.69 | 0.16 |  | 0.00 | 0.03 | 0.00 | 0.00 |
| Sablefish | 0.00 | 0.12 | 0.42 | 0.40 | 0.96 | 0.94 | 0.78 | 0.93 | 0.61 | 0.98 | 0.12 | 0.48 |
| Unknown | 0.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.49 | 0.04 | 0.02 |  |  |
| Yellowfin sole |  | 0.74 | 0.72 | 0.50 | 0.08 | 1.00 | 0.24 | 0.77 | 0.50 | 0.60 | 0.39 | 0.77 |
| All targets | 0.03 | 0.04 | 0.08 | 0.06 | 0.07 | 0.04 | 0.03 | 0.02 | 0.01 | 0.02 | 0.01 | 0.02 |


| Aleutian Islands <br> Target species/group | 1991 | 1992 | 1993 | 1994 | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Arrowtooth flounder | 1.00 |  |  |  |  |  |  |  |  |  |  |  |

Table 2.5b-Pacific cod discard rates by area, target species/group, and year for the period 2003-2004 (see Table 2.5a for the period 1991-2002; note that the IFQ halibut target does not exist in Table 2.5a). The discard rate is the ratio of discarded Pacific cod catch to total Pacific cod catch for a given area/target/year combination. An empty cell indicates that no Pacific cod were caught in that area/target/year combination. Note that the absolute amount of discards may be small even if the discard rate is large.

|  | Eastern Bering Sea |  | Aleutian Islands |  |
| :--- | :---: | :---: | :---: | :---: |
| Target species/group | 2003 | 2004 | 2003 | 2004 |
| Arrowtooth flounder | 0.01 | 0.00 |  |  |
| Atka mackerel | 0.02 | 0.00 | 0.03 | 0.02 |
| Flathead sole | 0.00 | 0.02 |  |  |
| Greenland turbot | 0.07 | 0.05 | 0.00 | 0.38 |
| IFQ halibut | 0.28 | 0.28 | 0.58 |  |
| Other flatfish | 0.02 | 0.00 |  |  |
| Other species | 0.02 | 0.04 | 0.00 | 0.01 |
| Pacific cod | 0.01 | 0.01 | 0.01 |  |
| Pollock | 0.00 | 0.02 |  |  |
| Rock sole | 0.08 | 0.03 | 0.11 | 0.02 |
| Rockfish | 0.00 | 0.00 | 0.00 | 0.06 |
| Sablefish | 0.44 | 0.03 | 0.37 |  |
| Unknown | 0.06 | 0.02 |  | 0.01 |
| Yellowfin sole | 0.02 | 0.01 | 0.01 | 0.0 |

Table 2.6a—EBS catch ( t ) of Pacific cod by year, gear, and period for the years 1964-1980. Because direct estimates of gear- and period-specific catches are not available for these years, the figures shown here are estimates derived by distributing each year's total catch according to the average proportion observed for each gear/period combination during the years 1981-1988.

| Year | Trawl Fishery |  |  |  | Longline Fishery |  |  |  | Pot Fishery |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: |
|  | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 |  |  |
| 1964 | 6007 | 2469 | 2759 | 744 | 105 | 1324 | 0 | 0 | 0 |  |  |
| 1965 | 6595 | 2711 | 3028 | 817 | 115 | 1453 | 0 | 0 | 0 |  |  |
| 1966 | 8154 | 3352 | 3744 | 1011 | 142 | 1797 | 0 | 0 | 0 |  |  |
| 1967 | 14366 | 5905 | 6597 | 1780 | 250 | 3166 | 0 | 0 | 0 |  |  |
| 1968 | 25942 | 10663 | 11913 | 3215 | 452 | 5718 | 0 | 0 | 0 |  |  |
| 1969 | 22559 | 9272 | 10359 | 2796 | 393 | 4972 | 0 | 0 | 0 |  |  |
| 1970 | 31404 | 12908 | 14421 | 3892 | 547 | 6922 | 0 | 0 | 0 |  |  |
| 1971 | 19289 | 7929 | 8858 | 2391 | 336 | 4252 | 0 | 0 | 0 |  |  |
| 1972 | 19223 | 7901 | 8827 | 2382 | 335 | 4237 | 0 | 0 | 0 |  |  |
| 1973 | 23918 | 9831 | 10984 | 2964 | 417 | 5272 | 0 | 0 | 0 |  |  |
| 1974 | 27985 | 11503 | 12851 | 3468 | 487 | 6168 | 0 | 0 | 0 |  |  |
| 1975 | 23096 | 9493 | 10606 | 2862 | 402 | 5091 | 0 | 0 | 0 |  |  |
| 1976 | 22617 | 9296 | 10386 | 2803 | 394 | 4985 | 0 | 0 | 0 |  |  |
| 1977 | 14935 | 6139 | 6858 | 1851 | 260 | 3292 | 0 | 0 | 0 |  |  |
| 1978 | 19710 | 8101 | 9051 | 2443 | 343 | 4344 | 0 | 0 | 0 |  |  |
| 1979 | 16131 | 6630 | 7407 | 1999 | 281 | 3555 | 0 | 0 | 0 |  |  |
| 1980 | 18387 | 7558 | 8444 | 2279 | 320 | 4053 | 0 | 0 | 0 |  |  |

Table 2.6b—EBS catch ( t ) of Pacific cod by year, gear, and period for the years 1981-2006. Period 3 catch values for 2006 are extrapolations based on the average values from the previous three years.

| Year | Trawl Fishery |  |  |  | Longline Fishery |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 | Period 1 | Period 2 | Period 3 |
| 1981 | 15067 | 14087 | 21486 | 1286 | 624 | 3942 | 0 | 0 | 0 |
| 1982 | 21742 | 18151 | 16348 | 363 | 475 | 2308 | 0 | 0 | 0 |
| 1983 | 40757 | 24300 | 22705 | 2941 | 748 | 2756 | 0 | 0 | 0 |
| 1984 | 48237 | 24964 | 25045 | 5012 | 2128 | 19508 | 0 | 0 | 0 |
| 1985 | 55673 | 28673 | 22310 | 13703 | 1710 | 21379 | 0 | 0 | 0 |
| 1986 | 59786 | 26598 | 22382 | 8895 | 438 | 17278 | 0 | 0 | 0 |
| 1987 | 64413 | 15604 | 21462 | 20947 | 723 | 26752 | 0 | 0 | 0 |
| 1988 | 127470 | 25662 | 47166 | 444 | 646 | 1385 | 90 | 51 | 160 |
| 1989 | 127459 | 16986 | 19798 | 3810 | 4968 | 5157 | 33 | 63 | 49 |
| 1990 | 101645 | 11402 | 10524 | 13171 | 16643 | 17299 | 0 | 986 | 395 |
| 1991 | 107979 | 15549 | 5863 | 25470 | 21472 | 29792 | 12 | 1042 | 2288 |
| 1992 | 59460 | 11840 | 5959 | 49696 | 24201 | 6276 | 2622 | 4632 | 258 |
| 1993 | 67148 | 5362 | 9280 | 49244 | 27 | 23 | 2073 | 24 | 0 |
| 1994 | 61009 | 5806 | 18115 | 57968 | 13 | 20585 | 4923 | 0 | 3113 |
| 1995 | 90366 | 8543 | 12047 | 68458 | 26 | 29180 | 12484 | 3469 | 3322 |
| 1996 | 78194 | 3126 | 10590 | 62011 | 26 | 26845 | 18143 | 6401 | 3462 |
| 1997 | 81313 | 3927 | 8684 | 70676 | 433 | 46290 | 14584 | 3576 | 3333 |
| 1998 | 45008 | 5603 | 10169 | 54234 | 18 | 30071 | 9022 | 2779 | 1432 |
| 1999 | 44904 | 3312 | 3686 | 55180 | 1923 | 24360 | 9346 | 1001 | 2052 |
| 2000 | 44508 | 4578 | 4730 | 40180 | 1375 | 40086 | 15742 | 0 | 107 |
| 2001 | 22849 | 7025 | 5781 | 38368 | 6700 | 45291 | 11645 | 442 | 4298 |
| 2002 | 37008 | 9554 | 4503 | 50024 | 12132 | 38113 | 10852 | 401 | 3799 |
| 2003 | 34515 | 9986 | 3079 | 53156 | 11032 | 42773 | 15452 | 74 | 6586 |
| 2004 | 42181 | 12407 | 3197 | 56050 | 10459 | 43183 | 12560 | 521 | 4388 |
| 2005 | 45014 | 6664 | 926 | 53556 | 12773 | 46665 | 12147 | 0 | 4957 |
| 2006 | 46045 | 5966 | 926 | 51072 | 14564 | 46665 | 14333 | 0 | 4957 |

Table 2.7a--Length frequencies of Pacific cod in the pre-1989 trawl fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| r. S | N 12 | 34 | 4 | 5 | 6 | 78 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 2425 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19741 | 800 | 0 | 0 | 0 | 0 | 00 | 0 | 1 | 4 | 0 | 7 | 9 | 13 | 8 | 5 | 1 | 2 |  | 1 | 4 |  | 10 |
| 19743 | 1000 | 0 | 0 | 0 | 0 | 00 | 0 | 0 |  | 3 | 13 | 16 | 39 | 19 | 3 | 4 | 1 | 0 | 0 |  |  | 00 |
| 19751 | 1600 | 0 | 0 | 0 | 0 | 04 | 10 | 16 | 19 | 20 | 33 | 31 | 29 | 30 | 30 | 15 | 9 | 3 | 0 | 2 | 2 | 0 |
| 19752 | 800 | 0 | 0 | 0 | 0 | 01 | 2 | 3 | 1 | 1 | 1 | 0 | 2 | 2 | 3 | 16 | 13 | 1 | 0 |  |  |  |
| 19771 | 1900 | 00 | 0 | 0 | 0 | 39 | 36 | 79 | 35 | 9 | 15 | 26 | 25 | 53 | 32 | 17 | 11 | 4 | 1 | 3 | 0 | 0 |
| 19772 | 4200 | 0 | 0 | 0 | 0 | 00 | 6 | 12 | 22 | 39 | 40 | 273 | 331 | 367 | 355 | 188 | 104 | 38 | 12 |  |  | 00 |
| 19773 | 1200 | 0 | 0 | 0 | 0 | 00 | 3 | 7 | 22 | 33 | 13 | 10 | 7 | 10 | 15 | 12 | 6 | 1 | 1 | 3 |  | 0 |
| 19781 | 2300 | 01 | 1 | 1 | 0 | 0 | 3 | 16 | 19 | 73 | 220 | 103 | 29 | 19 | 13 | 4 | 5 | 4 | 0 |  | 2 | 00 |
| 19783 | 5600 | 0 | 0 | 0 | 6 | $35 \quad 79$ | 37 | 21 | 19 | 5 | 62 | 387 | 999 | 882 | 337 | 159 | 81 | 37 | 13 | 2 | 0 | 0 |
| 19791 | 3900 | $0 \quad 0$ | 0 | 0 | 0 | 21 | 45 | 94 | 204 | 315 | 29 | 77 | 122 | 147 | 144 | 37 | 5 | 4 | 3 |  |  | 00 |
| 19793 | 2700 | $0 \quad 0$ | 0 | 0 | 0 | 03 | 5 | 24 | 74 | 150 | 220 | 78 | 38 | 47 | 58 | 31 | 14 | 4 | 0 | 0 | 0 | 11 |
| 19801 | 9100 | 0 | 0 | 0 | 0 | 236 | 75 | 235 | 635 | 1014 | 1560 | 1038 | 971 | 714 | 497 | 632 | 485 | 197 | 86 | 9 |  | 52 |
| 19802 | 900 | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 | 0 | 1 | 16 | 45 | 8 | 3 | 0 | 0 | 0 | 0 |  |  | 0 |
| 19803 | 1800 | 0 | 0 | 0 | 0 | 01 | 0 | 0 | 9 | 16 | 33 | 78 | 69 | 53 | 29 | 6 | 8 | 6 | 2 |  |  | 0 |
| 19811 | 1200 | 0 | 0 | 0 | 0 | 00 | 0 | 0 | 0 | 8 | 33 | 44 | 40 | 22 | 5 | 0 | 0 | 0 | 0 |  |  | 00 |
| 19812 | 6900 | 01 | 1 | 2 | 3 | $10 \quad 71$ | 398 | 675 | 423 | 365 | 1109 | 1006 | 448 | 152 | 34 | 13 | 1 | 0 | 0 | 0 | 0 | 00 |
| 19813 | 3800 | 0 | 0 | 0 | 0 | 02 | 1 | 0 | 2 | 7 | 21 | 11 | 315 | 353 | 284 | 179 | 103 | 27 | 13 | 7 | 2 | 00 |
| 19821 | 6300 | 0 | 0 | 0 | 2 | 6 | 58 | 113 | 64 | 73 | 294 | 386 | 518 | 729 | 731 | 534 | 241 | 104 | 51 | 1 | 21 | 33 |
| 19822 | 1700 | 0 | 0 | 0 | 0 | 00 | 0 | 2 |  | 0 | 22 | 18 | 26 | 50 | 48 | 40 | 34 | 21 | 6 | 5 | 3 |  |
| 19823 | 4100 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 27 | 70 | 143 | 215 | 196 | 302 | 346 | 215 | 90 | 18 | 9 | 5 | 10 |
| 198 | 10100 | 0 | 0 | 0 | 1 | 50 | 154 | 93 | 95 | 176 | 492 | 758 | 1626 | 2344 | 2071 | 1307 | 644 | 211 | 77 | 36 | 21 | 12 |
| 19832 | 3700 | 0 | 0 | 0 | 1 | 01 | 4 | 15 | 42 | 71 | 77 | 81 | 200 | 284 | 248 | 186 | 83 | 28 | 6 | 3 | 4 | 0 |
| 19833 | 12000 | 0 | 0 | 1 | 15 | 2426 | 15 | 8 | 35 | 205 | 421 | 508 | 1450 | 1996 | 2482 | 2430 | 2220 | 1546 | 742 | 272 | 64 | 21 |
| 19841 | 12801 | 21 | 1 | 0 | 15 | 194401 | 367 | 220 | 105 | 223 | 709 | 779 | 1264 | 2262 | 3195 | 2930 | 2027 | 1039 | 434 | 144 | 24 | 13 |
| 19842 | 9701 | 451 | 51 | 201 | 206 | 313556 | 455 | 357 | 339 | 305 | 679 | 695 | 891 | 1109 | 959 | 817 | 597 | 453 | 312 | 120 | 41 | 8 |
| 19843 | 6500 | $0 \quad 0$ | 0 | 0 | 7 | $21 \quad 15$ | 114 | 434 | 372 | 190 | 140 | 126 | 235 | 375 | 502 | 506 | 437 | 363 | 210 | 92 |  | 1 |
| 19851 | 17300 | 20 | 0 | 4 | 0 | 239 | 116 | 257 | 720 | 1752 | 2234 | 1079 | 1388 | 2440 | 4999 | 5563 | 4288 | 2630 | 1385 | 594 |  | 6723 |
| 19852 | 3500 | 0 | 0 | 0 | 0 | 3 | 24 | 74 | 68 | 19 | 04 | 556 | 66 | 35 | 39 | 58 | 46 | 23 | 9 |  |  | 2 |
| 19853 | 4100 | 0 | 0 | 0 | 0 | 01 | 0 | 5 | 43 | 04 | 389 | 168 | 98 | 63 | 144 | 212 | 187 | 148 | 76 | 39 |  | 0 |
| 19861 | 169041 | 168 | 8 | 34 | 60 | 118249 | 635 | 761 | 683 | 783 | 2228 | 3560 | 3287 | 2095 | 2631 | 3469 | 3357 | 2442 | 1346 |  | 168 | 5817 |
| 19862 | 2600 | 0 | 0 | 0 | 0 | 72 | 2 | 3 | 5 | 7 | 15 | 62 | 92 | 72 | 67 | 95 | 98 | 84 | 46 | 30 | 8 | 4 |
| 19863 | 3800 | 0 | 0 | 0 | 0 | 21 | 13 | 15 | 25 | 24 | 69 | 111 | 153 | 184 | 209 | 156 | 179 | 133 | 92 |  | 22 | 4 |
| 19871 | 20200 | 313 |  | 15 | 58 | 192440 | 477 | 592 | 1161 | 2054 | 3898 | 2890 | 3326 | 5470 | 5461 | 4306 | 3650 | 3106 | 1953 | 1076 |  | 19863 |
| 19872 | 7400 | 0 | 0 | 0 | 1 | 25 | 9 | 4 | 8 | 22 | 16 | 04 | 333 | 592 | 974 | 1093 | 720 | 525 | 385 | 248 |  | 6825 |
| 19873 | 14100 | 0 | 0 | 0 | 0 | $0 \quad 0$ | 6 | 10 | 56 | 60 | 198 | 929 | 1639 | 1957 | 2591 | 3113 | 2678 | 2055 | 1930 | 1548 |  | 30653 |
| 1988 | 31010 | 11 | 1 | 6 | 29 | 92580 | 1448 | 1956 | 2185 | 4311 | 11135 | 10599 | 10194 | 9103 | 10096 | 12012 | 10395 | 5807 | 3010 | 1686 |  | 34692 |
| 19883 | 5400 | $0 \quad 0$ | 0 | 0 | 0 | $0 \quad 0$ | 5 | 0 | 13 | 52 | 257 | 326 | 284 | 348 | 348 | 373 | 332 | 305 | 166 | 56 | 20 | 6 |

Table 2.7b--Length frequencies of Pacific cod in the 1989-1999 trawl fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| Yr.S N12 | 3 | 45 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1989124100 | 3 | 31 | 0 | 28 | 217 | 494 | 795 | 720 | 954 | 3110 | 4341 | 4654 | 5664 | 7033 | 8561 | 8246 | 6265 | 3826 | 1867 | 91938 | 144 |
| 198931800 | 0 | 01 | 6 | 7 | 13 | 32 | 53 | 48 | 30 | 82 | 32 |  |  | 0 |  |  |  | 0 | 0 | 0 |  |
| 1990125300 | 3 | 414 | 8430 |  | 708 | 942 | 885 | 712 | 536 | 1141 | 2564 | 4397 | 7314 | 9868 | 10274 | 9356 |  | 4171 | 2251 | 266477 | 167 |
| 199028800 | 0 | 0 | 0 | 0 | 1 |  | 2 | 5 | 14 | 69 | 68 | 22 | 834 | 1200 | 1191 | 288 | 042 | 582 | 420 | 8477 |  |
| 199031000 | 0 | 00 | 0 | 0 | 0 |  |  |  |  |  |  |  |  | 39 | 10 | 13 | 7 | 6 |  | $0 \quad 0$ |  |
| 1991129601 | 5 | 615 | 70 | 457 | 24 | 25 | 224 | 283 | 704 | 5124 | 6055 | 459 | 9063 | 12143 | 12515 | 0775 | 7626 | 5003 | 2893 | 509759 |  |
| 199123310 | 1 | 12 | 2 | 5 | 7 | 11 | 20 | 16 | 16 | 0 | 94 | 166 | 142 | 135 | 146 | 92 | 87 | 77 | 30 | 31 |  |
| 2125803 |  | 1521 | 6720 |  | 625 | 278 | 577 | 2356 | 132 | 4 | 722 | 052 | 90 | 6847 | 6025 | 551 |  | 2723 | 624 | 910400 | 179 |
| 1993126400 | 5 | 823 | 5625 |  | 1142 | 1629 | 1723 | 421 | 7656 | 1477 | 881 | 9348 | 6655 | 4288 | 3408 | 2771 |  | 1334 | 840 | 479211 |  |
| 19932700 | 0 | 30 | 0 | 1 | 3 |  |  |  |  | 0 |  | 8 |  |  |  |  |  | 0 | 0 | $0 \quad 0$ |  |
| 199331000 | 0 | 00 | 1 | 0 | 0 | 0 | 2 | 3 | 2 | 21 | 17 | 15 | 12 | 10 |  | 2 |  | 1 | 1 | $0 \quad 2$ |  |
| 1994131401 | 2 | 42 | 1066 | 610 | 147 | 791 | 26 | 1929 | 63 | 9871 | 4218 | 14269 | 1410 | 1301 | 9353 | 5752 | 408 | 2074 | 165 | 677340 | 177 |
| 19 |  | 2838 | 9115 |  | 281 | 361 | 664 | 2571 | 781 | 9060 | 5920 | 982 | 042 | 8517 | 6661 | 382 |  | 1332 | 735 | 367191 |  |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 26 | 24 | 65 | 9 | 40 | 17 | 36 | 63 | 67 | 91 | 57 | 51 | 41 |  |
| 199613141 |  | 2526 | 49 |  | 48 | 295 | 080 | 083 | 2352 | 8627 | 14582 | 13606 | 9698 | 10723 | 11524 | 9049 |  | 3376 | 027 | 44 53 | 294 |
| 622200 | 0 | 00 | 0 | 2 |  |  | 10 |  | 35 | 110 | 149 | 94 | 28 | 12 |  |  | 3 | 2 | 0 | 0 0 |  |
| 53500 | 0 | 00 | 0 | 1 | 0 | 3 | 1 | 13 | 50 | 24 | 127 | 268 | 318 | 283 | 277 | 86 |  | 405 | 58 | 19992 |  |
| 19971306 |  | 8098 | 6930 |  | 05 | 2289 | 2311 | 1741 | 1976 | 7253 | 8302 | 11127 | 15435 | 15210 | 10904 | 6382 |  | 2558 | 1334 | 664289 | 125 |
| 199721701 | 0 | 45 | 1 | 4 | 4 | 8 | 8 | 12 | 13 | 31 | 42 | 38 | 34 | 20 | 24 | 15 | 8 | 2 | 1 | $0 \quad 0$ |  |
| 1998131201 | 7 | 4 | 1474 |  | 448 | 1360 | 1037 | 925 | 162 | 5621 | 6684 | 7605 | 11929 | 17524 | 18355 | 11124 | 5592 | 2697 | 1589 | 876390 | 170 |
| 199825400 | 0 | 0 | 0 | 1 | 1 | 37 | 66 | 81 | 63 | 95 | 425 | 588 | 526 | 382 | 237 | 95 | 78 | 20 | 20 | 412 |  |
| 199834600 | 0 | 00 | 0 | 1 | 2 | 3 | 14 | 38 | 67 | 169 | 253 | 355 | 285 | 246 | 178 | 144 | 115 | 69 | 47 | 4727 |  |
| 1999119030 | 1 | 65 | 810 |  | 409 | 383 | 338 |  | 2416 | 4722 | 2851 | 2939 | 3653 | 4745 | 4814 | 3466 |  | 1092 | 536 | 219103 |  |
| 199921200 | 0 | 0 | 0 | 0 | 0 |  |  |  | 5 | 29 | 23 | 27 | 25 | 14 | 7 | 8 | 2 | 2 | 1 | $0 \quad 0$ |  |
| 199932700 | 0 | 00 | 0 | 0 | 1 | 1 | 0 | 1 | 15 | 40 | 87 | 118 | 93 | 100 | 78 | 61 | 52 | 40 | 24 | 20 |  |

