# 6. Assessment of the Rex Sole Stock in the Gulf of Alaska 

By

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## Executive Summary

Changes in the Input Data

1) The fishery catch and length compositions for 2010 and 2011 (through Sept. 24, 2011) were incorporated in the model.
2) The 2009 fishery catch and length compositions were updated.
3) The 2011 GOA groundfish survey biomass estimate and length composition data were added to the model. Survey biomass declined from 124,744 tin 2009 to $95,134 \mathrm{t}$ in 2011. Survey biomass estimates and length compositions were recalculated by the RACE GOA Groundfish Survey for all survey years.
4) Survey age compositions for two years (1999 and 2009) were added to the model.

## Changes in the Assessment Model

No changes were made in the assessment model. Because estimates of $\mathrm{F}_{40 \%}$ and $\mathrm{F}_{35 \%}$ (required for Tier 3 calculations) from the assessment model are considered unreliable while estimates of current and projected biomass are considered reliable, harvest specifications are based on Tier 5 calculations using estimated "adult biomass" from an age-structured assessment model (rather than survey biomass)

## Changes in the Assessment Results

1. Based on Tier 5 calculations, $\mathrm{F}_{\mathrm{ABC}}$ was found to correspond to a harvest level of $0.128 \mathrm{yr}^{-1}$, while $\mathrm{F}_{\mathrm{OFL}}$ corresponded to a harvest level of $0.170 \mathrm{yr}^{-1}$.
2. Using the age-structured assessment model and best estimates of actual catches in 2011 and 2012, "adult biomass" was estimated to be $87,162 \mathrm{t}$ in 2012 and 85,528 in 2013. Estimates of adult biomass were calculated by applying the rex sole maturity curve to estimates of biomass-at-age.
3. Using estimates of adult biomass from the age-structured assessment model, based on our best estimates for harvest levels in 2011-12, the recommended ABC for 2012 is $9,612 \mathrm{t}$ and the recommended ABC for 2013 is $9,432 \mathrm{t}$.
4. The OFL for 2012 is $12,561 \mathrm{t}$ and the OFL for 2013 is $12,326 \mathrm{t}$.

The area apportionments, based on the 2011 GOA Groundfish Survey, corresponding to the recommended ABCs are:

|  |  | West |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Southeast |  |  |  |  |  |
|  | Western | Central | Yakutat | Outside | Total |
| Area Apportionment | $13.6 \%$ | $66.7 \%$ | $8.7 \%$ | $11.0 \%$ | $100.0 \%$ |
| 2012 ABC (t) | 1,307 | 6,412 | 836 | 1,057 | 9,612 |
| 2013 ABC (t) | 1,283 | 6,291 | 821 | 1,037 | 9,432 |

A summary of the recommended ABCs from the 2011 assessment, relative to the 2010 SAFE projections, is as follows:

| Quantity | As estimated or specified last year (2010) |  | As estimated or specified this year (2011) |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 2011 | 2012 | 2012 | 2013 |
| M (natural mortality) | 0.17 | 0.17 | 0.17 | 0.17 |
| Specified/recommended tier | 5 | 5 | 5 | 5 |
| Biomass (adult; t) | 86,729 | 85,203 | 87,162 | 85,528 |
| $F_{\text {OFL }}=\mathrm{M}$ | 0.170 | 0.170 | 0.170 | 0.170 |
| max $F_{\text {ABC }}=0.75 * \mathrm{M}$ | 0.128 | 0.128 | 0.128 | 0.128 |
| recommended $F_{\text {ABC }}$ | 0.128 | 0.128 | 0.128 | 0.128 |
| OFL (t) | 12,499 | 12,279 | 12,561 | 12,326 |
| max ABC (t) | 9,565 | 9,396 | 9,612 | 9,432 |
| ABC (t) | 9,565 | 9,396 | 9,612 | 9,432 |
| Status | As determi | (2010) for: | As determi | (2011) for: |
|  | 2009 | 2010 | 2010 | 2011 |
| Overfishing | no | n/a | no | n/a |

## $\underline{\text { Plan Team Summary Tables }}$

| Species | Year | Biomass $^{\mathbf{1}}$ | OFL $^{\mathbf{2 , 3}}$ | ABC $^{\mathbf{2 , 3}}$ | TAC $^{2,3}$ | Catch $^{\mathbf{4}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2010 | 88,221 | 12,714 | 9,729 | 9,729 | 3,636 |
| Rex sole | 2011 | 86,974 | 12,499 | 9,565 | 9,565 | 2,594 |
|  | 2012 | 87,162 | 12,561 | 9,612 |  |  |
|  | 2013 | 85,528 | 12,326 | 9,432 |  |  |

${ }^{1}$ Adult biomass from the assessment model.
${ }^{2}$ http://www.fakr.noaa.gov/sustainablefisheries/specs10_11/goa_table1.pdf
${ }^{3}$ http://www.fakr.noaa.gov/sustainablefisheries/specs11_12/goa_table1.pdf
${ }^{4}$ As of Sept. 24, 2011.

| Stock/ <br> Assemblage | Area | 2011 <br> OFL $^{\mathbf{1}}$ | ABC $^{\mathbf{1}}$ | TAC $^{\mathbf{1}}$ | Catch $^{\mathbf{2}}$ | $\mathbf{2 0 1 2}$ <br> $\mathbf{O F L}^{\mathbf{3}}$ | $\mathbf{A B C}^{\mathbf{3}}$ | $\mathbf{2 0 1 3}$ <br> $\mathbf{O F L}^{\mathbf{3}}$ | $\mathbf{A B C}^{\mathbf{3}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W | -- | 1,517 | 1,517 | 105 | -- | 1,307 | - | 1,283 |
|  | C | -- | 6,294 | 6,294 | 2,488 | -- | 6,412 | -- | 6,291 |
| Rex sole | WYAK | - | 868 | 868 | 1 | - | 1,057 | - | 1,038 |
|  | SEO | -- | 886 | 886 | 0 | -- | 836 | -- | 820 |
|  | Total | 12,499 | 9,565 | 9,565 | 2,594 | 12,561 | 9,612 | 12,326 | 9,432 |

${ }^{1} h t t p: / / w w w . f a k r . n o a a . g o v / s u s t a i n a b l e f i s h e r i e s / s p e c s 11 \_12 / g o a \_t a b l e 1 . p d f ~$
${ }^{2}$ As of Sept. 24, 2011.
${ }^{3}$ Based on Tier 5 calculations using adult biomass estimates from the assessment model.

## SSC Comments Specific to the Rex Sole Assessments

SSC comment: The SSC requests that the next assessment re-evaluate the assumed age-length transition matrix to determine how it influences the estimated fishery selection curve. Also, the next assessment should provide analyses of mechanisms...that might account for the large differences between the survey and the fishery selection curves.

Author response: B. Matta of AFSC's Age and Growth Program has found potential differences in growth patterns for rex sole between the eastern portion of the Gulf of Alaska and the western and central portions, with individuals growing more slowly and attaining smaller maximum sizes in the eastern Gulf. While this result may have important implications for stock structure, the analysis is not yet complete. In addition, this year the Age and Growth Program completed processing several years worth of survey age data and, for the first time, fishery age data. Age composition data based on the new survey ages have been incorporated into this assessment, but the current assessment model does not incorporate fishery age compositions. The principal assessment author will make it a top priority to use the new age data to reevaluate the age-length conversion matrices used in the assessment. Unfortunately, he was not able to complete this analysis in time for inclusion in this assessment.

SSC comment: The SSC requests that the next assessment provide likelihood profiles or similar analyses that illustrate the consistency of the model fits to the various input data sources.

Author response: Posterior density plots for model parameters and other estimated quantities have been developed based on MCMC integration, replacing the limited number of likelihood profiles provided in the previous assessment. A number of these, both for individual parameters/quantities as well as for time series, have been incorporated into the current assessment. While these appear to address the overall issue of consistency of the model fits with respect to the input data, they do not address the issue of consistency of model fits with respect to individual data sources. Further guidance from the SSC on this issue would be greatly appreciated.

## SSC Comments on Assessments in General

SSC request: The SSC requested that the next round of assessments consider the possible use of ADF\&G bottom trawl survey data to expand the spatial and depth coverage.

Author response: The current assessment model does not accommodate surveys from multiple sources. We are (still) developing a new assessment model that will incorporate surveys from multiple sources as one of its new features. When completed, this new model will allow us to explore the utility of using the ADF\&G bottom trawl survey data in future assessments.

## Introduction

Rex sole (Glyptocephalus zachirus) is a right-eyed flatfish occurring from southern California to the Bering sea and ranging from shallow water ( $<100 \mathrm{~m}$ ) to about 800 meters depth (Mecklenburg et al., 2002). They are most abundant at depths between 100 and 200 m and are found fairly uniformly throughout the Gulf of Alaska (GOA).

Rex sole appear to exhibit latitudinal changes in growth rates and size at sexual maturity. Abookire (2006) found marked differences in growth rates and female size at maturity between stocks in the GOA and off the coast of Oregon. Size at sexual maturity was greater for fish in the GOA than in Oregon, as was size-at-age. However, these trends offset each other such that age-at-maturity was similar between the two regions.

Rex sole are batch spawners with a protracted spawning season in the GOA (Abookire, 2006). The spawning season for rex sole spans at least 8 months, from October to May. Eggs are fertilized near the sea bed, become pelagic, and probably require a few weeks to hatch (Hosie et al. 1977). Hatched eggs produce pelagic larvae that are about 6 mm in length and are thought to spend about a year in a pelagic stage before settling out to the bottom as 5 cm juveniles.

Rex sole are benthic feeders, preying primarily on amphipods, polychaetes, and some shrimp.

## Management units and stock structure

In 1993 rex sole was split out of the deep-water management category because of concerns regarding the Pacific ocean perch bycatch in the rex sole target fishery. The stock within the GOA is managed as a unit stock but with area-specific ABC and TAC apportionments to avoid the potential for localized depletion. Little is known on the stock structure of this species, although this is an area.

## Fishery

Rex sole in the Gulf of Alaska are caught in a directed fishery using bottom trawl gear. Fishing seasons are driven by seasonal halibut PSC apportionments, with approximately 7 months of fishing occurring between January and November. Catches of rex sole occur primarily in the Western and Central management areas in the gulf (statistical areas 610 and $620+630$, respectively). Recruitment to the fishery begins at about age 5 .

Catch is currently reported for rex sole by management area (Table 6.1, Fig. 6.1). Catches for rex sole were estimated from 1982 to 1994 by multiplying the deepwater flatfish catch by the fraction of rex sole in the observed catch. Historically, catches of rex sole have exhibited decadal-scale trends. Catches increased from a low of 93 t in 1986 to a high of $5,874 \mathrm{t}$ in 1996, then declined to about $3,000 \mathrm{t}$ thereafter. The 2009 catch $(4,753 \mathrm{t})$ was the largest since 1996. Catches have subsequently declined the past two years and is now more similar to the longterm average. In 2010 the catch was $3,636 \mathrm{t}$ and in 2011 it was 2,594 t (as of Sept. 24; 2011).

Based on observer data, the catch of rex sole is widely distributed along the outer margin of the continental shelf in the central and western portions of the Gulf (Figures 6.2-3). The spatial pattern of catches has been reasonably consistent over the past three years, with persistent areas of catches occurring off the Shumagin Islands, and the southwest tip and Cape Barnabas regions of Kodiak Island. Most of the catch is taken in the first and second quarters of the year.

The rex sole resource has been moderately harvested in recent years (Table 6.2). The fishery catches in 2009 and 2010 each represented between $40-50 \%$ of the rex sole ABC in that year. As of Sept. 24, catch in 2011 was less than $30 \%$ of the 2011 ABC.

Estimates of retained and discarded catch (t) in the rex sole fishery since 1995 were calculated from discard rates observed from at-sea sampling and industry reported retained catch (Table 6.2a). Retention of rex sole is high and has generally been over $95 \%$.

## Data

## Fishery Data

This assessment used fishery catches from 1982 through 24 September, 2011 (Table 6.1, Fig. 6.1), as well as estimates of the proportion of individuals caught by length group and sex for the years 1982-2011 (as of Sept. 24; Table 6.3). Thanks to recent work by the Alaska Fisheries Science Center’s (AFSC) Age \& Growth Program, two years of fishery age composition data is also available now, but the current assessment model does not incorporate fishery age data. Direct incorporation of fishery age data in the assessment awaits completion of a new assessment model. Sample sizes for the size (and age) compositions are shown in Table 6.4a.

## Survey Data

Because rex sole are often taken incidentally in target fisheries for other species, CPUE from commercial fisheries seldom reflects trends in abundance for this species. It is therefore necessary to use fisheryindependent survey data to assess the condition of this stock.

This assessment used estimates of total biomass for rex sole in the Gulf of Alaska from triennial (19841999) and biennial (2001-2011) groundfish surveys conducted by the AFSC's Resource Assessment and Conservation Engineering (RACE) division to provide an index of population abundance (Table 6.5, Fig. 6.4). Although survey depth coverage has been inconsistent for depth strata $>500 \mathrm{~m}$ (Table 6.5a), the fraction of the rex sole stock occurring in these depth strata is typically small (Table 6.5b), so we have not attempted to correct the survey estimates of total biomass for missing depth strata. We have, however, corrected the 2001 survey estimate of total biomass, because the eastern section of the Gulf was not sampled that year. We estimated the average stock biomass occurring in the unsampled area from the 1993, 1996 and 1999 surveys and expanded the 2001 estimate to correct for the missing area. As is evident from Fig. 6.4, survey biomass has fluctuated on decadal time scales. From an initial low of $\sim 60,000 \mathrm{t}$ in 1984, estimated biomass increased to a high of almost 100,000 t in 1990, then declined during the 1990s to slightly above $70,000 \mathrm{t}$. Subsequently, survey biomass increased once again and was above 100,000 $t$ in the 2005-2009 period. Survey biomass from the 2011 groundfish survey was $95,134 \mathrm{t}$, representing a $24 \%$ decline from the 2009 value ( $124,744 \mathrm{t}$ ), which was the largest in the time series. However, the 2011 survey biomass estimate is above the longterm average ( $\sim 86,000 \mathrm{t}$ ).

Estimates of the total number of individuals by length group from each RACE GOA groundfish survey (Table 6.6) were also incorporated into the assessment, as were estimates of total population numbers-at-age (Table 6.7). Survey age compositions were available for all survey years except for 2011 (1984, 1987, 1990, 1993, 1996, 1999, 2001, 2003, 2005, 2007 and 2009), although the age composition for 1990 was excluded from the model because the underlying ages are probably biased low due to the age reading technique (surface age reading) originally used to process the otoliths. Because age compositions were calculated from age-length data using the corresponding size compositions, size compositions were de-weighted in the model likelihood for years where age composition data was available to avoid double counting. Survey size composition data was fully weighted in the model likelihood only for years when age compositions were unavailable (1990 and 2011). Sample sizes for the survey-related data sources are given in Table 6.4b.

Data on individual growth was incorporated in the assessment using sex-specific age-length conversion matrices (Table 6.8a, b). These matrices have been used in previous assessments (Turnock et al., 2005; Stockhausen et al., 2007; Stockhausen et al. 2009). Sex-specific weight-at-age relationships and female maturity schedules used in previous assessments (Turnock et al., 2005; Stockhausen et al., 2007;
Stockhausen et al. 2009) were also used in this assessment (Table 6.9)
To summarize, the following data was incorporated in the assessment:

| Source | type | years |
| :--- | :--- | :--- |
| Fishery | catch | $1982--2011$ |
|  | size compositions | $1982-1984,1990-2011$ |
|  | biomass | $1984-1999$ (triennial); <br> $2001-2011$ (biennial) |
|  | size compositions | $1984-1999$ (triennial); <br> $2001-2011 ~(b i e n n i a l) ~$ |
|  | age compositions | $1984,1987,1993,1999 ;$ <br> $2001-2009$ (biennial) |

## Analytic Approach

Several alternative model configurations have been considered in previous assessments (Turnock et al., 2005; Stockhausen et al., 2007; Stockhausen et al., 2009). For this assessment, due to time constraints, we adopted the approach endorsed by the GOA Plan Team for this stock at the November 2009 Plan Team Meetings in Seattle. Consequently, we have developed harvest recommendations for the GOA rex sole stock using a Tier 5 approach ( $\mathrm{F}_{\mathrm{OFL}}=\mathrm{M}, \mathrm{F}_{\mathrm{ABC}}=0.75 \cdot \mathrm{M}$ ) applied to estimates of adult biomass from a Tier 3-type age-structured assessment model.

## Model structure

Current stock levels were estimated for 2011 and projected for 2012-2013 using the "base" model formulation as in 2009: a split-sex, age-structured model with parameters evaluated in a maximum likelihood context. The model structure (Appendix A) was developed following Fournier and Archibald's (1982) methods, with many similarities to Methot (1990). We implemented the model using automatic differentiation software developed as a set of libraries under C++ (ADModel Builder). ADModel Builder can estimate a large number of parameters in a non-linear model using automatic differentiation software extended from Greiwank and Corliss (1991) and developed into C++ class libraries. This software provides the derivative calculations needed for finding the minimum of an objective function via a quasiNewton function minimization routine (e.g., Press et al. 1992). It also gives simple and rapid access to these routines and provides the ability to estimate the variance-covariance matrix for all parameters of interest.

Age classes included in the model run from age 3 to 20. Age at recruitment was set at 3 years in the model due to the small number of fish caught at younger ages. The oldest age class in the model, age 20, serves as a plus group in the model; the maximum age of rex sole based on otolith age determinations has been estimated at 27 years (Turnock et al., 2005). Details of the population dynamics and estimation equations, description of variables and likelihood components are presented in Appendix A (Tables A.1, A.2, and A.3). Model parameters that are typically fixed are presented in Table A.4. A total of 89 parameters were estimated in the model (Table A.5).

## Parameters estimated independently

Model parameters related to natural mortality, growth, weight, maturity and survey catchability were fixed in the final model (Table A.4).

## Natural mortality

As in the previous full assessment (Turnock et al., 2005), natural mortality ( $M$ ) was fixed at $0.17 \mathrm{yr}^{-1}$ for both sexes in all age classes. This value was based on maximum observed age of 27 years for rex sole (Turnock et al., 2005).

Growth
The model estimates size compositions using fixed sex-specific age-length conversion matrices (Table 6.8). The distribution of size-at-age was assumed to be normally-distributed, with mean size-at-age modeled using the standard von Bertalanffy growth equation (Table 6.9, Fig. 6.6a):

$$
L_{t}=L_{\mathrm{inf}}\left(1-e^{-k\left(t-t_{0}\right)}\right)
$$

and age-varying variance. Sex-specific parameter values for the von Bertalanffy equation were estimated from mean length-at-age data collected during the 1984, 1987, 1990, 1993 and 1996 groundfish surveys (Turnock et al., 2005). The estimated values are

| Sex | $\mathbf{L}_{\infty}$ | $\mathbf{k}$ | $\mathbf{t}_{\mathbf{0}}$ |
| :--- | :---: | :---: | :---: |
| Males | 39.5 | 0.38 | 0.79 |
| Females | 44.9 | 0.31 | 0.69 |

Coefficients of variation (CVs) for length-at-age were also estimated from the survey data, and varied linearly from 0.13 for age 3 to 0.08 for age $20+$ (Turnock et al., 2005) for both sexes.