Table 2.7c--Length frequencies of Pacific cod in the post-1999 trawl fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| Yr. S | N | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2000 | 1 | 183 | 0 | 0 | 0 | 2 | 2 | 6 | 60 | 174 | 157 | 229 | 508 | 965 | 2756 | 3992 | 4293 | 3995 | 3965 | 4098 | 3229 | 2219 | 1353 | 750 | 383 | 182 | 85 |
| 2000 | 2 | 17 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 2 | 4 | 13 | 18 | 40 | 76 | 66 | 31 | 14 | 4 | 3 | 6 | 1 | 0 | 0 | 0 | 0 |
| 2000 | 3 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 13 | 20 | 12 | 8 | 6 | 3 | 0 | 1 | 1 | 0 | 0 | 0 |
| 2001 | 1 | 130 | 0 | 0 | 2 | 1 | 3 | 4 | 8 | 29 | 87 | 158 | 103 | 155 | 887 | 1372 | 1853 | 2785 | 2985 | 2416 | 1535 | 1115 | 679 | 424 | 186 | 93 | 39 |
| 2001 | 2 | 52 | 0 | 0 | 5 | 12 | 10 | 14 | 5 | 10 | 23 | 57 | 91 | 81 | 211 | 507 | 430 | 379 | 344 | 233 | 124 | 60 | 59 | 30 | 30 | 14 | 2 |
| 2001 | 3 | 36 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 8 | 12 | 8 | 21 | 33 | 80 | 108 | 202 | 206 | 202 | 169 | 101 | 59 | 26 | 13 | 6 | 5 | 1 |
| 2002 | 1 | 154 | 0 | 0 | 0 | 5 | 12 | 25 | 71 | 315 | 518 | 514 | 453 | 571 | 1671 | 1935 | 2020 | 3362 | 4287 | 3786 | 2177 | 1011 | 478 | 219 | 102 | 44 | 24 |
| 2002 | 2 | 67 | 0 | 0 | 0 | 6 | 8 | 3 | 12 | 68 | 201 | 263 | 305 | 288 | 415 | 593 | 740 | 524 | 387 | 229 | 175 | 124 | 57 | 35 | 18 | 8 | 3 |
| 2002 | 3 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 10 | 9 | 60 | 143 | 245 | 250 | 346 | 264 | 226 | 240 | 203 | 182 | 118 | 54 | 26 | 11 | 3 |
| 2003 | 1 | 157 | 0 | 0 | 2 | 4 | 1 | 2 | 5 | 82 | 266 | 333 | 355 | 647 | 1786 | 1867 | 2066 | 2749 | 3703 | 4240 | 3047 | 1799 | 895 | 379 | 170 | 78 | 30 |
| 2003 | 2 | 91 | 0 | 1 | 0 | 1 | 2 | 3 | 9 | 24 | 44 | 141 | 217 | 266 | 683 | 1106 | 1035 | 1046 | 1069 | 886 | 704 | 509 | 281 | 116 | 37 | 18 | 7 |
| 2003 | 3 | 44 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 7 | 31 | 94 | 190 | 200 | 221 | 225 | 246 | 237 | 232 | 156 | 92 | 35 | 7 | 1 |
| 2004 | 1 | 139 | 0 | 1 | 1 | 0 | 1 | 0 | 4 | 56 | 216 | 332 | 316 | 282 | 1295 | 2226 | 2490 | 2955 | 2947 | 2331 | 1575 | 998 | 535 | 339 | 197 | 113 | 48 |
| 2004 | 2 | 82 | 0 | 0 | 2 | 4 | 1 | 12 | 57 | 106 | 103 | 93 | 106 | 154 | 301 | 582 | 730 | 791 | 734 | 665 | 623 | 586 | 461 | 290 | 169 | 66 | 16 |
| 2004 | 3 | 39 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 14 | 55 | 103 | 198 | 169 | 134 | 133 | 164 | 219 | 161 | 94 | 61 | 11 | 1 |
| 2005 | 1 | 151 | 0 | 0 | 0 | 1 | 9 | 10 | 15 | 108 | 255 | 339 | 298 | 277 | 954 | 1579 | 2362 | 3351 | 4138 | 3778 | 2360 | 1496 | 809 | 466 | 220 | 90 | 34 |
| 2005 | 2 | 57 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 5 | 10 | 27 | 46 | 69 | 127 | 228 | 307 | 340 | 372 | 396 | 415 | 371 | 255 | 162 | 73 | 39 | 14 |
| 2005 | 3 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 8 | 16 | 20 | 17 | 22 | 28 | 21 | 16 | 14 | 13 | 2 | 5 | 1 | 0 |
| 2006 | 1 | 151 | 0 | 0 | 1 | 4 | 13 | 7 | 11 | 60 | 194 | 313 | 342 | 373 | 1219 | 1877 | 2123 | 2542 | 3157 | 3364 | 2784 | 2040 | 1238 | 621 | 285 | 83 | 55 |
| 2006 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 9 | 6 | 9 | 16 | 37 | 28 | 28 | 35 | 36 | 37 | 51 | 50 | 42 | 13 | 7 | 0 |  |

Table 2.8a--Length frequencies of Pacific cod in the pre-1989 longline fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| Yr. | S | N | 1 | 23 | 4 | 5 | 6 | 7 | $8 \quad 9$ | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1978 | 1 | 54 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 23 | 124 | 623 | 812 | 435 | 269 | 216 | 160 | 110 | 58 | 36 | 7 | 7 | 0 | 0 |
| 1978 | 2 | 59 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 2 | 78 | 444 | 1093 | 783 | 436 | 328 | 170 | 64 | 30 | 6 | 1 | 1 | 0 |
| 1978 | 3 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 54 | 344 | 719 | 770 | 275 | 94 | 49 | 32 | 16 | 7 | 2 | 0 | 0 |
| 1979 | 1 | 99 | 0 | 0 | 0 | 0 | 0 | 0 | 883 | 377 | 683 | 434 | 337 | 1135 | 2126 | 2432 | 1356 | 465 | 233 | 128 | 56 | 27 | 3 | 6 | 0 |
| 1979 | 2 | 48 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 14 | 49 | 90 | 155 | 93 | 302 | 604 | 628 | 274 | 74 | 33 | 4 | 3 | 3 | 0 | 0 |
| 1979 | 3 | 53 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 11 | 51 | 252 | 263 | 195 | 401 | 705 | 605 | 220 | 44 | 11 | 9 | 2 | 0 | 0 |
| 1980 | 1 | 51 | 0 | 0 | 0 | 0 | 0 | 0 | 05 | 15 | 66 | 212 | 591 | 604 | 320 | 182 | 199 | 244 | 111 | 36 | 11 | 4 | 0 | 0 | 0 |
| 1980 | 2 | 37 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 29 | 169 | 334 | 293 | 185 | 148 | 140 | 67 | 17 | 4 | 2 | 0 | 0 | 0 |
| 1980 | 3 | 54 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 18 | 235 | 558 | 679 | 652 | 350 | 194 | 138 | 76 | 25 | 5 | 0 | 1 | 0 |
| 1981 | 1 | 47 | 0 | 00 | 0 | 5 | 18 | 7 | $7 \quad 10$ | 0 | 18 | 48 | 285 | 496 | 448 | 335 | 197 | 153 | 89 | 70 | 36 | 9 | 4 | 0 | 0 |
| 1981 | 2 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 02 | 1 | 8 | 29 | 88 | 160 | 265 | 292 | 228 | 108 | 35 | 32 | 24 | 3 | 1 | 0 | 0 |
| 1981 | 3 | 36 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 8 | 86 | 230 | 318 | 300 | 220 | 89 | 29 | 15 | 2 | 0 | 1 | 0 |
| 1982 | 1 | 42 | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 9 | 13 | 18 | 131 | 184 | 266 | 334 | 314 | 211 | 101 | 61 | 44 | 31 | 0 | 1 | 1 |
| 1982 | 2 | 32 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 42 | 17 | 98 | 190 | 128 | 161 | 130 | 117 | 74 | 38 | 11 | 5 | 3 | 2 | 0 |
| 1982 | 3 | 69 | 0 | 00 | 0 | 0 | 0 | 0 | 10 | 1 | 14 | 33 | 92 | 235 | 460 | 773 | 1149 | 1066 | 614 | 235 | 77 | 27 | 6 | 2 | 2 |
| 1983 | 1 | 132 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 16 | 48 | 170 | 1116 | 1525 | 2035 | 2732 | 3421 | 3065 | 1838 | 792 | 334 | 163 | 88 | 36 | 7 |
| 1983 | 2 | 61 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 14 | 13 | 91 | 319 | 383 | 504 | 623 | 675 | 505 | 355 | 150 | 50 | 18 | 10 | 0 |
| 1983 | 3 | 98 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 4 | 28 | 129 | 459 | 1162 | 1260 | 1544 | 1776 | 1561 | 991 | 476 | 148 | 37 | 9 | 6 |
| 1984 | 1 | 122 | 0 | 0 | 0 | 0 | 1 | 0 | 619 | 40 | 41 | 46 | 416 | 800 | 1323 | 2414 | 3163 | 3015 | 2012 | 1015 | 437 | 155 | 70 | 24 | 6 |
| 1984 | 2 | 100 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 7 | 14 | 17 | 102 | 376 | 750 | 1602 | 2167 | 1873 | 1405 | 891 | 567 | 203 | 59 | 16 | 3 |
| 1984 | 3 | 289 | 0 | 00 | 0 | 0 | 0 | 1 | 314 | 55 | 293 | 764 | 1721 | 2467 | 6595 | 12255 | 15779 | 15982 | 12816 | 8397 | 4192 | 1528 | 407 | 91 | 24 |
| 1985 | 1 | 189 | 0 | 0 | 0 | 0 | 0 | 0 | 112 | 34 | 186 | 550 | 1367 | 958 | 1828 | 3877 | 7018 | 8009 | 5977 | 3362 | 1591 | 537 | 175 | 44 | 7 |
| 1985 | 2 | 73 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 28 | 246 | 368 | 206 | 418 | 775 | 1000 | 823 | 590 | 429 | 245 | 105 | 23 | 2 |
| 1985 | 3 | 362 | 0 | 00 | 0 | 0 | 1 | 0 | 0 | 23 | 116 | 605 | 5449 | 16095 | 14240 | 10594 | 17780 | 24998 | 19637 | 11586 | 6071 | 2786 | 920 | 215 | 51 |
| 1986 | 1 | 182 | 0 | 00 | 0 | 0 | 0 | 0 | 830 | 81 | 121 | 385 | 1765 | 3055 | 3578 | 3014 | 3739 | 5900 | 5622 | 3348 | 1554 | 654 | 237 | 63 | 13 |
| 1986 | 2 | 37 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 15 | 94 | 247 | 306 | 175 | 162 | 205 | 104 | 60 | 24 | 13 | 0 | 0 |
| 1986 | 3 | 325 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 18 | 158 | 616 | 2233 | 5154 | 14368 | 23612 | 20725 | 10897 | 10483 | 9006 | 4991 | 2308 | 881 | 326 | 85 |
| 1987 | 1 | 304 | 0 | 00 | 0 | 0 | 2 | 0 | 518 | 88 | 425 | 1362 | 4950 | 5219 | 8337 | 14661 | 16709 | 12862 | 11421 | 9132 | 4689 | 1828 | 519 | 180 | 31 |
| 1987 | 2 | 17 | 0 | 00 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 24 | 25 | 55 | 79 | 51 | 28 | 11 | 13 | 3 |  | 0 | 0 |
| 1987 | 3 | 420 | $0 \quad 0$ | $0 \quad 0$ | 3 | 0 | 0 | 03 | 39 | 30 | 147 | 593 | 4503 | 18418 | 29582 | 24338 | 25914 | 28336 | 20972 | 10694 | 6630 | 3800 | 1532 | 414 | 134 |

Table 2.8b--Length frequencies of Pacific cod in the 1990-1999 longline fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| S | 5678 | $9 \quad 10$ | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 199011370000 | $\begin{array}{llll}0 & 0 & 0\end{array}$ | 0 | 4 | 12 | 162 | 780 | 1688 | 2789 | 3515 | 3071 | 2482 | 1587 | 1215 | 721 | 480 | 217 | 92 |
| 9022730000 | $\begin{array}{llll}0 & 0 & 6\end{array}$ | 2355 | 131 | 225 | 69 | 2356 | 5901 | 10124 | 12987 | 12636 | 1020 | 7291 | 5046 | 323 | 168 |  |  |
| 199032460000 | $\begin{array}{lllll}0 & 0 & 1 & 3\end{array}$ | 12 | 18 | 56 | 47 | 1624 | 5113 | 278 | 11494 | 815 | 8609 | 5752 | 3410 | 028 | 1269 | 58 | 277 |
| 199112260000 | $\begin{array}{llll}0 & 0 & 0 & 4\end{array}$ | 1429 | 113 | 299 | 1018 | 2340 | 4652 | 31 | 647 | 8943 | 6336 | 565 | 2133 | 03 | 739 | 298 | 30 |
| 199122600000 | $\begin{array}{llll}0 & 0 & 0 & 2\end{array}$ | 10 | 32 | 137 | 742 | 079 | 500 | 7909 | 10294 | 11451 | 10371 | 8410 | 5876 | 3153 | 1759 | 787 |  |
| 199132910001 | 183338 | 58107 | 185 | 396 | 533 | 3750 | 6541 | 10028 | 12271 | 086 | 1243 | 9961 | 6816 | 3885 | 24 | 1012 | 18 |
| 199213660002 | 540 | 85291 | 1131 | 3030 | 9842 | 13958 | 1456 | 640 | 96 | 195 | 1312 | 37 | 592 | 3322 | 814 | 790 |  |
| 199223130000 | 21 | 64161 | 492 | 1076 | 5712 | 11463 | 11718 | 67 | 13825 | 11925 | 8736 | 695 | 522 | 359 | 2221 | 121 |  |
| 199231420000 | $\begin{array}{llll}0 & 0 & 0 & 1\end{array}$ | 19 | 52 | 154 | 65 | 375 | 564 | 390 | 741 | 2412 | 1943 | 1595 | 126 | 897 | 565 | 298 | 06 |
| 199313480010 | 1470 | 14 | 409 | 3110 | 9108 | 17361 | 321 | 20379 | 14246 | 10806 | 984 | 5839 | 358 | 1934 | 827 | 412 | 10 |
| 1993260000 | 00 | 00 | 0 | 0 | 1 |  |  | 3 | 7 | 3 | 5 | 8 |  | 2 | ) |  |  |
| 199413910003 | 122028 | 78183 | 490 | 00 | 6598 | 16608 | 23 | 676 | 84 | 22 | 7682 | 388 | 2661 | 1679 | 824 | 412 |  |
| 199432030000 | $\begin{array}{llll}0 & 0 & 2\end{array}$ | 1157 | 140 | 236 | 745 | 2605 | 5322 | 663 | 939 | 760 | 3892 | 08 | 1192 | 745 | 449 | 260 |  |
| 19951372 | 13 | 55 | 53 |  | 57 |  | 23425 | 86 | 42 |  | 99 | 940 | 6 | 712 | 363 | 93 |  |
| 19952 | $0 \begin{array}{llll}0 & 0 & 0\end{array}$ | $0 \quad 0$ | 0 | 0 | 0 | 4 | 10 | 19 | 15 | 8 | 1 | 12 | 9 | 5 | 4 | 0 |  |
| 199532580000 | $\begin{array}{llll}0 & 0 & 221\end{array}$ | $24 \quad 47$ | 210 | 515 | 2857 | 11 | 799 | 242 | 22 | 10274 | 7535 | 520 | 8 | 39 | 804 | 387 |  |
| 199613850000 | 14 | 49 | 521 | 1512 | 8322 | 18562 | 24 | 27 | 26151 | 48 | 11119 | 5431 | 2445 | 1204 | 669 | 296 |  |
| 199632710000 | 0 1 0 | 1483 | 180 | 390 | 1523 | 326 | 11475 | 14333 | 11034 | 119 | 7117 | 75 | 4074 | 225 | 1041 | 424 | 2 |
| 199714130000 | 417 | 7920 | 521 | 1612 | 8223 | 16931 | 29230 | 38525 | 33998 | 19962 | 0360 | 475 | 7 | 38 | 557 | 194 | 67 |
| 19972100000 | $0 \begin{array}{llll}0 & 0 & 0\end{array}$ | 00 |  | 1 | 4 | 6 | 18 | 18 | 15 | 16 | 5 | 8 | 6 | 4 | 4 | 3 |  |
| 199733800012 | 132089 | 160288 | 621 | 73 | 4814 | 358 | 15198 | 20854 | 26965 | 25031 | 17322 | 992 | 6073 | 3677 | 977 | 55 | 361 |
| 199813540000 | 433 | 34287 | 731 | 2105 | 6959 | 11401 | 16819 | 24275 | 25752 | 18 | 0 | 49 | 2 | 021 | 418 | 149 | 72 |
| 1998280000 | $0 \quad 0$ | 11 |  | 6 | 22 |  |  | 7 | 4 |  | 2 |  | 0 | 2 |  | 0 |  |
| 19983416 | 32224645 | 88333 | 1755 | 3717 | 8601 | 13692 | 20625 | 25081 | 28930 | 26157 | 19469 | 12038 | 6147 | 323 | 1938 | 810 | 349 |
| 199912760002 | 0422 | 54208 | 1211 | 3601 | 8274 | 26 | 798 | 11745 | 11498 | 679 | 6032 | 3123 | 1326 | 30 | 268 | 133 | 85 |
| 19992960000 | $\begin{array}{lll}0 & 0\end{array}$ | 613 | 44 | 130 | 943 | 1449 | 16 | 218 | 1260 | 111 | 757 | 531 | 305 | 62 | 69 | 34 | 12 |
| 199932171000 | $0 \quad 11027$ | 61112 | 361 | 666 | 3565 | 7611 | 6607 | 6008 | 6435 | 5846 | 4012 | 2777 | 1598 | 836 | 371 | 187 | 93 |

Table 2.8c--Length frequencies of Pacific cod in the post-1999 longline fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

|  |  |  |  |  |  | 7 |  |  | 10 |  | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2000 |  | 230 | 2 |  |  | 5 | 15 | 47 | 175 | 600 | 1419 | 88 | 9645 | 9944 | 82 | 55 | 4738 | 2758 | 1403 | 687 | 275 | 0 | 5116 |
| 2000 |  |  |  |  |  | 0 |  |  | 3 | 6 | 19 |  | 29 | 1058 | 1597 | 1244 | 69 | 81 | 408 | 13 | 135 | 52 | 26 |
|  |  | 294 |  |  |  | 2 |  |  | 39 |  | 1166 |  | 9291 | 15 | 19 | 15 | 8759 | 5366 | 8 | 1934 | 1055 | 497 | 18783 |
| 2001 |  |  |  |  |  | 2 | 22 | 108 | 320 | 469 | 1018 | 4563 | 8794 | 1211 | 1358 | 10723 | 66 | 2800 | 1399 | 97 | 376 | 164 | 84 |
|  |  |  | 00 |  |  | 2 |  |  |  |  |  |  |  |  |  | 3560 | 2460 | 1007 | 455 | 207 |  | 64 | 25 |
|  |  |  |  |  |  | 5 | 17 | 43 | 176 | 873 | 18 |  |  | 1466 | 168 | 16030 | 1200 | 5533 | 2478 | 1156 |  | 373 | 150 |
| 2002 |  | 261 | 125 | 557 | 14 | 20 | 85 | 183 |  |  |  |  |  | 10 | 14 |  |  | 3051 | 6 | 51 |  | 230 | 49 |
| 2002 |  |  |  |  | 2 | 3 | 27 | 74 |  | 327 |  |  |  | 5121 | 5225 | 51 | 3549 | 70 | 8 | 40 | 21 | 04 | 57 |
| 2002 |  |  | 2 | 000 | 2 | 17 | 49 | 159 |  | 1 | 2 | 7432 | 12 | 13 | 1358 | 12757 |  |  | 2857 | 1239 |  | 220 | 97 |
| 2003 |  | 3 | 002 |  |  | 9 | 40 | 233 | 28 | 146 | 362 | 10 |  | 16 | 15 | 13262 | 952 |  | 2150 | 75 | 312 | 23 | 53 |
| 2003 |  | 192 | 000 | 000 |  | 1 |  | 10 |  |  | 571 |  | 5601 | 7083 | 668 | 5653 | 4235 | 2311 | 1204 | 532 |  | 02 | 31 |
| 2003 |  | 3 |  |  |  | 0 |  | 23 | 91 | 269 | 118 |  | 13 | 18 | 18 | 16 | 12 |  | 4265 | 189 | 740 | 87 | 16 |
| 2004 |  | 27 | 0 |  | 0 | 2 | 8 | 55 | 15 |  | 954 |  | 109 | 16 | 18 | 12 |  | 33 | 1713 | 03 | 268 | 12 | 43 |
|  |  | 180 | 000 |  |  |  |  | 10 | 2 | 46 |  |  |  | 52 |  |  |  |  |  | 873 |  | 22 | 48 |
|  |  | 29 | 0 | 011 | 3 | 2 | 2 | 49 | 162 |  | 92 |  |  | 129 | 17 | 16 | 12 | 7120 | 40 | 214 | 938 | 356 | 111 |
|  |  | 2 | 000 |  | 2 | 5 | 27 | 84 | 199 |  | 11 |  |  | 9408 |  |  |  |  | 1517 | 699 |  | 00 | 24 |
| 20 |  |  | 000 | 0 |  | 0 | 5 |  | 22 |  | 25 |  |  | 4075 | 52 | 6464 | 634 | 49 | 3257 | 178 | 96 | 364 | 89 |
|  |  | 29 | 0 | 214 |  | 2 |  | 60 | 149 | 395 | 101 |  | 79 | 1152 | 1232 | 1226 | 11 | 1001 | 6756 | 3633 | 198 |  | 249 |
| 200 |  | 23 | 00 | 0 |  | 0 | 10 | 37 | 116 | 270 | 57 | 28 | 722 | 91 | 94 | 8979 |  | 516 | 2524 | 98 | 409 | 164 | 55 |
| 20 |  | 13 | 000 | 000 | 0 | 0 |  |  | 7 | 53 | 15 | 615 | 148 | 21 | 22 | 226 | 199 | 1820 | 1838 | 147 | 95 | 70 | 175 |
| 2006 | 3 |  | 000 | 000 | 0 | 0 | 0 | 0 | 3 | 28 | 48 | 137 | 305 | 510 | 914 | 916 | 1027 | 633 | 490 | 376 | 268 | 154 | 67 |

Table 2.9a--Length frequencies of Pacific cod in the 1989-1999 pot fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| Yr. S | N1234567 | 8 | 910 | 1011 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19902 | 390000000 | 0 | 0 | 00 | 1 | 7 | 43 | 75 | 141 | 234 | 296 | 227 | 230 | 139 | 82 | 45 | 3 | 3 |
| 19903 | 760000000 | 0 | 0 | 0 | 1 | 3 | 10 | 116 | 512 | 1149 | 1146 | 1360 | 701 | 391 | 260 | 109 | 12 |  |
| 19912 | 820000000 | 0 | 0 | $0 \quad 0$ | 0 | 39 | 237 | 545 | 975 | 1298 | 1315 | 991 | 681 | 329 | 189 | 59 | 16 |  |
| 19913 | 850000000 | 0 | 0 | $0 \quad 0$ | 6 | 26 | 149 | 382 | 712 | 1193 | 1508 | 1424 | 911 | 491 | 262 | 124 | 45 | 21 |
| 1992 | 1150000010 | 1 | 0 | 86 | 2 | 158 | 365 | 685 | 1747 | 3468 | 2854 | 1768 | 966 | 608 | 321 | 203 | 81 | 9 |
| 19922 | 1370000001 | 2 | 53 | 3482 | 275 | 1412 | 2419 | 2362 | 2726 | 2928 | 2376 | 1560 | 1066 | 685 | 425 | 196 | 96 | 37 |
| 19923 | 570000000 | 0 | 22 | 2168 | 117 | 472 | 751 | 618 | 458 | 364 | 192 | 66 | 71 | 40 | 29 | 11 | 2 | 0 |
| 19931 | 1120000000 | 0 | 0 | $0 \quad 9$ | 29 | 350 | 923 | 1763 | 2384 | 2259 | 1983 | 1278 | 757 | 441 | 265 | 111 | 50 | 18 |
| 1994 | 1700000000 | 0 | 0 | 121 | 126 | 758 | 3052 | 4939 | 5352 | 5172 | 3937 | 2601 | 1415 | 801 | 457 | 268 | 122 | 31 |
| 19943 | 800000000 | 0 | 5 | 310 | 25 | 152 | 576 | 1095 | 1255 | 1050 | 808 | 601 | 364 | 229 | 136 | 71 | 39 | 16 |
| 1995 | 2200000000 | 0 | 1 | 448 | 251 | 1255 | 3298 | 7553 | 10763 | 9549 | 6607 | 4013 | 2228 | 1338 | 782 | 377 | 163 | 76 |
| 19952 | 1170000000 | 0 | 0 | 04 | 33 | 399 | 793 | 1579 | 2527 | 2468 | 1815 | 1343 | 982 | 672 | 479 | 295 | 152 | 67 |
| 19953 | 1020000000 | 0 | 0 | 10 | 10 | 225 | 676 | 1158 | 1822 | 2056 | 1587 | 1041 | 749 | 504 | 296 | 156 | 80 | 11 |
| 19961 | 2710000035 | 111 | 1439 | 3984 | 251 | 2216 | 6984 | 11621 | 13669 | 13851 | 10926 | 6386 | 3538 | 1904 | 1173 | 638 | 285 | 103 |
| 19962 | 1310000000 | 0 | 1 | 14 | 23 | 243 | 1085 | 2558 | 3219 | 2777 | 2079 | 1602 | 1361 | 928 | 642 | 362 | 193 | 103 |
| 19963 | 1060000000 | 0 | 0 | 25 | 25 | 176 | 463 | 982 | 1875 | 1950 | 1478 | 1042 | 979 | 921 | 686 | 408 | 212 | 79 |
| 19971 | 2120000100 | 1 | 315 | 1539 | 83 | 667 | 2211 | 5414 | 10164 | 11131 | 7151 | 3776 | 1877 | 1043 | 633 | 398 | 187 | 84 |
| 19972 | 1230000000 | 1 | 5 | 27 | 22 | 219 | 858 | 1793 | 3043 | 3852 | 2334 | 1129 | 683 | 423 | 312 | 180 | 105 | 56 |
| 19973 | 1080000000 | 1 | 0 | 17 | 22 | 163 | 453 | 972 | 1661 | 2418 | 2501 | 1431 | 702 | 473 | 390 | 266 | 140 | 69 |
| 19981 | 1850000100 | 0 | 2 | 419 | 105 | 855 | 1837 | 3037 | 6318 | 8140 | 6974 | 3987 | 1684 | 650 | 424 | 183 | 90 | 44 |
| 19982 | 980000000 | 0 | 0 | $0 \quad 5$ | 12 | 175 | 615 | 1121 | 1527 | 1977 | 1864 | 1172 | 578 | 290 | 198 | 82 | 39 | 19 |
| 19983 | 560000000 | 1 | 0 | 38 | 10 | 67 | 235 | 356 | 534 | 603 | 615 | 380 | 167 | 91 | 39 | 40 | 23 | 16 |
| 19991 | 1310000001 | 2 | 2 | 615 | 100 | 855 | 1329 | 2031 | 3154 | 3317 | 2836 | 1784 | 971 | 458 | 255 | 96 | 46 | 20 |
| 19992 | 300000000 | 0 | 0 | $0 \quad 2$ | 7 | 52 | 105 | 117 | 174 | 152 | 121 | 69 | 48 | 28 | 9 | 4 | 3 | 2 |
| 19993 | 610000000 | 0 | 0 | 13 | 16 | 93 | 456 | 637 | 500 | 560 | 543 | 295 | 235 | 141 | 90 | 49 | 36 | 29 |

Table 2.9b--Length frequencies of Pacific cod in the post-1999 pot fisheries, by year, season (S), and length bin. $\mathrm{N}=$ input sample size.

| Yr. | N | N | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2000 | 1 | 144 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 1 | 12 | 97 | 867 | 2375 | 3670 | 3687 | 3366 | 3027 | 1827 | 1028 | 482 | 215 | 95 | 34 | 17 |
| 2000 | 3 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 26 | 62 | 67 | 27 | 13 | 3 | 0 | 1 | 0 | 0 | 0 | 0 |
| 2001 | 1 | 121 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 13 | 230 | 791 | 2183 | 3947 | 3814 | 2027 | 848 | 341 | 190 | 106 | 48 | 17 | 7 |
| 2001 | 2 | 21 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 45 | 77 | 103 | 93 | 61 | 33 | 5 | 7 | 4 | 6 | 1 |
| 2001 | 3 | 80 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 2 | 1 | 4 | 3 | 11 | 130 | 528 | 1140 | 1279 | 1374 | 977 | 435 | 263 | 160 | 98 | 40 | 19 | 6 |
| 2002 | 1 | 105 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 15 | 131 | 600 | 1554 | 2841 | 2843 | 1781 | 765 | 297 | 134 | 78 | 39 | 20 | 14 |
| 2002 | 2 | 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 3 | 20 | 45 | 72 | 65 | 62 | 41 | 32 | 9 | 12 | 2 | 0 | 1 |
| 2002 | 3 | 79 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 5 | 26 | 165 | 578 | 1237 | 1348 | 1021 | 726 | 482 | 311 | 144 | 107 | 43 | 19 | 11 |
| 2003 | 1 | 111 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 3 | 6 | 54 | 267 | 766 | 1391 | 2203 | 2788 | 2360 | 1328 | 655 | 259 | 97 | 41 | 19 | 9 |
| 2003 | 3 | 88 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 2 | 19 | 236 | 920 | 1472 | 1403 | 1227 | 951 | 607 | 449 | 291 | 153 | 60 | 26 | 4 |
| 2004 | 1 | 94 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 15 | 253 | 895 | 1493 | 1870 | 1709 | 1185 | 668 | 375 | 188 | 105 | 43 | 17 | 5 |
| 2004 | 2 | 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 22 | 53 | 56 | 60 | 46 | 32 | 15 | 16 | 18 | 3 | 0 | 0 |
| 2004 | 3 | 77 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 | 2 | 2 | 4 | 18 | 105 | 510 | 1039 | 1101 | 904 | 660 | 420 | 389 | 330 | 188 | 127 | 47 | 14 |
| 2005 | 1 | 86 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 6 | 13 | 119 | 426 | 1088 | 1709 | 1600 | 1117 | 554 | 309 | 180 | 118 | 62 | 26 | 4 |
| 2005 | 3 | 80 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 2 | 7 | 14 | 114 | 434 | 1006 | 1308 | 1117 | 719 | 459 | 374 | 294 | 207 | 155 | 92 | 41 |
| 2006 | 1 | 107 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 8 | 218 | 767 | 1610 | 2346 | 2473 | 1783 | 1022 | 610 | 324 | 175 | 104 | 44 | 19 |
| 2006 | 3 | 56 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 68 | 252 | 557 | 557 | 546 | 357 | 248 | 189 | 131 | 96 | 84 | 47 | 27 |

Table 2.10a--Length frequencies of Pacific cod in the 1979-1981 EBS shelf trawl survey, by year and length bin. $\mathrm{N}=$ input sample size.

| S | N 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19792 | 100 | 5 | 44 | 186 | 374 | 457 | 694 | 1764 | 2393 | 1884 | 1171 | 61 | 202 | 0 | 44 | 51 | 29 |  | 0 | 3 |  |  | 0 | 0 |  |
| 882 | 1000 | 6 | 85 | 241 | 82 | 42 | 224 | 87 | 29 | 1320 | 1542 | 2062 | 1364 | 893 | 333 | 100 | 33 | 31 | 19 | 6 | 2 |  |  |  |  |
| 19812 | 100 | 20 | 156 | 330 | 278 | 32 | 100 | 330 | 653 | 724 | 511 | 1063 | 1396 | 1746 | 1215 | 812 | 398 | 156 | 39 | 27 | 13 | 1 | 0 | 0 |  |

Table 2.10b--Length frequencies of Pacific cod in the post-1981 EBS shelf trawl survey, by year
and length bin. $\mathrm{N}=$ input sample size.

| Yr. | S N | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 617 | 718 | 19 | 20 | 21 | 2 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 19822 | 2103 | 17 | 97 | 234 | 148 | 37 | 28 | 132 | 403 | 766 | 750 | 416 | 520 | 1512 | 1326 | 1288 | 1178 | 8 | 4774 | 210 | 90 | 29 | 9 | 0 |  |
| 1983 | 2115 | 393 | 1396 | 1289 | 622 | 147 | 32 | 135 | 370 | 551 | 380 | 209 | 393 | 1367 | 1289 | 1 | 12 | 921 | 65 | 25 | 1 | 31 | 19 | 1 |  |
| 19842 | 2 | 70 | 129 | 82 | 142 | 282 | 920 | 1653 | 1712 | 1041 | 485 | 249 | 261 | 536 | 579 | 864 | 61 | 0 | 590 |  | 173 | 94 | 38 |  |  |
| 1985 | 130 | 162 | 540 | 964 | 1537 | 1761 | 664 | 298 | 595 | 880 | 942 | 154 | 1528 | 1879 | 678 | 480 | 543 | 687 | 67 |  | 253 | 111 | 3817 | 5 |  |
| 1986 | 2124 | 15 | 465 | 50 | 154 | 14 | 92 | 1775 | 1908 | 1585 | 1083 | 553 | 425 | 1069 | 38 | 1203 | 628 |  | 453 | 370 | 264 | 9 | 74 |  |  |
| 1987 | 2103 | 18 | 69 | 250 | 398 | 267 | 185 | 440 | 899 | 779 | 606 | 617 | 957 | 1478 | 827 | 598 | 65 | 632 | 413 | 211 | 166 | 71 | 4916 | 7 |  |
| 1988 | 10 | 8 | 49 | 76 | 88 | 109 | 233 | 279 | 84 | 64 | 62 | 49 | 660 | 1418 | 1306 | 14 | 84 | 570 |  |  | 244 | 74 | 3225 | 7 |  |
| 1989 | 2100 | 24 | 154 | 298 | 205 | 70 | 34 | 82 | 87 | 139 | 348 | 339 | 366 | 871 | 1193 | 1294 | 143 | 945 | 858 | 666 |  | 47 |  |  |  |
| 19 |  | 201 | 4 | 699 | 35 | 133 | 122 | 249 | 29 | 321 | 276 | 175 | 123 | 194 | 2 | 346 | 419 | 283 | 6 | 182 | 128 | 2 | 33261 |  |  |
| 1991 | 285 | 131 | 389 | 432 | 369 | 229 | 272 | 620 | 898 | 932 | 631 | 346 | 193 | 301 | 312 | 250 | 21 | 207 | 7 | 10 | 12 | 49 | 2022 | 7 |  |
| 19 | 98 | 18 | 45 | 517 | 6 | 556 | 435 | 854 | 1075 | 856 | 542 | 451 | 622 | 915 | 546 | 242 | 222 | 176 | 6103 | 97 | 86 | 51 | 37281 |  |  |
| 19 | 102 | 114 | 92 | 1088 | 98 | 677 | 213 | 247 | 614 | 847 | 666 | 48 | 615 | 1071 | 665 | 399 | 26 | 230 | - 85 | 62 | 48 | 37 | 2023 |  |  |
| 19 | 2-118 | 19 | 14 | 29 | 36 | 32 | 445 | 95 | 1922 | 2081 | 1121 | 4 | 52 | 1216 | 96 | 1059 | 920 | 565 | 5 | 92 | 46 | 4 | 60162 |  |  |
| 19 | 96 | 30 | 73 | 135 | 208 | 77 | 17 | 460 | 691 | 79 | 70 | 06 | 233 | 60 | 616 | 434 | 484 | 326 | 3 | 2 | 84 | 40 | 27 | 9 |  |
| 199 | -97 | 14 | 65 | 16 | 19 | 11 | 10 | 35 | 69 | 67 | 52 | 4 | 74 | 477 | 1404 | 908 | 49 | 288 | 237 | 48 | 109 | 71 | 25 | 7 |  |
| 19 | -96 | 9 | 473 | 60 | 728 | 507 | 140 | 215 | 48 | 628 | 45 | 40 | 39 | 919 | 809 | 842 | 58 | 436 | 6215 | 105 | 60 | 40 | 2610 |  |  |
| 1998 | $2-98$ | 30 | 262 | 334 | 74 | 46 | 311 | 1151 | 1837 | 1396 | 65 | 379 | 367 | 659 | 458 | 378 | 39 | 333 | 24 | 132 | 64 | 33 | $29 \quad 9$ |  |  |
| 19 | -10 | 7 | 33 | 286 | 113 | 141 | 41 | 760 | 874 | 67 | 71 | 169 | 164 | 1854 | 768 | 4 | 447 | 337 | 2 | 132 | 89 | 62 | 37 | 7 |  |
| 2000 | 112 | 175 | 918 | 1310 | 505 | 54 | 141 | 488 | 785 | 604 | 56 | 749 | 958 | 1720 | 1419 | 894 | 53 | 266 | 188 | 99 | 79 | 57 | 3319 |  |  |
| 20 | $2-141$ | 95 | 646 | 1828 | 2113 | 1010 | 408 | 903 | 1990 | 2543 | 1614 | 705 | 486 | 1192 | 1277 | 1077 | 818 | 514 | 257 | 123 | 71 | 34 | 2214 | 4 |  |
| 2002 | 2-111 | 31 | 190 | 374 | 352 | 105 | 209 | 664 | 1459 | 1449 | 1005 | 792 | 1216 | 1578 | 878 | 609 | 545 | 367 | 208 | 103 | 49 | 19 | 1615 | 3 |  |
| 2003 | -111 | 19 | 283 | 633 | 774 | 682 | 489 | 182 | 252 | 682 | 83 | 974 | 1192 | 1974 | 218 | 770 | 516 | 340 | 261 | 142 | 86 | 35 | 14 | 1 |  |
| 2004 | 2-104 | 24 | 275 | 483 | 562 | 318 | 218 | 484 | 729 | 931 | 979 | 712 | 578 | 806 | 925 | 844 | 71 | 474 | 28 | 211 | 111 | 82 | 3415 | 5 |  |
| 20052 | -106 | 5 | 153 | 589 | 891 | 1017 | 1051 | 488 | 419 | 576 | 72 | 652 | 632 | 859 | 702 | 518 | 525 | 5490 | 355 | 288 | 180 | 102 | 4621 | 7 |  |
| 20062 | 2110 | 478 | 1288 | 1076 | 885 | 317 | 165 | 266 | 605 | 754 | 867 | 707 | 533 | 729 | 856 | 643 | 494 | 435 | 5321 | 259 | 238 | 144 | 76351 |  |  |

Table 2.11—Age composition estimates from the 1994 and 1996-2005 EBS shelf bottom trawl surveys (expressed as numbers per 10,000 ).

| Age | 1994 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1 | 536 | 32 | 2355 | 664 | 715 | 2240 | 2598 | 794 | 1487 | 1421 | 1834 |
| 2 | 4015 | 2306 | 1841 | 4546 | 1992 | 1162 | 2469 | 1869 | 1633 | 1622 | 2567 |
| 3 | 1844 | 2469 | 1737 | 2020 | 3090 | 1675 | 2052 | 3105 | 2546 | 2805 | 1880 |
| 4 | 1259 | 3568 | 1610 | 1137 | 2409 | 2476 | 941 | 2444 | 2212 | 1301 | 1387 |
| 5 | 1241 | 941 | 1225 | 589 | 806 | 1563 | 915 | 734 | 1220 | 1333 | 619 |
| 6 | 837 | 541 | 898 | 596 | 575 | 595 | 703 | 575 | 412 | 908 | 837 |
| 7 | 195 | 144 | 227 | 284 | 266 | 108 | 236 | 390 | 291 | 346 | 478 |
| 8 | 50 | 0 | 81 | 140 | 103 | 120 | 56 | 65 | 151 | 177 | 240 |
| 9 | 20 | 0 | 9 | 22 | 36 | 28 | 14 | 18 | 33 | 62 | 104 |
| 10 | 1 | 0 | 10 | 0 | 0 | 26 | 9 | 5 | 3 | 11 | 16 |
| 11 | 2 | 0 | 6 | 2 | 7 | 7 | 6 | 0 | 3 | 14 | 39 |
| $12+$ | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 7 | 0 | 0 |

Table 2.12a-Biomass, standard error, $95 \%$ confidence interval (CI), and population numbers of Pacific cod estimated by NMFS' annual bottom trawl survey of the EBS shelf, 1979-1981. All figures except population numbers are expressed in metric tons. Population numbers are expressed in terms of individual fish.

| Year | Biomass | Standard Error | Lower 95\% CI | Upper 95\% CI | Numbers |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1979 | 754,314 | 97,844 | 562,539 | 946,089 | $1,530,429,650$ |
| 1980 | 905,344 | 87,898 | 733,063 | $1,077,624$ | $1,084,147,540$ |
| 1981 | $1,034,629$ | 123,849 | 791,885 | $1,277,373$ | $794,619,624$ |

Table 2.12b-Biomass, standard error, $95 \%$ confidence interval (CI), and population numbers of Pacific cod estimated by NMFS' annual bottom trawl survey of the EBS shelf, 1982-2006. All figures except population numbers are expressed in metric tons. Population numbers are expressed in terms of individual fish.

| Year | Biomass | Standard Error | Lower 95\% CI | Upper 95\% CI | Numbers |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1982 | $1,012,856$ | 73,588 | 867,151 | $1,158,562$ | $583,715,842$ |
| 1983 | $1,185,419$ | 120,868 | 941,146 | $1,429,692$ | $751,066,723$ |
| 1984 | $1,048,595$ | 63,643 | 922,583 | $1,174,608$ | $680,914,697$ |
| 1985 | $1,001,108$ | 55,845 | 890,536 | $1,111,681$ | $841,108,075$ |
| 1986 | $1,117,774$ | 69,604 | 979,957 | $1,255,590$ | $838,123,105$ |
| 1987 | $1,106,621$ | 68,682 | 970,630 | $1,242,612$ | $728,956,963$ |
| 1988 | 959,000 | 76,265 | 807,996 | $1,110,004$ | $508,065,276$ |
| 1989 | 836,177 | 62,981 | 711,475 | 960,878 | $292,210,905$ |
| 1990 | 691,255 | 51,455 | 589,375 | 793,136 | $423,835,267$ |
| 1991 | 517,209 | 38,158 | 441,657 | 592,761 | $488,861,768$ |
| 1992 | 551,369 | 45,780 | 460,725 | 642,013 | $601,795,262$ |
| 1993 | 690,535 | 54,380 | 582,862 | 798,208 | $851,863,422$ |
| 1994 | $1,368,120$ | 250,044 | 868,032 | $1,868,209$ | $1,237,758,281$ |
| 1995 | $1,003,096$ | 91,739 | 821,453 | $1,184,740$ | $757,657,482$ |
| 1996 | 890,793 | 87,552 | 717,439 | $1,064,146$ | $609,304,214$ |
| 1997 | 604,881 | 69,250 | 466,382 | 743,380 | $487,429,700$ |
| 1998 | 558,419 | 45,182 | 468,960 | 647,879 | $537,278,347$ |
| 1999 | 583,891 | 50,621 | 483,662 | 684,120 | $500,915,139$ |
| 2000 | 528,466 | 43,037 | 443,253 | 613,679 | $481,358,109$ |
| 2001 | 833,626 | 76,247 | 681,133 | 986,119 | $985,568,802$ |
| 2002 | 618,680 | 69,082 | 480,516 | 756,845 | $566,471,072$ |
| 2003 | 595,826 | 62,099 | 471,628 | 720,024 | $499,925,561$ |
| 2004 | 596,464 | 35,191 | 526,787 | 666,142 | $424,075,921$ |
| 2005 | 603,788 | 43,150 | 517,488 | 690,089 | $452,075,840$ |
| 2006 | 517,698 | 28,341 | 461,583 | 573,813 | $393,993,981$ |

Table 2.13--Length frequencies of Pacific cod in the 2002-2004 EBS slope trawl survey, by year and length bin.
$\mathrm{N}=$ input sample size.

| Yr. | S | N | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 2002 | 2 | 22 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 5 | 18 | 69 | 105 | 86 | 62 | 55 | 39 | 21 | 7 | 1 | 0 | 0 | 0 | 0 |
| 2004 | 2 | 23 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 2 | 32 | 94 | 114 | 128 | 93 | 44 | 10 | 7 | 3 | 2 | 0 | 0 | 0 |

Table 2.14a--Length frequencies of Pacific cod in the Japanese longline survey, by year and length bin.
$\mathrm{N}=$ input sample size.

| Yr. S | N | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1982 | 2 | 177 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 | 50 | 271 | 125 | 1445 | 2048 | 4054 | 4321 | 7416 | 4607 | 4400 | 1252 | 748 | 316 | 205 | 35 | 14 |
| 1983 | 2 | 170 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 17 | 37 | 514 | 2527 | 3062 | 4174 | 4691 | 4504 | 4104 | 2922 | 1419 | 509 | 231 | 86 | 17 |
| 1984 | 2 | 177 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 23 | 60 | 64 | 415 | 2000 | 3033 | 5608 | 6407 | 5018 | 3601 | 2452 | 1552 | 736 | 258 | 70 | 12 |
| 1985 | 2 | 193 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 7 | 90 | 508 | 3292 | 3157 | 2270 | 3822 | 5245 | 6049 | 5015 | 3205 | 2272 | 1302 | 699 | 164 | 17 |
| 1986 | 2 | 212 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 31 | 113 | 232 | 1598 | 4740 | 8267 | 8191 | 5006 | 4183 | 4372 | 3464 | 2518 | 1252 | 675 | 268 | 31 |
| 1987 | 2 | 210 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 13 | 104 | 1415 | 3570 | 4593 | 7155 | 9199 | 6638 | 3606 | 2650 | 2354 | 1590 | 846 | 339 | 64 |
| 1988 | 2 | 109 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 20 | 44 | 318 | 987 | 2114 | 2691 | 2146 | 1437 | 928 | 482 | 256 | 226 | 129 | 44 | 7 |
| 1989 | 2 | 140 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 34 | 183 | 825 | 1619 | 2611 | 3538 | 3735 | 2901 | 1788 | 1053 | 642 | 396 | 212 | 63 |
| 1990 | 2 | 89 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 59 | 292 | 701 | 1257 | 1655 | 1635 | 1008 | 650 | 314 | 168 | 85 | 29 | 11 |
| 1991 | 2 | 80 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 3 | 10 | 58 | 388 | 638 | 1053 | 1254 | 1130 | 712 | 496 | 313 | 164 | 101 | 24 | 11 |
| 1992 | 2 | 77 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 28 | 361 | 1024 | 967 | 1089 | 1116 | 656 | 344 | 192 | 109 | 52 | 35 | 11 | 8 |
| 1993 | 2 | 87 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 28 | 391 | 1051 | 1404 | 1996 | 1411 | 662 | 318 | 137 | 72 | 47 | 31 | 17 | 5 |
| 1994 | 2 | 100 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 16 | 176 | 716 | 1409 | 2645 | 2648 | 1532 | 526 | 164 | 67 | 32 | 20 | 8 | 5 |

Table 2.14b--Length frequencies of Pacific cod in the U.S. longline survey, by year and length bin.
$\mathrm{N}=$ input sample size.

| Yr. | N | N | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1997 | 2 | 90 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 11 | 93 | 499 | 982 | 1531 | 2344 | 1691 | 598 | 186 | 87 | 50 | 36 | 10 | 3 |
| 1999 | 2 | 91 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 11 | 82 | 735 | 1107 | 1386 | 1712 | 1619 | 913 | 420 | 133 | 49 | 26 | 6 | 10 | 3 |
| 2001 | 2 | 96 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 26 | 216 | 681 | 1495 | 2077 | 2332 | 1490 | 522 | 169 | 63 | 21 | 10 | 9 | 1 |
| 2003 | 2 | 90 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 9 | 68 | 661 | 1694 | 1842 | 1905 | 1185 | 484 | 164 | 82 | 31 | 18 | 5 | 2 | 0 |
| 2005 | 2 | 74 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 26 | 123 | 522 | 855 | 1137 | 1195 | 902 | 448 | 207 | 72 | 34 | 4 | 2 | 1 |

Table 2.15-Japanese and U.S. longline survey abundance indices. Mean = average catch (in numbers of fish) per station. $\mathrm{CV}=$ coefficient of variation.

|  | Japanese Survey |  |
| ---: | ---: | ---: |
| Year | Mean | CV |
| 1982 | 315.31 | 0.14 |
| 1983 | 258.56 | 0.12 |
| 1984 | 250.72 | 0.10 |
| 1985 | 928.66 | 0.13 |
| 1986 | 1086.31 | 0.11 |
| 1987 | 540.19 | 0.17 |
| 1988 | 365.16 | 0.12 |
| 1989 | 1252.28 | 0.12 |
| 1990 | 686.03 | 0.15 |
| 1991 | 567.25 | 0.13 |
| 1992 | 208.00 | 0.19 |
| 1993 | 328.81 | 0.15 |
| 1994 | 439.16 | 0.18 |


|  | U.S. Survey |  |
| :---: | ---: | ---: |
| Year | Mean | CV |
| 1997 | 758.45 | 0.21 |
| 1999 | 718.73 | 0.24 |
| 2001 | 777.18 | 0.19 |
| 2003 | 733.73 | 0.22 |
| 2005 | 510.55 | 0.21 |

Table 2.16-Summary of key parameter estimates and objective function values from last year, from last year's model updated with new data (Model 0), and from eight alternative models.