## Weight at length

Weight-at-length was modeled using the equation $W=a L^{b}$, with $L$ in centimeters and $W$ in grams. The parameter values for this equation, estimated from survey data, are

| Sex | a | b |
| :--- | :---: | :---: |
| Males | $1.0770 \mathrm{E}-06$ | 3.30571 |
| Females | $4.7933 \mathrm{E}-07$ | 3.44963 |
| Combined | $5.9797 \mathrm{E}-07$ | 3.41049 |

and are the same as used in the previous assessment. Weight-at-age (Table 6.9, Fig. 6.6b) was estimated using the weight-length relationship and the age-length conversion matrices.

## Maturity

Abookire (2006) modeled female rex sole size-at-maturity using a logistic model, obtaining a value for size at $50 \%$ maturity of 351.7 mm with a slope of $0.0392 \mathrm{~mm}^{-1}$. About half of the maturity samples were obtained from fishery catches and half from research trawls during 2000-2001. Using the mean length-atage relationship estimated from the 1984-1996 survey data, the age at $50 \%$-maturity was estimated at 5.6 years, (Table 6.9, Fig. 6.6). Estimates of mean size-at-age for the maturity samples were similar to those for mean size-at-age estimated from the survey data (Turnock et al., 2005).

Survey catchability
For the assessment, survey catchability ( $Q$ in Table A.1) was fixed at 1.

## Parameters estimated conditionally

A total of 89 parameters were estimated in the final model (Table A.5), including parameters on the recruitment of rex sole to the population (48 parameters total, including ones determining the initial age composition) and values related to annual fishing mortality (31 parameters total). The separable age component of fishing mortality was modeled using ascending logistic functions estimated separately for males and females (4 parameters total). The same approach was also used to estimate relative agespecific survey selectivity (4 parameters total). We also estimated the Tier 3 values for $\mathrm{F}_{\mathrm{ABC}}$ and $\mathrm{F}_{\mathrm{OFL}}$ : F40\% and F35\% (2 parameters).

Annual recruitment to the age 3 year class was parameterized in the model using one parameter for the log-scale mean recruitment and 47 parameters for the annual log-scale deviation from the mean. Recruitments were estimated back to 1965 to provide an initial age distribution for the model in its starting year (1982). In an analogous fashion, fully-recruited fishing mortality was parameterized in the model using one parameter for the log-scale mean and 30 parameters for the annual log-scale deviation from the mean.

Parameters in the model were selected based on minimizing an objective function equivalent to a negative log-likelihood function; hence, the parameter estimates are maximum likelihood estimates. Components that contribute to the overall (negative log) likelihood include those related to observed fishery catches, fishery size compositions, survey biomass estimates, survey size compositions, survey age composition, and recruitment deviations (Table A.3). The observed fishery catch was assumed to have a lognormal error structure, as was estimated survey biomass. The recruitment deviation parameters were incorporated directly into the overall likelihood via three components: "early" recruitment, "ordinary" recruitment and "late" recruitment (Table A.3). The "early" recruitment component incorporated deviations from 1965 to 1981 (i.e., prior to the modeled age structure), "ordinary" recruitment incorporated deviations from 1982-2008 and "late" recruitment incorporated deviations from 2009-2011. All three components were formulated assuming a lognormal error structure. The size and age compositions were assumed to be drawn from different sex-specific multinomial distributions. If this assumption were strictly correct, then the number of individuals contributing to each composition would be the appropriate corresponding sample size. However, because fish of the same size and age tend to be found together, size and age compositions tend to be overdispersed with respect to actual multinomial distributions. Also, the use of high sample sizes can lead to numerical problems in estimating the model parameters. Previous experience indicates that using a uniform sample size of 200 for compositions with more than 200 individuals provides an adequately simple solution to the problem of assigning sample sizes. Thus, a sample size of 200 was used for fully-weighted compositions (all age compositions and size compositions from years with no corresponding age compositions) and 1 for de-weighted compositions (size compositions with corresponding age compositions).

Different weights can be assigned to each likelihood component to increase or decrease the relative degree of model fit to the data underlying the respective component; a larger weight induces a closer fit to a given likelihood component. Typically, a relatively large weight (e.g., 20) is applied to the catch component while smaller weights (e.g., 1) are applied to the survey biomass, recruitment, and size and age composition components. This reflects a belief that total catch data are reasonably well known (smaller variance) compared with the other types of data. For the recruitment components, larger weights applied to a component force the deviations contributing to that component closer to zero (and thus force recruitment closer to the geometric mean over the years that contribute to the component). The weights used in this assessment are given in Table 6.10.

Initial values for the estimable parameters were set as listed in Table 6.11. To test whether resulting model solutions were indeed global, rather than local, maximum on the likelihood surface, we started the
two model cases using several different parameter sets. Most runs converged to the same (maximum) likelihood value and parameter estimates, providing evidence that the original solution was indeed the global maximum. The scores associated with different components of the maximum likelihood (e..g, fishery catch, survey biomass) are given in Table 6.12.

## Final parameter estimates

The estimated maximum likelihood parameter values are given in Table 6.13 for all model parameters.

## Model evaluation

Model estimates of fishery catch closely matched the observed values (Table 6.14 and Figure 6.7). This, however, is expected because a large weighting factor (20) was placed on the catch biomass component in the likelihood function. The model did not fit observed survey biomass values as closely as it did the catch (Table 6.14 and Figure 6.8), but the fit does appear reasonable.

Model fits to fishery size compositions and survey size and age compositions were quite similar to those obtained in the last full assessment (Stockhausen et al., 2009). For the most part, the model fit the fishery size compositions reasonably well, although not in 1982-1984 and 1988 (Figures 6.9a, b). Excluding these notable years, the model tended to slightly underestimate the peak and overestimate the width of the size compositions, particularly when the observed size composition was dominated by a single size class and thus sharply peaked-females in 1993, for example (Figure 6.9a). The smoothing inherent in using an age-length conversion matrix to convert age classes to size classes in the model precludes close fits to peaked size compositions: the peak will be underestimated and the tails will be overestimated. The slight bias in these fits might be improved by reducing the currently-assumed age-to-length conversion variance, but it may also be indicative of an interaction between fishery selectivity and age/size segregation in the stock.

The model's substantial misfits to the observed fishery size compositions in 1982-1984 (Figures 6.9a, b) suggest that fishery size selection patterns changed between 1982-1984 (years when the foreign fleets operated) and subsequent to 1990 (implementation of the domestic-only fishery). The average observed size caught was 35.6 cm in 1982-1984 and 40.4 cm since 1990. However, the model assumes that fishery selection is constant over the entire model time period (1982-2011) and is much more influenced by the data from the post-1990 era than by the 1982-1984 data. Finally, the model's substantial misfit in to the observed fishery size compositions in 1998 is caused by a secondary peak and exceedingly heavy right tail in the observed size composition in both sexes. This feature is unique in the 25 -year time series of observed fishery size compositions; it is unclear what might have been different in the fishery in 1998 to have caused this unique occurrence.

On the whole, the model fits to survey size compositions were better than those to fishery size compositions (Figures 6.10a, b). As with the fishery size compositions, the model fits to the survey size compositions were poorest when the observed size compositions were sharply peaked (e.g., 1984-1990, Figure 6.10a). Also, the model tends to overestimate abundance at large sizes ( $>40 \mathrm{~cm}$ ) for both sexes, although the effect is more consistent for males. Although this may indicate a bias in the current age/size conversion matrix used in the model, it may also indicate somewhat higher natural mortality than is assumed in the model ( $0.17 \mathrm{yr}^{-1}$ ) or a decrease in survey selectivity at larger ages/sizes-the latter cannot be accommodated in the current model configuration because selectivity is modeled using ascending logistic functions.

As in the previous full assessment (Stockhausen et al., 2009), the model fit survey age compositions "marginally well". The model fits to age compositions that were included in the previous assessment are very similar to the fits obtained in that assessment. One reason for this marginal performance may be that
recruitment in the model to the youngest age class (age 3) is assumed to have an equal sex ratio, while several of the survey age compositions exhibit substantial differences in sex ratio in the early age classes. For example, the 1987 survey age composition indicates six times more male age-5's (1982 year class) than females. However, there are no dynamics in the model that allow it to fit this type of discrepancy well, so the model ends up underestimating the proportion of males and overestimating the proportion of females in this case. It is also possible that these types of differences by sex in the observed age compositions indicate more overdispersion in the age sampling than is currently assumed. In this latter case, a different weighting scheme for the individual age composition-based on number of hauls, for example, rather than number of individuals-may improve the fits somewhat.

## Results

The estimated selectivity curves for the fishery and survey indicate that the fishery generally catches older flathead sole than the survey (Figures 6.12, 6.13). For the fishery, age at $95 \%$ selection was 12.4 years for females and 13.5 years for males. For the survey, the ages at $95 \%$ selection were younger: 6.3 years for females and 5.1 years for males. The rates of increase in the selectivity curves at $50 \%$ selection ( $\beta$ ) were reasonably steep (>1 yr ${ }^{-1}$ ) and similar between males and females, the fishery and the survey. Examination of the marginal posterior distributions from MCMC integration for these parameters indicates that they were well-estimated in the model, except for the value of the $\beta$ parameter for female selectivity in the fishery (Figure 6.13). Although the uncertainty associated with this latter parameter is fairly large, it has little impact on the resulting selectivity function-the actual curve approximates a knife-edge selectivity curve over the range of values indicated.

The model also estimates other population variables of interest, such as time series of total biomass, spawning biomass, recruitment and fully-selected fishing mortality. In this assessment, total biomass is represented by age $3+$ biomass whereas spawning biomass is female spawning biomass. Model quantities such as median recruitment, median fishing mortality, total biomass, spawning biomass and recruitment all seem to be reasonably well-estimated, based on examination of MCMC posterior distributions (Figures 6.14-15). The maximum likelihood estimates of these quantities tends to be very close to the posterior modes, with median F (Fig. 6.17, upper right plot) and estimates of recent recruitment (20052009; Figure 6.15, lower plot) being exceptions to this. Median recruitment was estimated at 47.6 million individuals. Median fishing mortality was estimated at $0.015 \mathrm{yr}^{-1}$. Total biomass in 2011 was estimated at $117,000 \mathrm{t}$, spawning stock biomass at $52,600 \mathrm{t}$ and recruitment at 49.4 million.

Model estimates suggest that age 3+ biomass generally underwent a decadal-scale oscillation, with total biomass increasing from 76,500 $t$ in 1982 to $99,100 \mathrm{t}$ in 1991 followed by a decline to $74,500 \mathrm{t}$ in 1998 and a subsequent increase to $119,700 \mathrm{t}$ in 2009 (Table 6.15, Figures 6.15-16). The estimate for 2011 is $116,900 \mathrm{t}$, slightly smaller than the maximum in 2009. In years where they overlap, the estimated age $3+$ biomass in this assessment is almost identical to that estimated in previous assessments (Table 6.15, Figure 6.16). The time series for estimated female spawning biomass underwent a progression similar to that of total biomass, but lagging the timing of the peaks and valleys in total biomass by 2 years (Table 6.16, Figures 6.15-16). The estimated 2011 spawning biomass is the largest in the time series ( $52,600 \mathrm{t}$ ). As with total biomass, spawning biomass estimated in this assessment is almost identical to that from previous assessments in years where they overlap (Table 6.16, Figure 6.16).

Model estimates of annual recruitment (age 3 numbers) ranged from a low of 28.9 million individuals in 1984 to a high of 114.7 million in 2008 (Table 6.17 and Figure 6.17). Prior to 1999 recruitment was generally below the long-term average ( 51.2 million) while it has generally been higher since 1999. In 2011, recruitment was estimated below the long-term average, but this is expected because of the structure of the recruitment likelihood component. Results from the current assessment are generally similar to those estimated in the previous assessment, particularly prior to 2006 (Table 6.17, Figure 6.17).

However, the last 3-5 recruitment estimates are highly uncertain, as is reflected in the variation between assessments.

Marginal posterior distributions based on MCMC integration are shown in Figure 6.18 for the Tier 3 quantities $\mathrm{F}_{35 \%}$ and $\mathrm{F}_{40 \%}, \mathrm{~B}_{35 \%}$ and $\mathrm{B}_{40 \%}$, and max ABC and OFL for 2012 as calculated using Tier 3a rules. The distributions for $\mathrm{F}_{35 \%}$ and $\mathrm{F}_{40 \%}$ indicate that, as expected from previous assessments, these quantities are highly uncertain and consequently rex sole does not qualify as a Tier 3 stock. This uncertainty results from a combination of relatively young age-at-maturity for rex sole ( 5.6 years) and selection by the fishery at relatively old ages (Figure 6.12), making spawner-per-recruit calculations insensitive to the overall level of fishing mortality (i.e., no matter how high F is, all fish caught by the fishery have already spawned several times).

## Reference fishing mortality rates

Because $\mathrm{F}_{35 \%}$ and $\mathrm{F}_{40 \%}$ are highly uncertain, Tier 3 considerations cannot be used to set reference fishing mortality rates and make harvest specifications for the GOA rex sole stock. In 2009, the GOA Plan Team decided that reference rates and harvest specifications for rex sole should be set using Tier 5 considerations. For Tier 5 stocks, reference fishing mortality rates are given by $\mathrm{F}_{\mathrm{OFL}}=\mathrm{M}$ (the rate of natural mortality) and max $\mathrm{F}_{\mathrm{ABC}}=0.75 \cdot \mathrm{M}$. Consequently, values for the reference fishing mortality rates for GOA rex sole are $\mathrm{F}_{\mathrm{OFL}}=0.17 \mathrm{yr}^{-1}$ and $\mathrm{F}_{\mathrm{ABC}}=0.128 \mathrm{yr}^{-1}$.

## Acceptable Biological Catch and Overfishing Level

In 2009, the GOA Plan Team decided that reference rates and harvest specifications for rex sole should be set using Tier 5 considerations. For Tier 5 stocks, harvest specifications are given by $O F L=F_{O F L} \cdot \bar{B}$ and $A B C=F_{A B C} \cdot \bar{B}$, where $\bar{B}$ is an estimate of stock biomass. For most Tier 5 stocks, the estimate of survey biomass for the stock from the most recent groundfish survey is used as $\bar{B}$. For rex sole, however, the GOA Plan Team determined that estimates of "adult" biomass (i.e., total biomass-at-age weighted by the fraction mature-at-age) from the assessment model provided more appropriate estimates of stock biomass than the groundfish survey and should be used for setting harvest specifications. Estimating adult biomass in the assessment model for 2012 and 2013 requires predictions of the total catch taken in 2011 and 2012. Because the 2011 fishery is not yet complete, we estimated the total catch taken in 2011 (3,448 t) using the average catch over the last 5 years. We assumed the same catch would be taken in 2012, as well. Using these values and the estimated numbers-at-age at the start of 2011 from the assessment model, we projected the stock ahead and calculated adult biomass $\left(B_{A}\right)$ at the start of 2012 and 2013 ( $87,162 \mathrm{t}$ and $85,528 \mathrm{t}$, respectively). We then calculated appropriate $\bar{B}$ 's for 2012 and 2013 using the Baranov catch equation

$$
\bar{B}=\frac{\left(1-e^{-Z}\right)}{Z} \cdot B_{A}
$$

where $\mathrm{Z}=\mathrm{M}+\mathrm{F}$ and F was $\mathrm{F}_{\mathrm{ABC}}$ or $\mathrm{F}_{\text {OfL }}$.
The estimated ABCs for 2012 and 2013 are 9,612 t and 9,432 t, respectively, while the estimated OFLs are $12,561 \mathrm{t}$ and $12,326 \mathrm{t}$.

## Area allocation of harvests

TACs for rex sole in the Gulf of Alaska are divided among four smaller management areas (Western, Central, West Yakutat and Southeast Outside). As in the previous assessment, the area-specific ABCs for rex sole in the GOA are divided up over the four management areas by applying the fraction of the most recent survey biomass estimated for each area (relative to the total over all areas) to the 2012 and 2013 ABCs. The area-specific allocations for 2012 and 2013 are:

|  |  | West |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Southeast |  |  |  |  |  |
|  | Western | Central | Yakutat | Outside | Total |
| Area Apportionment | $13.6 \%$ | $66.7 \%$ | $8.7 \%$ | $11.0 \%$ | $100.0 \%$ |
| 2012 ABC (t) | 1,307 | 6,412 | 836 | 1,057 | 9,612 |
| 2013 ABC (t) | 1,283 | 6,291 | 821 | 1,037 | 9,432 |

## Ecosystem Considerations

## Ecosystem effects on the stock

Prey availability/abundance trends
Based on results from an ecosystem model for the Gulf of Alaska (Aydin et al., 2007), rex sole in the Gulf of Alaska occupy an intermediate trophic level (Fig. 6.19). Polychaetes, euphasiids, and miscellaneous worms were the most important prey for rex sole in the Gulf of Alaska (Fig. 6.20).. Other major prey items included benthic amphipods, polychaetes, and shrimp (Livingston and Goiney, 1983; Yang, 1993; Yang and Nelson, 2000). Little to no information is available to assess trends in abundance for the major benthic prey species of rex sole.

## Predator population trends

Important predators on rex sole include longnosed skate and arrowtooth flounder (Fig. 6.21). The flatfish-directed fishery constitutes the second-largest known source of mortality on rex sole. However, unexplained mortality is the second largest component of mortality.

The longnose skate population appears to be stable. Arrowtooth flounder are currently the most abundant groundfish in the Gulf of Alaska, and have steadily increased in abundance since the early 1970's (Turnock et al., 2003b). Although the continued increase in abundance of arrowtooth flounder is cause for some concern, the abundance of rex sole has actually increased in recent years, as well. Increased predation by arrowtooth may be limiting the potential rate of increase of rex sole under current conditions, but it does not appear to represent a threat to the stock.

## Fishery effects on ecosystem

Catches of rex sole are widely distributed in the Gulf of Alaska over the past few years (Figure 6.2). The ecosystem effects of this spatial distribution of fishing activity are unknown.

Prohibited species such as halibut, salmon, and crab are also taken to some extent in the rex sole-directed fishery (Table 6.18). In 2011 (through September), the overall prohibited species catch (PSC) rate for Bairdi crab was 6,102 individuals, which accounted for $6.8 \%$ of the total Bairdi PSC. No king crab or opilio crab were caught in the rex sole fishery. The halibut PSC in the rex sole fishery was 172 t -less than half that in 2010 ( 388 t ). This accounted for $3.9 \%$ of the total PSC for halibut in 2011. The salmon PSC in the rex sole fishery was 2,300 Chinook and 93 non-Chinook in 2010. This accounted for $4.2 \%$ of the total Chinook PSC and $4.6 \%$ of the total non-Chinook PSC in 2010. o information was available at the time this document was compiled for 2011.