| Item | Last Yr. | Model0 | ModelA1 | ModelA2 | ModelB1 | ModelB2 | ModelC1 | ModelC2 | ModelD1 | ModelD2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline survey data: | excluded | excluded | excluded | excluded | excluded | excluded | included | included | included | included |
| Selectivity function: | dbl. log. | dbl. log. | bl. log. | dbl. log. | dbl. nor | dbl. nor. | dbl. log. | dbl. log. | dbl. nor. | dbl. nor. |
| Weight assigned to log priors: | 1.0 | 1.0 | 1.0 | 0.5 | 1.0 | 0.5 | 1.0 | 0.5 | 1.0 | 0.5 |
| sigmaR | 0.69 | 0.62 | 0.63 | 0.65 | 0.62 | 0.61 | 0.61 | 0.61 | 0.59 | 0.59 |
| $\ln$ (post76 Rmed) | 13.50 | 13.53 | 13.60 | 13.69 | 13.62 | 13.61 | 13.51 | 13.53 | 13.51 | 13.49 |
| $\ln$ (pre77 Rmed)-ln(post76 Rmed) | -1.31 | -1.19 | -1.21 | -1.19 | -1.18 | -1.18 | -1.23 | -1.22 | -1.21 | -1.21 |
| pre82 shelf trawl surv. catchability | 1.00 | 1.00 | 0.90 | 0.87 | 0.97 | 0.99 | 1.10 | 1.19 | 1.24 | 1.28 |
| post81 shelf trawl surv. catchability | 1.00 | 1.00 | 0.61 | 0.55 | 0.57 | 0.59 | 0.67 | 0.66 | 0.68 | 0.70 |
| post81 shelf trawl surv. sel. at 90 cm | 0.55 | 0.64 | 0.95 | 0.93 | 1.00 | 1.00 | 0.98 | 0.99 | 1.00 | 1.00 |
| pre82 shelf trawl surv. abund. $\ln$ (like) | n/a | 0.92 | 0.71 | 0.30 | 0.40 | 0.42 | 0.64 | 0.33 | 0.36 | 0.36 |
| post81 shelf trawl surv. abund. ln(like) | 46.38 | 49.71 | 47.57 | 48.12 | 45.29 | 45.03 | 51.15 | 50.50 | 47.91 | 47.78 |
| slope trawl surv. abund. $\ln$ (like) | 0.59 | 0.28 | 0.29 | 0.27 | 0.22 | 0.24 | 0.31 | 0.31 | 0.25 | 0.26 |
| Japan longline surv. abund. $\ln$ (like) | n/a | $\mathrm{n} / \mathrm{a}$ | n /a | n /a | n /a | n /a | 142.96 | 144.48 | 145.17 | 144.13 |
| U.S. longline surv. abund. $\ln$ (like) | n/a | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | n/a | n/a | n/a | 0.80 | 0.82 | 0.87 | 0.95 |
| Jan-May trawl fishery size comp. ln(like) | 296.21 | 273.60 | 270.37 | 265.18 | 266.07 | 266.46 | 275.08 | 270.60 | 272.81 | 273.41 |
| Jun-Dec trawl fishery size comp. $\ln$ (like) | 323.57 | 448.09 | 445.95 | 443.72 | 440.95 | 440.87 | 442.23 | 440.70 | 438.74 | 438.38 |
| longline fishery size comp. $\ln$ (like) | 571.88 | 587.42 | 585.81 | 578.79 | 489.07 | 488.12 | 587.37 | 583.33 | 479.77 | 478.73 |
| pot fishery size comp. $\ln$ (like) | 217.37 | 204.86 | 202.91 | 200.10 | 172.44 | 172.13 | 203.32 | 200.69 | 169.61 | 169.53 |
| pre82 shelf trawl surv. size comp. $\ln$ (like) | n/a | 36.96 | 35.97 | 34.80 | 35.18 | 35.19 | 42.11 | 40.72 | 40.65 | 40.68 |
| post81 shelf trawl surv. size comp. $\ln$ (like) | 246.86 | 177.66 | 173.33 | 175.54 | 188.30 | 187.76 | 179.23 | 179.31 | 194.69 | 194.47 |
| slope trawl surv. size comp. $\ln$ (like) | 3.65 | 3.54 | 3.83 | 3.36 | 1.17 | 1.16 | 3.30 | 2.87 | 1.13 | 1.13 |
| Japan longline surv. size comp. $\ln$ (like) | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | n /a | 103.83 | 104.18 | 89.62 | 89.21 |
| U.S. longline surv. size comp. $\ln$ (like) | n/a | $\mathrm{n} / \mathrm{a}$ | n/a | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | n/a | 26.87 | 27.04 | 24.01 | 23.73 |
| post81 shelf trawl surv. age comp. $\ln$ (like) | 70.20 | 88.31 | 87.84 | 92.55 | 94.91 | 94.17 | 83.70 | 84.89 | 90.08 | 89.52 |
| post81 shelf trawl surv. size-at-age $\ln$ (like) | 204.14 | 411.36 | 386.15 | 376.58 | 420.83 | 422.62 | 397.94 | 394.23 | 433.58 | 435.53 |
| recruitment $\ln ($ like $)$ | 29.48 | 30.17 | 30.22 | 29.37 | 30.04 | 30.18 | 31.32 | 31.09 | 31.54 | 31.62 |
| log priors | 106.56 | 79.87 | 87.48 | 115.76 | 18.59 | 19.85 | 87.76 | 108.82 | 24.07 | 26.35 |
| log posterior (weighted sum of the above) | 2,116.89 | 2,392.74 | 2,358.42 | 2,306.55 | 2,203.47 | 2,194.25 | 2,659.93 | 2,610.49 | 2,484.86 | 2,472.59 |

Table 2.17-Summary of management-related quantities as estimated last year, as estimated using last year's model updated with new data (Model 0), and as estimated by eight alternative models. Results in normal font correspond to outputs from the SS2 assessment model, and results in bold font correspond to outputs from the standard projection model.

| Item | Last Yr. | Model0 | ModelA1 | ModelA2 | ModelB1 | ModelB2 | ModelC1 | ModelC2 | ModelD1 | ModelD2 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline survey data: | excluded |  |  |  |  |  |  |  |  |  |
| Selectivity function: | dbl. log. | dbl. log. | dbl. log. | dbl. log. | dbl. no | dbl. nor. | dbl. log. | dbl. log. | dbl. nor. | dbl. nor. |
| Weight assigned to log priors: | 1.0 | 1.0 | 1.0 | 0.5 | 1.0 | 0.5 | 1.0 | 0.5 | 1.0 | 0.5 |
| BSAI total biomass 2005 | 1,073,901 | 1,130,968 | 1,264,702 | 1,425,857 | 1,318,952 | 1,291,488 | 1,108,749 | 1,134,506 | 1,119,163 | 1,099,939 |
| BSAI total biomass 2006 | n/a | 979,815 | 1,110,143 | 1,257,202 | 1,169,095 | 1,145,140 | 966,200 | 990,648 | 982,092 | 965,839 |
| BSAI female spawning biomass 2005 | 333,028 | 362,914 | 411,634 | 475,793 | 428,882 | 417,148 | 349,905 | 359,992 | 352,192 | 343,696 |
| BSAI female spawning biomass 2006 | 278,665 | 322,305 | 370,004 | 429,826 | 388,571 | 378,228 | 311,854 | 321,839 | 315,130 | 307,879 |
| BSAI female spawning biomass 2007 | 246,178 | 247,860 | 289,029 | 337,705 | 306,790 | 298,766 | 241,216 | 249,661 | 245,508 | 240,106 |
| BSAI female spawning biomass 2008 | 224,295 | 221,620 | 252,250 | 285,396 | 266,134 | 260,735 | 219,645 | 225,409 | 223,904 | 220,351 |
| Proportion of B100\% in 2005 | 0.40 | 0.43 | 0.46 | 0.48 | 0.46 | 0.46 | 0.42 | 0.43 | 0.43 | 0.42 |
| Proportion of B100\% in 2006 | 0.34 | 0.38 | 0.41 | 0.44 | 0.42 | 0.42 | 0.38 | 0.38 | 0.38 | 0.38 |
| Proportion of B100\% in 2007 | 0.33 | 0.34 | 0.37 | 0.40 | 0.38 | 0.38 | 0.33 | 0.34 | 0.34 | 0.34 |
| Proportion of B100\% in 2008 | 0.30 | 0.30 | 0.32 | 0.34 | 0.33 | 0.33 | 0.30 | 0.31 | 0.31 | 0.31 |
| BSAI ABC 2006 (Council adopted) | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 | 194,000 |
| BSAI maxABC 2007 (from model) | 148,297 | 132,544 | 162,172 | 193,186 | 176,482 | 172,080 | 129,126 | 135,242 | 133,536 | 130,660 |
| BSAI maxABC 2008 (from model) | 121,390 | 104,447 | 121,971 | 135,858 | 130,876 | 129,146 | 105,690 | 108,732 | 109,609 | 108,618 |
| rel. change in ABC (2006 to 2007) | -0.24 | -0.32 | -0.16 | 0.00 | -0.09 | -0.11 | -0.33 | -0.30 | -0.31 | -0.33 |
| rel. change in ABC (2007 to 2008) | -0.18 | -0.21 | -0.25 | -0.30 | -0.26 | -0.25 | -0.18 | -0.20 | -0.18 | -0.17 |
| BSAI OFL 2006 (Council adopted) | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 | 230,000 |
| BSAI OFL 2007 (from model) | 176,135 | 156,515 | 190,609 | 225,845 | 206,861 | 201,918 | 152,651 | 159,669 | 157,630 | 154,438 |
| BSAI OFL 2008 (from model) | 144,805 | 123,719 | 143,934 | 159,624 | 154,061 | 152,162 | 125,285 | 128,756 | 129,736 | 128,709 |
| rel. change in OFL (2006 to 2007) | -0.23 | -0.32 | -0.17 | -0.02 | -0.10 | -0.12 | -0.34 | -0.31 | -0.31 | -0.33 |
| rel. change in OFL (2007 to 2008) | -0.18 | -0.21 | -0.24 | -0.29 | -0.26 | -0.25 | -0.18 | -0.19 | -0.18 | -0.17 |

Table 2.18—Estimates of Pacific cod fishing mortality rates, expressed on an annual time scale (Model B1). Empty cells indicate that recorded catch was negligible or that no catch was recorded.

|  | Trawl |  |  | Longline |  |  | Pot |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 | Sea. 1 | Sea. 2 | Sea. 3 |
| 1964 | 0.018 | 0.007 | 0.008 | 0.003 | 0.000 | 0.004 |  |  |  |
| 1965 | 0.020 | 0.008 | 0.008 | 0.003 | 0.000 | 0.005 |  |  |  |
| 1966 | 0.024 | 0.009 | 0.010 | 0.003 | 0.000 | 0.006 |  |  |  |
| 1967 | 0.042 | 0.016 | 0.018 | 0.006 | 0.001 | 0.011 |  |  |  |
| 1968 | 0.080 | 0.032 | 0.037 | 0.011 | 0.002 | 0.021 |  |  |  |
| 1969 | 0.079 | 0.032 | 0.036 | 0.011 | 0.002 | 0.022 |  |  |  |
| 1970 | 0.129 | 0.054 | 0.063 | 0.018 | 0.003 | 0.038 |  |  |  |
| 1971 | 0.102 | 0.041 | 0.047 | 0.014 | 0.002 | 0.029 |  |  |  |
| 1972 | 0.119 | 0.048 | 0.055 | 0.017 | 0.003 | 0.034 |  |  |  |
| 1973 | 0.168 | 0.070 | 0.080 | 0.024 | 0.004 | 0.048 |  |  |  |
| 1974 | 0.221 | 0.097 | 0.110 | 0.032 | 0.005 | 0.063 |  |  |  |
| 1975 | 0.206 | 0.091 | 0.100 | 0.030 | 0.005 | 0.056 |  |  |  |
| 1976 | 0.206 | 0.091 | 0.096 | 0.031 | 0.005 | 0.054 |  |  |  |
| 1977 | 0.120 | 0.049 | 0.049 | 0.019 | 0.002 | 0.028 |  |  |  |
| 1978 | 0.110 | 0.045 | 0.044 | 0.018 | 0.002 | 0.024 |  |  |  |
| 1979 | 0.060 | 0.024 | 0.023 | 0.009 | 0.001 | 0.013 |  |  |  |
| 1980 | 0.042 | 0.017 | 0.016 | 0.007 | 0.001 | 0.009 |  |  |  |
| 1981 | 0.022 | 0.020 | 0.027 | 0.002 | 0.001 | 0.005 |  |  |  |
| 1982 | 0.022 | 0.019 | 0.016 | 0.000 | 0.001 | 0.002 |  |  |  |
| 1983 | 0.034 | 0.021 | 0.019 | 0.003 | 0.001 | 0.002 |  |  |  |
| 1984 | 0.037 | 0.020 | 0.020 | 0.004 | 0.002 | 0.017 |  |  |  |
| 1985 | 0.044 | 0.023 | 0.018 | 0.012 | 0.002 | 0.020 |  |  |  |
| 1986 | 0.050 | 0.022 | 0.018 | 0.008 | 0.000 | 0.017 |  |  |  |
| 1987 | 0.056 | 0.013 | 0.018 | 0.021 | 0.001 | 0.027 |  |  |  |
| 1988 | 0.114 | 0.023 | 0.042 | 0.001 | 0.001 | 0.002 |  |  |  |
| 1989 | 0.125 | 0.016 | 0.019 | 0.004 | 0.005 | 0.005 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.110 | 0.012 | 0.011 | 0.013 | 0.019 | 0.020 |  | 0.001 | 0.000 |
| 1991 | 0.135 | 0.020 | 0.008 | 0.030 | 0.029 | 0.042 | 0.000 | 0.001 | 0.003 |
| 1992 | 0.090 | 0.018 | 0.009 | 0.074 | 0.040 | 0.010 | 0.004 | 0.008 | 0.000 |
| 1993 | 0.110 | 0.009 | 0.014 | 0.079 | 0.000 | 0.000 | 0.004 | 0.000 |  |
| 1994 | 0.095 | 0.009 | 0.026 | 0.086 | 0.000 | 0.030 | 0.008 |  | 0.005 |
| 1995 | 0.133 | 0.013 | 0.017 | 0.096 | 0.000 | 0.043 | 0.018 | 0.006 | 0.005 |
| 1996 | 0.116 | 0.005 | 0.015 | 0.087 | 0.000 | 0.039 | 0.027 | 0.010 | 0.005 |
| 1997 | 0.122 | 0.006 | 0.013 | 0.100 | 0.000 | 0.070 | 0.022 | 0.006 | 0.005 |
| 1998 | 0.073 | 0.009 | 0.016 | 0.084 | 0.000 | 0.049 | 0.015 | 0.005 | 0.002 |
| 1999 | 0.074 | 0.006 | 0.006 | 0.089 | 0.003 | 0.039 | 0.016 | 0.002 | 0.004 |
| 2000 | 0.078 | 0.008 | 0.007 | 0.060 | 0.002 | 0.056 | 0.025 |  | 0.000 |
| 2001 | 0.037 | 0.011 | 0.008 | 0.052 | 0.009 | 0.059 | 0.017 | 0.001 | 0.006 |
| 2002 | 0.056 | 0.014 | 0.006 | 0.065 | 0.016 | 0.049 | 0.015 | 0.001 | 0.005 |
| 2003 | 0.051 | 0.014 | 0.004 | 0.068 | 0.015 | 0.055 | 0.021 | 0.000 | 0.009 |
| 2004 | 0.062 | 0.018 | 0.005 | 0.072 | 0.014 | 0.058 | 0.017 | 0.001 | 0.006 |
| 2005 | 0.070 | 0.010 | 0.001 | 0.074 | 0.019 | 0.071 | 0.017 |  | 0.008 |
| 2006 | 0.079 | 0.011 | 0.002 | 0.081 | 0.026 | 0.085 | 0.023 |  | 0.009 |

Table 2.19-Estimates of Pacific cod regime-specific median recruitments and recruitment deviations (Model B1). Deviations are expressed as the difference between the logarithm of annual recruitment at age 0 and the logarithm of median recruitment for the respective environmental regime.

| Year | $\ln$ (Median Recruitment) | Annual Deviation |
| ---: | ---: | ---: |
| 1964 | 12.443 | -0.366 |
| 1965 | 12.443 | -0.447 |
| 1966 | 12.443 | -0.512 |
| 1967 | 12.443 | -0.520 |
| 1968 | 12.443 | -0.406 |
| 1969 | 12.443 | -0.157 |
| 1970 | 12.443 | -0.222 |
| 1971 | 12.443 | -0.295 |
| 1972 | 12.443 | -0.157 |
| 1973 | 12.443 | 0.433 |
| 1974 | 12.443 | 1.343 |
| 1975 | 12.443 | -0.941 |
| 1976 | 12.443 | 2.296 |
| 1977 | 13.623 | 0.861 |
| 1978 | 13.623 | 0.398 |
| 1979 | 13.623 | 0.330 |
| 1980 | 13.623 | -0.407 |
| 1981 | 13.623 | 0.263 |
| 1982 | 13.623 | 0.795 |
| 1983 | 13.623 | -0.576 |
| 1984 | 13.623 | 0.575 |
| 1985 | 13.623 | -0.489 |
| 1986 | 13.623 | -0.553 |
| 1987 | 13.623 | -0.776 |
| 1988 | 13.623 | 0.286 |
| 1989 | 13.623 | 0.514 |
| 1990 | 13.623 | -0.035 |
| 1991 | 13.623 | 0.308 |
| 1992 | 13.623 | 0.384 |
| 1993 | 13.623 | -0.603 |
| 1994 | 13.623 | -0.187 |
| 1995 | 13.623 | 0.369 |
| 1996 | 13.623 | 0.484 |
| 1997 | 13.623 | -0.126 |
| 1998 | 13.623 | 0.198 |
| 1999 | 13.623 | 0.407 |
| 2000 | 13.623 | -0.254 |
| 2001 | 13.623 | -0.387 |
| 2002 | 13.623 | -0.462 |
| 2003 | 13.623 | -0.546 |
| 2004 | 13.623 | -0.809 |
| 2005 | 13.623 | -0.009 |
|  |  |  |

Table 2.20-Estimates of Pacific cod selectivity parameters (Model B1). The first column lists the years defining the era for which the parameter values in that row are applicable. The eras for the commercial fisheries are 1964-1988, 1989-1999, and 2000-2006 (no eras per se are defined for the surveys, although separate shelf bottom trawl surveys are defined for the years prior to 1982 and after 1981). The second column lists the particular parameter being described. Four parameters define the shape of the selectivity function: the size at which selectivity first reaches a value of 1.0 ("peak location"), the logit transform of the region (within the range from peak location to the maximum length in the model) over which selectivity remains at a value of 1.0 ("logit(peak width)"), the $\log$ of the variance term in the ascending curve (" $\ln ($ asc. variance)"), and the log of the variance term in the descending curve (" $\ln$ (des. variance)"). See text for further description of these parameters and how they are used to define the selectivity function. The remaining columns correspond to the fishery or survey to which the values are applicable, using the following notation: TWL1 = January-May trawl fishery, TWL2 $=$ June-December trawl fishery, LGL = longline fishery, POT = pot fishery, SRV1 = pre-1982 shelf trawl survey, SRV2 = post-1981 shelf trawl survey, and SRV3 = slope trawl survey.

| Years | Parameter | TWL1 | TWL2 | LGL | POT |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1964-1988 | peak location | 76.519 | 80.324 | 72.835 |  |
| 1989-1999 | peak location | 79.197 | 78.482 | 69.926 | 70.626 |
| 2000-2006 | peak location | 81.159 | 83.026 | 66.386 | 66.059 |
| 1964-1988 | logit(peak width) | -8.007 | -0.040 | -3.641 |  |
| 1989-1999 | logit(peak width) | -1.872 | 1.742 | -0.446 | 0.040 |
| 2000-2006 | logit(peak width) | -7.936 | 1.391 | -2.728 | -7.983 |
| 1964-1988 | $\ln$ (asc. variance) | 6.329 | 6.419 | 5.526 |  |
| 1989-1999 | $\ln$ (asc. variance) | 6.373 | 6.303 | 5.336 | 5.135 |
| 2000-2006 | $\ln$ (asc. variance) | 6.283 | 6.558 | 5.294 | 4.746 |
| 1964-1988 | $\ln$ (des. variance) | 6.327 | 5.674 | 6.284 |  |
| 1989-1999 | $\ln$ (des. variance) | 5.986 | 3.946 | 6.011 | 5.541 |
| 2000-2006 | $\ln$ (des. variance) | 6.427 | 3.800 | 6.729 | 7.392 |
| Years | Parameter | SRV1 | SRV2 | SRV3 |  |
| n/a | peak location | 40.245 | 45.071 | 55.825 |  |
| $\mathrm{n} / \mathrm{a}$ | $\operatorname{logit}$ (peak width) | -8.842 | 3.678 | -1.388 |  |
| $\mathrm{n} / \mathrm{a}$ | $\ln$ (asc. variance) | 5.257 | 7.103 | 4.225 |  |
| $\mathrm{n} / \mathrm{a}$ | $\ln$ (des. variance) | 7.034 | 2.670 | 5.555 |  |

Table 2.21a-Schedules of Pacific cod selectivities at length in the commercial fisheries as defined by final parameter estimates (Model B1). Lengths (cm) correspond to mid-points of size bins. Len. $=$ length, $\mathrm{FOR}=1964-1988, \mathrm{DOM}=1989-1999, \mathrm{NEW}=2000-2006$.

|  | Jan-May Trawl |  |  | Fishery | Jun-Dec Trawl Fishery |  |  | Longline Fishery |  | Pot Fishery |  |
| ---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Len. | FOR | DOM | NEW | FOR | DOM | NEW | FOR | DOM | NEW | DOM | NEW |
| 10.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 13.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 16.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 19.5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 22.5 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 25.5 | 0.01 | 0.01 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 28.5 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 31.5 | 0.03 | 0.02 | 0.01 | 0.02 | 0.02 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| 34.5 | 0.04 | 0.03 | 0.02 | 0.03 | 0.03 | 0.04 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| 37.5 | 0.07 | 0.05 | 0.03 | 0.05 | 0.05 | 0.05 | 0.01 | 0.01 | 0.02 | 0.00 | 0.00 |
| 40.5 | 0.10 | 0.08 | 0.05 | 0.08 | 0.07 | 0.08 | 0.02 | 0.02 | 0.03 | 0.00 | 0.00 |
| 43.5 | 0.14 | 0.11 | 0.07 | 0.11 | 0.11 | 0.11 | 0.03 | 0.03 | 0.07 | 0.01 | 0.01 |
| 47.5 | 0.22 | 0.18 | 0.12 | 0.17 | 0.17 | 0.17 | 0.08 | 0.09 | 0.17 | 0.04 | 0.05 |
| 52.5 | 0.36 | 0.30 | 0.22 | 0.28 | 0.29 | 0.27 | 0.19 | 0.23 | 0.38 | 0.14 | 0.20 |
| 57.5 | 0.52 | 0.45 | 0.35 | 0.43 | 0.45 | 0.40 | 0.39 | 0.48 | 0.67 | 0.36 | 0.53 |
| 62.5 | 0.70 | 0.62 | 0.52 | 0.60 | 0.63 | 0.55 | 0.65 | 0.77 | 0.93 | 0.68 | 0.90 |
| 67.5 | 0.86 | 0.79 | 0.71 | 0.76 | 0.80 | 0.71 | 0.89 | 0.97 | 1.00 | 0.94 | 1.00 |
| 72.5 | 0.97 | 0.93 | 0.87 | 0.91 | 0.94 | 0.85 | 1.00 | 1.00 | 0.98 | 1.00 | 0.97 |
| 77.5 | 1.00 | 1.00 | 0.98 | 0.99 | 1.00 | 0.96 | 0.97 | 1.00 | 0.91 | 1.00 | 0.92 |
| 82.5 | 0.94 | 1.00 | 1.00 | 1.00 | 1.00 | 1.00 | 0.86 | 1.00 | 0.80 | 1.00 | 0.85 |
| 87.5 | 0.81 | 0.94 | 0.94 | 1.00 | 1.00 | 1.00 | 0.70 | 0.96 | 0.65 | 1.00 | 0.75 |
| 92.5 | 0.63 | 0.78 | 0.81 | 1.00 | 1.00 | 1.00 | 0.51 | 0.81 | 0.51 | 0.91 | 0.65 |
| 97.5 | 0.46 | 0.57 | 0.65 | 0.90 | 1.00 | 1.00 | 0.35 | 0.61 | 0.37 | 0.68 | 0.54 |
| 102.5 | 0.30 | 0.37 | 0.48 | 0.68 | 0.90 | 0.85 | 0.21 | 0.40 | 0.25 | 0.42 | 0.44 |