Bycatch of non-target species in the rex sole fishery tends be highly variable between years, at least when expressed as a percentage of the total observed bycatch in the FMP by non-target species group (Table 6.19). In 2010, the rex sole fishery accounted for more than $10 \%$ of the bycatch of four species groups: corals and bryozoans (10.3\%), unidentified invertebrates (14.3\%), miscellaneous invertebrates (e.g., worms) ( $100 \%$ ) and unidentified polychaetes (100\%). In 2009, by contrast, the fishery reportedly accounted for over $10 \%$ of total bycatch in 19 species groups, including three of the four species groups caught in 2010 (miscellaneous worms were not caught in 2009). The fishery has had no bycatch of birds and has accounted for less than $10 \%$ of bycatch in all shark and skate species groups over the time frame
analyzed (2003-2011), except for other skates (2003, 2006, 2009). The rex sole fishery has played a substantial role in bycatch of forage fish (capelin, eulachon, sandlance) in certain years, accounting for over $50 \%$ of capelin bycatch in 2008 and 2009 and almost $20 \%$ of eulachon bycatch in 2009.

Over the past five years, the rex sole-directed fishery caught more arrowtooth flounder than any other non-prohibited FMP species, including rex sole (Table 6.20). Rex sole was the second most-caught species in the directed fishery. Only small amounts of arrowtooth were retained (typically 10-20\%), while generally more than $98 \%$ of rex sole was retained. Catches of other non-prohibited species in the rex sole fishery were typically less than $20 \%$ of the rex sole catch.

Effects of discards and offal production on the ecosystem are unknown for the rex sole fishery.

## Data gaps and research priorities

The AFSC's Age and Growth Program has made substantial progress in processing survey age data for rex sole in the Gulf of Alaska. While this information has been incorporated in the current stock assessment in the form of survey age compositions, age information also enters the assessment in the form of age-length conversion matrices estimated outside the assessment model. The matrices currently used in the assessment are now several years old. One of our top priorities for the next assessment is to use the newly-available age data to revise growth schedules for GOA rex sole and reassess these agelength conversion matrices. In addition, we are working to incorporate such estimation directly into the assessment model, rather than performing it outside the model. This approach will also allow us to incorporate ageing error into the model structure.

Although the AFSC's Age and Growth Program has made substantial progress in processing survey age data for rex sole in the Gulf of Alaska, only two years of fishery age data has been processed. Additional age data (both survey and fishery) should improve future stock assessments by allowing improved estimates of individual growth and age-length transition matrices, and by filling in missing years with age composition data.

We will also investigate potential growth rate differences for rex sole between the eastern Gulf and the central/western Gulf. Although little catch is taken from the eastern Gulf, divergent growth patterns may have management implications for the stock as they may influence the perceived productivity of the stock.

Finally, further modeling research should address the use of length-based approaches to fishery and survey selectivity in the assessment model, as well as alternative forms for the selectivity function. The utility of potential environmental predictors of recruitment (e.g., temperature) should also be investigated. We also plan to revisit the estimates used for natural mortality in the model.

## Summary

| Tier | 5 |  |
| :--- | :---: | :---: |
| Reference mortality rates |  |  |
| $M$ | 0.17 |  |
|  |  |  |
| Fishing rates |  |  |
| $F_{\text {OFL }}$ | 0.170 |  |
| $F_{A B C}$ (maximum permissible) | 0.128 |  |
| $F_{A B C}$ (recommended) | 0.128 |  |
| Projected biomass |  |  |
| Adult biomass (t) | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ |
|  | 87,162 | 85,528 |
| Harvest limits |  |  |
| OFL (t) | $\mathbf{2 0 1 2}$ | $\mathbf{2 0 1 3}$ |
| ABC (maximum permissible; t) | 12,561 | 12,326 |
| ABC (recommended; t) | 9,612 | 9,432 |

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## Tables

Table 6.1. Annual catch of rex sole in the Gulf of Alaska, from 1982 to 2011. 2011 catch is through Sept. 24.

| year | total catch <br> (t) | Western Gulf | Central Gulf | West Yakutat | Southeast |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 959 |  |  |  | -- |
| 1983 | 595 |  |  |  | -- |
| 1984 | 365 |  |  |  | -- |
| 1985 | 154 |  |  |  | -- |
| 1986 | 93 |  |  |  | -- |
| 1987 | 1,151 |  |  |  | -- |
| 1988 | 1,192 |  |  |  | -- |
| 1989 | 599 |  |  |  | -- |
| 1990 | 1,269 |  |  |  | -- |
| 1991 | 4,636 |  |  |  | -- |
| 1992 | 3,000 |  |  |  | -- |
| 1993 | 3,000 |  |  |  | -- |
| 1994 | 3,673 |  |  |  | -- |
| 1995 | 4,021 |  |  |  | -- |
| 1996 | 5,874 |  |  |  | -- |
| 1997 | 3,294 |  |  |  | -- |
| 1998 | 2,669 |  |  |  | -- |
| 1999 | 3,060 |  |  |  | -- |
| 2000 | 3,591 |  |  |  | -- |
| 2001 | 2,940 |  |  |  | -- |
| 2002 | 2,941 |  |  |  | -- |
| 2003 | 3,485 | 767 | 2,716 | 1 | 1 |
| 2004 | 1,464 | 526 | 936 | 0 | 0 |
| 2005 | 2,176 | 576 | 1,600 | 0 | 0 |
| 2006 | 3,294 | 350 | 2,944 | 0 | 0 |
| 2007 | 2,852 | 413 | 2,438 | 1 | 0 |
| 2008 | 2,703 | 185 | 2,518 | 0 | 0 |
| 2009 | 4,753 | 342 | 4,410 | 1 | 0 |
| 2010 | 3,636 | 134 | 3,500 | 2 | 0 |
| 2011 | 2,594 | 105 | 2,488 | 1 | 0 |

Table 6.2a. Time series of recent reference points (ABC, OFL), TACs, total catch and retention rates for rex sole.

| Year | ABC (t) | TAC (t) | OFL (t) | Total <br> Catch (t) | \% <br> Retained |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | 11,210 | 9,690 | 13,091 | 4,021 | $90 \%$ |
| 1996 | 11,210 | 9,690 | 13,091 | 5,874 | $95 \%$ |
| 1997 | 9,150 | 9,150 | 11,920 | 3,294 | $92 \%$ |
| 1998 | 9,150 | 9,150 | 11,920 | 2,669 | $97 \%$ |
| 1999 | 9,150 | 9,150 | 11,920 | 3,060 | $96 \%$ |
| 2000 | 9,440 | 9,440 | 12,300 | 3,591 | $97 \%$ |
| 2001 | 9,440 | 9,440 | 12,300 | 2,940 | $95 \%$ |
| 2002 | 9,470 | 9,470 | 12,320 | 2,941 | $95 \%$ |
| 2003 | 9,470 | 9,470 | 12,320 | 3,485 | $95 \%$ |
| 2004 | 12,650 | 12,650 | 16,480 | 1,464 | $93 \%$ |
| 2005 | 12,650 | 12,650 | 16,480 | 2,176 | $91 \%$ |
| 2006 | 9,200 | 9,200 | 12,000 | 3,294 | $95 \%$ |
| 2007 | 9,100 | 9,100 | 11,900 | 2,852 | $98 \%$ |
| 2008 | 9,132 | 9,132 | 11,933 | 2,703 | $97 \%$ |
| 2009 | 8,996 | 8,996 | 11,756 | 4,753 | $99 \%$ |
| 2010 | 9,729 | 9,729 | 12,714 | 3,636 | $98 \%$ |
| 2011 | 9,565 | 9,565 | 12,499 | 2,594 | $97 \%$ |

Table 6.2b. Status of the rex sole fishery in recent years.

| Year | Dates | Status |
| :---: | :---: | :---: |
| 2005 | Jan 20 | open |
|  | Mar 23 | halibut bycatch status |
|  | Apr 1 | open |
|  | Apr 8 | halibut bycatch status |
|  | Apr 24 | open |
|  | May 3 | halibut bycatch status |
|  | Jul 5 | open |
|  | Jul 24 | halibut bycatch status |
|  | Sep 1 | open |
|  | Sep 4 | halibut bycatch status |
|  | Sep8 | open |
|  | Sep 10 | halibut bycatch status |
|  | Oct 1 | open |
|  | Oct 1 | halibut bycatch status |
| 2006 | Jan 20 | open |
|  | Apr 27 | halibut bycatch status |
|  | Jul 1 | open |
|  | Sep 5 | halibut bycatch status |
|  | Oct 1 | open |
|  | Oct 8 | halibut bycatch status |
| 2007 | Jan 20 | open |
|  | May 17 | halibut bycatch status |
|  | Jul 1 | open |
|  | Aug 10 | halibut bycatch status |
|  | Sep 1 | open |
|  | Sep 8 | halibut bycatch status |
|  | Oct 1 | open |
|  | Oct 15 | halibut bycatch status |
|  | Oct 22 | open |
| 2008 | Jan 20 | open |
|  | Apr 21 | halibut bycatch status |
|  | Jul 1 | open |
|  | Sep 9 | A80 vessels subject to sideboard limits |
|  | Sep 11 | halibut bycatch status |
|  | Oct 1 | open |
|  | Nov 6 | halibut bycatch status |
|  | Nov 16 | open |
| 2009 | Jan 20 | open |
|  | Mar 3 | halibut bycatch status |
|  | Apr 1 | open |
|  | Apr 23 | halibut bycatch status |
|  | Jul 1 | open |


| Year | Dates | Status |
| :--- | :--- | :--- |
| 2010 | Jan 20 | open |
|  | Apr 28 |  |
| halibut bycatch status |  |  |
| open |  |  |$|$| 2011 | Jan 20 | open |
| :--- | :--- | :--- |
|  | Apr 22 | halibut bycatch status |
|  | Jul 1 | open |
|  | Jul 1 | Rockfish Program CV <br> Coop.s and Limited <br> Access on halibut |

Table 6.3a. Annual fishery size compositions for female rex sole. The 2011 composition is based on observer reports through Sept. 24.

|  | Length cutpoints (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 9 | 11 | 13 | 15 | 17 | 19 | 21 | 23 | 25 | 27 | 29 | 31 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51 | 53 | 55 |
| 1982 | 0.0001 | 0.0000 | 0.0010 | 0.0018 | 0.0000 | 0.0004 | 0.0003 | 0.0004 | 0.0005 | 0.0015 | 0.0057 | 0.0362 | 0.1111 | 0.1040 | 0.0590 | 0.0332 | 0.0153 | 0.0039 | 0.0010 | 0.0009 | 0.0005 | 0.0002 | 0.0001 | 0.0000 |
| 1983 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0046 | 0.0185 | 0.0386 | 0.0974 | 0.1097 | 0.0788 | 0.0216 | 0.0031 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1984 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0101 | 0.0067 | 0.0236 | 0.0471 | 0.0404 | 0.0640 | 0.0101 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1985 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1986 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1987 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1988 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1989 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0011 | 0.0003 | 0.0055 | 0.0050 | 0.0151 | 0.0225 | 0.0291 | 0.0366 | 0.0491 | 0.0530 | 0.0525 | 0.0531 | 0.0426 | 0.0188 | 0.0097 | 0.0053 | 0.0013 | $\begin{array}{llllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0011 & 0.0003 & 0.0055 & 0.0050 & 0.0151 & 0.0225 & 0.0291 & 0.0366 & 0.0491 & 0.0530 & 0.0525 & 0.0531 & 0.0426 & 0.0188 & 0.0097 & 0.0053 & 0.0013 \\ 0.0000 & 0.050\end{array}$ 1991 1993 1993

1994 1994 1995 1996
1998
1999
2000
2001
2002
2003
2004
2005
2006
2007
2008
2009
2010


 $\begin{array}{lllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0000 & 0.0005 & 0.0006 & 0.0015 & 0.0029 & 0.0092 & 0.0244 & 0.0476 & 0.0865 & 0.1066 & 0.0954 & 0.0765 & 0.0439 & 0.0212 & 0.0106 & 0.0045 & 0.0017\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0002 & 0.0002 & 0.0000 & 0.0006 & 0.0006 & 0.0004 & 0.0008 & 0.0015 & 0.0025 & 0.0075 & 0.0098 & 0.0137 & 0.0315 & 0.0653 & 0.0960 & 0.1218 & 0.1187 & 0.0799 & 0.0402 & 0.0237 & 0.0102 & 0.0050\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0001 & 0.0001 & 0.0001 & 0.0003 & 0.0003 & 0.0007 & 0.0012 & 0.0038 & 0.0082 & 0.0213 & 0.0449 & 0.0791 & 0.1058 & 0.1068 & 0.0781 & 0.0462 & 0.0276 & 0.0136 & 0.0103\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0000 & 0.0002 & 0.0003 & 0.0007 & 0.0030 & 0.0025 & 0.0047 & 0.0074 & 0.0126 & 0.0172 & 0.0279 & 0.0381 & 0.0451 & 0.0623 & 0.0761 & 0.0720 & 0.0621 & 0.0349 & 0.0158 & 0.0061\end{array}$


 $\begin{array}{lllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0003 & 0.0009 & 0.0020 & 0.0026 & 0.0040 & 0.0072 & 0.0187 & 0.0448 & 0.0701 & 0.0790 & 0.0893 & 0.0856 & 0.0689 & 0.0348 & 0.0236 & 0.0121\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0004 & 0.0005 & 0.0007 & 0.0018 & 0.0070 & 0.0174 & 0.0303 & 0.0548 & 0.0711 & 0.0810 & 0.0849 & 0.0795 & 0.0494 & 0.0304 & 0.0167 & 0.0099\end{array}$ $\begin{array}{lllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0003 & 0.0008 & 0.0016 & 0.0038 & 0.0081 & 0.0146 & 0.0309 & 0.0526 & 0.0597 & 0.0689 & 0.0656 & 0.0475 & 0.0315 & 0.0168 & 0.0081 & 0.0051\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0019 & 0.0010 & 0.0029 & 0.0038 & 0.0318 & 0.0318 & 0.0760 & 0.0914 & 0.0943 & 0.0568 & 0.0298 & 0.0173 & 0.0087 & 0.0096 & 0.0029\end{array}$
 $\begin{array}{llllllllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0021 & 0.0103 & 0.0206 & 0.0351 & 0.0557 & 0.0804 & 0.0701 & 0.0990 & 0.0680 & 0.0351 & 0.0206 & 0.0124 & 0.0124 & 0.0041\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0006 & 0.0000 & 0.0011 & 0.0022 & 0.0011 & 0.0067 & 0.0134 & 0.0324 & 0.0687 & 0.1067 & 0.0983 & 0.0771 & 0.0508 & 0.0268 & 0.0145 & 0.0056 & 0.0011\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0012 & 0.0024 & 0.0048 & 0.0143 & 0.0308 & 0.0565 & 0.0826 & 0.0721 & 0.0719 & 0.0496 & 0.0255 & 0.0155 & 0.0037 & 0.0037 & 0.0007\end{array}$



[^0]Table 6.3b. Annual fishery size compositions for male rex sole. The 2011 composition is based on observer reports through Sept. 24.

|  | Length cutpoints (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| year | 9 | 11 | 13 | 15 | 17 | 19 | 21 | 23 | 25 | 27 | 29 | 31 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51 | 53 | 55 |
| 1982 | 0.0000 | 0.0003 | 0.0022 | 0.0022 | 0.0009 | 0.0006 | 0.0017 | 0.0006 | 0.0022 | 0.0056 | 0.0227 | 0.0968 | 0.2051 | 0.1560 | 0.0822 | 0.0342 | 0.0082 | 0.0011 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1983 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0015 | 0.0015 | 0.0031 | 0.0185 | 0.0371 | 0.0526 | 0.0680 | 0.1963 | 0.1901 | 0.0541 | 0.0046 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1984 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0067 | 0.0572 | 0.1313 | 0.3502 | 0.2088 | 0.0370 | 0.0034 | 0.0034 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 1985 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1986 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1987 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1988 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1989 | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | - | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- | -- |
| 1990 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0019 | 0.0006 | 0.0023 | 0.0055 | 0.0086 | 0.0177 | 0.0322 | 0.0536 | 0.1082 | 0.1467 | 0.1283 | 0.0622 | 0.0202 | 0.0057 | 0.0013 | 0.0006 | 0.0010 | 0.0008 | 0.0000 | $\begin{array}{llllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0019 & 0.0006 & 0.0023 & 0.0055 & 0.0086 & 0.0177 & 0.0322 & 0.0536 & 0.1082 & 0.1467 & 0.1283 & 0.0622 & 0.0202 & 0.0057 & 0.0013 & 0.0006 & 0.0010 & 0.0008 & 0.0000\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0001 & 0.0002 & 0.0009 & 0.0025 & 0.0045 & 0.0078 & 0.0089 & 0.0259 & 0.0649 & 0.1251 & 0.1349 & 0.0664 & 0.0253 & 0.0066 & 0.0025 & 0.0010 & 0.0004 & 0.0002 & 0.0002\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0003 & 0.0008 & 0.0015 & 0.0020 & 0.0054 & 0.0127 & 0.0239 & 0.0498 & 0.0812 & 0.1053 & 0.0774 & 0.0423 & 0.0216 & 0.0086 & 0.0046 & 0.0019 & 0.0010 & 0.0019\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0001 & 0.0001 & 0.0002 & 0.0004 & 0.0013 & 0.0028 & 0.0087 & 0.0219 & 0.0590 & 0.1195 & 0.1214 & 0.0766 & 0.0246 & 0.0059 & 0.0016 & 0.0012 & 0.0006 & 0.0000 & 0.0000\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0001 & 0.0003 & 0.0005 & 0.0006 & 0.0013 & 0.0028 & 0.0084 & 0.0314 & 0.0751 & 0.1260 & 0.1150 & 0.0634 & 0.0244 & 0.0093 & 0.0035 & 0.0015 & 0.0008 & 0.0002 & 0.0001\end{array}$ $\begin{array}{llllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0002 & 0.0010 & 0.0000 & 0.0015 & 0.0010 & 0.0019 & 0.0077 & 0.0160 & 0.0292 & 0.0502 & 0.0701 & 0.0805 & 0.0541 & 0.0249 & 0.0151 & 0.0081 & 0.0025 & 0.0017 & 0.0000 & 0.0000 \\ 0.0 .000\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0001 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0007 & 0.0010 & 0.0032 & 0.0051 & 0.0080 & 0.0249 & 0.0522 & 0.0786 & 0.0990 & 0.0850 & 0.0475 & 0.0192 & 0.0079 & 0.0035 & 0.0029 & 0.0012 & 0.0003\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0001 & 0.0007 & 0.0022 & 0.0038 & 0.0057 & 0.0101 & 0.0185 & 0.0421 & 0.0636 & 0.0846 & 0.0959 & 0.0898 & 0.0561 & 0.0230 & 0.0069 & 0.0034 & 0.0006 & 0.0003 & 0.0001\end{array}$

 $\begin{array}{lllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0004 & 0.0011 & 0.0023 & 0.0045 & 0.0076 & 0.0186 & 0.0356 & 0.0589 & 0.0763 & 0.0832 & 0.0838 & 0.0508 & 0.0225 & 0.0068 & 0.0043 & 0.0022 & 0.0008 & 0.0006\end{array}$ $\begin{array}{llllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0002 & 0.0005 & 0.0008 & 0.0014 & 0.0026 & 0.0051 & 0.0050 & 0.0118 & 0.0189 & 0.0386 & 0.0626 & 0.0694 & 0.0603 & 0.0534 & 0.0367 & 0.0178 & 0.0064 & 0.0020 & 0.0023 & 0.0008 & 0.0002\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllll} \\ 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0009 & 0.0020 & 0.0052 & 0.0083 & 0.0210 & 0.0419 & 0.0554 & 0.0718 & 0.0781 & 0.0758 & 0.0428 & 0.0250 & 0.0066 & 0.0049 & 0.0020 & 0.0014 & 0.0011\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0004 & 0.0000 & 0.0002 & 0.0004 & 0.0005 & 0.0027 & 0.0048 & 0.0115 & 0.0319 & 0.0665 & 0.0801 & 0.0867 & 0.0711 & 0.0430 & 0.0297 & 0.0170 & 0.0066 & 0.0011 & 0.0009 & 0.0014 & 0.0004\end{array}$

 $\begin{array}{llllllllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0059 & 0.0261 & 0.0571 & 0.1084 & 0.1328 & 0.0891 & 0.0429 & 0.0151 & 0.0084 & 0.0042 & 0.0008 & 0.0000 & 0.0000 & 0.0000\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0041 & 0.0041 & 0.0247 & 0.0330 & 0.1031 & 0.1010 & 0.0866 & 0.0701 & 0.0247 & 0.0165 & 0.0021 & 0.0021 & 0.0000 & 0.0000 & 0.0000\end{array}$
 $\begin{array}{lllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0003 & 0.0012 & 0.0009 & 0.0027 & 0.0090 & 0.0218 & 0.0611 & 0.0956 & 0.1181 & 0.0949 & 0.0745 & 0.0391 & 0.0279 & 0.0102 & 0.0047 & 0.0011 & 0.0008 & 0.0002\end{array}$ $\begin{array}{llllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0007 & 0.0028 & 0.0047 & 0.0203 & 0.0448 & 0.0946 & 0.1528 & 0.1374 & 0.0729 & 0.0294 & 0.0058 & 0.0011 & 0.0009 & 0.0001 & 0.0001 & 0.0000\end{array}$ $\begin{array}{lllllllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0001 & 0.0000 & 0.0012 & 0.0048 & 0.0117 & 0.0446 & 0.0893 & 0.1112 & 0.1415 & 0.1122 & 0.0561 & 0.0196 & 0.0090 & 0.0009 & 0.0009 & 0.0007 & 0.0004 & 0.0004\end{array}$


Table 6.4a. Sample sizes from the domestic fishery.

| year |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1990 | 74 | 7438 | 2482 | 3693 |  |  |  |  | 165 |
| 1991 | 257 | 18652 | 4724 | 4339 |  |  |  |  | 262 |
| 1992 | 220 | 19586 | 8045 | 6420 |  |  |  |  | 300 |
| 1993 | 372 | 25972 | 9067 | 7293 |  |  |  |  | 79 |
| 1994 | 328 | 19756 | 6935 | 6038 |  |  |  |  | 158 |
| 1995 | 257 | 11868 | 3282 | 1897 |  |  |  |  | 209 |
| 1996 | 277 | 18548 | 8212 | 6474 |  |  |  |  | 53 |
| 1997 | 193 | 10305 | 4962 | 5070 |  |  |  |  |  |
| 1998 | 213 | 10509 | 4609 | 3313 |  |  |  |  | 35 |
| 1999 | 393 | 8294 | 4466 | 3816 |  |  |  |  | 669 |
| 2000 | 347 | 7435 | 4484 | 2881 |  |  |  |  | 368 |
| 2001 | 194 | 3546 | 1949 | 1594 |  |  |  |  | 243 |
| 2002 | 320 | 5790 | 3110 | 2667 |  |  |  |  | 345 |
| 2003 | 352 | 6414 | 2662 | 3706 |  |  |  |  | 643 |
| 2004 | 62 | 1039 | 484 | 555 |  |  |  |  | 101 |
| 2005 | 71 | 1205 | 615 | 590 |  |  |  |  | 163 |
| 2006 | 37 | 501 | 256 | 229 |  |  |  |  | 150 |
| 2007 | 140 | 2261 | 1189 | 1057 | 44 | 192 | 109 | 82 | 277 |
| 2008 | 159 | 2677 | 1205 | 1459 |  |  |  |  | 297 |
| 2009 | 230 | 4189 | 1992 | 2114 | 73 | 344 | 166 | 177 | 486 |
| 2010 | 152 | 2892 | 1241 | 1651 |  |  |  |  | 350 |
| 2011 | 143 | 2859 | 1235 | 1621 |  |  |  |  | 413 |

Table 6.4b. Sample sizes from the GOA groundfish survey.

| year | $\frac{\text { biomass }}{\text { total }}$ hauls | hauls | Size co total indiv.s | sitions <br> females | males | hauls | total <br> indiv.s | composit females | males | otoliths collected |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 929 | 310 | 16927 | 6739 | 7191 | 5 | 233 | 155 | 78 | 233 |
| 1987 | 783 | 105 | 11577 | 5364 | 5998 | 5 | 189 | 102 | 87 | 823 |
| 1990 | 708 | 237 | 14387 | 7593 | 6793 | 27 | 270 | 156 | 114 | 550 |
| 1993 | 775 | 374 | 19100 | 9943 | 8166 | 29 | 332 | 193 | 139 | 341 |
| 1996 | 807 | 517 | 14496 | 6768 | 7718 | 77 | 370 | 212 | 158 | 383 |
| 1999 | 764 | 469 | 11652 | 5408 | 6204 | 51 | 381 | 196 | 174 | 487 |
| 2001 | 489 | 278 | 7675 | 3861 | 3814 | 130 | 668 | 383 | 284 | 682 |
| 2003 | 809 | 520 | 17833 | 8778 | 9028 | 95 | 596 | 328 | 266 | 602 |
| 2005 | 839 | 551 | 19233 | 9393 | 9806 | 102 | 588 | 310 | 278 | 600 |
| 2007 | 820 | 514 | 17305 | 8606 | 8555 | 55 | 416 | 220 | 196 | 424 |
| 2009 | 823 | 555 | 19933 | 9969 | 9941 | 100 | 484 | 267 | 217 | 496 |
| 2011 | 670 | 414 | 12871 | 6634 | 6166 |  |  |  |  | 523 |

Table 6.5. Biomass estimates ( t ) for GOA rex sole from the NMFS groundfish trawl surveys. Note that the Eastern Gulf (West Yakutat + Southeast) was not surveyed in 2001.
a) Biomass by NPFMC regulatory area. "Max Depth" is the maximum depth stratum surveyed.

| Year | Western <br> Gulf | Central <br> Gulf | West <br> Yakutat | Southeast | Total | Std. Dev | Max <br> Depth <br> (m) |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 6,672 | 40,688 | 9,209 | 4,102 | 60,670 | 6,023 | 1000 |
| 1987 | 8,801 | 39,722 | 11,160 | 4,144 | 63,826 | 5,906 | 1000 |
| 1990 | 6,765 | 75,147 | 12,745 | 3,569 | 98,225 | 10,731 | 500 |
| 1993 | 10,700 | 55,310 | 15,761 | 5,140 | 86,911 | 6,211 | 500 |
| 1996 | 9,419 | 43,778 | 9,855 | 9,705 | 72,757 | 5,301 | 500 |
| 1999 | 12,755 | 42,750 | 10,138 | 9,326 | 74,969 | 8,655 | 1000 |
| 2001 | 9,571 | 41,687 | 0 | 0 | 51,258 | 4,404 | 500 |
| 2003 | 13,265 | 57,973 | 10,566 | 18,093 | 99,897 | 7,559 | 700 |
| 2005 | 12,768 | 60,600 | 11,539 | 16,351 | 101,257 | 8,195 | 1000 |
| 2007 | 11,614 | 76,490 | 5,914 | 9,758 | 103,776 | 9,646 | 1000 |
| 2009 | 19,780 | 82,091 | 11,318 | 11,555 | 124,744 | 9,608 | 1000 |
| 2011 | 12,964 | 63,490 | 8,296 | 10,385 | 95,134 | 7,211 | 700 |

b) Biomass by depth stratum.

| year | Depth range (m) |  |  |  |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{1 - 1 0 0}$ |  |  |  |  |  | $\mathbf{1 0 1 - 2 0 0}$ | $\mathbf{2 0 1 - 3 0 0}$ | $\mathbf{3 0 1 - 5 0 0}$ | $\gg 500$ |
| 1984 | 3,987 | 37,040 | 13,083 | 5,161 | 1,399 |  |  |  |  |  |
| 1987 | 5,691 | 40,244 | 14,508 | 1,812 | 1,572 |  |  |  |  |  |
| 1990 | 15,460 | 59,833 | 21,791 | 1,140 | -- |  |  |  |  |  |
| 1993 | 11,233 | 54,064 | 16,995 | 4,619 | -- |  |  |  |  |  |
| 1996 | 10,403 | 43,419 | 14,929 | 4,006 | -- |  |  |  |  |  |
| 1999 | 14,682 | 40,239 | 15,766 | 3,841 | 440 |  |  |  |  |  |
| 2001 | 7,742 | 29,206 | 11,045 | 3,265 | -- |  |  |  |  |  |
| 2003 | 17,529 | 58,787 | 19,094 | 4,017 | 470 |  |  |  |  |  |
| 2005 | 14,786 | 65,060 | 16,731 | 4,535 | 146 |  |  |  |  |  |
| 2007 | 9,081 | 71,514 | 18,368 | 4,504 | 309 |  |  |  |  |  |
| 2009 | 16,017 | 79,662 | 25,032 | 2,980 | 1,054 |  |  |  |  |  |
| 2011 | 11,969 | 53,199 | 25,171 | 4,342 | 454 |  |  |  |  |  |

Table 6.6. Survey size compositions (in 1000’s) for rex sole.

## a) Females.

| year | Length bin cutpoints (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 9 | 11 | 13 | 15 | 17 | 19 | 21 | 23 | 25 | 27 | 29 | 31 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 |
| 1984 | 0 | 0 | 0 | 0 | 3 | 23 | 64 | 215 | 1,363 | 2,473 | 4,418 | 6,919 | 9,932 | 14,852 | 14,810 | 12,028 | 8,941 | 5,979 | 3,820 | 1,963 |
| 1987 | 0 | 0 | 63 | 0 | 16 | 267 | 394 | 977 | 1,484 | 2,432 | 4,290 | 5,623 | 11,104 | 12,967 | 10,274 | 8,841 | 8,183 | 6,566 | 4,052 | 2,441 |
| 1990 | 0 | 0 | 0 | 40 | 342 | 655 | 1,428 | 2,222 | 1,940 | 3,214 | 4,695 | 7,845 | 9,513 | 12,188 | 13,151 | 18,024 | 18,356 | 15,896 | 10,378 | 4,263 |
| 1993 | 0 | 14 | 17 | 87 | 292 | 495 | 634 | 536 | 876 | 1,512 | 3,361 | 5,114 | 8,930 | 11,925 | 13,904 | 13,509 | 13,518 | 11,324 | 9,504 | 6,524 |
| 1996 | 9 | 33 | 219 | 326 | 757 | 1,359 | 1,241 | 1,609 | 2,571 | 3,452 | 4,295 | 5,588 | 6,901 | 8,140 | 8,485 | 8,170 | 9,412 | 9,445 | 9,244 | 6,813 |
| 1999 | 22 | 38 | 163 | 538 | 1,034 | 2,131 | 2,431 | 3,180 | 3,935 | 6,402 | 7,864 | 7,557 | 9,026 | 8,958 | 9,481 | 7,987 | 8,173 | 7,666 | 6,709 | 4,597 |
| 2001 | 31 | 84 | 187 | 384 | 1,158 | 2,340 | 2,718 | 2,681 | 4,197 | 4,781 | 5,099 | 6,946 | 8,045 | 6,026 | 5,387 | 6,187 | 5,686 | 5,143 | 4,965 | 3,802 |
| 2003 | 92 | 381 | 1,024 | 1,272 | 2,137 | 4,563 | 5,881 | 6,974 | 10,071 | 13,522 | 15,947 | 18,050 | 19,262 | 20,527 | 17,516 | 13,654 | 10,210 | 8,028 | 5,780 | 3,432 |
| 2005 | 0 | 142 | 414 | 1,774 | 1,978 | 2,153 | 2,312 | 3,379 | 5,195 | 8,560 | 12,017 | 15,689 | 17,296 | 18,967 | 19,285 | 17,500 | 14,443 | 10,950 | 6,497 | 3,478 |
| 2007 | 71 | 0 | 339 | 1,597 | 3,732 | 4,960 | 6,524 | 5,212 | 6,500 | 6,988 | 9,776 | 11,966 | 11,982 | 12,662 | 14,063 | 16,091 | 16,975 | 12,802 | 9,308 | 5,594 |
| 2009 | 53 | 23 | 221 | 652 | 1,456 | 3,036 | 2,975 | 5,417 | 8,567 | 10,358 | 12,935 | 17,624 | 19,703 | 18,683 | 18,998 | 17,240 | 18,564 | 13,445 | 11,119 | 6,506 |
| 2011 | 0 | 0 | 13 | 171 | 415 | 1,410 | 1,645 | 2,050 | 3,236 | 5,646 | 9,630 | 13,974 | 14,622 | 14,859 | 16,027 | 15,471 | 14,489 | 12,286 | 8,675 | 5,167 |

## b) Males.

D) Males.

| year | Length bin cutpoints (cm) | $\mathbf{9}$ | $\mathbf{1 1}$ | $\mathbf{1 3}$ | $\mathbf{1 5}$ | $\mathbf{1 7}$ | $\mathbf{1 9}$ | $\mathbf{2 1}$ | $\mathbf{2 3}$ | $\mathbf{2 5}$ | $\mathbf{2 7}$ | $\mathbf{2 9}$ | $\mathbf{3 1}$ | $\mathbf{3 3}$ | $\mathbf{3 5}$ | $\mathbf{3 7}$ | $\mathbf{3 9}$ | $\mathbf{4 1}$ | $\mathbf{4 3}$ | $\mathbf{4 5}$ | $\mathbf{4 7}$ |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1984 | 0 | 0 | 7 | 7 | 14 | 186 | 193 | 581 | 2,282 | 5,131 | 7,178 | 12,590 | 16,570 | 12,058 | 6,898 | 4,049 | 2,174 | 713 | 339 | 34 |  |
| 1987 | 0 | 0 | 0 | 51 | 72 | 1,110 | 1,787 | 3,320 | 4,530 | 4,742 | 7,920 | 11,331 | 16,219 | 14,296 | 11,421 | 8,456 | 4,223 | 1,620 | 374 | 49 |  |
| 1990 | 0 | 0 | 34 | 98 | 393 | 970 | 2,091 | 2,119 | 3,476 | 4,874 | 9,361 | 15,469 | 18,820 | 21,143 | 20,689 | 12,464 | 7,418 | 3,668 | 1,185 | 101 |  |
| 1993 | 11 | 0 | 21 | 206 | 334 | 1,103 | 1,042 | 1,430 | 2,121 | 4,303 | 7,026 | 11,695 | 17,235 | 19,454 | 16,650 | 12,095 | 6,609 | 2,937 | 750 | 239 |  |
| 1996 | 48 | 42 | 164 | 741 | 952 | 1,691 | 1,694 | 2,722 | 4,901 | 7,640 | 10,058 | 13,671 | 15,125 | 13,728 | 10,597 | 7,770 | 4,953 | 2,305 | 1,257 | 672 |  |
| 1999 | 47 | 130 | 215 | 598 | 1,761 | 3,858 | 4,594 | 4,306 | 6,834 | 9,562 | 10,477 | 14,753 | 16,055 | 14,203 | $12, \mathbf{2 5 4}$ | 8,654 | 6,217 | 4,018 | 1,966 | 518 |  |
| 2001 | 0 | 63 | 111 | 687 | 1,889 | 2,123 | 3,178 | 3,794 | 4,403 | 6,706 | 6,814 | 7,223 | 6,420 | 5,471 | 6,494 | 6,258 | 5,961 | 3,394 | 1,114 | 223 |  |
| 2003 | 56 | 449 | 998 | 1,809 | 2,698 | 5,226 | 8,479 | 11,194 | 13,354 | 18,595 | 22,049 | 28,362 | 26,513 | 21,152 | 13,636 | 8,689 | 6,163 | 3,406 | 883 | 374 |  |
| 2005 | 146 | 36 | 599 | 947 | 1,828 | 2,490 | 3,152 | 5,558 | 9,966 | 16,061 | 19,653 | 23,121 | 23,995 | 20,691 | 15,771 | 11,622 | 6,786 | 2,896 | 990 | 390 |  |
| 2007 | 42 | 34 | 85 | 946 | 2,633 | 4,161 | 5,677 | 5,412 | 6,108 | 10,053 | 15,187 | 17,473 | 19,137 | 17,813 | 20,016 | 16,052 | 9,341 | 5,389 | 1,981 | 592 |  |
| 2009 | 88 | 40 | 129 | 588 | 1,301 | 3,287 | 4,279 | 7,063 | 10,505 | 14,658 | 20,590 | 26,509 | 31,402 | 28,897 | 23,860 | 15,937 | 7,437 | 2,940 | 1,360 | 373 |  |
| 2011 | 0 | 0 | 183 | 144 | 416 | 1,607 | 2,640 | 3,784 | 5,922 | 9,424 | 12,848 | 16,447 | 20,091 | 19,265 | 15,441 | 10,836 | 6,937 | 3,722 | 1,380 | 300 |  |

Table 6.7. Survey age compositions (in 1000's) for rex sole. The 1990 age composition was not used in the assessment model because the ages were probably underestimated due to the ageing technique (surface age reading) used.

b) Males.

| year | $\begin{array}{\|c\|} \hline \text { Age bins } \\ 3 \\ \hline \end{array}$ | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 0 | 11,696 | 14,404 | 5,642 | 3,556 | 9,754 | 1,817 | 3,148 | 1,956 | 1,734 | 996 | 4,661 | 0 | 3,497 | 1,976 | 2,588 | 0 | 665 |
| 1987 | 1,580 | 6,167 | 13,342 | 7,596 | 12,547 | 8,620 | 2,872 | 8,290 | 3,110 | 12,694 | 1,142 | 516 | 1,260 | 780 | 0 | 0 | 0 | 0 |
| 1990 | 11,896 | 24,450 | 52,356 | 26,245 | 8,333 | 688 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 2,389 | 4,926 | 24,642 | 24,216 | 15,868 | 6,658 | 11,018 | 6,722 | 5,870 | 708 | 59 | 1,302 | 436 | 0 | 0 | 0 | 0 | 191 |
| 1996 | 4,905 | 7,138 | 17,588 | 14,689 | 11,723 | 7,396 | 12,189 | 7,105 | 7,346 | 4,370 | 3,153 | 340 | 991 | 0 | 0 | 0 | 0 | 314 |
| 1999 | 10,511 | 21,674 | 18,335 | 7,933 | 9,685 | 2,523 | 6,637 | 8,995 | 2,050 | 5,053 | 4,287 | 3,364 | 2,167 | 237 | 5,444 | 0 | 1,731 | 2,907 |
| 2001 | 7,688 | 20,414 | 9,910 | 9,597 | 2,622 | 1,619 | 1,008 | 853 | 1,644 | 1,534 | 614 | 1,801 | 377 | 1,031 | 522 | 489 | 450 | 5,128 |
| 2003 | 15,432 | 18,937 | 22,860 | 38,205 | 25,693 | 10,704 | 11,233 | 4,742 | 8,781 | 3,591 | 5,540 | 5,426 | 2,027 | 2,122 | 3,482 | 2,351 | 916 | 6,685 |
| 2005 | 7,794 | 19,670 | 30,953 | 29,080 | 16,589 | 7,876 | 6,651 | 6,396 | 4,657 | 3,923 | 4,912 | 3,258 | 3,449 | 5,568 | 1,312 | 1,011 | 80 | 6,808 |
| 2007 | 6,681 | 20,353 | 7,348 | 20,081 | 19,098 | 19,882 | 15,149 | 6,522 | 2,449 | 0 | 962 | 962 | 3,996 | 2,876 | 3,616 | 2,326 | 744 | 13,033 |
| 2009 | 3,900 | 37,596 | 30,549 | 36,463 | 11,415 | 17,214 | 7,216 | 14,795 | 8,122 | 10,780 | 3,234 | 1,500 | 2,746 | 480 | 2,830 | 0 | 2,676 | 6,773 |