Table 2.21b-Schedules of Pacific cod selectivities at length in the bottom trawl surveys as defined by final parameter estimates (Model B1). Lengths (cm) correspond to lower bounds of size bins.

|  | Shelf Survey |  |  |
| ---: | ---: | ---: | ---: |
| Length | pre-1982 | post-1981 | Slope |
| 10.5 | 0.01 | 0.37 | 0.00 |
| 13.5 | 0.02 | 0.44 | 0.00 |
| 16.5 | 0.05 | 0.51 | 0.00 |
| 19.5 | 0.11 | 0.58 | 0.00 |
| 22.5 | 0.19 | 0.66 | 0.00 |
| 25.5 | 0.32 | 0.73 | 0.00 |
| 28.5 | 0.49 | 0.80 | 0.00 |
| 31.5 | 0.67 | 0.86 | 0.00 |
| 34.5 | 0.84 | 0.91 | 0.00 |
| 37.5 | 0.96 | 0.95 | 0.01 |
| 40.5 | 1.00 | 0.98 | 0.03 |
| 43.5 | 0.99 | 1.00 | 0.11 |
| 47.5 | 0.95 | 1.00 | 0.36 |
| 52.5 | 0.88 | 1.00 | 0.85 |
| 57.5 | 0.77 | 1.00 | 1.00 |
| 62.5 | 0.65 | 1.00 | 1.00 |
| 67.5 | 0.52 | 1.00 | 0.98 |
| 72.5 | 0.40 | 1.00 | 0.82 |
| 77.5 | 0.29 | 1.00 | 0.57 |
| 82.5 | 0.21 | 1.00 | 0.32 |
| 87.5 | 0.14 | 1.00 | 0.15 |
| 92.5 | 0.09 | 1.00 | 0.06 |
| 97.5 | 0.06 | 1.00 | 0.02 |
| 102.5 | 0.03 | 1.00 | 0.00 |

Table 2.22-Schedules of Pacific cod length ( cm ), proportion mature, and weight ( kg ) by season and age as estimated by Model B1. Pop. = population, Sea. $1=$ Jan-Jun, Sea. $2=$ Jul-Aug, Sea. $3=$ Sep-Dec, Beg. $=$ beginning of season, Mid. = middle of season, SDev. $=$ standard deviation, Mat. = proportion mature, Twl. = trawl fishery, Lgl. = longline fishery, pot = pot fishery, shelf $=$ shelf survey, slope $=$ slope survey.

|  |  | Length |  |  |  | Pop. Weight |  | Fishery Weight |  |  | Survey Wt. |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sea. | Age | Beg. | Mid. | S.Dev. | Mat. | Beg. | Mid. | Twl. | Lgl. | Pot | Shelf | Slope |
| 1 | 1 | 11.10 | 13.79 | 3.54 | 0.00 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 | 0.02 | 0.07 |
| 1 | 2 | 23.44 | 25.85 | 4.87 | 0.01 | 0.13 | 0.17 | 0.24 | 0.31 | 0.42 | 0.18 | 0.45 |
| 1 | 3 | 34.47 | 36.61 | 5.80 | 0.05 | 0.43 | 0.53 | 0.73 | 0.87 | 1.08 | 0.55 | 1.01 |
| 1 | 4 | 44.32 | 46.23 | 6.42 | 0.18 | 0.98 | 1.13 | 1.45 | 1.56 | 1.78 | 1.13 | 1.52 |
| 1 | 5 | 53.11 | 54.82 | 6.80 | 0.38 | 1.75 | 1.94 | 2.35 | 2.32 | 2.49 | 1.94 | 2.13 |
| 1 | 6 | 60.97 | 62.50 | 7.00 | 0.59 | 2.72 | 2.95 | 3.36 | 3.17 | 3.26 | 2.95 | 2.93 |
| 1 | 7 | 67.99 | 69.35 | 7.07 | 0.75 | 3.86 | 4.12 | 4.44 | 4.14 | 4.19 | 4.12 | 3.84 |
| 1 | 8 | 74.25 | 75.47 | 7.05 | 0.85 | 5.12 | 5.40 | 5.58 | 5.24 | 5.31 | 5.40 | 4.75 |
| 1 | 9 | 79.85 | 80.94 | 6.97 | 0.90 | 6.47 | 6.76 | 6.76 | 6.42 | 6.54 | 6.76 | 5.66 |
| 1 | 10 | 84.85 | 85.82 | 6.83 | 0.93 | 7.86 | 8.16 | 7.96 | 7.63 | 7.83 | 8.14 | 6.59 |
| 1 | 11 | 89.31 | 90.18 | 6.67 | 0.95 | 9.27 | 9.56 | 9.16 | 8.86 | 9.13 | 9.49 | 7.53 |
| 1 | 12 | 93.30 | 94.08 | 6.49 | 0.96 | 10.65 | 10.94 | 10.36 | 10.09 | 10.43 | 10.73 | 8.47 |
| 1 | 13 | 96.86 | 97.55 | 6.71 | 0.97 | 11.98 | 12.24 | 11.48 | 11.20 | 11.64 | 11.72 | 9.20 |
| 1 | 14 | 100.04 | 100.66 | 6.90 | 0.98 | 13.16 | 13.38 | 12.52 | 12.26 | 12.75 | 12.49 | 9.87 |
| 2 | 1 | 16.41 | 17.96 | 3.54 | n/a | 0.05 | 0.05 | 0.07 | 0.10 | 0.15 | 0.06 | 0.17 |
| 2 | 2 | 28.19 | 29.57 | 4.87 | n/a | 0.27 | 0.27 | 0.36 | 0.48 | 0.64 | 0.28 | 0.65 |
| 2 | 3 | 38.71 | 39.94 | 5.80 | n/a | 0.71 | 0.71 | 0.92 | 1.11 | 1.33 | 0.72 | 1.19 |
| 2 | 4 | 48.10 | 49.21 | 6.42 | n/a | 1.38 | 1.38 | 1.69 | 1.83 | 2.04 | 1.38 | 1.72 |
| 2 | 5 | 56.49 | 57.48 | 6.80 | n/a | 2.26 | 2.26 | 2.63 | 2.61 | 2.75 | 2.26 | 2.39 |
| 2 | 6 | 63.99 | 64.87 | 7.00 | n/a | 3.33 | 3.33 | 3.69 | 3.49 | 3.57 | 3.33 | 3.23 |
| 2 | 7 | 70.68 | 71.47 | 7.07 | n/a | 4.54 | 4.54 | 4.83 | 4.50 | 4.56 | 4.54 | 4.13 |
| 2 | 8 | 76.66 | 77.36 | 7.05 | n/a | 5.86 | 5.86 | 6.04 | 5.63 | 5.71 | 5.86 | 5.03 |
| 2 | 9 | 82.00 | 82.63 | 6.97 | n/a | 7.23 | 7.23 | 7.32 | 6.81 | 6.96 | 7.23 | 5.94 |
| 2 | 10 | 86.77 | 87.33 | 6.83 | n/a | 8.63 | 8.63 | 8.62 | 8.03 | 8.26 | 8.61 | 6.85 |
| 2 | 11 | 91.03 | 91.53 | 6.67 | n/a | 10.03 | 10.03 | 9.89 | 9.25 | 9.56 | 9.92 | 7.78 |
| 2 | 12 | 94.83 | 95.28 | 6.49 | n/a | 11.39 | 11.39 | 11.04 | 10.46 | 10.84 | 11.09 | 8.72 |
| 2 | 13 | 98.23 | 98.63 | 6.71 | n/a | 12.64 | 12.64 | 11.98 | 11.56 | 12.02 | 12.00 | 9.43 |
| 2 | 14 | 101.27 | 101.62 | 6.90 | n/a | 13.71 | 13.71 | 12.76 | 12.59 | 13.09 | 12.70 | 10.08 |
| 3 | 1 | 19.48 | 21.48 | 3.54 | n/a | 0.10 | 0.10 | 0.13 | 0.19 | 0.26 | 0.10 | 0.30 |
| 3 | 2 | 30.93 | 32.72 | 4.87 | n/a | 0.37 | 0.37 | 0.50 | 0.67 | 0.87 | 0.39 | 0.84 |
| 3 | 3 | 41.16 | 42.75 | 5.80 | n/a | 0.88 | 0.88 | 1.14 | 1.33 | 1.56 | 0.89 | 1.36 |
| 3 | 4 | 50.29 | 51.72 | 6.42 | n/a | 1.62 | 1.62 | 1.97 | 2.07 | 2.27 | 1.62 | 1.91 |
| 3 | 5 | 58.45 | 59.72 | 6.80 | n/a | 2.56 | 2.56 | 2.95 | 2.87 | 3.00 | 2.56 | 2.63 |
| 3 | 6 | 65.73 | 66.87 | 7.00 | n/a | 3.68 | 3.68 | 4.03 | 3.78 | 3.85 | 3.68 | 3.49 |
| 3 | 7 | 72.24 | 73.26 | 7.07 | n/a | 4.93 | 4.93 | 5.19 | 4.82 | 4.89 | 4.93 | 4.38 |
| 3 | 8 | 78.05 | 78.96 | 7.05 | n/a | 6.26 | 6.26 | 6.42 | 5.96 | 6.07 | 6.26 | 5.27 |
| 3 | 9 | 83.24 | 84.05 | 6.97 | n/a | 7.65 | 7.65 | 7.71 | 7.15 | 7.33 | 7.64 | 6.17 |
| 3 | 10 | 87.88 | 88.60 | 6.83 | n/a | 9.05 | 9.05 | 9.00 | 8.37 | 8.63 | 9.00 | 7.08 |
| 3 | 11 | 92.02 | 92.67 | 6.67 | n/a | 10.44 | 10.44 | 10.23 | 9.58 | 9.92 | 10.27 | 8.00 |
| 3 | 12 | 95.72 | 96.30 | 6.49 | n/a | 11.77 | 11.77 | 11.33 | 10.79 | 11.19 | 11.38 | 8.93 |
| 3 | 13 | 99.02 | 99.54 | 6.71 | n/a | 12.98 | 12.98 | 12.22 | 11.87 | 12.35 | 12.22 | 9.62 |
| 3 | 14 | 101.97 | 102.43 | 6.90 | $\mathrm{n} / \mathrm{a}$ | 13.98 | 13.98 | 12.96 | 12.87 | 13.37 | 12.88 | 10.26 |

Table 2.23-Time series of EBS (not expanded to BSAI) Pacific cod female spawning biomass for the years 1977-2006 as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1, 1977-2006 (note that the entries labeled "Last Year's Values" do not correspond to the values given in last year's SAFE report, because the values given in last year's SAFE report corresponded to the authors' preferred model, not the model chosen by the Plan Team and SSC). The columns labeled "L95\%CI" and "U95\%CI" represent the lower and upper bounds of the $95 \%$ confidence interval.

|  | Last Year's Values |  |  | This Year's Values |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Sp. Bio. | L95\%CI | U95\%CI | Sp. Bio. | L95\%CI | U95\%CI |
| 1977 | 32,871 | 23,372 | 42,369 | 56,590 | 39,103 | 74,077 |
| 1978 | 48,058 | 36,761 | 59,354 | 78,325 | 57,381 | 99,269 |
| 1979 | 76,760 | 60,268 | 93,252 | 114,795 | 87,281 | 142,309 |
| 1980 | 134,915 | 109,848 | 159,982 | 181,760 | 144,358 | 219,162 |
| 1981 | 243,335 | 205,705 | 280,965 | 290,795 | 239,529 | 342,061 |
| 1982 | 381,235 | 330,376 | 432,094 | 424,045 | 357,133 | 490,957 |
| 1983 | 501,700 | 441,912 | 561,488 | 544,850 | 465,244 | 624,456 |
| 1984 | 567,600 | 504,904 | 630,296 | 613,850 | 527,493 | 700,207 |
| 1985 | 577,950 | 516,926 | 638,974 | 630,500 | 542,936 | 718,064 |
| 1986 | 565,500 | 508,045 | 622,955 | 622,950 | 537,388 | 708,512 |
| 1987 | 564,550 | 510,416 | 618,684 | 619,300 | 537,036 | 701,564 |
| 1988 | 564,450 | 513,092 | 615,808 | 607,300 | 529,182 | 685,418 |
| 1989 | 543,900 | 495,219 | 592,582 | 564,850 | 491,797 | 637,903 |
| 1990 | 513,600 | 468,179 | 559,021 | 516,550 | 449,321 | 583,779 |
| 1991 | 456,835 | 415,863 | 497,807 | 454,815 | 394,277 | 515,353 |
| 1992 | 375,875 | 339,795 | 411,955 | 378,065 | 324,193 | 431,937 |
| 1993 | 337,610 | 305,129 | 370,091 | 344,165 | 295,331 | 392,999 |
| 1994 | 346,000 | 315,330 | 376,670 | 351,985 | 306,049 | 397,921 |
| 1995 | 354,910 | 325,102 | 384,718 | 360,540 | 315,910 | 405,170 |
| 1996 | 344,020 | 314,354 | 373,686 | 350,860 | 306,281 | 395,439 |
| 1997 | 333,220 | 303,174 | 363,266 | 343,040 | 297,689 | 388,391 |
| 1998 | 296,725 | 266,672 | 326,778 | 314,645 | 268,605 | 360,685 |
| 1999 | 275,280 | 245,114 | 305,446 | 308,685 | 261,600 | 355,770 |
| 2000 | 266,385 | 235,573 | 297,197 | 319,535 | 270,639 | 368,431 |
| 2001 | 268,275 | 236,733 | 299,817 | 342,440 | 291,318 | 393,562 |
| 2002 | 275,295 | 243,594 | 306,996 | 366,965 | 314,358 | 419,572 |
| 2003 | 277,895 | 246,138 | 309,652 | 376,425 | 323,431 | 429,419 |
| 2004 | 284,915 | 252,345 | 317,485 | 376,585 | 323,995 | 429,175 |
| 2005 | 283,075 | 249,153 | 316,997 | 360,260 | 308,790 | 411,730 |
| 2006 | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | 326,400 | 276,697 | 376,103 |
|  |  |  |  |  |  |  |

Table 2.24-Time series of EBS (not expanded to BSAI) Pacific cod age 0 recruitment (1000s of fish) as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1, 1977-2005 (note that the entries labeled "Last Year's Values" do not correspond to the values given in last year's SAFE report, because the values given in last year's SAFE report corresponded to the authors' preferred model, not the model chosen by the Plan Team and SSC). The columns labeled "L95\%CI" and "U95\%CI" represent the lower and upper bounds of the $95 \%$ confidence interval for each cohort.

|  | Last Year's Values |  |  |  | This Year's Values |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| Year | Recruits | L95\%CI | U95\%CI | Recruits | L95\%CI | U95\%CI |  |
| 1977 | $2,087,960$ | $1,727,781$ | $2,523,294$ | $1,611,960$ | $1,292,760$ | $2,009,960$ |  |
| 1978 | 522,535 | 312,677 | 873,249 | $1,014,290$ | 755,490 | $1,361,690$ |  |
| 1979 | $1,074,910$ | 834,512 | $1,384,544$ | 947,821 | 723,421 | $1,241,821$ |  |
| 1980 | 370,327 | 233,561 | 587,207 | 453,442 | 302,942 | 678,742 |  |
| 1981 | 482,648 | 339,877 | 685,403 | 886,610 | 704,310 | $1,116,110$ |  |
| 1982 | $1,637,790$ | $1,407,769$ | $1,905,306$ | $1,508,730$ | $1,280,230$ | $1,778,030$ |  |
| 1983 | 315,147 | 205,383 | 483,561 | 383,242 | 263,542 | 557,342 |  |
| 1984 | $1,494,730$ | $1,285,365$ | $1,738,179$ | $1,210,830$ | $1,030,230$ | $1,423,130$ |  |
| 1985 | 428,535 | 314,820 | 583,336 | 418,040 | 315,040 | 554,740 |  |
| 1986 | 286,273 | 206,672 | 396,524 | 392,177 | 299,587 | 513,377 |  |
| 1987 | 200,418 | 134,291 | 298,974 | 313,653 | 227,433 | 432,553 |  |
| 1988 | 658,175 | 544,584 | 795,467 | 906,898 | 766,798 | $1,072,598$ |  |
| 1989 | $1,224,710$ | $1,061,143$ | $1,413,498$ | $1,139,520$ | 975,220 | $1,331,520$ |  |
| 1990 | 657,983 | 532,483 | 813,062 | 658,085 | 534,385 | 810,485 |  |
| 1991 | 640,898 | 524,260 | 783,476 | 926,882 | 787,582 | $1,090,882$ |  |
| 1992 | $1,031,550$ | 898,553 | $1,184,225$ | $1,000,980$ | 858,580 | $1,166,980$ |  |
| 1993 | 280,836 | 212,685 | 370,814 | 373,064 | 285,674 | 487,164 |  |
| 1994 | 351,743 | 280,394 | 441,241 | 565,069 | 460,369 | 693,569 |  |
| 1995 | 627,883 | 531,606 | 741,596 | 985,921 | 844,021 | $1,151,721$ |  |
| 1996 | 878,950 | 767,880 | $1,006,078$ | $1,106,130$ | 960,530 | $1,273,830$ |  |
| 1997 | 411,017 | 340,031 | 496,831 | 600,909 | 500,609 | 721,309 |  |
| 1998 | 631,846 | 539,514 | 739,979 | 830,782 | 710,482 | 971,382 |  |
| 1999 | 943,613 | 820,365 | $1,085,367$ | $1,023,880$ | 890,480 | $1,177,280$ |  |
| 2000 | 693,481 | 586,035 | 820,616 | 528,671 | 442,611 | 631,471 |  |
| 2001 | 300,762 | 234,407 | 385,904 | 462,633 | 381,223 | 561,433 |  |
| 2002 | 411,992 | 323,510 | 524,669 | 429,282 | 342,522 | 537,982 |  |
| 2003 | 272,626 | 193,079 | 384,942 | 394,653 | 298,673 | 521,453 |  |
| 2004 | 435,093 | 279,269 | 677,917 | 303,430 | 193,130 | 476,630 |  |
| 2005 | $n / a$ | $n / a$ | $\mathrm{n} / \mathrm{a}$ | 675,083 | 448,783 | $1,015,383$ |  |
|  |  |  |  |  |  |  |  |
|  |  |  |  |  |  |  |  |

Table 2.25-Time series of EBS Pacific cod catch divided by age $3+$ biomass as estimated last year under the Plan Team's and SSC's preferred model and this year under Model B1, 1977-2006 (note that the entries labeled "Last Year's Values" do not correspond to the values given in last year's SAFE report, because the values given in last year's SAFE report corresponded to the authors' preferred model, not the model chosen by the Plan Team and SSC). The last entry in each column is based on partial catches for the respective year, because the year was/is still in progress at the time of the assessment.

| Year | Last Year's Values | This Year's Values |
| ---: | ---: | ---: |
| 1977 | 0.16 | 0.11 |
| 1978 | 0.18 | 0.12 |
| 1979 | 0.08 | 0.05 |
| 1980 | 0.05 | 0.04 |
| 1981 | 0.05 | 0.04 |
| 1982 | 0.04 | 0.04 |
| 1983 | 0.06 | 0.05 |
| 1984 | 0.08 | 0.07 |
| 1985 | 0.08 | 0.08 |
| 1986 | 0.08 | 0.07 |
| 1987 | 0.09 | 0.08 |
| 1988 | 0.12 | 0.12 |
| 1989 | 0.12 | 0.11 |
| 1990 | 0.13 | 0.13 |
| 1991 | 0.17 | 0.17 |
| 1992 | 0.14 | 0.14 |
| 1993 | 0.12 | 0.11 |
| 1994 | 0.15 | 0.14 |
| 1995 | 0.19 | 0.18 |
| 1996 | 0.19 | 0.18 |
| 1997 | 0.23 | 0.21 |
| 1998 | 0.17 | 0.15 |
| 1999 | 0.16 | 0.13 |
| 2000 | 0.17 | 0.13 |
| 2001 | 0.16 | 0.12 |
| 2002 | 0.17 | 0.13 |
| 2003 | 0.18 | 0.14 |
| 2004 | 0.19 | 0.16 |
| 2005 | 0.21 | 0.17 |
| 2006 | $\mathrm{n} / \mathrm{a}$ | 0.19 |

Table 2.26-Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in 2007-2019 (Scenarios 1 and 2), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 176482 | 176482 | 176482 | 176482 | 0 |
| 2008 | 130874 | 130876 | 130876 | 130879 | 2 |
| 2009 | 112823 | 112928 | 112951 | 113155 | 110 |
| 2010 | 119780 | 121578 | 121951 | 125418 | 1885 |
| 2011 | 137288 | 149589 | 152036 | 175137 | 12907 |
| 2012 | 142863 | 179072 | 184434 | 237710 | 31529 |
| 2013 | 139315 | 198970 | 202011 | 271842 | 43446 |
| 2014 | 138648 | 209000 | 209525 | 288995 | 48424 |
| 2015 | 136339 | 213074 | 212460 | 297573 | 49877 |
| 2016 | 134168 | 214000 | 212948 | 295017 | 50477 |
| 2017 | 130585 | 215360 | 212735 | 296951 | 50374 |
| 2018 | 132485 | 214179 | 212568 | 296138 | 49711 |
| 2019 | 134790 | 212231 | 212703 | 294603 | 49481 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 306790 | 306790 | 306790 | 306790 | 0 |
| 2008 | 266129 | 266133 | 266134 | 266140 | 3 |
| 2009 | 246408 | 246580 | 246616 | 246949 | 180 |
| 2010 | 247692 | 249682 | 250055 | 253788 | 2028 |
| 2011 | 259440 | 269530 | 271219 | 288794 | 9938 |
| 2012 | 264890 | 291408 | 295471 | 337235 | 24702 |
| 2013 | 263146 | 306330 | 312557 | 380843 | 39215 |
| 2014 | 262217 | 314329 | 322391 | 406002 | 47908 |
| 2015 | 260860 | 318157 | 327356 | 426511 | 51489 |
| 2016 | 258451 | 319368 | 329367 | 424301 | 52803 |
| 2017 | 256819 | 322538 | 330002 | 428346 | 52984 |
| 2018 | 257637 | 321676 | 330283 | 428222 | 52615 |
| 2019 | 259754 | 320427 | 330705 | 424271 | 52600 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.330 | 0.330 | 0.330 | 0.330 | 0.000 |
| 2008 | 0.284 | 0.284 | 0.284 | 0.284 | 0.000 |
| 2009 | 0.261 | 0.261 | 0.261 | 0.262 | 0.000 |
| 2010 | 0.263 | 0.265 | 0.265 | 0.270 | 0.002 |
| 2011 | 0.276 | 0.287 | 0.289 | 0.309 | 0.011 |
| 2012 | 0.282 | 0.312 | 0.314 | 0.345 | 0.020 |
| 2013 | 0.280 | 0.329 | 0.323 | 0.345 | 0.023 |
| 2014 | 0.279 | 0.338 | 0.325 | 0.345 | 0.023 |
| 2015 | 0.278 | 0.343 | 0.327 | 0.345 | 0.024 |
| 2016 | 0.275 | 0.344 | 0.327 | 0.345 | 0.025 |
| 2017 | 0.273 | 0.345 | 0.327 | 0.345 | 0.025 |
| 2018 | 0.274 | 0.345 | 0.327 | 0.345 | 0.025 |
| 2019 | 0.276 | 0.345 | 0.327 | 0.345 | 0.024 |