Table 6．8．Age－length conversion matrices for rex sole．Values at a row／column combination correspond to the fraction of individuals at the age indicated by the row that fall into the length group indicated by the column．

## a）Females．

| ${ }_{\text {age }}$ | ${ }_{\text {ate }}$ |  | 15 |  | 19 | ${ }^{21}$ | ${ }^{23}$ | 25 | 27 | 29 | ${ }^{31}$ | ${ }^{33}$ | 35 | ${ }^{37}$ | 39 | 41 | ${ }^{43}$ | 45 | 47 | 49 | 51 | 53 | 55 | 57 | 59 | 61 | ${ }^{3}$ | 65 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\underbrace{0}_{\substack{0.0022 \\ 0.0001}}$ | $\underbrace{0.00133}_{0} 0$ | ${ }_{\text {0，0．0899 }}^{0.0024}$ | $\underbrace{\text { a }}_{\substack{0.1528 \\ 0.017}}$ | ${ }_{\substack{0.2504 \\ 0.0351}}$ |  |  | ${ }^{0.00722}$ | ${ }_{\substack{0.0123 \\ 0.2096}}^{\text {0，}}$ | ${ }_{0}^{0.1561}$ | ${ }_{0}^{0.0865}$ | ${ }^{0.0356}$ | 0.0109 | 0.0025 | 0.0004 | 0.0001 |  | ${ }_{0}^{0}$ | ${ }_{0}^{0}$ |  | ${ }_{0}^{0}$ |  | 0 | ${ }_{0}^{0}$ | ${ }_{0}^{0}$ |  | ${ }_{0}^{0}$ |
| 5 |  |  | 0.000 | 0 | ${ }_{0}^{0.0013}$ | ${ }_{0}^{0.0051}$ | ${ }_{0}^{0.0164}$ | ${ }_{0}^{0.0451}$ | ${ }_{0}^{0.0058}$ |  | ${ }_{0}^{0.18511}$ | ${ }_{\substack{0.1872}}^{0.0182}$ | ${ }_{\substack{0.1554}}^{0.154}$ | ${ }_{0}^{0.1015}$ | ${ }_{0.0532}^{0.0025}$ | ${ }_{\text {coin }}^{\substack{0.0022}}$ | ${ }_{\text {cole }}^{\substack{0.0074}}$ | 0.002 | ${ }^{\text {0．0004 }}$ | 01 |  | 0 |  | 0 |  | 0 | \％ | ${ }_{0}$ |
|  | 0 |  | － |  | ${ }_{\text {a }}^{0.0001}$ | ${ }_{0}^{0.0005}$ | $\underbrace{\substack{\text { a }}}_{\substack{0.0045 \\ 0.0017}}$ | ${ }_{\substack{0 \\ 0.0052}}^{0.0052}$ | $\underbrace{0.093}_{\substack{0.033 \\ 0.014}}$ | ${ }_{\substack{0.0653 \\ 0.032}}^{0.093}$ | ${ }_{\text {a }}^{0} 0.1093$ | ${ }_{\substack{0 \\ 0.1113 \\ 0.107}}$ | $\underbrace{0.1731}$ | $\substack{0.1639 \\ 0.162}_{\substack{\text { 0，}}}$ | ${ }^{0.1283}$ | ${ }_{\substack{0.0331 \\ 0.133}}^{0.041}$ | ${ }^{0.0045}$ | ${ }^{0.0197}$ | ${ }^{0.0072}$ | ${ }^{0.0022}$ | ${ }^{0.0005}$ | ${ }^{0.0001}$ |  |  | 0 | $\bigcirc$ | 0 |  |
| 8 | 0 | 。 | $0$ | $0$ | 0.0001 | 0.0002 | 0．0009 | 0．0029 |  | 0.0195 | 0.0045 | 0.0715 | ${ }_{0}^{0.1084}$ | ${ }_{0}^{0.1408}$ | ${ }_{0.1566}$ | ${ }_{0.1492}$ | ${ }_{0.1217}$ | ${ }_{0}^{0.0851}$ |  | ${ }_{0}^{0.0261}$ | ${ }_{\text {a }}^{0.014}$ | ${ }_{\text {a }}^{0.00012}$ | ${ }_{\text {cole }}^{0.0014}$ | ${ }_{0}^{0.0004}$ | 0.0001 |  | $\bigcirc$ |  |
|  | 0 |  | $0$ | $0$ |  |  |  |  |  |  |  |  |  |  | 56 | ${ }^{0.1521}$ |  |  |  |  | 0.0211 |  |  |  |  | ${ }^{0.0001}$ | $\bigcirc$ |  |
| 10 <br> 11 <br> 11 | 0 0 | \％ | ： | ！ | $\bigcirc$ | ${ }_{0}^{0.0001}$ | ${ }_{0}^{0.0004}$ | 0．0013 | ${ }_{\substack{0.0038 \\ 0.003}}^{\substack{\text { a }}}$ | ${ }_{\text {a }}^{\substack{0.0097 \\ 0.0078}}$ |  | ${ }_{\substack{0 \\ 0.0356}}^{0.024}$ | ${ }_{\substack{0.0715 \\ 0.062}}^{0.0}$ | ${ }_{\substack{0 \\ 0 \\ 0.094}}^{0.094}$ | ${ }_{0}^{0.11241}$ | ${ }_{\substack{0.1488 \\ 0.144}}^{0.14}$ | ${ }_{\substack{0 \\ 0.143535}}^{0.145}$ | ${ }_{\substack{0 \\ 0.12792}}^{0.129}$ | ${ }_{\substack{0.0076 \\ 0.093}}^{0.09}$ | ${ }_{\text {a }}^{556}$ | S05 | ${ }_{\substack{0.0146 \\ 0.0197}}^{0.0}$ | ${ }_{\substack{0 \\ 0.0061}}^{0.0061}$ | ${ }_{0}^{0.0022}$ | ${ }_{\substack{0}}^{0.0007}$ | ${ }_{\substack{0 \\ 0.0002}}^{0.003}$ | 0 |  |
|  | 0 | 。 | 。 | 0 | 。 |  | ${ }^{0.000}$ | 0.0008 | ${ }_{0}^{0.0025}$ | 0.0067 | 0.0155 | ${ }_{0}^{0.0313}$ | 0.0556 | 863 | 0.1173 | 0.1395 | ${ }^{0.1452}$ | ${ }_{0}^{0.1323}$ | ${ }_{0}^{0.1056}$ | \％ 37 | 451 | ${ }_{0}^{0.0241}$ | ${ }_{0}^{0.013}$ | 0．0046 | ${ }_{0}^{0.0017}$ |  | 0.0001 |  |
| ${ }_{13}^{13}$ | $\bigcirc$ | \％ | 0 | 0 | 0 | 0.001 | 0．0022 |  | 0.0022 |  | 0.0139 | 0.0235 | ${ }^{0.0513}$ | 0.0808 | 0.1118 | ${ }^{0.1357}$ |  | ${ }^{0.11348}$ | 0.1104 |  |  |  |  |  | ${ }^{0.0021}$ | ${ }^{0.0007}$ | 2 |  |
| ${ }_{15}^{14}$ | ${ }_{0}^{0}$ | O | $\bigcirc$ | 0 | $\bigcirc$ |  | 0．002 | ${ }^{0.0000}$ | ${ }_{\text {dor }}^{0.002}$ | ${ }_{\text {0，}}^{0.0055}$ | 121 | 2066 | ${ }_{\substack{0.0483 \\ 0.042}}^{0.056}$ | ${ }^{0.0077}$ | ${ }^{0.1078}$ | ${ }^{1337}$ | ${ }^{0.1435}$ | 1333 | ${ }_{\substack{0.1137}}^{0116}$ | （134 | ${ }^{\text {a } 564}$ | 304 |  | （066 | 0．0025 | ${ }^{0.0008}$ | （0003 | （001 |
| ${ }_{16}^{15}$ | 0 | ${ }_{0}$ | \％ | \％ | 0 | $\bigcirc$ | 0．0002 | ${ }^{0} 0.00006$ | ${ }_{0}^{0.0019}$ | ${ }_{0}^{0.0049}$ | ${ }_{0}^{0.0116}$ | ${ }_{0}^{0.02223}$ | ${ }^{\text {a }}$ | ${ }_{0}^{0.0722}$ | ${ }_{0}^{0.1029}$ | ${ }_{\substack{0 \\ 0.12885}}^{0.1059}$ | ${ }_{\substack{0}}^{0.1427}$ | ${ }_{0}^{0.1377}$ | ${ }_{0}^{0.1176}$ | ${ }_{0}^{0.0083}$ | ${ }_{\text {a }}^{\substack{0.0584}}$ | ${ }^{0.0324}$ 0．034 | ${ }_{0}^{0.0174}$ | ${ }_{0}^{0.0078}$ | ${ }_{0}^{0.0039}$ | ${ }_{0}^{0.0011}$ | ${ }^{\text {0．0．003 }}$ | ${ }_{\text {a }}^{0.0001}$ |
| 17 | $\bigcirc$ | 0 | 0 | 0 |  | 0 | 0.0002 | ${ }^{0.00066}$ | ${ }^{0.0018}$ | S047 | 0.0013 | ${ }_{0}^{0.0237}$ | ${ }^{0.0436}$ | ${ }^{\text {0．07788 }}$ | ${ }^{0.1013}$ | ${ }_{0}^{0.1275}$ | 0.1414 | ${ }_{\text {O }}^{0} 0.1381$ | ${ }_{0}^{0.1187}$ | 0．0399 | ${ }^{0.0599}$ | ${ }^{0} 0.0352$ | ${ }^{0.00182}$ | ${ }^{\text {0．0033 }}$ | ${ }_{0}^{0.0033}$ | ${ }^{0} 0.0012$ | 004 | 0.0001 |
| ${ }_{19}^{18}$ | \％ | － | $\bigcirc$ | $\stackrel{0}{0}$ | $\bigcirc$ |  | （0，002 | 006 | （ent | ${ }^{0.00046}$ | ${ }^{0.0019}$ | ${ }_{\substack{0 \\ 0.0228}}^{0.0232}$ | ${ }_{\substack{0}}^{0.0429}$ | ${ }_{\substack{0.0698 \\ 0.0098}}^{0.0}$ | ${ }_{\text {a }}^{\text {0．0933 }}$ |  | ${ }_{\substack{0 \\ 0.140}}^{0.141}$ | ${ }_{\substack{0 \\ 0.1384}}^{0.1383}$ | ${ }_{\substack{0 \\ 0.125}}^{0.195}$ | ${ }_{\substack{0 \\ 0.0917}}^{\text {0．099 }}$ | ${ }_{\substack{\text { a }}}^{\text {0．061 }}$ | ${ }^{0.0367}$ | $\underbrace{0.0187}_{\substack{0.0192}}$ |  | ${ }_{\substack{0 \\ 0.0035}}^{0.0035}$ | $\underbrace{\text { 0．012 }}_{\substack{0 \\ 0.0013}}$ | ${ }_{\text {a }}^{0.0004}$ | ${ }_{\text {done }}^{\substack{0.0001}}$ |
| ${ }^{20}$ | 0 0 |  | 0 | － | 。 |  |  |  |  |  |  |  |  | 0065 | 209 | －1134 |  | －13 | 0.1204 | （090 | ．00 | ， 0371 | －0， |  |  |  |  | ${ }_{0}^{0.0002}$ |

## b）Males．

| age | ${ }^{\text {length cutpo }}$ | ${ }_{11}{ }^{\text {tis }}$（cm） | 13 | 15 | 17 | 19 | 21 | 23 | 25 | 27 | 29 | 31 | 33 | 35 | 37 | 39 | 41 | 43 | 45 | 47 | 49 | 51 | 53 | 55 | 57 | 59 | 61 | 63 | 65 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 0 | 0.0006 | 0.0047 | 0.0254 | 0.0873 | 0.1906 | 0.2649 | ${ }^{0.2344}$ | ${ }^{0.1321}$ | 0.0474 | 0.0108 | 0.0016 | 0.0001 |  |  | 0 |  |  | 0 |  |  | 0 |  | 0 | 0 | 0 | 0 | 0 |  |
| 4 | 0 | 0 | 0.0001 | 0.001 | 0.0052 | 0.0204 | 0.059 | 0.1251 | 0.1947 | ${ }^{0.2221}$ | 0.1857 | 0.1139 | ${ }^{0.0512}$ | 0.0169 | ${ }^{0.0041}$ | 0.0007 | 0.0001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | ${ }_{0} 0.0001$ | 0.0006 | 0.0029 | 0.0112 | 0.0331 | 0.076 | ${ }_{0}^{0.1353}$ | ${ }_{0} 0.1867$ | 0.1998 | ${ }^{0.1658}$ | 0.1067 | ${ }^{0.0533}$ | 0.0206 | ${ }^{0.0062}$ | 0.0014 | 0.0003 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0 | 0 | 0.0001 | 0.0007 | 0.003 | 0.0106 | 0.0299 | 0.0669 | 0.1192 | 0.169 | 0.1907 | 0.1712 | 0.1223 | 0.0695 | 0.0314 | 0.0113 | 0.0032 | 0.0007 | 0.0001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 7 | 0 | 0 | 0 | 0 | 0 | 0.0002 | 0.0011 | 0.0043 | 0.0139 | 0.036 | 0.0751 | 0.1261 | ${ }^{0.1706}$ | 0.1856 | 0.1626 | 0.1146 | 0.0651 | 0.0297 | 0.0109 | 0.0032 | 0.0008 | 0.0001 | 0 | 0 | 0 | 0 | 0 | 0 |  |
| 8 | 0 | 0 | 0 | 0 | 0 | 0.0001 | 0.0005 | 0.0021 | 0.0075 | 0.0217 | 0.0506 | 0.0956 | 0.1459 | 0.1798 | 0.179 | 0.144 | 0.0936 | 0.0491 | 0.0208 | 0.0071 | 0.002 | 0.0004 | 0.0001 | 0 | 0 | 0 | 0 | 0 |  |
| 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0002 | 0.0011 | 0.0045 | ${ }^{0.0143}$ | 0.0367 | ${ }^{0.0758}$ | ${ }_{0}^{0.1263}$ | 0.1699 | 0.1845 | ${ }^{0.1616}$ | 0.1143 | 0.0652 | 0.03 | 0.0112 | ${ }^{0.0033}$ | 0.0008 | 0.0002 | 0 | 0 | 0 | 0 | 0 |  |
| 10 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0001 | 0.0007 | 0.0029 | 0.0101 | 0.0281 | 0.0626 | 0.1121 | 0.1611 | 0.1861 | ${ }^{0.1727}$ | 0.1288 | 0.0772 | ${ }^{0.0372}$ | 0.0144 | 0.0045 | 0.0011 | 0.0002 | 0 | 0 | 0 | 0 | 0 |  |
| ${ }_{12}^{11}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0.0001 | ${ }^{0.0004}$ | ${ }^{0.002}$ | ${ }^{0.0075}$ | ${ }^{0.0225}$ | ${ }^{0.0535}$ | ${ }^{0.1017}$ | ${ }^{0.1544}$ | ${ }^{0.1869}$ | ${ }^{0.1806}$ | ${ }^{0.1393}$ | ${ }^{0.0857}$ | ${ }^{0.0421}$ | ${ }^{0.0165}$ | ${ }^{0.0052}$ | ${ }^{0.0013}$ | ${ }^{0.0003}$ | 0 | 0 | 0 | 0 | 0 |  |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | ${ }^{0.0003}$ | ${ }^{0.0014}$ | ${ }^{0.0057}$ | ${ }^{0.0184}$ | ${ }^{0.0468}$ | ${ }^{0.0941}$ | 0．1496 | ${ }_{\substack{0.1881 \\ 0.1898}}$ | ${ }_{0}^{0.1871} 0$ | ${ }_{0}^{0.1472}$ | ${ }_{0}^{0.0916} 0$ | ${ }^{0.0451} 0$ | ${ }_{\substack{0.0176 \\ 0.0178}}^{0}$ | ${ }_{0}^{0.0054} 0$ | ${ }_{\substack{0.0013 \\ 0.0012}}^{0}$ | ${ }_{0}^{0.0003}$ | 0 | ${ }_{0}$ | 0 | 0 | 0 |  |
| 14 | 0 | 0 | ${ }_{0}$ | 0 | 0 | 0 | ${ }_{0}$ | ${ }_{0}^{0.00001}$ | ${ }^{0.0007}$ | ${ }_{0}^{0.00355}$ | ${ }_{0}^{0.0129}$ | ${ }_{0}^{0.0374}$ | ${ }_{0}^{0.0834}$ | ${ }_{0}^{0.1442}$ | ${ }_{0.1923}^{0.1898}$ | ${ }_{0}^{0.1986}$ | ${ }_{0}^{0.1587}$ | ${ }_{0}^{0.00996}$ | ${ }_{0}^{0.04669}$ | ${ }_{0.0173}^{0.078}$ | ${ }_{0}^{0.0049}$ | ${ }_{0}^{0.0012}$ | ${ }_{0}^{0.0002}$ | ${ }_{0}^{0}$ | 0 | ${ }_{0}$ | 0 | ${ }_{0}$ |  |
| 15 | 0 |  | 0 | 0 | 0 | 0 | 0 | 0.0001 | 0.0005 | 0.0027 | 0.0109 | 0.0337 | ${ }^{0.0793}$ | ${ }_{0}^{0.1425}$ | 0.1953 | 0.2044 | 0.1632 | ${ }^{0.0994}$ | 0.0462 | 0.0164 | 0.0044 | ${ }_{0} 0.0009$ | 0.0001 | 0 | 0 | 0 | 0 | 0 |  |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0004 | ${ }^{0.0021}$ | 0.0092 | ${ }_{0} 0.3304$ | 0.0756 | 0.1414 | 0.1989 | 0.2103 | 0.1672 | 0.0999 | 0.0449 | 0.0151 | 0.0038 | 0.0007 | 0.0001 |  | 0 | 0 | 0 | 0 | 0 |
| 17 18 18 | ${ }_{0}$ | ${ }_{0}^{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}^{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}^{0}$ | ${ }_{0}^{0.00002}$ | －0．0016 | ${ }_{0}^{0.0076}$ | ${ }_{0}^{0.0273}$ | ${ }_{0}^{0.072} 0$ | 0.1406 0.14 | ${ }^{0.203}$ | ${ }_{0}^{0.21266}$ | 0.17799 0.1743 | 0.0996 <br> 0.0988 <br> 0.0 | 0.043 0.0406 | ${ }_{0}^{0.0137} 0$ | ${ }_{0}^{0.0032}$ | ${ }_{0}^{0.0006}$ | 0.0001 0 | ${ }_{0}$ | ${ }_{0}^{0}$ | 0 | ${ }_{0}^{0}$ | ${ }_{0}$ | ${ }_{0}$ |
| 19 19 | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}^{0.00001}$ | ${ }_{0}^{0.00009}$ | ${ }_{0}^{0.0051}$ | ${ }_{0}^{0.02416}$ | ${ }_{0.065}^{0.0655}$ | ${ }^{0.1393}$ | ${ }_{0}^{0.2123}$ | ${ }_{0}^{0.23321}$ | ${ }_{0}^{0.1775}$ | ${ }_{0}^{0.0973}$ | ${ }_{0}^{0.04068}$ | ${ }_{0}^{0.0121}$ | ${ }_{0}^{0.0026}$ | ${ }_{0}^{0.00003}$ | ${ }_{0}$ | ${ }_{0}$ | ${ }_{0}$ | 0 | ${ }_{0}$ | ${ }_{0}$ |  |
| 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0004 | 0.0032 | 0.0164 | 0.0577 | 0.1378 | 0.2232 | 0.2453 | 0.1829 | 0.0926 | 0.0318 | 0.0074 | 0.0012 | 0.0001 | 0 | 0 | 0 | － | 0 | 0 |  |

Table 6.9. Age-specific schedules for rex sole in the Gulf of Alaska. The maturity ogive is based on Abookire (2006).

|  | Length (cm) |  | Weight (g) |  | Maturity |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Age | Males | Females | Males | Females | ogive |
| 3 | 22.44 | 22.96 | 31.52 | 23.74 | 0.0083 |
| 4 | 27.84 | 28.81 | 64.22 | 51.93 | 0.0763 |
| 5 | 31.52 | 33.10 | 96.88 | 83.82 | 0.3073 |
| 6 | 34.05 | 36.24 | 124.95 | 114.66 | 0.6037 |
| 7 | 35.77 | 38.55 | 147.12 | 141.86 | 0.7901 |
| 8 | 36.95 | 40.24 | 163.77 | 164.52 | 0.8796 |
| 9 | 37.76 | 41.48 | 175.89 | 182.70 | 0.9224 |
| 10 | 38.31 | 42.39 | 184.53 | 196.91 | 0.9444 |
| 11 | 38.68 | 43.06 | 190.60 | 207.82 | 0.9566 |
| 12 | 38.94 | 43.55 | 194.84 | 216.08 | 0.9639 |
| 13 | 39.12 | 43.91 | 197.77 | 222.29 | 0.9685 |
| 14 | 39.24 | 44.18 | 199.79 | 226.93 | 0.9715 |
| 15 | 39.32 | 44.37 | 201.18 | 230.37 | 0.9736 |
| 16 | 39.38 | 44.51 | 202.14 | 232.92 | 0.9749 |
| 17 | 39.42 | 44.61 | 202.79 | 234.80 | 0.9759 |
| 18 | 39.44 | 44.69 | 203.24 | 236.19 | 0.9766 |
| 19 | 39.46 | 44.75 | 203.55 | 237.21 | 0.9771 |
| 20 | 39.47 | 44.79 | 203.76 | 237.96 | 0.9775 |

Table 6.10. Likelihood multiplier settings for all model cases.

| Fishery |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| catch | size <br> compositions | biomass | size <br> compositions | age <br> compositions | early | ordinary | late | Recruitment |
| :---: |
| 20 |

Table 6.11. Initial parameter values. Subscripts for recruitment deviations ( $\tau$ ) run from 1965 to 2011, with the subscript increasing moving across, then down. Subscripts for fishing mortality deviations ( $\varepsilon$ ) run from 1982 to 2011 in the same manner.

| Recruitment |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\overline{\operatorname{lom} R_{0}}$ | 17 |  |  |  |  |  |  |  |  |  |
| $\tau_{t}$ |  | 1965-2011: |  |  |  | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 |  |  |  |  |  |  |  |  |  |
| Fishing mortality |  |  |  |  |  |  |  |  |  |  |
| $\overline{\ln F}$ | 0 |  |  |  |  |  |  |  |  |  |
| $\varepsilon_{t}$ | 1982-2011: |  | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 |  |  |  |  |  |  |  |  |  |
| Fishery Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.4 | 0.4 |  |  |  |  |  |  |  |  |
| $\mathbf{A}_{50}$ | 5 | 5 |  |  |  |  |  |  |  |  |
| scale par. | -- | 0 |  |  |  |  |  |  |  |  |
| Survey Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.8 | 0.4 |  |  |  |  |  |  |  |  |
| $\mathbf{A}_{50}$ | 4 | 4 |  |  |  |  |  |  |  |  |
| scale par. | -- | 0 |  |  |  |  |  |  |  |  |

Table 6.12. Likelihood components for the model.

| likelihood component | Value |
| :---: | :---: |
| ordinary recruitment | -6.52 |
| 'late" recruitment | -0.01 |
| "early" recruitment | -2.24 |
| fishery catch | -0.16 |
| fishery size | -707.06 |
| survey biomass | -9.81 |
| survey size composition | -21.48 |
| survey age | -273.51 |

Table 6.13. Final parameter estimates. Subscripts for recruitment deviations ( $\tau$ ) run from 1965 to 2011, with the subscript increasing moving across, then down. Subscripts for fishing mortality deviations ( $\varepsilon$ ) run from 1982 to 2011 in the same manner.

| Recruitment |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\overline{\tau_{t}} \overline{\tau_{0}}$ | 16.98683 |  |  |  |  |  |  |  |  |  |
|  |  |  | 1965-2011: |  | -1.0525 | -0.4049 | -0.4453 | -0.3883 | -0.1999 | -0.2549 |
|  | 0.0714 | -0.1345 | -0.2511 | -0.1805 | -0.1926 | -0.3289 | -0.2538 | -0.0304 | 0.0149 | -0.1315 |
|  | -0.4037 | 0.0070 | -0.0226 | -0.4995 | 0.1863 | 0.2457 | 0.3221 | 0.4078 | 0.2712 | 0.2774 |
|  | 0.0745 | -0.2610 | -0.1183 | -0.4342 | -0.3105 | -0.5325 | -0.0956 | 0.1433 | 0.3429 | 0.6783 |
|  | 0.5575 | 0.5479 | 0.4983 | 0.2975 | -0.0787 | 0.5855 | 0.4343 | 0.8780 | 0.0454 | 0.0830 |
|  | 0.0355 |  |  |  |  |  |  |  |  |  |
| Fishing mortality |  |  |  |  |  |  |  |  |  |  |
| $\overline{\ln F}$ | -2.694578 |  |  |  |  |  |  |  |  |  |
| $\varepsilon_{t}$ | 1982-2011: | -0.6254 | -1.0893 | -1.5621 | -2.4128 | -2.9254 | -0.5241 | -0.4777 | -1.1408 | -0.4166 |
|  | 0.9129 | 0.5208 | 0.4421 | 0.5939 | 0.6484 | 0.9769 | 0.4784 | 0.2955 | 0.4394 | 0.6398 |
|  | 0.5007 | 0.5522 | 0.8092 | 0.0068 | 0.3390 | 0.6638 | 0.4225 | 0.2673 | 0.7918 | 0.4995 |
|  | 0.3734 |  |  |  |  |  |  |  |  |  |


| Fishery Selectivity |  |  |
| :--- | :---: | :---: |
| females |  |  |$\quad$ males


| Survey Selectivity |  |  |
| :--- | :---: | :---: |
|  | females | males |
| slope | 1.3461 | 1.9828 |
| $\mathbf{A}_{\mathbf{5 0}}$ | 4.12 | 3.65 |
| scale par. | -- | 0.0000 |

Table 6.14. Model-estimated fishery catch and survey biomass.

| year | catch (t) |  |  | survey biomass (t) |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | estimate | std dev | observed | estimate |  |  |
| std dev | observed |  |  |  |  |  |
| 1982 | 1,015 | 159 | 959 | 70,758 | 3,776 |  |
| 1983 | 636 | 99 | 595 | 70,830 | 3,639 |  |
| 1984 | 400 | 62 | 365 | 71,526 | 3,525 | 60,670 |
| 1985 | 175 | 27 | 154 | 72,153 | 3,403 |  |
| 1986 | 108 | 17 | 93 | 73,404 | 3,276 |  |
| 1987 | 1,205 | 188 | 1,151 | 76,098 | 3,181 | 63,826 |
| 1988 | 1,247 | 194 | 1,192 | 79,026 | 3,127 |  |
| 1989 | 641 | 99 | 599 | 82,888 | 3,107 |  |
| 1990 | 1,326 | 207 | 1,269 | 87,474 | 3,109 | 98,225 |
| 1991 | 4,762 | 750 | 4,636 | 90,727 | 3,107 |  |
| 1992 | 3,053 | 478 | 3,000 | 89,477 | 3,067 |  |
| 1993 | 2,855 | 433 | 3,000 | 88,399 | 2,983 | 86,911 |
| 1994 | 3,422 | 514 | 3,673 | 86,082 | 2,884 |  |
| 1995 | 3,704 | 553 | 4,021 | 82,223 | 2,770 |  |
| 1996 | 5,112 | 742 | 5,874 | 77,501 | 2,647 | 72,757 |
| 1997 | 3,060 | 458 | 3,294 | 71,392 | 2,545 |  |
| 1998 | 2,539 | 383 | 2,669 | 68,246 | 2,456 |  |
| 1999 | 2,847 | 426 | 3,060 | 6,443 | 2,416 | 74,969 |
| 2000 | 3,268 | 484 | 3,591 | 68,872 | 2,448 |  |
| 2001 | 2,650 | 391 | 2,940 | 72,891 | 2,558 | 71,326 |
| 2002 | 2,622 | 385 | 2,941 | 79,576 | 2,746 |  |
| 2003 | 3,197 | 477 | 3,485 | 86,652 | 2,961 | 99,897 |
| 2004 | 1,437 | 219 | 1,464 | 92,469 | 3,193 |  |
| 2005 | 2,159 | 333 | 2,176 | 98,157 | 3,397 | 101,255 |
| 2006 | 3,271 | 507 | 3,294 | 100,740 | 3,555 |  |
| 2007 | 2,865 | 446 | 2,852 | 101,700 | 3,684 | 103,776 |
| 2008 | 2,747 | 429 | 2,703 | 104,290 | 3,885 |  |
| 2009 | 4,945 | 791 | 4,753 | 108,390 | 4,249 | 124,744 |
| 2010 | 3,783 | 601 | 3,636 | 109,550 | 4,839 |  |
| 2011 | 3,389 | 523 | 2,594 | 109,150 | 5,207 | 95,134 |

Table 6.15. Total (age $3^{+}$) population biomass estimated in this year's model, compared with estimates from previous assessments.

| year | Age 3+ Biomass (1000's t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\mathbf{2 0 1 1}$ Assessment |  | $\mathbf{2 0 0 9}$ Assessment |  | $\mathbf{2 0 0 7}$ Assessment |  |
| estimate | std dev | estimate | std dev | estimate | std dev |  |
| 1982 | 76.5 | 3.9 | 79.0 | 4.0 | 74.6 | 3.8 |
| 1983 | 77.1 | 3.7 | 79.4 | 3.9 | 74.2 | 3.7 |
| 1984 | 77.1 | 3.6 | 79.2 | 3.7 | 73.2 | 3.6 |
| 1985 | 78.4 | 3.5 | 80.1 | 3.6 | 73.0 | 3.4 |
| 1986 | 80.9 | 3.4 | 82.7 | 3.5 | 73.8 | 3.3 |
| 1987 | 84.6 | 3.3 | 86.5 | 3.4 | 76.4 | 3.3 |
| 1988 | 88.4 | 3.2 | 90.3 | 3.3 | 81.6 | 3.3 |
| 1989 | 92.2 | 3.2 | 94.0 | 3.4 | 87.7 | 3.3 |
| 1990 | 96.6 | 3.2 | 98.1 | 3.4 | 94.5 | 3.4 |
| 1991 | 99.1 | 3.2 | 100.4 | 3.3 | 99.1 | 3.5 |
| 1992 | 96.4 | 3.2 | 97.4 | 3.3 | 97.4 | 3.5 |
| 1993 | 94.5 | 3.1 | 95.3 | 3.2 | 95.2 | 3.4 |
| 1994 | 91.4 | 3.0 | 92.2 | 3.1 | 91.7 | 3.3 |
| 1995 | 87.2 | 2.9 | 87.8 | 2.9 | 86.0 | 3.1 |
| 1996 | 82.0 | 2.8 | 82.5 | 2.8 | 79.4 | 3.0 |
| 1997 | 76.4 | 2.7 | 77.0 | 2.7 | 74.4 | 2.8 |
| 1998 | 74.5 | 2.6 | 76.1 | 2.7 | 72.2 | 2.7 |
| 1999 | 75.3 | 2.6 | 78.3 | 2.8 | 72.9 | 2.8 |
| 2000 | 79.2 | 2.7 | 82.5 | 2.9 | 73.7 | 2.9 |
| 2001 | 84.5 | 2.8 | 87.6 | 3.1 | 79.1 | 3.1 |
| 2002 | 91.3 | 3.1 | 94.0 | 3.3 | 85.6 | 3.4 |
| 2003 | 98.1 | 3.3 | 100.6 | 3.6 | 95.3 | 3.8 |
| 2004 | 103.0 | 3.6 | 104.9 | 4.0 | 101.5 | 4.4 |
| 2005 | 106.7 | 3.8 | 108.6 | 4.3 | 106.9 | 4.9 |
| 2006 | 110.1 | 4.0 | 113.8 | 4.8 | 108.7 | 5.2 |
| 2007 | 112.0 | 4.2 | 116.1 | 5.3 | 108.2 | 5.4 |
| 2008 | 116.9 | 4.6 | 117.6 | 5.7 |  |  |
| 2009 | 119.7 | 5.1 | 117.9 | 6.0 |  |  |
| 2010 | 118.6 | 5.6 |  |  |  |  |
| 2011 | 116.9 | 5.7 |  |  |  |  |

Table 6.16. Female spawning biomass estimated in this year's model, compared with estimates from previous assessments.

| year | Female Spawning Stock Biomass (1000's t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2011 Assessment estimate std dev |  | 2009 Assessment |  | 2007 Assessment |  |
| 1982 | 34.8 | 1.9 | 36.0 | 2.0 | 34.2 | 1.9 |
| 1983 | 34.9 | 1.9 | 36.1 | 1.9 | 34.1 | 1.9 |
| 1984 | 35.1 | 1.8 | 36.2 | 1.9 | 34.1 | 1.8 |
| 1985 | 35.6 | 1.7 | 36.6 | 1.8 | 34.1 | 1.7 |
| 1986 | 36.0 | 1.7 | 37.0 | 1.8 | 34.1 | 1.7 |
| 1987 | 36.6 | 1.6 | 37.5 | 1.7 | 34.0 | 1.6 |
| 1988 | 37.3 | 1.6 | 38.0 | 1.6 | 33.8 | 1.5 |
| 1989 | 38.6 | 1.6 | 39.4 | 1.6 | 34.7 | 1.5 |
| 1990 | 40.8 | 1.5 | 41.7 | 1.6 | 37.4 | 1.5 |
| 1991 | 42.8 | 1.5 | 43.6 | 1.6 | 40.7 | 1.6 |
| 1992 | 42.6 | 1.5 | 43.3 | 1.6 | 42.1 | 1.6 |
| 1993 | 43.1 | 1.5 | 43.6 | 1.6 | 43.5 | 1.7 |
| 1994 | 42.8 | 1.5 | 43.2 | 1.5 | 43.6 | 1.7 |
| 1995 | 41.4 | 1.5 | 41.6 | 1.5 | 41.9 | 1.6 |
| 1996 | 39.4 | 1.4 | 39.5 | 1.4 | 39.2 | 1.5 |
| 1997 | 36.2 | 1.3 | 36.4 | 1.4 | 35.3 | 1.5 |
| 1998 | 34.3 | 1.3 | 34.3 | 1.3 | 32.6 | 1.4 |
| 1999 | 32.9 | 1.2 | 33.1 | 1.3 | 31.2 | 1.3 |
| 2000 | 32.2 | 1.2 | 32.8 | 1.3 | 30.7 | 1.3 |
| 2001 | 32.4 | 1.2 | 33.7 | 1.3 | 30.7 | 1.3 |
| 2002 | 34.6 | 1.3 | 36.2 | 1.4 | 31.9 | 1.4 |
| 2003 | 37.9 | 1.4 | 39.3 | 1.5 | 34.4 | 1.5 |
| 2004 | 41.1 | 1.5 | 42.2 | 1.6 | 38.0 | 1.6 |
| 2005 | 45.1 | 1.6 | 46.0 | 1.8 | 43.3 | 1.9 |
| 2006 | 47.8 | 1.7 | 48.6 | 1.9 | 47.3 | 2.2 |
| 2007 | 48.7 | 1.8 | 49.5 | 2.1 | 48.8 | 2.4 |
| 2008 | 49.4 | 1.9 | 50.7 | 2.3 |  |  |
| 2009 | 50.6 | 2.0 | 52.3 | 2.6 |  |  |
| 2010 | 51.3 | 2.3 |  |  |  |  |
| 2011 | 52.6 | 2.6 |  |  |  |  |

Table 6.17. Age 3 recruitment (in millions) estimated in this year's model, compared with estimates from previous assessments.