Table 2.27—Projections for BSAI Pacific cod catch ( t ), spawning biomass $(\mathrm{t})$, and fishing mortality under the assumption that $F=1 / 2$ max $F_{A B C}$ in 2007-2019 (Scenario 3), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 93127 | 93127 | 93127 | 93127 | 0 |
| 2008 | 83988 | 83989 | 83989 | 83991 | 1 |
| 2009 | 80014 | 80078 | 80092 | 80217 | 67 |
| 2010 | 86845 | 87945 | 88173 | 90292 | 1152 |
| 2011 | 99701 | 106595 | 106831 | 115299 | 5033 |
| 2012 | 105824 | 118961 | 121142 | 142404 | 12597 |
| 2013 | 105656 | 129048 | 132226 | 168277 | 20735 |
| 2014 | 106975 | 136625 | 139901 | 183197 | 25577 |
| 2015 | 106646 | 140876 | 144613 | 194583 | 27648 |
| 2016 | 106650 | 143599 | 147145 | 196444 | 28470 |
| 2017 | 105368 | 146017 | 148414 | 199035 | 28642 |
| 2018 | 107034 | 146524 | 149134 | 198258 | 28359 |
| 2019 | 109316 | 146552 | 149686 | 198024 | 28186 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 312884 | 312884 | 312884 | 312884 | 0 |
| 2008 | 300303 | 300306 | 300307 | 300314 | 4 |
| 2009 | 293991 | 294165 | 294202 | 294540 | 182 |
| 2010 | 301388 | 303412 | 303791 | 307587 | 2062 |
| 2011 | 317522 | 327934 | 329764 | 348209 | 10391 |
| 2012 | 328519 | 357557 | 362400 | 408849 | 27631 |
| 2013 | 332456 | 386405 | 393385 | 474562 | 47710 |
| 2014 | 337674 | 409392 | 418470 | 525680 | 62885 |
| 2015 | 339846 | 427673 | 436960 | 563421 | 71627 |
| 2016 | 340821 | 439588 | 449732 | 583264 | 76108 |
| 2017 | 343889 | 449275 | 458344 | 593009 | 78075 |
| 2018 | 345309 | 457062 | 464311 | 603655 | 78523 |
| 2019 | 353270 | 460440 | 468698 | 605962 | 78583 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.165 | 0.165 | 0.165 | 0.165 | 0.000 |
| 2008 | 0.158 | 0.158 | 0.158 | 0.158 | 0.000 |
| 2009 | 0.155 | 0.155 | 0.155 | 0.155 | 0.000 |
| 2010 | 0.159 | 0.160 | 0.160 | 0.162 | 0.001 |
| 2011 | 0.168 | 0.172 | 0.171 | 0.172 | 0.002 |
| 2012 | 0.172 | 0.172 | 0.172 | 0.172 | 0.000 |
| 2013 | 0.172 | 0.172 | 0.172 | 0.172 | 0.001 |
| 2014 | 0.172 | 0.172 | 0.172 | 0.172 | 0.001 |
| 2015 | 0.172 | 0.172 | 0.172 | 0.172 | 0.001 |
| 2016 | 0.172 | 0.172 | 0.172 | 0.172 | 0.002 |
| 2017 | 0.172 | 0.172 | 0.172 | 0.172 | 0.002 |
| 2018 | 0.172 | 0.172 | 0.172 | 0.172 | 0.002 |
| 2019 | 0.172 | 0.172 | 0.172 | 0.172 | 0.001 |

Table 2.28—Projections for BSAI Pacific cod catch ( t ), spawning biomass $(\mathrm{t})$, and fishing mortality under the assumption that $F=$ the 2002-2006 average in 2007-2019 (Scenario 4), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 142355 | 142355 | 142355 | 142355 | 0 |
| 2008 | 126101 | 126101 | 126101 | 126101 | 0 |
| 2009 | 117417 | 117444 | 117450 | 117505 | 29 |
| 2010 | 122044 | 122899 | 123077 | 124720 | 889 |
| 2011 | 132185 | 138739 | 139992 | 151943 | 6669 |
| 2012 | 135310 | 154971 | 158116 | 189284 | 18537 |
| 2013 | 133744 | 167377 | 171783 | 222896 | 29384 |
| 2014 | 134544 | 175892 | 180526 | 240880 | 34986 |
| 2015 | 134146 | 180743 | 185403 | 253582 | 36915 |
| 2016 | 134045 | 182597 | 187696 | 253732 | 37506 |
| 2017 | 132500 | 185365 | 188662 | 255484 | 37455 |
| 2018 | 133651 | 185381 | 189147 | 255445 | 37063 |
| 2019 | 136148 | 184942 | 189621 | 252896 | 36958 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 309351 | 309351 | 309351 | 309351 | 0 |
| 2008 | 279042 | 279045 | 279046 | 279053 | 4 |
| 2009 | 259241 | 259418 | 259456 | 259801 | 186 |
| 2010 | 257319 | 259396 | 259785 | 263681 | 2117 |
| 2011 | 267007 | 277736 | 279568 | 298358 | 10594 |
| 2012 | 272975 | 302415 | 307063 | 353114 | 27451 |
| 2013 | 273220 | 325462 | 331859 | 410662 | 45685 |
| 2014 | 275830 | 341799 | 350431 | 448735 | 57979 |
| 2015 | 274931 | 354149 | 362960 | 477048 | 63997 |
| 2016 | 275960 | 363226 | 370775 | 484384 | 66597 |
| 2017 | 276880 | 368476 | 375540 | 493276 | 67447 |
| 2018 | 278403 | 371516 | 378662 | 493772 | 67318 |
| 2019 | 282597 | 373292 | 381018 | 495003 | 67174 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2008 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2009 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2010 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2011 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2012 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2013 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2014 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2015 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2016 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2017 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2018 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |
| 2019 | 0.260 | 0.260 | 0.260 | 0.260 | 0.000 |

Table 2.29—Projections for BSAI Pacific cod catch $(\mathrm{t})$, spawning biomass ( t ), and fishing mortality under the assumption that $F=0$ in 2007-2019 (Scenario 5), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0 | 0 | 0 | 0 | 0 |
| 2008 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 0 | 0 | 0 | 0 |
| 2010 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 0 | 0 | 0 | 0 |
| 2012 | 0 | 0 | 0 | 0 | 0 |
| 2013 | 0 | 0 | 0 | 0 | 0 |
| 2014 | 0 | 0 | 0 | 0 | 0 |
| 2015 | 0 | 0 | 0 | 0 | 0 |
| 2016 | 0 | 0 | 0 | 0 | 0 |
| 2017 | 0 | 0 | 0 | 0 | 0 |
| 2018 | 0 | 0 | 0 | 0 | 0 |
| 2019 | 0 | 0 | 0 | 0 | 0 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 319108 | 319108 | 319108 | 319108 | 0 |
| 2008 | 340897 | 340900 | 340901 | 340908 | 4 |
| 2009 | 362981 | 363159 | 363197 | 363541 | 186 |
| 2010 | 393148 | 395229 | 395619 | 399520 | 2120 |
| 2011 | 430208 | 441064 | 442932 | 462007 | 10741 |
| 2012 | 464007 | 495258 | 500089 | 549150 | 29146 |
| 2013 | 490667 | 551014 | 558557 | 647290 | 53081 |
| 2014 | 514955 | 602199 | 612237 | 743683 | 75333 |
| 2015 | 533675 | 645942 | 658038 | 818593 | 92243 |
| 2016 | 547035 | 683346 | 695057 | 881167 | 103831 |
| 2017 | 561236 | 711336 | 723966 | 915900 | 111149 |
| 2018 | 573798 | 736203 | 746163 | 943696 | 115212 |
| 2019 | 590222 | 749743 | 763066 | 967657 | 117293 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2008 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2012 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2013 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2014 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2015 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2017 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2018 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2019 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 2.30—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=F_{\text {OFL }}$ in 2007-2019 (Scenario 6), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 206861 | 206861 | 206861 | 206861 | 0 |
| 2008 | 142693 | 142695 | 142695 | 142699 | 2 |
| 2009 | 119744 | 119862 | 119887 | 120116 | 124 |
| 2010 | 127325 | 129345 | 129765 | 133664 | 2121 |
| 2011 | 146585 | 160382 | 163159 | 189113 | 14640 |
| 2012 | 151534 | 191551 | 198795 | 269243 | 37767 |
| 2013 | 146437 | 210632 | 218508 | 306107 | 52538 |
| 2014 | 145624 | 219162 | 225469 | 320973 | 57253 |
| 2015 | 142783 | 221312 | 227147 | 326327 | 58156 |
| 2016 | 139461 | 220826 | 226482 | 324021 | 58583 |
| 2017 | 135291 | 219871 | 225559 | 322843 | 58321 |
| 2018 | 137372 | 220287 | 224847 | 321089 | 57609 |
| 2019 | 139397 | 219695 | 224882 | 321543 | 57581 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 304427 | 304427 | 304427 | 304427 | 0 |
| 2008 | 254137 | 254140 | 254141 | 254148 | 3 |
| 2009 | 231568 | 231739 | 231776 | 232107 | 179 |
| 2010 | 232263 | 234243 | 234614 | 238329 | 2018 |
| 2011 | 243572 | 253572 | 255244 | 272651 | 9843 |
| 2012 | 247954 | 273966 | 277888 | 317987 | 24020 |
| 2013 | 245253 | 286928 | 292226 | 355132 | 36622 |
| 2014 | 244018 | 292270 | 298808 | 373316 | 42774 |
| 2015 | 241907 | 294603 | 301037 | 383590 | 44536 |
| 2016 | 239695 | 294503 | 301151 | 382626 | 44926 |
| 2017 | 237445 | 294710 | 300643 | 383097 | 44692 |
| 2018 | 238495 | 294042 | 300312 | 383892 | 44187 |
| 2019 | 239929 | 293153 | 300484 | 378550 | 44136 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.395 | 0.395 | 0.395 | 0.395 | 0.000 |
| 2008 | 0.326 | 0.326 | 0.326 | 0.326 | 0.000 |
| 2009 | 0.295 | 0.295 | 0.295 | 0.296 | 0.000 |
| 2010 | 0.296 | 0.299 | 0.299 | 0.304 | 0.003 |
| 2011 | 0.311 | 0.325 | 0.327 | 0.351 | 0.013 |
| 2012 | 0.317 | 0.353 | 0.357 | 0.413 | 0.028 |
| 2013 | 0.314 | 0.371 | 0.370 | 0.416 | 0.034 |
| 2014 | 0.312 | 0.378 | 0.374 | 0.416 | 0.036 |
| 2015 | 0.309 | 0.381 | 0.376 | 0.416 | 0.037 |
| 2016 | 0.306 | 0.381 | 0.375 | 0.416 | 0.038 |
| 2017 | 0.303 | 0.381 | 0.375 | 0.416 | 0.039 |
| 2018 | 0.304 | 0.380 | 0.375 | 0.416 | 0.038 |
| 2019 | 0.306 | 0.379 | 0.375 | 0.416 | 0.038 |

Table 2.31—Projections for BSAI Pacific cod catch ( t ), spawning biomass ( t ), and fishing mortality under the assumption that $F=\max F_{A B C}$ in each year 2007-2008 and $F=F_{O F L}$ thereafter (Scenario 7), with random variability in future recruitment.

| Catch Projections |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 176482 | 176482 | 176482 | 176482 | 0 |
| 2008 | 130874 | 130876 | 130876 | 130879 | 2 |
| 2009 | 133192 | 133315 | 133342 | 133581 | 129 |
| 2010 | 134346 | 136413 | 136843 | 140834 | 2170 |
| 2011 | 149805 | 163733 | 166529 | 192716 | 14741 |
| 2012 | 152493 | 192578 | 199770 | 270355 | 37664 |
| 2013 | 146466 | 210589 | 218429 | 305669 | 52431 |
| 2014 | 145442 | 218884 | 225206 | 320735 | 57226 |
| 2015 | 142603 | 221080 | 226946 | 326118 | 58158 |
| 2016 | 139336 | 220761 | 226365 | 323904 | 58588 |
| 2017 | 135227 | 219828 | 225499 | 322806 | 58324 |
| 2018 | 137342 | 220253 | 224818 | 321060 | 57610 |
| 2019 | 139379 | 219679 | 224868 | 321535 | 57581 |
| Spawning Biomass Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 306790 | 306790 | 306790 | 306790 | 0 |
| 2008 | 266129 | 266133 | 266134 | 266140 | 3 |
| 2009 | 244994 | 245164 | 245201 | 245532 | 179 |
| 2010 | 239659 | 241635 | 242006 | 245713 | 2015 |
| 2011 | 247245 | 257229 | 258899 | 276275 | 9828 |
| 2012 | 249480 | 275451 | 279372 | 319381 | 23995 |
| 2013 | 245780 | 287365 | 292703 | 355548 | 36628 |
| 2014 | 244160 | 292402 | 298934 | 373476 | 42805 |
| 2015 | 241926 | 294614 | 301067 | 383800 | 44565 |
| 2016 | 239684 | 294507 | 301161 | 382692 | 44945 |
| 2017 | 237432 | 294708 | 300649 | 383100 | 44704 |
| 2018 | 238486 | 294043 | 300316 | 383899 | 44193 |
| 2019 | 239926 | 293155 | 300485 | 378553 | 44139 |
| Fishing Mortality Projections |  |  |  |  |  |
| Year | L90\%CI | Median | Mean | U90\%CI | Std. Dev. |
| 2007 | 0.330 | 0.330 | 0.330 | 0.330 | 0.000 |
| 2008 | 0.284 | 0.284 | 0.284 | 0.284 | 0.000 |
| 2009 | 0.313 | 0.314 | 0.314 | 0.314 | 0.000 |
| 2010 | 0.306 | 0.309 | 0.309 | 0.314 | 0.003 |
| 2011 | 0.316 | 0.330 | 0.332 | 0.356 | 0.013 |
| 2012 | 0.319 | 0.355 | 0.359 | 0.415 | 0.027 |
| 2013 | 0.314 | 0.371 | 0.371 | 0.416 | 0.034 |
| 2014 | 0.312 | 0.378 | 0.375 | 0.416 | 0.036 |
| 2015 | 0.309 | 0.381 | 0.376 | 0.416 | 0.037 |
| 2016 | 0.306 | 0.381 | 0.375 | 0.416 | 0.038 |
| 2017 | 0.303 | 0.381 | 0.375 | 0.416 | 0.039 |
| 2018 | 0.304 | 0.380 | 0.375 | 0.416 | 0.038 |
| 2019 | 0.306 | 0.379 | 0.375 | 0.416 | 0.038 |

Table 2.32a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod trawl fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS Pacific cod trawl fishery |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sculpin | 1508 | 1365 | 893 | 1280 | 749 | 925 | 0.22 | 0.26 | 0.20 | 0.23 | 0.12 | 0.12 |  |  |  |  |  |  |  |  |  |  |  |  |
| Skates | 678 | 676 | 946 | 981 | 583 | 1303 | 0.04 | 0.04 | 0.07 | 0.06 | 0.03 | 0.05 |  |  |  |  |  |  |  |  |  |  |  |  |
| Shark | 0 | 0 | 0 | 9 | 2 | 3 | 0.00 | 0.00 | 0.00 | 0.15 | 0.09 | 0.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.04 | 0.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sleepershk | 8 | 33 | 4 | 0 | 12 | 10 | 0.03 | 0.10 | 0.01 | 0.00 | 0.02 | 0.01 |  |  |  |  |  |  |  |  |  |  |  |  |
| Octopus | 29 | 19 | 17 | 68 | 17 | 30 | 0.14 | 0.13 | 0.13 | 0.19 | 0.09 | 0.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| Squid | 7 | 1 | 0 | 2 | 4 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Smelts | 1 | 0 | 1 | 0 | 0 | 0 | 0.03 | 0.00 | 0.03 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.00 | 0.00 | 0.00 | 0.71 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.03 | 0.00 | 0.00 | 0.01 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sandfish | 0 | 0 | 3 | 0 | 0 | 1 | 0.27 | 0.08 | 0.91 | 0.02 | 0.05 | 0.36 |  |  |  |  |  |  |  |  |  |  |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.90 | 0.01 |  |  |  |  |  |  |  |  |  |  |  |  |
| Grenadier | 1 | 6 | 0 | 3 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Otherfish | 231 | 232 | 195 | 302 | 220 | 157 | 0.16 | 0.21 | 0.20 | 0.24 | 0.18 | 0.14 |  |  |  |  |  |  |  |  |  |  |  |  |
| Crabs | 10 | 6 | 5 | 8 | 3 | 6 | 0.03 | 0.03 | 0.05 | 0.06 | 0.02 | 0.04 |  |  |  |  |  |  |  |  |  |  |  |  |
| Starfish | 133 | 63 | 83 | 109 | 57 | 98 | 0.02 | 0.02 | 0.03 | 0.03 | 0.01 | 0.02 |  |  |  |  |  |  |  |  |  |  |  |  |
| Jellyfish | 948 | 213 | 416 | 413 | 112 | 93 | 0.11 | 0.03 | 0.06 | 0.04 | 0.03 | 0.05 |  |  |  |  |  |  |  |  |  |  |  |  |
| Invertunid | 1 | 9 | 3 | 11 | 1 | 51 | 0.00 | 0.02 | 0.02 | 0.01 | 0.00 | 0.05 |  |  |  |  |  |  |  |  |  |  |  |  |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.10 | 0.09 | 0.01 | 0.06 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Sponge | 73 | 34 | 39 | 28 | 9 | 13 | 0.23 | 0.09 | 0.22 | 0.30 | 0.05 | 0.08 |  |  |  |  |  |  |  |  |  |  |  |  |
| Anemone | 14 | 5 | 18 | 10 | 6 | 9 | 0.08 | 0.05 | 0.11 | 0.03 | 0.03 | 0.03 |  |  |  |  |  |  |  |  |  |  |  |  |
| Tunicate | 6 | 10 | 0 | 67 | 5 | 1 | 0.00 | 0.01 | 0.00 | 0.06 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Benthinv | 25 | 18 | 11 | 23 | 6 | 12 | 0.04 | 0.03 | 0.05 | 0.06 | 0.01 | 0.03 |  |  |  |  |  |  |  |  |  |  |  |  |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| echinoderm | 13 | 4 | 13 | 13 | 20 | 14 | 0.31 | 0.20 | 0.54 | 0.33 | 0.50 | 0.46 |  |  |  |  |  |  |  |  |  |  |  |  |
| Coral | 0 | 0 | 0 | 4 | 0 | 0 | 0.02 | 0.01 | 0.04 | 0.37 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.07 | 0.03 | 0.01 | 0.00 | 0.01 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  |  |  |  |

Table 2.32b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod trawl fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( $t$ ) of each species group taken as bycatch in the EBS Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 14 | 4 | 9 | 0.01 | 0.00 | 0.01 |
| Birds | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 |
| Bivalves | 1 | 10 | 0 | 0.05 | 0.52 | 0.03 |
| Brittle star unidentified | 1 | 1 | 0 | 0.02 | 0.03 | 0.00 |
| Capelin |  | 0 |  |  | 0.02 |  |
| Corals Bryozoans | 1 | 1 | 0 | 0.28 | 0.25 | 0.06 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 62 | 27 | 1 | 0.27 | 0.30 | 0.02 |
| Eulachon |  | 0 | 0 |  | 0.00 | 0.00 |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 4 | 2 | 1 | 0.43 | 0.40 | 0.23 |
| Grenadier | 14 | 9 | 0 | 0.01 | 0.00 | 0.00 |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 5 | 3 | 1 | 0.04 | 0.05 | 0.01 |
| Invertebrate unidentified | 5 | 4 | 0 | 0.01 | 0.01 | 0.00 |
| Lanternfishes (myctophidae) |  | 0 |  |  | 0.07 |  |
| Large Sculpins | 547 | 1422 | 897 | 0.39 | 0.32 | 0.22 |
| Misc crabs | 7 | 3 | 2 | 0.13 | 0.09 | 0.07 |
| Misc crustaceans | 0 | 0 | 0 | 0.24 | 0.20 | 0.07 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 174 | 152 | 149 | 0.35 | 0.30 | 0.31 |
| Misc inverts (worms etc) | 0 | 0 | 0 | 0.07 | 0.02 | 0.00 |
| Octopus | 14 | 44 | 12 | 0.10 | 0.12 | 0.05 |
| Other osmerids | 0 | 0 |  | 0.01 | 0.09 |  |
| Other Sculpins | 854 | 95 | 58 | 0.22 | 0.18 | 0.12 |
| Pacific Sand lance | 0 | 0 | 0 | 0.45 | 0.40 | 0.59 |
| Pandalid shrimp | 0 | 0 | 0 | 0.15 | 0.18 | 0.01 |
| Polychaete unidentified |  | 0 | 0 |  | 0.01 | 0.08 |
| Scypho jellies | 727 | 699 | 391 | 0.11 | 0.10 | 0.06 |
| Sea anemone unidentified | 14 | 16 | 12 | 0.10 | 0.09 | 0.12 |
| Sea pens whips | 0 | 1 | 0 | 0.01 | 0.05 | 0.01 |
| Sea star | 118 | 91 | 81 | 0.03 | 0.03 | 0.03 |
| Shark | 10 | 29 | 11 | 0.03 | 0.08 | 0.05 |
| Skate | 1010 | 1355 | 570 | 0.06 | 0.07 | 0.03 |
| Snails | 14 | 13 | 3 | 0.07 | 0.05 | 0.02 |
| Sponge unidentified | 3 | 7 | 3 | 0.01 | 0.08 | 0.04 |
| Squid | 5 | 4 | 1 | 0.00 | 0.00 | 0.00 |
| Stichaeidae | 0 | 0 | 0 | 0.12 | 0.07 | 0.14 |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 11 | 10 | 12 | 0.36 | 0.43 | 0.48 |

Table 2.33a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod longline fishery, 1997-2002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod longline fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS P. cod longline fishery |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 706 | 931 | 821 | 801 | 1142 | 1383 | 0.11 | 0.18 | 0.18 | 0.14 | 0.19 | 0.18 |
| Skates | 12961 | 12808 | 9178 | 11578 | 11932 | 17507 | 0.77 | 0.70 | 0.69 | 0.68 | 0.66 | 0.66 |
| Shark | 27 | 48 | 18 | 47 | 17 | 22 | 0.50 | 0.40 | 0.11 | 0.78 | 0.70 | 0.48 |
| Salmonshk | 0 | 1 | 1 | 0 | 1 | 10 | 0.00 | 0.05 | 0.04 | 0.01 | 0.05 | 0.22 |
| Dogfish | 4 | 5 | 5 | 8 | 11 | 8 | 1.00 | 0.90 | 0.99 | 0.98 | 0.83 | 0.92 |
| Sleepershk | 67 | 114 | 99 | 114 | 240 | 250 | 0.24 | 0.34 | 0.35 | 0.33 | 0.37 | 0.30 |
| Octopus | 15 | 15 | 13 | 29 | 15 | 76 | 0.07 | 0.10 | 0.10 | 0.08 | 0.08 | 0.19 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.60 | 0.00 | 0.80 | 0.00 | 0.00 |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 | 0.56 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 437 | 604 | 356 | 364 | 162 | 336 | 0.15 | 0.12 | 0.08 | 0.09 | 0.07 | 0.06 |
| Otherfish | 43 | 27 | 38 | 38 | 71 | 122 | 0.03 | 0.03 | 0.04 | 0.03 | 0.06 | 0.11 |
| Crabs | 1 | 0 | 0 | 1 | 1 | 3 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 |
| Starfish | 136 | 141 | 250 | 132 | 319 | 384 | 0.02 | 0.04 | 0.08 | 0.04 | 0.08 | 0.08 |
| Jelly fish | 5 | 7 | 24 | 2 | 2 | 5 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 10 | 12 | 1 | 6 | 10 | 11 | 0.01 | 0.02 | 0.01 | 0.01 | 0.01 | 0.01 |
| seapen/whip | 2 | 2 | 4 | 3 | 6 | 41 | 0.83 | 0.79 | 0.87 | 0.63 | 0.79 | 0.95 |
| Sponge | 1 | 1 | 2 | 1 | 0 | 5 | 0.00 | 0.00 | 0.01 | 0.01 | 0.00 | 0.03 |
| Anemone | 76 | 58 | 123 | 200 | 115 | 195 | 0.42 | 0.51 | 0.73 | 0.58 | 0.55 | 0.59 |
| Tunicate | 1 | 1 | 0 | 2 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 7 | 5 | 10 | 11 | 12 | 12 | 0.01 | 0.01 | 0.04 | 0.03 | 0.02 | 0.03 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 1.00 | 0.00 |
| echinoderm | 1 | , | 3 | 0 | 0 | 0 | 0.02 | 0.00 | 0.11 | 0.00 | 0.00 | 0.01 |
| Coral | 1 | 0 | 0 | 3 | 1 | 2 | 0.07 | 0.02 | 0.04 | 0.30 | 0.01 | 0.03 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 26 | 33 | 17 | 24 | 13 | 13 | 0.98 | 0.86 | 0.81 | 0.97 | 0.88 | 0.96 |

Table 2.33b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod hook-and-line (including jigs) fishery, 2003-2005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod hook-and-line fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 19972002.

|  | Byatch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Birds | 6 | 6 | 2 | 0.93 | 0.93 | 0.44 |
| Bivalves | 4 | 6 | 5 | 0.36 | 0.33 | 0.68 |
| Brittle star unidentified | 0 | 0 | 0 | 0.01 | 0.00 | 0.01 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 1 | 1 | 1 | 0.23 | 0.23 | 0.30 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 4 | 8 | 16 | 0.02 | 0.09 | 0.25 |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier | 1 | 16 | 91 | 0.01 | 0.08 | 0.08 |
| Greenlings | 3 | 1 | 1 | 0.28 | 0.23 | 0.20 |
| Grenadier | 221 | 202 | 158 | 0.08 | 0.10 | 0.12 |
| Gunnels |  | 0 | 0 |  | 1.00 | 1.00 |
| Hermit crab unidentified | 1 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Invertebrate unidentified | 14 | 2 | 3 | 0.02 | 0.00 | 0.01 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 194 | 1087 | 865 | 0.14 | 0.24 | 0.21 |
| Misc crabs | 1 | 1 | 9 | 0.01 | 0.02 | 0.24 |
| Misc crustaceans | 0 | 0 | 0 | 0.02 | 0.00 | 0.43 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 44 | 58 | 26 | 0.09 | 0.12 | 0.05 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.00 | 0.01 |
| Octopus | 41 | 37 | 20 | 0.30 | 0.10 | 0.08 |
| Other osmerids |  |  | 0 |  |  | 0.00 |
| Other Sculpins | 993 | 234 | 163 | 0.25 | 0.44 | 0.33 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified | 0 | 0 | 0 | 0.13 | 0.01 | 0.64 |
| Scypho jellies | 16 | 4 | 1 | 0.00 | 0.00 | 0.00 |
| Sea anemone unidentified | 79 | 94 | 69 | 0.58 | 0.53 | 0.69 |
| Sea pens whips | 6 | 10 | 19 | 0.86 | 0.84 | 0.88 |
| Sea star | 288 | 288 | 202 | 0.07 | 0.10 | 0.08 |
| Shark | 140 | 146 | 128 | 0.50 | 0.42 | 0.55 |
| Skate | 13519 | 13863 | 13219 | 0.74 | 0.75 | 0.78 |
| Snails | 5 | 6 | 6 | 0.03 | 0.02 | 0.05 |
| Sponge unidentified | 3 | 1 | 2 | 0.01 | 0.01 | 0.02 |
| Squid | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Stichaeidae | 0 |  |  | 0.05 |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |

Table 2.34a-Bycatch of nontarget and "other" species taken in the EBS Pacific cod pot fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year.

|  | Bycatch in EBS Pacific cod pot fishery |  |  |  |  |  | Proportion of total EBS catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 351 | 267 | 438 | 494 | 315 | 384 | 0.05 | 0.05 | 0.10 | 0.09 | 0.05 | 0.05 |
| Skates | 1 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Octopus | 79 | 95 | 80 | 199 | 140 | 254 | 0.38 | 0.65 | 0.64 | 0.56 | 0.75 | 0.65 |
| Squid | 0 | 0 | 0 | 0 | 1 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  | 0.00 | 0.00 | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 27 | 44 | 32 | 12 | 48 | 23 | 0.02 | 0.04 | 0.03 | 0.01 | 0.04 | 0.02 |
| Crabs | 1 | 1 | 4 | 2 | 1 | 2 | 0.00 | 0.00 | 0.04 | 0.01 | 0.01 | 0.01 |
| Starfish | 64 | 14 | 15 | 35 | 31 | 11 | 0.01 | 0.00 | 0.01 | 0.01 | 0.01 | 0.00 |
| Jellyfish | 11 | 1 | 16 | 0 | 6 | 2 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sponge | 0 | 0 | 0 | 0 | 0 | 1 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Anemone | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 8 | 3 | 4 | 11 | 4 | 9 | 0.01 | 0.01 | 0.02 | 0.03 | 0.01 | 0.02 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| echinoderm | 1 | 0 | 0 | 2 | 1 | 0 | 0.02 | 0.02 | 0.02 | 0.04 | 0.02 | 0.01 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |

Table 2.34b-Bycatch of nontarget and "other" species taken in the EBS Pacific cod pot fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the EBS Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total EBS catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the EBS during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Byatch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0.01 | 0.00 | 0.01 |
| Bivalves | 0 | 0 | 0 | 0.01 | 0.02 | 0.01 |
| Brittle star unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 0 |  | 0 | 0.01 |  | 0.01 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 |  |  | 0.00 |  |  |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 0 | 0 | 0.06 | 0.07 | 0.14 |
| Grenadier |  |  |  |  |  |  |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Invertebrate unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 122 | 191 | 109 | 0.09 | 0.04 | 0.03 |
| Misc crabs | 0 | 1 | 1 | 0.01 | 0.02 | 0.04 |
| Misc crustaceans | 0 | 0 |  | 0.00 | 0.01 |  |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 30 | 13 | 14 | 0.06 | 0.03 | 0.03 |
| Misc inverts (worms etc) |  |  |  |  |  |  |
| Octopus | 49 | 57 | 187 | 0.35 | 0.15 | 0.76 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 133 | 13 | 2 | 0.03 | 0.03 | 0.00 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified |  |  |  |  |  |  |
| Scypho jellies | 2 | 1 | 3 | 0.00 | 0.00 | 0.00 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Sea pens whips | 0 |  |  | 0.00 |  |  |
| Sea star | 41 | 30 | 27 | 0.01 | 0.01 | 0.01 |
| Shark |  |  |  |  |  |  |
| Skate | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Snails | 7 | 1 | 2 | 0.04 | 0.00 | 0.02 |
| Sponge unidentified | 1 | 1 | 0 | 0.00 | 0.01 | 0.00 |
| Squid |  |  | 1 |  |  | 0.00 |
| Stichaeidae |  |  |  |  |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 1 | 1 | 0 | 0.04 | 0.06 | 0.01 |

Table 2.35a-Bycatch of nontarget and "other" species taken in the AI Pacific cod trawl fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

|  | Bycatch in AI Pacific cod trawl fishery |  |  |  |  |  | Proportion of total AI catch |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 107 | 146 | 131 | 257 | 102 | 131 | 0.14 | 0.14 | 0.14 | 0.18 | 0.06 | 0.12 |
| Skates | 37 | 95 | 38 | 72 | 49 | 97 | 0.04 | 0.08 | 0.05 | 0.04 | 0.02 | 0.14 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.03 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 4 | 0 | 0 | 0.00 | 0.00 | 0.00 | 1.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.04 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.01 | 0.01 |
| Octopus | 2 | 2 | 9 | 2 | 1 | 9 | 0.06 | 0.05 | 0.04 | 0.03 | 0.03 | 0.38 |
| Squid | 1 | 0 | 0 | 1 | 2 | 4 | 0.01 | 0.01 | 0.01 | 0.07 | 0.30 | 0.25 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.95 | 0.00 | 1.00 | 1.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | . 00 |  | 1.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 9 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 6 | 38 | 29 | 25 | 26 | 15 | 0.04 | 0.14 | 0.09 | 0.12 | 0.11 | 0.07 |
| Crabs | 1 | 1 | 0 | 0 | 1 | 2 | 0.13 | 0.44 | 0.27 | 0.22 | 0.42 | 0.88 |
| Starfish | 2 | 3 | 5 | 5 | 5 | 5 | 0.12 | 0.15 | 0.29 | 0.20 | 0.17 | 0.46 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.17 | 0.00 | 0.99 | 0.01 | 0.44 |
| Invertunid | 0 | 2 | 3 | 6 | 2 | 0 | 0.00 | 0.03 | 0.34 | 0.40 | 0.36 | 0.02 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.85 | 0.23 | 0.54 | 0.33 | 0.08 | 0.16 |
| Sponge | 4 | 52 | 15 | 15 | 13 | 28 | 0.02 | 0.47 | 0.10 | 0.21 | 0.18 | 0.16 |
| Anemone | 0 | 0 | 1 | 0 | 0 | 0 | 0.09 | 0.08 | 0.41 | 0.17 | 0.05 | 0.17 |
| Tunicate | 0 | 0 | 0 | 0 | 1 | 0 | 0.63 | 0.75 | 0.08 | 0.58 | 0.40 | 0.07 |
| Benthinv | 4 | 3 | 1 | 2 | 3 | 6 | 0.90 | 0.68 | 0.16 | 0.73 | 0.76 | 0.92 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 1 | 1 | 1 | 1 | 2 | 0.16 | 0.26 | 0.23 | 0.35 | 0.44 | 0.75 |
| Coral | 2 | 8 | 2 | 8 | 3 | 11 | 0.07 | 0.48 | 0.03 | 0.24 | 0.15 | 0.52 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.05 | 0.00 | 0.11 | 0.19 | 0.10 |
| Birds | 0 | 1 | 0 | 0 | 0 | 0 | 0.02 | 0.11 | 0.02 | 0.04 | 0.01 | 0.16 |

Table 2.35b-Bycatch of nontarget and "other" species taken in the AI Pacific cod trawl fishery, 20032005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod trawl fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.05 | 0.16 | 0.37 |
| Birds | 0 | 0 | 0 | 0.21 | 0.01 | 0.38 |
| Bivalves | 15 | 1 | 0 | 0.99 | 0.92 | 0.81 |
| Brittle star unidentified |  | 0 | 0 |  | 0.05 | 0.01 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 24 | 11 | 12 | 0.40 | 0.35 | 0.24 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 | 1 | 0 | 0.08 | 0.51 | 0.00 |
| Eulachon |  |  | 0 |  |  | 0.68 |
| Giant Grenadier |  |  |  |  |  |  |
| Greenlings | 1 | 0 | 0 | 0.66 | 0.05 | 0.01 |
| Grenadier |  | 4 | 0 |  | 0.01 | 0.00 |
| Gunnels |  |  |  |  |  |  |
| Hermit crab unidentified | 0 | 0 | 0 | 0.80 | 0.98 | 0.09 |
| Invertebrate unidentified | 0 | 0 | 0 | 0.09 | 0.00 | 0.02 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 78 | 159 | 88 | 0.37 | 0.23 | 0.18 |
| Misc crabs | 1 | 1 | 0 | 0.73 | 0.59 | 0.52 |
| Misc crustaceans | 0 | 0 | 0 | 0.99 | 0.29 | 0.98 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 28 | 15 | 19 | 0.23 | 0.10 | 0.12 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.29 | 1.00 |
| Octopus | 6 | 5 | 3 | 0.36 | 0.28 | 0.40 |
| Other osmerids |  |  |  |  |  |  |
| Other Sculpins | 122 | 1 | 3 | 0.31 | 0.01 | 0.04 |
| Pacific Sand lance | 0 |  | 0 | 1.00 |  | 1.00 |
| Pandalid shrimp | 0 | 0 | 0 | 0.06 | 0.01 | 0.03 |
| Polychaete unidentified |  | 0 | 0 |  | 0.13 | 0.97 |
| Scypho jellies | 0 | 0 | 1 | 0.17 | 0.49 | 0.44 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.61 | 0.31 | 0.32 |
| Sea pens whips | 0 | 0 | 0 | 0.34 | 0.91 | 0.42 |
| Sea star | 5 | 3 | 2 | 0.49 | 0.27 | 0.17 |
| Shark | 0 | 2 | 2 | 0.01 | 0.43 | 0.10 |
| Skate | 72 | 76 | 65 | 0.13 | 0.09 | 0.11 |
| Snails | 1 | 1 | 0 | 0.52 | 0.50 | 0.21 |
| Sponge unidentified | 24 | 18 | 22 | 0.30 | 0.13 | 0.28 |
| Squid | 3 | 2 | 1 | 0.10 | 0.11 | 0.07 |
| Stichaeidae |  |  | 0 |  |  | 0.00 |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 1 | 1 | 0 | 0.40 | 0.43 | 0.15 |

Table 2.36a-Bycatch of nontarget and "other" species taken in the AI Pacific cod longline fishery, 19972002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod longline fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

| Species group | Bycatch in AI Pacific cod longline fishery |  |  |  |  |  |  | Proportion of total AI catch |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| Sculpin | 334 | 597 | 356 | 662 | 1004 | 214 | 0.43 | 0.55 | 0.37 | 0.47 | 0.63 | 0.19 |
| Skates | 338 | 727 | 473 | 1397 | 2184 | 246 | 0.39 | 0.64 | 0.59 | 0.77 | 0.87 | 0.35 |
| Shark | 0 | 1 | 0 | 0 | 0 | 0 | 0.78 | 0.04 | 0.05 | 0.03 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 1 | 0 | 0.96 | 0.55 | 0.84 | 0.85 | 0.31 | 0.54 |
| Sleepershk | 0 | 0 | 1 | 0 | 1 | 2 | 0.00 | 0.00 | 0.02 | 0.00 | 0.03 | 0.49 |
| Octopus | 10 | 21 | 9 | 13 | 21 | 8 | 0.27 | 0.47 | 0.05 | 0.20 | 0.51 | 0.32 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0.00 |  | . 00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | . 00 | 0.00 |
| Grenadier | 397 | 83 | 215 | 151 | 6 | 88 | 0.14 | 0.05 | 0.07 | 0.05 | 0.00 | 0.03 |
| Otherfish | 2 | 5 | 2 | 6 | 10 | 3 | 0.02 | 0.02 | 0.01 | 0.03 | 0.04 | 0.01 |
| Crabs | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.01 | 0.01 | 0.04 | 0.00 |
| Starfish | 3 | 7 | 4 | 13 | 16 | 3 | 0.22 | 0.41 | 0.28 | 0.51 | 0.59 | 0.25 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Invertunid | 0 | 1 | 0 | 1 | 0 | 0 | 0.00 | 0.01 | 0.02 | 0.06 | 0.08 | 0.02 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.21 | 0.44 | 0.54 | 0.92 | 0.56 |
| Sponge | 0 | 4 | 3 | 11 | 4 | 1 | 0.00 | 0.04 | 0.02 | 0.15 | 0.06 | 0.00 |
| Anemone | 0 | 0 | 1 | 1 | 0 | 1 | 0.34 | 0.57 | 0.32 | 0.59 | 0.47 | 0.69 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 | 0.24 | 0.00 | 0.00 |
| Benthinv | 0 | 0 | 0 | 0 | 0 | 0 | 0.02 | 0.00 | 0.02 | 0.06 | 0.04 | 0.03 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 0 | 0 | 0 | 0 | 0 | 0.10 | 0.04 | 0.00 | 0.09 | 0.04 | 0.02 |
| Coral | 0 | 1 | 2 | 6 | 3 | 1 | 0.02 | 0.03 | 0.04 | 0.17 | 0.16 | 0.03 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.09 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 |
| Birds | 2 | 2 | 2 | 2 | 1 | 0 | 0.75 | 0.45 | 0.55 | 0.66 | 0.48 | 0.16 |

Table 2.36b-Bycatch of nontarget and "other" species taken in the AI Pacific cod hook-and-line (including jigs) fishery, 2003-2005. The first part of the table ("Bycatch") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod hook-and-line fishery, broken down by year. The second part of the table ("Proportion of total") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year. Note that the list of nontarget species groups used for 2003-2005 differs from that used for 1997-2002.

|  | Catch (t) |  |  | Proportion of total |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Species group | 2003 | 2004 | 2005 | 2003 | 2004 | 2005 |
| Benthic urochordata | 0 | 0 | 0 | 0.09 | 0.00 | 0.01 |
| Birds | 0 | 0 | 0 | 0.03 | 0.21 | 0.29 |
| Bivalves | 0 | 0 | 0 | 0.00 | 0.02 | 0.18 |
| Brittle star unidentified | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Capelin |  |  |  |  |  |  |
| Corals Bryozoans | 1 | 1 | 0 | 0.01 | 0.05 | 0.01 |
| Deep sea smelts (bathylagidae) |  |  |  |  |  |  |
| Eelpouts | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Eulachon |  |  |  |  |  |  |
| Giant Grenadier | 0 | 0 | 0 | 0.30 | 0.00 | 0.00 |
| Greenlings | 0 | 0 | 0 | 0.08 | 0.16 | 0.02 |
| Grenadier | 46 | 8 | 0 | 0.01 | 0.01 | 0.00 |
| Gunnels |  |  | 0 |  |  | 0.00 |
| Hermit crab unidentified | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Invertebrate unidentified | 0 | 1 | 0 | 0.00 | 0.12 | 0.03 |
| Lanternfishes (myctophidae) |  |  |  |  |  |  |
| Large Sculpins | 28 | 133 | 91 | 0.14 | 0.19 | 0.18 |
| Misc crabs | 0 | 0 | 0 | 0.00 | 0.01 | 0.01 |
| Misc crustaceans | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 |
| Misc deep fish |  |  |  |  |  |  |
| Misc fish | 1 | 3 | 1 | 0.01 | 0.02 | 0.00 |
| Misc inverts (worms etc) |  | 0 | 0 |  | 0.00 | 0.00 |
| Octopus | 8 | 8 | 4 | 0.54 | 0.49 | 0.55 |
| Other osmerids |  |  | 0 |  |  | 0.00 |
| Other Sculpins | 31 | 63 | 1 | 0.08 | 0.41 | 0.01 |
| Pacific Sand lance |  |  |  |  |  |  |
| Pandalid shrimp |  |  |  |  |  |  |
| Polychaete unidentified | 0 | 0 | 0 | 1.00 | 0.00 | 0.03 |
| Scypho jellies | 0 | 0 | 0 | 0.01 | 0.00 | 0.00 |
| Sea anemone unidentified | 0 | 0 | 0 | 0.24 | 0.23 | 0.58 |
| Sea pens whips | 0 | 0 | 0 | 0.46 | 0.09 | 0.15 |
| Sea star | 1 | 6 | 3 | 0.10 | 0.47 | 0.25 |
| Shark | 0 | 0 | 0 | 0.01 | 0.08 | 0.02 |
| Skate | 105 | 402 | 245 | 0.20 | 0.48 | 0.43 |
| Snails | 0 | 0 | 0 | 0.01 | 0.03 | 0.05 |
| Sponge unidentified | 2 | 5 | 2 | 0.02 | 0.04 | 0.03 |
| Squid |  | 0 |  |  | 0.00 |  |
| Stichaeidae | 0 |  |  | 0.00 |  |  |
| Surf smelt |  |  |  |  |  |  |
| Urchins dollars cucumbers | 0 | 0 | 0 | 0.02 | 0.11 | 0.01 |

Table 2.37-Bycatch of nontarget and "other" species taken in the AI Pacific cod pot fishery, 1997-2002. The first part of the table ("Bycatch in...") shows the amount ( t ) of each species group taken as bycatch in the AI Pacific cod pot fishery, broken down by year. The second part of the table ("Proportion of...") shows the same quantity expressed relative to the total AI catch (taken in all target categories with all gears) of that species group in that year. An empty cell in the second part of the table indicates that no catch of that group was observed in the AI during that year.

Bycatch in AI Pacific cod pot fishery

| Species group | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Sculpin | 7 | 12 | 221 | 211 | 42 | 0 | 0.01 | 0.01 | 0.23 | 0.15 | 0.03 | 0.00 |
| Skates | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shark | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Salmonshk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |  |
| Dogfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Sleepershk | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Octopus | 24 | 18 | 182 | 47 | 17 | 0 | 0.62 | 0.40 | 0.90 | 0.75 | 0.41 | 0.00 |
| Squid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Smelts | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Gunnel | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0.00 |  | 0.00 |  |
| Sticheidae | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Sandfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 |  |  | 0.00 |  |  |
| Lanternfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 |  |  |  |  |
| Sandlance | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  | 0.00 | 0.00 |
| Grenadier | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Otherfish | 0 | 0 | 7 | 1 | 4 | 0 | 0.00 | 0.00 | 0.02 | 0.01 | 0.02 | 0.00 |
| Crabs | 0 | 0 | 1 | 1 | 0 | 0 | 0.00 | 0.06 | 0.51 | 0.61 | 0.31 | 0.00 |
| Starfish | 0 | 0 | 1 | 1 | 0 | 0 | 0.00 | 0.00 | 0.05 | 0.05 | 0.00 | 0.00 |
| Jellyfish | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Invertunid | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 |
| seapen/whip | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.07 | 0.00 | 0.00 |
| Sponge | 0 | 0 | 0 | 4 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.06 | 0.00 | 0.00 |
| Anemone | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.01 | 0.00 | 0.00 | 0.00 | 0.00 |
| Tunicate | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Benthinv | 0 | 0 | 1 | 0 | 0 | 0 | 0.00 | 0.01 | 0.09 | 0.12 | 0.00 | 0.00 |
| Snails | 0 | 0 | 0 | 0 | 0 | 0 |  |  |  |  |  |  |
| echinoderm | 0 | 0 | 1 | 1 | 0 | 0 | 0.01 | 0.00 | 0.20 | 0.18 | 0.00 | 0.00 |
| Coral | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Shrimp | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 |
| Birds | 0 | 0 | 0 | 0 | 0 | 0 | 0.00 | 0.00 | 0.02 | 0.00 | 0.00 | 0.00 |

Table 2.38-Summary of major results for the stock assessment of Pacific cod in the BSAI region.

| Tier | 3 b |
| :--- | ---: |
| Reference mortality rates |  |
| $M$ | 0.37 |
| $F_{40 \%}$ | 0.34 |
| $F_{355}$ | 0.42 |
| Equilibrium spawning biomass |  |
| $B_{35 \%}$ | $280,000 \mathrm{t}$ |
| $B_{40 \%}$ | $320,000 \mathrm{t}$ |
| $B_{100 \%}$ | $800,000 \mathrm{t}$ |
| Projected biomass for 2007 |  |
| $\quad$ Spawning (at max FABC) | $307,000 \mathrm{t}$ |
| Age 3+ | $960,000 \mathrm{t}$ |
| ABC for 2007 |  |
| $F_{A B C}$ (maximum permissible) | 0.33 |
| $F_{A B C}$ (recommended) | 0.33 |
| ABC (maximum permissible) | $176,000 \mathrm{t}$ |
| ABC (recommended) | $176,000 \mathrm{t}$ |
| Overfishing level for 2007 |  |
| Fishing Mortality | 0.39 |
| Catch | $207,000 \mathrm{t}$ |



Figure 2.1-Maps showing each 400 square kilometer cell with at least 3 observed hauls/sets containing Pacific cod in 2005, by gear type, overlaid against NMFS 3-digit statistical areas.


Figure 2.2-Comparison of numbers of age 0 EBS Pacific cod under last year's model updated with new data (Model 0 ) and eight alternative models.


Figure 2.3-Comparison of female spawning biomass of EBS Pacific cod under last year's model updated with new data (Model 0 ) and eight alternative models.


Figure 2.4 - Comparison of projected female spawning biomass of BSAI Pacific cod under last year's model updated with new data (Model 0) and eight alternative models.


Figure 2.5- Comparison of estimated selectivity at length in the EBS shelf bottom trawl survey under last year's model updated with new data (Model 0 ) and eight alternative models.


Figure 2.6-Selectivity at length (cm, evaluated at midpoints of length bins) as estimated by Model B1.


Figure 2.7-Biomass time trends (age 3+ biomass, female spawning biomass, survey biomass) of EBS Pacific cod as estimated by Model B1.


Figure 2.8-Time series of EBS Pacific cod recruitment at age 0 , with $95 \%$ confidence intervals, as estimated by Model B1.


Figure 2.9-Age 0 recruitment versus female spawning biomass for Pacific cod during the years 19772005 as estimated by Model B1, with Ricker stock-recruitment curve (for illustrative purposes only).


Figure 2.10-Trajectory of Pacific cod fishing mortality and female spawning biomass as estimated by Model B1, 1977-present. Because Pacific cod is a key prey of Steller sea lions, harvests of Pacific cod would be restricted to incidental catch in the event that spawning biomass fell below $B_{20 \%}$.