| Year | $\mathbf{2 0 1 1}$ Assessment |  | $\mathbf{2 0 0 9}$ Assessment |  | 2007 Assessment |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | estimate | std dev | estimate | std dev | estimate | std dev |
| 1982 | 48.0 | 6.5 | 49.1 | 6.6 | 43.3 | 5.9 |
| 1983 | 46.6 | 6.5 | 46.5 | 6.6 | 37.0 | 5.3 |
| 1984 | 28.9 | 5.5 | 28.9 | 5.5 | 23.2 | 4.5 |
| 1985 | 57.4 | 8.5 | 54.8 | 8.5 | 44.4 | 7.0 |
| 1986 | 61.0 | 8.8 | 66.9 | 9.4 | 49.6 | 7.0 |
| 1987 | 65.8 | 9.4 | 67.1 | 9.8 | 64.7 | 8.5 |
| 1988 | 71.7 | 8.9 | 71.3 | 9.4 | 100.3 | 11.1 |
| 1989 | 62.5 | 7.0 | 61.5 | 7.3 | 72.2 | 8.8 |
| 1990 | 62.9 | 6.6 | 60.9 | 6.8 | 69.5 | 8.2 |
| 1991 | 51.4 | 5.9 | 53.4 | 6.5 | 53.5 | 7.0 |
| 1992 | 36.7 | 4.8 | 34.2 | 5.1 | 25.5 | 4.7 |
| 1993 | 42.4 | 5.2 | 42.4 | 5.8 | 31.1 | 5.7 |
| 1994 | 30.9 | 4.4 | 34.4 | 5.3 | 34.1 | 6.5 |
| 1995 | 35.0 | 4.6 | 31.1 | 5.0 | 19.5 | 4.9 |
| 1996 | 28.0 | 4.2 | 29.8 | 5.3 | 22.8 | 7.2 |
| 1997 | 43.3 | 5.3 | 46.2 | 6.9 | 68.3 | 19.8 |
| 1998 | 55.0 | 5.9 | 67.5 | 7.6 | 39.7 | 16.2 |
| 1999 | 67.2 | 6.7 | 78.5 | 7.6 | 67.1 | 13.1 |
| 2000 | 93.9 | 8.4 | 81.9 | 8.2 | 47.4 | 16.2 |
| 2001 | 83.3 | 8.0 | 79.4 | 8.4 | 120.7 | 20.4 |
| 2002 | 82.5 | 8.4 | 82.3 | 9.2 | 74.9 | 23.9 |
| 2003 | 78.5 | 8.4 | 83.1 | 9.8 | 120.0 | 21.8 |
| 2004 | 64.2 | 8.1 | 61.1 | 9.4 | 48.9 | 17.2 |
| 2005 | 44.1 | 7.1 | 47.5 | 9.3 | 48.1 | 5.8 |
| 2006 | 85.6 | 11.3 | 115.8 | 21.1 | 46.7 | 5.8 |
| 2007 | 73.6 | 11.8 | 59.3 | 7.4 | 58.7 | 7.1 |
| 2008 | 114.7 | 19.4 | 58.6 | 7.5 |  |  |
| 2009 | 49.9 | 6.1 | 57.6 | 7.2 |  |  |
| 2010 | 51.8 | 6.4 |  |  |  |  |
| 2011 | 49.4 | 6.1 |  |  |  |  |
|  |  |  |  |  |  |  |

Table 6.18. Prohibited species catch (PSC) in the rex sole target fishery.
a). Crab.

| year | PSC in target fishery (\#) |  |  |  |  | fraction of total PSC |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | King Crab |  |  | Tanner Crab |  | King Crab |  |  | Tanner Crab |  |
|  | Blue | Golden | Red | Bairdi | Opilio | Blue | Golden | Red | Bairdi | Opilio |
| 2003 | 0 | 0 | 0 | 28,780 | 0 | 0.0\% | 0.0\% | 0.0\% | 19.5\% | 0.0\% |
| 2004 | 0 | 0 | 0 | 9,014 | 0 | 0.0\% | 0.0\% | 0.0\% | 12.5\% | 0.0\% |
| 2005 | 0 | 0 | 0 | 7,949 | 0 | 0.0\% | 0.0\% | 0.0\% | 3.4\% | 0.0\% |
| 2006 | 0 | 0 | 0 | 73,530 | 0 | 0.0\% | 0.0\% | 0.0\% | 17.9\% | 0.0\% |
| 2007 | 0 | 0 | 0 | 45,272 | 0 | 0.0\% | 0.0\% | 0.0\% | 9.2\% | 0.0\% |
| 2008 | 0 | 0 | 0 | 48,204 | 0 | 0.0\% | 0.0\% | 0.0\% | 12.9\% | 0.0\% |
| 2009 | 0 | 54 | 0 | 140,364 | 0 | 0.0\% | 1.6\% | 0.0\% | 54.0\% | 0.0\% |
| 2010 | 0 | 0 | 0 | 14,266 | 0 | 0.0\% | 0.0\% | 0.0\% | 5.4\% | 0.0\% |
| 2011 | 0 | 0 | 0 | 6,102 | 0 | 0.0\% | 0.0\% | 0.0\% | 6.8\% | 0.0\% |

b). Halibut.

| Year | directed fishery <br> halibut PSC (kg) | \% total halibut <br> PSC |
| :---: | :---: | :---: |
| 2003 | 393,373 | $7.2 \%$ |
| 2004 | 304,274 | $5.0 \%$ |
| 2005 | 86,281 | $1.8 \%$ |
| 2006 | 208,398 | $3.7 \%$ |
| 2007 | 60,735 | $1.3 \%$ |
| 2008 | 173,430 | $2.9 \%$ |
| 2009 | 435,047 | $9.2 \%$ |
| 2010 | 387,523 | $8.6 \%$ |
| 2011 | 171,575 | $3.9 \%$ |

c). Salmon. (PSC for2011 unavailable at time of document preparation).

| Year | Chinook <br> fraction of <br> total |  | PSC (\#) | Non-Chinook <br> fraction of <br> total |
| :---: | :---: | :---: | :---: | :---: |
| 2003 | 2,900 | $18.3 \%$ | 520 | $4.9 \%$ |
| 2004 | 494 | $2.8 \%$ | 1,049 | $18.1 \%$ |
| 2005 | 525 | $2.4 \%$ | 98 | $1.5 \%$ |
| 2006 | 1,445 | $7.5 \%$ | 557 | $12.4 \%$ |
| 2007 | 715 | $1.8 \%$ | 663 | $19.0 \%$ |
| 2008 | 0 | $0.0 \%$ | 140 | $5.9 \%$ |
| 2009 | 1,909 | $24.2 \%$ | 413 | $16.2 \%$ |
| 2010 | 2,300 | $4.2 \%$ | 93 | $4.6 \%$ |

Table 6.19. Catch of nontarget species in the rex sole target fishery, expressed as the fraction of species catch by all fisheries in the FMP.

| Nontarget Species Group | 2011 | 2010 | 2009 | 2008 | $\begin{aligned} & \text { Year } \\ & 2007 \\ & \hline \end{aligned}$ | 2006 | 2005 | 2004 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Benthic urochordata | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 28.8\% | 0.3\% | 48.9\% |
| Birds | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Bivalves | 0.5\% | 0.0\% | 9.6\% | 9.9\% | 0.0\% | 0.0\% | 5.4\% | 8.4\% | 8.6\% |
| Brittle star unidentified | 4.0\% | 0.1\% | 15.1\% | 3.8\% | 7.1\% | 0.0\% | 0.2\% | 0.0\% | 0.0\% |
| Capelin | 0.0\% | 0.0\% | 51.0\% | 95.5\% | -- | 0.0\% | 0.0\% | 0.0\% | 17.3\% |
| Corals Bryozoans | 3.1\% | 10.3\% | 13.5\% | 0.0\% | 6.7\% | 0.0\% | 0.0\% | 0.0\% | 17.8\% |
| Dark Rockfish | 0.0\% | 0.0\% | 0.0\% | 0.0\% | -- | -- | -- | -- | -- |
| Eelpouts | 2.6\% | 9.8\% | 19.3\% | 0.0\% | 0.0\% | 0.0\% | 0.3\% | 0.5\% | 11.0\% |
| Eulachon | 0.0\% | 5.5\% | 11.5\% | 2.9\% | 4.4\% | 0.0\% | 1.9\% | 0.0\% | 9.9\% |
| Giant Grenadier | 3.6\% | 8.9\% | 21.5\% | 3.2\% | 5.2\% | 8.6\% | 3.6\% | 0.0\% | 0.0\% |
| Greenlings | 0.0\% | 8.4\% | 10.5\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 0.0\% |
| Grenadier | 11.2\% | 0.0\% | 0.0\% | 4.0\% | 0.0\% | 0.0\% | 0.0\% | 0.4\% | 7.8\% |
| Gunnels | -- | -- | -- | 12.9\% | -- | 0.0\% | -- | -- | 0.0\% |
| Hermit crab unidentified | 4.9\% | 4.3\% | 11.7\% | 4.6\% | 5.8\% | 15.6\% | 4.8\% | 0.0\% | 10.2\% |
| Invertebrate unidentified | 0.0\% | 14.3\% | 17.0\% | 5.9\% | 0.5\% | 0.0\% | 0.0\% | 0.3\% | 9.0\% |
| Lanternfishes (myctophidae) | -- | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Large Sculpins | 1.8\% | 3.4\% | 7.9\% | 1.5\% | 3.2\% | 5.8\% | 3.1\% | 3.3\% | 7.8\% |
| Misc crabs | 0.2\% | 5.7\% | 14.2\% | 4.5\% | 5.6\% | 12.6\% | 3.9\% | 0.2\% | 8.3\% |
| Misc crustaceans | 10.0\% | 0.0\% | 0.0\% | -- | -- | -- | 64.7\% | 0.0\% | 65.1\% |
| Misc deep fish | -- | -- | -- | 0.0\% | -- | -- | -- | -- | -- |
| Misc fish | 2.1\% | 3.5\% | 8.5\% | 2.5\% | 3.4\% | 4.1\% | 1.9\% | 2.5\% | 5.7\% |
| Misc inverts (worms etc) | 50.5\% | 100.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | -- | -- |
| Octopus | 0.2\% | 0.0\% | 0.0\% | 0.0\% | 0.2\% | 0.0\% | 0.0\% | 0.0\% | 0.2\% |
| Other osmerids | 4.4\% | 7.7\% | 16.1\% | 4.2\% | 6.5\% | 0.0\% | 0.0\% | 0.0\% | 5.2\% |
| Other Sculpins | 4.1\% | 6.3\% | 11.2\% | 3.4\% | 4.1\% | 8.8\% | 4.3\% | 0.3\% | 7.6\% |
| Pacific Sand lance | 0.0\% | -- | 0.0\% | 0.0\% | -- | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Pandalid shrimp | 4.0\% | 6.4\% | 18.8\% | 4.3\% | 5.8\% | 8.7\% | 2.9\% | 0.0\% | 10.4\% |
| Polychaete unidentified | 0.0\% | 100.0\% | 100.0\% | 0.0\% | 0.0\% | -- | 40.4\% | -- | 0.0\% |
| Scypho jellies | 2.4\% | 3.6\% | 12.6\% | 0.0\% | 5.3\% | 0.0\% | 2.2\% | 0.0\% | 5.2\% |
| Sea anemone unidentified | 3.2\% | 3.7\% | 9.3\% | 0.5\% | 4.2\% | 0.0\% | 3.0\% | 4.5\% | 7.2\% |
| Sea pens whips | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.3\% | 16.3\% |
| Sea star | 2.6\% | 4.8\% | 9.5\% | 3.3\% | 4.7\% | 4.9\% | 3.0\% | 3.4\% | 7.7\% |
| Shark, Other | 2.0\% | 0.0\% | 0.5\% | 0.0\% | 0.1\% | 0.9\% | 0.6\% | 8.8\% | 4.1\% |
| Shark, Pacific sleeper | 0.4\% | 0.5\% | 1.8\% | 0.1\% | 0.1\% | 2.6\% | 1.1\% | 1.4\% | 3.7\% |
| Shark, salmon | 0.0\% | 0.2\% | 0.0\% | 0.0\% | 0.7\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% |
| Shark, spiny dogfish | 0.2\% | 2.0\% | 0.3\% | 0.0\% | 0.0\% | 0.6\% | 1.3\% | 0.8\% | 3.0\% |
| Skate, Alaska | 0.1\% | 0.1\% | -- | -- | -- | -- | -- | -- | -- |
| Skate, Aleutian | 0.6\% | -- | -- | -- | -- | -- | -- | -- | -- |
| Skate, Big | 3.3\% | 4.3\% | 7.0\% | 2.2\% | 5.8\% | 8.9\% | 4.7\% | 2.9\% | -- |
| Skate, Longnose | 4.0\% | 4.9\% | 7.4\% | 4.1\% | 6.1\% | 5.9\% | 2.2\% | 2.5\% | 0.0\% |
| Skate, Other | 3.2\% | 6.3\% | 11.4\% | 5.2\% | 8.9\% | 10.0\% | 6.4\% | 5.9\% | 10.0\% |
| Skate, Whiteblotched | 0.0\% | -- | -- | -- | -- | -- | -- | -- | -- |
| Snails | 3.2\% | 5.9\% | 9.5\% | 4.3\% | 4.6\% | 9.2\% | 3.8\% | 4.7\% | 8.0\% |
| Sponge unidentified | 3.0\% | 5.6\% | 12.4\% | 0.0\% | 10.0\% | 0.0\% | 6.1\% | 0.0\% | 9.7\% |
| Squid | 0.2\% | 0.6\% | 0.7\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.6\% |
| Stichaeidae | 0.0\% | 14.1\% | 21.8\% | 22.8\% | 13.7\% | 0.0\% | 17.5\% | 0.0\% | 34.7\% |
| Surf smelt | -- | -- | -- | 0.0\% | -- | -- | 0.0\% | 0.0\% | -- |
| urchins dollars cucumbers | 4.0\% | 7.7\% | 15.1\% | 3.7\% | 4.8\% | 0.0\% | 4.8\% | 0.4\% | 7.7\% |

Table 6.20. Catch of non-prohibited species in the rex sole target fishery. The species accounting for the two largest totals are highlighted.

| Species | 2011 |  | 2010 |  | 2009 |  | 2008 |  | 2007 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total <br> (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total <br> (t) | $\begin{gathered} \text { \% } \\ \text { retained } \end{gathered}$ |
| Atka mackerel | 4 | 99\% | 225 | 83\% | 225 | 83\% | 0 | 0\% | 1 | 89\% |
| arrowtooth flounder | 1,790 | 19\% | 5,628 | 10\% | 6,207 | 9\% | 2,501 | 12\% | 3,108 | 8\% |
| big skate | 106 | 84\% | 214 | 83\% | 264 | 85\% | 70 | 96\% | 74 | 99\% |
| deep water flatfish | 47 | 7\% | 269 | 7\% | 321 | 6\% | 227 | 3\% | 68 | 0\% |
| flathead sole | 178 | 94\% | 497 | 93\% | 629 | 94\% | 283 | 81\% | 264 | 92\% |
| longnose skate | 44 | 94\% | 76 | 93\% | 82 | 94\% | 36 | 97\% | 24 | 97\% |
| northern rockfish | 12 | 39\% | 37 | 38\% | 37 | 39\% | 12 | 0\% | 12 | 0\% |
| all sharks, squid, sculpin, octopus | -- | -- | 31 | 1\% | 36 | 2\% | 9 | 0\% | 15 | 0\% |
| Pacific cod | 155 | 87\% | 557 | 86\% | 592 | 85\% | 238 | 96\% | 409 | 88\% |
| pelagic rockfish complex | 11 | 78\% | 35 | 89\% | 42 | 91\% | 5 | 94\% | 31 | 94\% |
| pollock | 118 | 83\% | 550 | 70\% | 615 | 72\% | 70 | 95\% | 110 | 99\% |
| POP | 291 | 25\% | 399 | 34\% | 420 | 32\% | 76 | 2\% | 68 | 10\% |
| rex sole | 1,073 | 98\% | 3,142 | 99\% | 3,401 | 99\% | 1,091 | 98\% | 1,556 | 100\% |
| rougheye | 3 | 92\% | 10 | 27\% | 10 | 29\% | 14 | 41\% | 4 | 94\% |
| other rockfish | 1 | 37\% | 3 | 9\% | 3 | 9\% | 1 | 0\% | 0 | 0\% |
| sablefish | 29 | 91\% | 122 | 93\% | 125 | 93\% | 35 | 76\% | 42 | 83\% |
| shallow water flatfish | 11 | 93\% | 32 | 88\% | 46 | 92\% | 12 | 82\% | 10 | 100\% |
| shortraker | 9 | 78\% | 20 | 62\% | 21 | 62\% | 4 | 71\% | 4 | 92\% |
| thornyheads | 27 | 95\% | 52 | 99\% | 54 | 97\% | 29 | 100\% | 24 | 95\% |
| unidentified skates | 21 | 28\% | 50 | 66\% | 60 | 63\% | 22 | 56\% | 103 | 50\% |
| octopus | 0 | 8\% | -- | -- | -- | -- | -- | -- | -- | -- |
| sculpin | 3 | 6\% | -- | -- | -- | -- | -- | -- | -- | -- |
| USRK (???) | 6 | 0\% | -- | -- | -- | -- | -- | -- | -- | -- |

## Figures



Figure 6.1. Fishery catches for GOA rex sole, 1982-2011. Catch for 2011 is through Sept. 24.


Figure 6.2. Spatial patterns of fishery catches for GOA rex sole, 2009-2011.


Figure 6.3. Spatial patterns of observed fishery catches by for GOA rex sole from 2010 and the first three quarters of 2011.


Figure 6.4. GOA survey biomass for rex sole. Error bars represent 95\% lognormal confidence intervals. The 2001 GOA survey did not survey the Eastern Gulf. The value shown here for 2001 has been corrected to account for this (see text).


Figure 6.5. Spatial patterns of CPUE for rex sole in the GOA groundfish surveys for 2007, 2009 and 2011.
a) Length-at-age.

b) Weight-at-age.

c) Maturity-at-age (females).


Figure 6.6. Age-specific schedules for GOA rex sole: females solid line, males dotted line.


Figure 6.7. Estimated and observed annual catches for GOA rex sole for the assessment model. Estimated catch $=$ dotted line with circles, observed catch $=$ solid line.


Fig. 6.8. Estimated and observed survey biomass for GOA rex sole for the assessment model. Estimated survey biomass = triangles, observed survey biomass = circles (error bars are approximate lognormal 95\% confidence intervals).


Figure 6.9a. Fits to female GOA rex sole fishery size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 6.9b. Fits to male GOA rex sole fishery size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 6.10a. Fit to the female GOA rex sole survey size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 6.10b. Fits to the male GOA rex sole survey size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 6.11a. Fits to the female survey GOA rex sole age composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure6.11b. Fit to the male survey GOA rex sole age composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 6.12. Estimated selectivity functions. Survey selectivities are plotted in red with a dotted line, fishery selectivities are plotted in black with a solid line. Male selectivity functions are plotted with a triangle symbol, female selectivity functions are plotted without a symbol.


Figure 6.13. Marginal posterior distributions based on MCMC integration for parameters related to fishery and survey age selectivity functions. " $\mathrm{a}_{50}$ " denotes the parameter for the age at which the logistic selectivity function is $50 \%$. " $\beta$ " is related to the slope of the selectivity function at age $=a_{50}$. The maximum likelihood estimate for each parameter is indicated by the appropriately-colored vertical line.


Figure 6.14. Marginal posterior distributions based on MCMC integration for: median recruitment (upper left), median fishing mortality (upper right), total (age 3+) biomass in 2011 (lower left), spawning biomass in 2011 (lower middle), and recruitment in 2011 (lower right). The maximum likelihood estimate for each quantity is indicated by the vertical line.


Figure 6.15. Time series plots of estimated total (age $3+$ ) biomass and spawning biomass (upper graph) and recruitment (lower graph). 99\% credibility intervals based on marginal posterior distributions from MCMC integration for parameters related to fishery (top row) and survey (bottom row) age selectivity functions. The solid lines indicate time series of maximum likelihood estimates.



Figure 6.16. Upper: : Comparison of total (age 3+) biomass estimates from the current assessment with results from the 2009 and 2007 assessments. Lower: Comparison of spawning biomass estimates from the current assessment with results from the 2009 and 2007 assessments.