# Attachment 2.1: Results from Ecosystem Models on the Role of Pacific Cod In the Eastern Bering Sea and Aleutian Islands Ecosystems 

Sarah Gaichas and Kerim Aydin

Pacific cod are important predators in the EBS and AI ecosystems. While they are managed similarly in both ecosystems, food web modeling suggests key differences in cod's ecosystem role in the AI and EBS. The first key difference between ecosystems relates to cod's relative density in its continental shelf habitats in each system: because the AI has a much smaller area of shelf relative to the EBS (and the Gulf of Alaska, GOA), the smaller survey biomass estimate of cod in this area translates into a higher density in tons per square kilometer relative to the density in the EBS (Figure 1, left panel). Although the density of cod differs between systems, the relative effects of fishing and predation mortality as estimated within food web models constructed for each ecosystem (Aydin et al in review) are similar between the AI, EBS, and GOA. Here, sources of mortality are compared against the total production of cod as estimated in the BSAI and GOA cod stock assessment models (see Annex 2.1.A, "Production rates," for detailed methods). The "unknown" mortality in Figure 1 (left) represents the difference between the stock assessment estimated cod production and the known sources of fishing and predation mortality. While nearly half of cod production as estimated by the stock assessment appears to be "unused" in all three ecosystems, it is also clear that cod have relatively more fishing mortality than predation mortality in all three ecosystems (Figure 1, right panel). This suggests that changing fishing mortality is likely to affect cod population trajectories; therefore, we may ask what ecosystem effects changes in cod mortality might cause in each ecosystem.

To determine the potential ecosystem effects of changing total cod mortality, we first examine the diet data collected for cod. Diet data are collected aboard NMFS bottom trawl surveys in both the EBS and AI ecosytems during the summer (May - August); this comparison uses diet data collected in the early 1990's in each ecosystem. In the EBS, 2436 cod stomachs were collected during the 1991 bottom trawl survey and used in this analysis. In the AI, a total of 1181 cod stomachs were collected between the 1991 and 1994 bottom trawl surveys ( $\mathrm{n}=659$ and 533, respectively) and used in this analysis. The diet compositions reported here reflect the size and spatial distribution of cod in each survey (see Annex 2.1.A, "Diet calculations" for detailed methods). While the diet compositions reported here most accurately reflect early 1990's conditions in the BSAI, it is possible to update this information and examine changes in cod diets over time; that more extensive analysis is planned for a future assessment.

Food habits data show that Pacific cod have an extremely varied diet in both ecosystems (Figure 2). In the EBS, pollock are a major diet item for cod ( $26 \%$ of diet), but in the AI Atka mackerel and sculpins are the predominant fish prey for $\operatorname{cod}(15 \%$ of diet each), with pollock comprising less than $5 \%$ of the diet. In both ecosystems, Pandalid and non-Pandalid (NP) shrimp and various crabs are important prey, but other major prey items differ by ecosystem and seem to relate to the relative importance of benthic and pelagic pathways in each ecosystem as discussed in Aydin et al (in review). Commerically important crab species such as snow crab (C. opilio) and tanner crab (C. bairdi) make up 9\% of cod diets in the EBS, but less than $3 \%$ in the AI, reflecting the stronger benthic energy flow in the EBS. In contrast, squids make up over $6 \%$ of cod diets in the AI, but are very small proportions of diets in the EBS, reflecting the stronger pelagic energy flow in the AI. Myctophids are also found in cod diets only in the AI, reflecting the oceanic nature of the food web there. Cod are clearly opportunistic predators in both ecosystems, feeding on a variety of fish and invertebrates, and scavenging as well. Fishery offal makes up $5-7 \%$ of cod diets in both systems, indicating that while fishing causes cod mortality, it also contributes to cod production (although much fishery offal comes from fisheries directed at pollock, not cod).

Using diet data for all predators of cod and consumption estimates for those predators, as well as fishery catch data, we next estimate the sources of cod mortality in the AI and EBS (see detailed methods in Annex 2.1.A). As described above, sources of mortality are compared against the total production of cod as estimated in the BSAI cod stock assessment model. Mortality sources for cod are similar when comparing fisheries, but different when comparing predators between the EBS and AI. In both ecosystems, the trawl and longline fisheries for cod were the largest mortality sources for cod in the early 1990s (Figure 3). The next largest source of cod mortality is the pollock trawl fishery in the EBS and the directed Atka mackerel ("Other groundfish") fishery in the AI, which retains incidentally caught cod. In the EBS, pollock predation ranks next, and in the AI, adult and juvenile Steller sea lion predation represents the largest single source of predation mortality for cod. Cod cannibalism is a significant source of cod mortality only in the EBS, and flatfish trawl fisheries round out the large cod mortality sources in that ecosystem. Therefore, we see groundfish-dominated predation mortality sources for cod in the EBS, but sea-lion dominated predation mortality in the AI.

After comparing the different diet compositions and mortality sources of cod in each ecosystem, we shift focus slightly to view cod within the context of the larger EBS and AI food webs (Figure 4). Visually, it is apparent that cod's direct trophic relationships in each ecosystem include a majority of species groups; there are few boxes not connected to cod. However, comparing these food webs show further differences in cod trophic relationships between ecosystems. In the EBS, the significant predators of cod (blue boxes joined by blue lines) include the cod fisheries, the pollock fishery, and resident seals (upper panel of
Figure 4). Significant prey of cod (green boxes joined by green lines) include the many species shown in Figure 2. Light blue boxes in the EBS food web represent species which are both predators and prey of cod at some stage of life, with the most significant predator/prey of cod being pollock. In contrast, there are no species groups in the AI which are both predator and prey to cod (Figure 4, lower panel).

We can investigate whether these differences in cod diet, mortality, and relationships between the EBS and AI might suggest different ecosystem roles for cod in these areas. We use the diet and mortality results integrated with information on uncertainty in the food web using the Sense routines (Aydin et al in review) and a perturbation analysis with each model food web to explore the ecosystem relationships of cod further. Two questions are important in determining the ecosystem role of cod: which species groups are cod important to, and which species groups are important to cod? First, the importance of cod to other groups within the EBS and AI ecosystems was assessed using a model simulation analysis where cod survival was decreased (mortality was increased) by a small amount, $10 \%$, over 30 years to determine the potential effects on other living groups. This analysis also incorporated the uncertainty in model parameters using the Sense routines, resulting in ranges of possible outcomes which are portrayed as $50 \%$ confidence intervals (boxes in Figure 5) and $95 \%$ confidence intervals (error bars in Figure 5). Species showing the largest median changes from baseline conditions are presented in descending order from left to right. Therefore, the largest change resulting from a $10 \%$ decrease in cod survival in both ecosystems is a decrease in adult cod biomass, as might have been expected from such a perturbation. However, the decrease in biomass resulting from the same perturbation is different between the EBS and AI: the $50 \%$ intervals range from a $7-11 \%$ decrease in the AI, to a $7-17 \%$ decrease in the EBS (Figure 5).
The simulated decrease in cod survival affects the fisheries for cod similarly in the EBS and AI. After the decreased adult cod biomass, the next largest effect of the perturbation predicted by the models is a decrease in the "biomass" (catch) of the pot, longline, and trawl fisheries targeting adult cod in the EBS (Figure 5, top panel). In the AI ecosystem model, adult sablefish are predicted to have a larger change from the cod manipulation than the fisheries, although the predicted increase in sablefish biomass is much more uncertain than the predicted decrease in fishery catch in the AI (bottom panel, Figure 5). We discuss the sablefish result in detail below; for this discussion, we note that the cod fisheries in the AI are behaving similarly to the cod fisheries in the EBS after the simulated decrease in cod survival. Since cod fisheries are extremely specialized predators of cod, it makes sense that they are most sensitive to changes in the survival of cod in each ecosystem. It is notable that none of the other predators of cod showed a
significant sensitivity to a $10 \%$ decrease in cod survival. Pollock and sea lions ranked highest as nonfishery mortality sources of cod in the EBS and AI, respectively, but neither of these species were predicted to have significant changes in biomass in either ecosystem in this analysis: neither EBS pollock nor AI sea lions showed enough change from the baseline condition to be included in the plots. While these predators may cause significant cod mortality in each system, this analysis suggests that none of them are dependent on cod to the extent that small changes in cod survival affect their biomass in a predictable manner. It may be that these predator species would react more strongly to larger changes in cod survival; this could be further analyzed with different perturbation analyses.

In contrast with the predators of cod, a $10 \%$ decrease in cod survival is predicted to change the biomass of some cod prey, and even some species not directly connected to cod. In the EBS, greenling biomass is predicted to increase as a result of the perturbation, as are tanner crab and king crab biomass, albeit wth less certainty (Figure 5, top panel). In the AI, a larger set of species appear to react more strongly to increases in cod mortality than in the other two systems: sablefish, rex sole, arrowtooth flounder, and sleeper sharks are all predicted to increase in biomass in addition to greenlings and small sculpins (Figure 5). Of these, only rex sole, greenlings and other sculpins are direct cod prey; the change in adult sablefish and adult arrowtrooth biomass apparently arises from reduced cod predation mortality on the juveniles of each species in the AI ecosystem model: cod cause $80 \%$ of juvenile sablefish and juvenile arrowtooth mortality in the AI model. Sleeper sharks are neither predators nor prey of cod in the AI, suggesting that decreased cod survival has strong indirect effects in this ecosystem. Some of these differences in species sensitivity to cod mortality arise from the differences in cod diet in each system, but it seems likely that the higher sensitivity of multiple species to cod in the AI may also be due to cod's higher biomass per unit area there relative to the EBS. This in turn suggests that in the AI there may be stronger potential ecosystem effects of cod fishing than in the other two systems.

To determine which groups were most important to cod in each ecosystem, we conducted the inverse of the analysis presented above. In this simulation, each species group in the ecosystem had survival reduced by $10 \%$ and the system was allowed to adjust over 30 years. The strongest median effects on EBS and AI adult cod are presented in Figure 6. The largest effect on adult cod was the reduction in biomass resulting from the reduced survival of juvenile cod, followed by the expected direct effect, reduced biomass of adult cod in response to reduced survival of adult cod, in both ecosystems (Figure 6). Beyond these direct single species effects, cod appear most sensitive in all ecosystems to bottom up effects from both pelagic and benthic production pathways (small phytoplankton and benthic detritus). However, the bottom up effect is most pronounced in the AI, where the upper $95 \%$ intervals for the percent change of cod indicate that cod biomass will almost certainly decrease as a result of decreased survival of small phytoplankton, benthic detritus, and large phytoplankton (Figure 6). In contrast, the EBS model prediction is that cod biomass is likely to decrease from decreased survival of small phytoplankton and benthic detritus, but the detritus $95 \%$ intervals cross the x axis indicating that no change is also a possible outcome.

While decreased survival of primary producers appears to hurt cod, there are few species groups in either ecosystem which appear to benefit cod through reduced survival. In other words, they have no obvious single competitor or predator supressing cod biomass in the AI or EBS. In general, reduced "survival" (lower catch) of fisheries means more cod in the EBS and AI. In the EBS, reduced survival of other sculpins may increase cod biomass to some extent (Figure 6), which may seem counterintuitive given that reduced cod survival appeared to increase other sculpin biomass in the AI (Figure 5). While adult cod eat other sculpins, other sculpins in turn eat juvenile cod in the EBS (Figure 7), likely accounting for the results shown in Figure 6.

The results of these perturbation analyses suggest that the regional level of management applied to Pacific cod should be modified to account for differences between ecosystems. The food web relationships of cod are demonstrably different between the EBS and AI ecosystems, where they are currently assessed and managed identically. The impacts of changing cod survival (and by extension, fishing mortality) differ by
ecosystem as well, with the impacts felt most strongly and with highest certainty in the AI ecosystem according to this analysis. Therefore, it seems that the cod fishery in the AI should be managed separately from that in the EBS to ensure that any potential ecosystem effects of changing fishing mortality might be monitored at the appropriate scale.

## References

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Figure 1. Comparative biomass density (left) and mortality sources (right) for Pacific cod in the AI, EBS, and GOA ecosystems. For the AI and GOA, biomass density (left) is the average biomass from early 1990s NMFS bottom trawl surveys divided by the total area surveyed. For the EBS, biomass density is the stock assessment estimated adult (age 3+) biomass for 1991 (Thompson and Dorn 2005) divided by the total area covered by the EBS bottom trawl survey. Total cod production (right) is derived from cod stock assessments for the early 1990's, and partitioned according to fishery catch data and predation mortality estimated from cod predator diet data (Aydin et al in review). See Annex 2.1.A for detailed methods.

AI P. Cod diet


| ther sculpins <br> andalidae <br> quids <br> ffal <br> olychaetes <br> . Pollock_Adu <br> yctophidae <br> isc. crabs <br> enthic Amphipods <br> airdi <br> lisc. fish shallow <br> th. managed forage elpouts |  |
| :---: | :---: |

Figure 2. Comparison of Pacific cod diet compositions for the EBS (top) and AI (bottom) ecosystems. Diets are estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI). See Annex 2.1.A for detailed methods.


Figure 3. Comparison of Pacific cod mortality sources for the EBS (top) and AI (bottom) ecosystems. Mortality sources reflect cod predator diets estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI), cod predator consumption rates estimated from stock assessments and other studies, and catch of cod by all fisheries in the same time periods (Aydin et al in review). See Annex 2.1.A for detailed methods.


Figure 4. Adult and juvenile cod in the EBS (top) and AI (bottom) food webs. Predators of cod are dark blue, prey of cod are green, and species that are both predators and prey of cod are light blue. Box size is proportional to biomass and lines between boxes represent the most significant energy flows.

BS P. Cod effects on other species


AI P. Cod effects on other species


Figure 5. Effect of changing cod survival on fishery catch (yellow) and biomass of other species (dark red): EBS (top) and AI (bottom), from a simulation analysis where cod survival was decreased by $10 \%$ and the rest of the ecosystem adjusted to this decrease for 30 years. Boxes show resulting percent change in the biomass of each species on the x axis after 30 years for $50 \%$ of feasible ecosystems, error bars show results for $95 \%$ of feasible ecosystems (see Aydin et al in review for detailed Sense methods).


Figure 6. Effect of reducing fisheries catch (yellow) and other species survival (dark red) on cod biomass: EBS (top) and AI (bottom), from a simulation analysis where survival of each X axis species group was decreased by $10 \%$ and the rest of the ecosystem adjusted to this decrease for 30 years. Boxes show resulting percent change in the biomass of adult cod after 30 years for $50 \%$ of feasible ecosystems, error bars show results for $95 \%$ of feasible ecosystems (see Aydin et al in review for detailed Sense methods).


Figure 7. Juvenile cod mortality sources: EBS (top) and AI (bottom). Mortality sources reflect juvenile cod predator diets estimated from stomach collections taken aboard NMFS bottom trawl surveys in 1991 (EBS) and in 1991-1994 (AI), cod predator consumption rates estimated from stock assessments and other studies, and catch of cod by all fisheries in the same time periods (Aydin et al in review). See Annex 2.1.A for detailed methods.

## Annex 2.1.A

## Diet composition calculations

## Notation:

$$
\begin{aligned}
& D C=\text { diet composition } \\
& W=\text { weight in stomach } \\
& n=\text { prey } \\
& p=\text { predator } \\
& s=\text { predator size class } \\
& h=\text { survey haul } \\
& r=\text { survey stratum } \\
& B=\text { biomass estimate } \\
& v=\text { survey } \\
& a=\text { assessment } \\
& R=\text { ration estimate }
\end{aligned}
$$

The diet composition for a species is calculated from stomach sampling beginning at the level of the individual survey haul (1), combining across hauls within a survey stratum (2), weighting stratum diet compositions by stratum biomass (3), and finally combining across predator size classes by weighting according to size-specific ration estimates and biomass from stock assessment estimated age structure (4). Ration calculations are described in detail below.

Diet composition (DC) of prey $n$ in predator $p$ of size $s$ in haul $h$ is the total weight of prey $n$ in all of the stomachs of predator p of size s in the haul divided by the sum over all prey in all of the stomachs for that predator size class in that haul:

$$
\begin{equation*}
D C_{n, p, s, h}=W_{n, p, s, h} / \sum_{n} W_{n, p, s, h} \tag{1}
\end{equation*}
$$

Diet composition of prey n in predator p of size s in survey stratum r is the average of the diet compositions across hauls within that stratum:

$$
\begin{equation*}
D C_{n, p, s, r}=\sum_{h} D C_{n, p, s, h} / h \tag{2}
\end{equation*}
$$

Diet composition of prey n in predator p of size s for the entire area t is the sum over all strata of the diet composition in stratum $r$ weighted by the survey biomass proportion of predator $p$ of size $s$ in stratum $r$ :

$$
\begin{equation*}
D C_{n, p, s, t}=\sum_{r} D C_{n, p, s, r} * B_{p, s, r}^{v} / \sum_{r} B_{p, s, r}^{v} \tag{3}
\end{equation*}
$$

Diet composition of prey n in predator p for the entire area t is the sum over all predator sizes of the diet composition for predator p of size s as weighted by the relative stock assessment biomass of predator size $s$ times the ration of predator $p$ of size $s$ :

$$
\begin{equation*}
D C_{n, p, t}=\sum_{s} D C_{n, p, s, t} * B_{p, s}^{a} * R_{p, s} / \sum_{s} B_{p, s}^{a} * R_{p, s} \tag{4}
\end{equation*}
$$

## Ration Calculations

Size specific ration (consumption rate) for each predator was determined by the method of fitting the generalized Von Bertalanffy growth equations (Essington et al. 2001) to weight-at-age data collected aboard NMFS bottom trawl surveys.

The generalized Von Bertalanffy growth equation assumes that both consumption and respiration scale allometrically with body weight, and change in body weight over time ( $\mathrm{dW} / \mathrm{dT}$ ) is calculated as follows (Paloheimo and Dickie 1965):

$$
\begin{equation*}
\frac{d W_{t}}{d t}=H \cdot W_{t}^{d}-k \cdot W_{t}^{n} \tag{5}
\end{equation*}
$$

Here, $W_{t}$ is body mass, $t$ is the age of the fish (in years), and $H, d, k$, and $n$ are allometric parameters. The term $H \cdot W_{t}^{d}$ is an allometric term for "useable" consumption over a year, in other words, the consumption (in wet weight) by the predator after indigestible portions of the prey have been removed and assuming constant caloric density between predator and prey. Total consumption $(Q)$ is calculated as $(1 / A) \cdot H \cdot W_{t}^{d}$, where $A$ is a scaling fraction between predator and prey wet weights that accounts for indigestible portions of the prey and differences in caloric density. The term $k \cdot W_{t}^{n}$ is an allometric term for the amount of biomass lost yearly as respiration.

Based on an analysis performed across a range of fish species, Essington et al. (2001) suggested that it is reasonable to assume that the respiration exponent $n$ is equal to 1 (respiration linearly proportional to body weight). In this case, the differential equation above can be integrated to give the following solution for weight-at-age:

$$
\begin{equation*}
W_{t}=W_{\infty} \cdot\left(1-e^{-k(1-d)\left(t-t_{0}\right)}\right)^{\frac{1}{1-d}} \tag{6}
\end{equation*}
$$

Where $W_{\infty}$ (asymptotic body mass) is equal to $(H / k)^{\frac{1}{1-d}}$, and $t_{0}$ is the weight of the organism at time $=0$. If the consumption exponent $d$ is set equal to $2 / 3$, this equation simplifies into the "specialized" von Bertalanffy length-at-age equation most used in fisheries management, with the "traditional" von Bertalanffy K parameter being equal to the $k$ parameter from the above equations divided by 3 .

From measurements of body weight and age, equation 2 can be used to fit four parameters ( $W_{\infty}, d, k$, and $t_{0}$ ) and the relationship between $W_{\infty}$ and the $H, k$, and $d$ parameters can then be used to determine the consumption rate $H \cdot W_{t}^{d}$ for any given age class of fish. For these calculations, weight-at-age data available and specific to the modeled regions were fit by minimizing the difference between $\log$ (observed) and $\log$ (predicted) body weights as calculated by minimizing negative log likelihood: observation error was assumed to be in weight but not aging. A process-error model was also examined but did not give significantly different results.

Initial fitting of 4-parameter models showed, in many cases, poor convergence to unique minima and shallow sum-of-squares surfaces: the fits suffered especially from lack of data at the younger age classes that would allow fitting to body weights near $\mathrm{t}=0$ or during juvenile, rapidly growing life stages. To counter this, the following multiple models were tested for goodness-of-fit:

1. All four parameters estimated by minimization;
2. $d$ fixed at $2 / 3$ (specialized von Bertalanffy assumption)
3. $d$ fixed at 0.8 (median value based on metaanalysis by Essington et al. 2001).
4. $t_{0}$ fixed at 0 .
5. $d$ fixed at $2 / 3$ with $t_{0}$ fixed at 0 , and $d$ fixed at 0.8 with $t_{0}$ fixed at 0 .

The multiple models were evaluated using Aikeike's Information Criterion, AIC (spreadsheet review). In general, the different methods resulted in a twofold range of consumption rate estimates; consistently, model \#3, $d$ fixed at 0.8 while the other three parameters were free, gave the most consistently good results using the AIC. In some cases model \#1 was marginally better, but in some cases, model \#1 failed to converge. The poorest fits were almost always obtained by assuming that d was fixed at $2 / 3$.

To obtain absolute consumption $(Q)$ for a given age class, the additional parameter $A$ is required to account for indigestible and otherwise unassimilated portions of prey. We noted that the range of indigestible percentage for a wide range of North Pacific zooplankton and fish summarized in Davis (2003) was between $5-30 \%$, with major zooplankton (copepods and euphasiids), as well as many forage fish, having a narrower range of indigestible percentages, generally between 10-20\%. Further, bioenergetics models, for example for walleye pollock (Buckley and Livingston), indicate that nitrogenous waste (excretion) and egestion resulted in an additional 20-30\% loss of consumed biomass. As specific bioenergetics models were not available for most species, we made a uniform assumption of a total non-respirative loss of $40 \%$ (from a range of $25-60 \%$ ) for all fish species, with a corresponding $A$ value of 0.6 .

Finally, consumption for a given age class was scaled to population-level consumption using the available numbers-at-age data from stock assessments, or using mortality rates from stock assessments and the assumption of an equilibrium age structure in cases where numbers-at-age reconstructions were not available.

## Production rates

Production per unit biomass $(P / B)$ and consumption per unit biomass $(Q / B=R$, ration above) for a given population depend heavily on the age structure, and thus mortality rate of that population. For a population with an equilibrium age structure, assuming exponential mortality and Von Bertalanffy growth, $P / B$ is in fact equal to total mortality $Z$ (Allen 1971) and $Q / B$ is equal to $(Z+3 K) / A$, where $K$ is Von Bertalanffy's $K$, and $A$ is a scaling factor for indigestible proportions of prey (Aydin 2004). If a population is not in equilibrium, $\mathrm{P} / \mathrm{B}$ may differ substantially from Z although it will still be a function of mortality.

For the Bering Sea, Aleutian Islands, and Gulf of Alaska ECOPATH models, P/B and $\mathrm{Q} / \mathrm{B}$ values depend on available mortality rates, which were taken from estimates or literature values used in single-species models of the region. It is noted that the single-species model assumptions of constant natural mortality are violated by definition in multispecies modeling; therefore, these estimates should be seen as "priors" to be input into the ECOPATH balancing procedures or other parameter-fitting (e.g. Bayesian) techniques.

Several methods were used to calculate $\mathrm{P} / \mathrm{B}$, depending on the level of data available. Proceeding from most data to least data, the following methods were used:

1. If a population is not in equilibrium, total production P for a given age class over the course of a year can be approximated as $\left(\mathrm{N}_{\mathrm{at}} \cdot \mathrm{W}_{\mathrm{at}}\right)$, where $\mathrm{N}_{\mathrm{at}}$ is the number of fish of a given age class in a given year, exponentially averaged to account for mortality throughout the year, and $W_{a t}$ is the change in body weight of that age class over that year. For a particular stock, if weight-at-age data existed for multiple years, and stock-assessment reconstructed numbers-at-age were also available, production was calculated by summing this equation over all assessed age classes. Walleye pollock P/B for both the EBS and GOA were calculated using this method: examining the components of this sum over the years showed that numbers-at-age variation was responsible for considerably more variability in overall $\mathrm{P} / \mathrm{B}$ than was weight-at-age variation.
2. If stock assessment numbers-at-age were available, but a time series of weight-at-age was not available and some weight-at-age data was available, the equation in (1), above, was used, however, the change in body weight over time was estimated using fits to the generalized Von Bertalanffy equations described in the consumption section, above.
3. If no stock assessment of numbers-at-age was available, the population was assumed to be in equilibrium, so that $\mathrm{P} / \mathrm{B}$ was taken to equal Z . In cases for many nontarget species, estimates of Z were not available so estimates of $M$ were taken from conspecifics with little assumed fishing mortality for this particular calculation.

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