Figure 6.17. Upper: Estimated age 3 recruitments of GOA rex sole with approximate $95 \%$ lognormal confidence intervals estimated from the model hessian. Horizontal line is mean recruitment. Lower: Comparison of recruitment estimates from the current assessment with results from the 2009 and 2007 assessments.


Figure 6.18. Marginal posterior distributions based on MCMC integration for several managementrelated quantities estimated in the model: $\mathrm{F}_{40 \%}$ and $\mathrm{F}_{35 \%}$,(upper left), $\mathrm{B}_{40 \%}$, and $\mathrm{B}_{35 \%}$ (upper right), and example Tier 3-based estimates of 2012 ABC, OFL (lower left). The value for each quantity associated with the maximum likelihood solution is indicated by the appropriately-colored vertical line. Note: $\mathrm{F}_{40 \%}$ and $\mathrm{F}_{35 \%}$ are not considered to be estimated reliably and so ABCs and OFLs for GOA rex sole are based on Tier 5 calculations, not those shown here.


Figure 6.19. Gulf of Alaska food web from the GOA ecosystem model (Aydin et al., 2007) highlighting rex sole links to predators (blue boxes and lines) and prey (green boxes and lines). Box size reflects relative standing stock biomass.


Figure 6.20. Diet composition for Gulf of Alaska rex sole from the GOA ecosystem model (Aydin et al., 2007).


Figure 6.21. Decomposition of natural mortality for Gulf of Alaska rex sole from the GOA ecosystem model (Aydin et al., 2007).

## Chapter 6 Appendix A: Model Equations

Table A.1. List of quantities and their definitions as used in the model.

| Quantity | Definition |
| :---: | :---: |
| T | number of years in the model. |
| A | number of age classes (18). |
| $L$ | number of length classes (29). |
| $T_{\text {min }}$ | model start year (1982). |
| $T_{\text {max }}$ | assessment year (2011). |
| $t$ | time index. |
| $a$ | age index ( $1 \leq a \leq A ; a=1$ corresponds to age at recruitment). |
| $x$ | sex index ( $1 \leq x \leq 2 ; 1=$ female, $2=$ male $)$. |
| l | length index ( $1 \leq l \leq L$; $l=1$ corresponds to minimum length class). |
| $\left\{t^{S}\right\}$ | set of years for which survey biomass data is available. |
| $\left\{t^{F, A}\right\}$ | set of years for which fishery age composition data is available. |
| $\left\{t^{F, L}\right\}$ | set of years for which fishery length composition data is available. |
| $\left\{t^{S, A}\right\}$ | set of years for which survey age composition data is available. |
| $\left\{t^{S, L}\right\}$ | set of years for which survey length composition data is available. |
| $L^{\chi}{ }_{l, a}$ | elements of length-age conversion matrix (proportion of $\operatorname{sex} x$ fish in age class $a$ that are in length class $l$ ). (fixed) |
| $w_{x, a}$ | mean body weight (kg) of sex $x$ fish in age group $a$. (fixed) |
| $\phi_{a}$ | proportion of females mature at age $a$. (fixed) |
| $\overline{\ln R_{0}}$ | mean value of log-transformed recruitment. (estimable) |
| $\tau_{t}$ | recruitment deviation in year $t$. (estimable) |
| $M_{x}$ | instantaneous natural mortality rate. (fixed) |
| $\overline{\ln F}$ | mean value of log-transformed fishing mortality. (estimable) |
| $\varepsilon_{t}$ | deviations in fishing mortality rate in year $t$. (estimable) |
| $R_{t}$ | recruitment in year $t$. |
| $N_{t, x, a}$ | number of fish of sex $x$ and age class $a$ in year $t$. |
| $C_{t, x, a}$ | catch (number) of fish of sex $x$ and age class $a$ in year $t$. |
| $p^{F, A}{ }_{t, x, a}$ | proportion of the total catch in year $t$ that is sex $x$ and in age class $a$. |
| $p^{F, L}{ }_{t, x, l}$ | proportion of the total catch in year $t$ that is sex $x$ and in length class $l$. |
| $p^{S, A}{ }_{\text {ct, }}{ }^{\text {ct,a }}$ | proportion of the survey biomass in year $t$ that is sex $x$ and in age group a. |
| $p^{s, L}{ }_{t, x, l}$ | proportion of the survey biomass in year $t$ that is sex $x$ and in age group a. |
| $C_{t}$ | total catch (yield) in tons in year $t$. |
| $F_{t, x, a}$ | instantaneous fishing mortality rate for sex $x$ and age group $a$ in year $t$. |
| $Z_{t, x, a}$ | instantaneous total mortality for sex $x$ and age group $a$ in year $t$. |
| $s^{F U}{ }^{\text {x,a }}$ | unnormalized fishery selectivity for sex $x$ and age group $a$. |
| $s^{S U}{ }_{\chi, a}$ | unnormalized survey selectivity for sex $x$ and age group $a$. |
| $s^{F N}{ }^{\text {x,a }}$ | normalized fishery selectivity for sex $x$ and age group $a$. |
| $s^{S N}{ }_{x, a}$ | normalized survey selectivity for sex $x$ and age group $a$. |

Table A.2. Model equations describing the model populations dynamics.

| Equation | Description |
| :---: | :---: |
| $\tau_{t} \sim N\left(0, \sigma_{R}^{2}\right)$ | Random deviate associated with recruitment. |
| $N_{t, x, 1}=R_{t}=\exp \left(\overline{\overline{\ln R_{0}}}+\tau_{t}\right)$ | Recruitment (assumed equal for males and females). |
| $N_{t+1, x, a+1}=N_{t, x, a} e^{-Z_{t, x, a}}$ | Numbers at age. |
| $N_{t+1, x, A}=N_{t, x, A-1} e^{-Z_{t, x, A-1}}+N_{t, x, A} e^{-Z_{t, x, A}}$ | Numbers in "plus" group. |
| $C_{t, x, a}=\frac{F_{t, x, a}}{Z_{t, x, a}}\left(1-e^{-z_{t, x, a}}\right) N_{t, x, a}$ | Catch at age (in numbers caught). |
| $C_{t}=\sum_{x=1}^{2} \sum_{a=1}^{A} w_{x, a} C_{t, x, a}$ | Total catch in tons (i.e., yield). |
| $F S B_{t}=\sum_{a=1}^{A} w_{1, a} \phi_{a} N_{t, 1, a}$ | Female spawning biomass. |
| $Z_{t, x, a}=F_{t, x, a}+M$ | Total mortality. |
| $F_{t, x, a}=s_{x, a}^{F} \cdot \exp \left(\overline{\ln F}+\varepsilon_{t}\right)$ | Fishing mortality. |
| $\varepsilon_{t} \sim N\left(0, \sigma_{F}^{2}\right)$ | Random deviate associated with fishing mortality. |
| $s_{x, a}^{F U}=\frac{1}{1+e^{\left(-b_{x}^{F}\left(a g e-50 A_{x}^{F}\right)\right)}}$ | Unnormalized fishery selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x, a}^{S U}=\frac{1}{1+e^{\left(-b_{x}^{S}\left(a g e-500_{x}^{S}\right)\right)}}$ | Unnormalized survey selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x, a}^{F N}=\exp \left(r_{x}^{F}\right) \frac{s_{x, a}^{F U}}{\max \left\{s_{1, a}^{F U}\right\}}$ | Normalized fishery selectivity. $r^{F}{ }_{1} \equiv 0$. |
| $s_{x, a}^{S N}=\exp \left(r_{x}^{S}\right) \frac{s_{x, a}^{S U}}{\max \left\{s_{1, a}^{S U}\right\}}$ | Normalized survey selectivity. $r^{\text {S }} \equiv 0$. |
| $N^{S}{ }_{t, x, a}=Q s_{x, a}^{S} N_{t, x, a}$ | Survey numbers for sex $x$, age $a$ at time $t$. |
| $S B_{t}=\sum_{x=1}^{2} \sum_{a=1}^{A} w_{x, a} N^{S}{ }_{t, x, a}$ | Total survey biomass. |
| $p_{t, x, a}^{F, A}=C_{t, x, a} / \sum_{x=1}^{2} \sum_{a=1}^{A} C_{t, x, a}$ | Proportion at age in the catch. |
| $p_{t, x, l}^{F, L}=\sum_{a=1}^{A} L_{l, a}^{X} \cdot p_{t, x, a}^{F, A}$ | Proportion at length in the catch. |
| $p_{t, x, a}^{S, A}=N_{t, x, a}^{S} / \sum_{x=1}^{2} \sum_{a=1}^{A} N_{t, x, a}^{S}$ | Proportion at age in the survey. |
| $p_{t, x, l}^{S, L}=\sum_{a=1}^{A} L_{l, a}^{X} \cdot p_{t, x, a}^{S, A}$ | Proportion at length in the survey. |

Table A.3. Likelihood components.

| Component | Description |
| :---: | :---: |
| $\sum_{t=1}^{T}\left[\log \left(C_{t}^{\text {obs }}\right)-\log \left(C_{t}\right)\right]^{2}$ | Catch; assumes a lognormal distribution. |
| $\sum_{t \in\left\{\left\{^{F, A}\right\}\right.} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{F, A, o b s} \cdot \log \left(p_{t, x, a}^{F, A}\right)-\text { offset }$ | Fishery age composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{\left\{^{F, L}\right.\right.} \sum_{x=1}^{2} \sum_{l=1}^{L} n_{t}^{\text {samp }} \cdot p_{t, x, l}^{F, L, o b s} \cdot \log \left(p_{t, x, x}^{F, L}\right)-\text { offset }$ | Fishery length composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{^{E}, a,\right\}} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{S, A, o b s} \cdot \log \left(p_{t, x, a}^{S, A}\right)-\text { offset }$ | Survey age composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{\left\{^{F, L},\right.\right.} \sum_{x=1}^{2} \sum_{l=1}^{L} n_{t}^{\text {samp }} \cdot p_{t, x, l}^{s, L, o b s} \cdot \log \left(p_{t, x, l}^{s, L}\right) \text { - offset }$ | Survey length composition; uses a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\text { offset } \left.=\sum_{t} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{o b s} \cdot \log \left(p_{t, x, a}^{o b s}\right)\right)$ | The offset constants for age composition components are calculated from the observed proportions and the sample sizes. A similar formula is used for length composition component offsets. |
| $\sum_{t \in\left\{t^{s}\right\}}\left[\frac{\log \left[\frac{S B_{t}^{\text {obs }}}{S B_{t}}\right]}{\sqrt{2} \cdot \text { s.d. }\left(\log \left(S B_{t}^{\text {obs }}\right)\right)}\right]^{2}$ | Survey biomass; assumes a lognormal distribution. |
| $\sum_{t=T_{\min }}^{T_{\text {max }}-3}\left(\tau_{t}\right)^{2}$ | Recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a log scale. |
| $\sum_{t=T_{\max }-2}^{T_{\text {max }}}\left(\tau_{t}\right)^{2}$ | "Late" recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a $\log$ scale. |
| $\sum_{t=T_{\min }-A+1}^{T_{\min }-1}\left(\tau_{t}\right)^{2}$ | "Early" recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a log scale. <br> Determines age composition at starting year of model. |

Table A.4. Parameters fixed in the model.

| Parameter | Description |
| :--- | :--- |
| $M_{x}=0.17$ | sex-specific natural mortality rate. |
| $Q=1.0$ | survey catchability. |
| $L_{l, a}^{x}$ | sex-specific length-at-age conversion matrix. |
| $w_{x, a}$ | sex-specific weight-at-age. |
| $\phi_{a}$ | proportion of females mature at age $a$. |

Table A.5. Parameters estimated in the accepted model. A total of 87 population parameters were estimated.

| Parameter | Subscript <br> range | Total no. of <br> Parameters | Description |
| :--- | :---: | :---: | :--- |
| $\ln \left(R_{0}\right)$ | NA | 1 | natural log of the geometric mean <br> value of age 3 recruitment. |
| $\tau_{t}$ | $T_{\min }-A+1 \leq t \leq T_{\max }$ | 47 | log-scale recruitment deviation in <br> year $t$. |
| $\ln \left(f_{0}\right)$ | NA | 1 | natural log of the geometric mean <br> value of fishing mortality. |
| $\varepsilon_{t}$ | $T_{\min } \leq t \leq T_{\max }$ | 30 | log-scale deviations in fishing <br> mortality rate in year $t$. |
| $r^{F}{ }_{2}$ | NA | not estimated | scaling from female to male fishery <br> selectivity (log-scale). |
| $b^{F}{ }_{x},{ }_{50} \mathrm{~A}^{F}{ }_{x}$ | $1 \leq x \leq 2$ | 4 | sex-specific selectivity parameters <br> (slope and age at $50 \%$ selected) for <br> the fishery. |
| $r^{S}$ | $S=1$ | not estimated | scaling from female to male survey <br> selectivity (log-scale). |
| $b^{S}{ }_{x},{ }_{50} \mathrm{~A}_{x}^{S}$ | $1 \leq x \leq 2$ <br> $S=1$ | 4 | sex-specific selectivity parameters <br> (slope and age at $50 \%$ selected) for <br> the survey. |

## Chapter 6 Appendix B: Supplemental Catch Data

In order to comply with the Annual Catch Limit (ACL) requirements, two new datasets have been generated to help estimate total catch and removals from NMFS stocks in Alaska.

The first dataset, non-commercial removals, estimates total removals that do not occur during directed groundfish fishing activities (Table 6B.1). This includes removals incurred during research, subsistence, personal use, recreational, and exempted fishing permit activities, but does not include removals taken in fisheries other than those managed under the groundfish FMP. These estimates represent additional sources of removals to the existing Catch Accounting System estimates. For the GOA rex sole stock, these estimates (currently available only for 2010) can be compared to research removals that have occurred in conjunction with the Gulf of Alaska Groundfish Surveys (Table 6B.2). Compared with the 2010 ABC ( $9,729 \mathrm{t}$ ), these non-commercial catches are miniscule ( $<0.2 \% \mathrm{ABC}$ ) and do not present a risk to the GOA rex sole stock.

The second dataset, the Halibut Fishery Incidental Catch Estimation (HFICE), is an estimate of the incidental catch of groundfish in the halibut IFQ fishery in Alaska, which is currently unobserved. To estimate removals in the halibut fishery, methods were developed by the HFICE working group and approved by the Gulf of Alaska and Bering Sea/Aleutian Islands Plan Teams and the Scientific and Statistical Committee of the North Pacific Fishery Management Council. A detailed description of the methods is available in Tribuzio et al. (2011).

These estimates are for total catch of groundfish species in the halibut IFQ fishery and do not distinguish between "retained" or "discarded" catch. These estimates should be considered a separate time series from the current CAS estimates of total catch. Because of potential overlaps HFICE removals should not be added to the CAS produced catch estimates. The overlap will apply when groundfish are retained or discarded during an IFQ halibut trip. IFQ halibut landings that also include landed groundfish are recorded as retained in eLandings and a discard amount for all groundfish is estimated for such landings in CAS. Discard amounts for groundfish are not currently estimated for IFQ halibut landings that do not also include landed groundfish. For example, catch information for a trip that includes both landed IFQ halibut and sablefish would contain the total amount of sablefish landed (reported in eLandings) and an estimate of discard based on at-sea observer information. Further, because a groundfish species was landed during the trip, catch accounting would also estimate discard for all groundfish species based on available observer information and following methods described in Cahalan et al. (2010). The HFICE method estimates all groundfish caught during a halibut IFQ trip and thus is an estimate of groundfish caught whether landed or discarded. This prevents simply adding the CAS total with the HFICE estimate because it would be analogous to counting both retained and discarded groundfish species twice. Further, there are situations where the HFICE estimate includes groundfish caught in State waters and this would need to be considered with respect to ACLs (e.g. Chatham Strait sablefish fisheries), although the extent to which this occurs for rex sole is unknown. Therefore, the HFICE estimates should be considered preliminary estimates for what is caught in the IFQ halibut fishery. Improved estimates of groundfish catch in the halibut fishery may become available following restructuring of the Observer Program in 2013.

The HFICE estimates of rex sole catch by the halibut fishery in the Gulf of Alaska are miniscule compared with recent ABC's for the GOA stock (Table 6B.3). Based on these values, the risk to the stock from the halibut IFQ fishery is nil.

## References:

Cahalan J., J. Mondragon., and J. Gasper. 2010. Catch Sampling and Estimation in the Federal Groundfish Fisheries off Alaska. NOAA Technical Memorandum NMFS-AFSC-205. 42 p.
Tribuzio, CA, S Gaichas, J Gasper, H Gilroy, T Kong, O Ormseth, J Cahalan, J DiCosimo, M Furuness, H Shen, K Green. 2011. Methods for the estimation of non-target species catch in the unobserved halibut IFQ fleet. August Plan Team document. Presented to the Joint Plan Teams of the North Pacific Fishery Management Council.

## Tables

Table 6B.1. Non-commercial use catches of rex sole in the Gulf of Alaska for 2010. Non-commercial use includes catches for research, recreation, subsistence, personal use and exempted fishing permits. The ABC for 2010 was $9,729 \mathrm{t}$.

| Source | Rex Sole (t) |
| :--- | :---: |
| 2010 Shelikof Acoustic Survey | 0.0 |
| 2010 Shumigans Acoustic Survey | 0.0 |
| large-mesh trawl | 5.5 |
| Scallop dredge | 0.0 |
| small-mesh trawl | 0.3 |
| Grand Total | $\mathbf{5 . 8}$ |

Table 6B.2. Research catches from the Gulf of Alaska Groundfish Surveys. The ABC for 2011 was 9,565 t.

| year | Research <br> Catch (t) |
| ---: | ---: |
| 1984 | 13.58 |
| 1987 | 17.02 |
| 1990 | 11.99 |
| 1993 | 12.53 |
| 1996 | 6.02 |
| 1999 | 4.73 |
| 2001 | 3.07 |
| 2003 | 6.39 |
| 2005 | 7.78 |
| 2007 | 8.52 |
| 2009 | 9.31 |
| 2011 | 5.77 |

Table 6B.3. HFICE estimated catches of rex sole in the Gulf of Alaska by the halibut fishery. The ABC for the GOA rex sole fishery is also listed for each year. The ABC for 2011 was $9,565 \mathrm{t}$.

|  | Rex sole (t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | ---: |
| Year | Western Gulf | Central Gulf | West Yakutat | Southeast | Total | ABC |
| 2001 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9,440 |
| 2002 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9,470 |
| 2003 | 0.0 | 1.6 | 0.0 | 0.0 | 1.6 | 9,470 |
| 2004 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12,650 |
| 2005 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 12,650 |
| 2006 | 0.1 | 0.0 | 0.0 | 0.0 | 0.1 | 9,200 |
| 2007 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9,100 |
| 2008 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9,132 |
| 2009 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 8,996 |
| 2010 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 9,729 |


[^0]:    $\begin{array}{llllllllllllllllllllllllllllllllll}0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0000 & 0.0005 & 0.0008 & 0.0041 & 0.0078 & 0.0228 & 0.0347 & 0.0639 & 0.0753 & 0.0734 & 0.0574 & 0.0318 & 0.0103 & 0.0063 & 0.0022 & 0.0025\end{array}$

