## 8. Assessment of the Flathead Sole Stock in the Gulf of Alaska

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## Executive Summary

## Changes in the Input Data

1) The fishery catch and length compositions for 2010 and 2011 (through Sept. 24, 2011) were incorporated in the model.
2) The 2009 fishery catch and length compositions were updated.
3) The 2011 GOA groundfish survey biomass estimate and length composition data were added to the model. Survey biomass increased from 225,377 t in 2009 to 235,639 t in 2011. Survey biomass estimates and length compositions were recalculated for all survey years.
4) Age compositions from the 2001 and 2009 groundfish surveys were added to the model.

## Changes in the Assessment Model

No changes were made to the structure of the assessment model.

## Changes in the Assessment Results

1. Based on the assessment model, the recommended Tier 3 a ABC, using an $\mathrm{F}_{40 \%}$ harvest level of 0.450 , is $47,407 \mathrm{t}$ for 2012 and $48,081 \mathrm{t}$ for 2013.
2. The OFL, based on an $F_{35 \%}$ harvest level of 0.593 , is $59,380 t$ for 2012 and 60,219 t for 2013.
3. Projected female spawning biomass is estimated at $104,301 \mathrm{t}$ for 2012 and $105,127 \mathrm{t}$ for 2013.
4. Total biomass (age 3+) is estimated at 292,189 t for 2012 and 286,274 t for 2013.

The area apportionments corresponding to the recommended ABCs from the assessment model are:

|  |  | West |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Southeast |  |  |  |  |  |
|  | Western | Central | Yakutat | Outside | Total |
| Area Apportionment | $32.3 \%$ | $54.5 \%$ | $9.6 \%$ | $3.6 \%$ | $100.0 \%$ |
| 2012 ABC $(\mathrm{t})$ | 15,300 | 25,838 | 4,558 | 1,711 | 47,407 |
| 2013 ABC $(\mathrm{t})$ | 15,518 | 26,205 | 4,623 | 1,735 | 48,081 |

A summary of important reference values from the preferred model for this assessment, relative to the 2010 SAFE projections, is as follows:

| Quantity | As estimated or specified last year (2010) |  | As estimated or specified this year (2011) |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 2011 | 2012 | 2012 | 2013 |
| M (natural mortality) | 0.2 | 0.2 | 0.2 | 0.2 |
| Specified/recommended tier | 3 a | 3a | 3a | 3a |
| Total biomass (Age 3+; t) | 325,357 | 321,355 | 292,189 | 286,274 |
| Female Spawning Biomass (t) | 113,406 | 115,427 | 104,301 | 105,127 |
| B 100\% | 111,884 | 111,884 | 103,868 | 103,868 |
| B $40 \%$ | 44,754 | 44,754 | 41,547 | 41,547 |
| B $35 \%$ | 39,159 | 39,159 | 36,354 | 36,354 |
| $\begin{aligned} & F_{\text {OFL }}=F_{35 \%} \\ & \text { max } F_{A B C}=F_{40 \%} \\ & \text { recommended } F_{A B C} \end{aligned}$ | 0.530 | 0.530 | 0.593 | 0.593 |
|  | 0.406 | 0.406 | 0.450 | 0.450 |
|  | 0.406 | 0.406 | 0.450 | 0.450 |
| OFL (t) | 61,412 | 63,202 | 59,380 | 60,219 |
| max ABC (t) | 49,133 | 50,591 | 47,407 | 48,081 |
| ABC (t) | 49,133 | 50,591 | 47,407 | 48,081 |
| Status | As determined last year (2010) for: |  | As determined this year (2011) for: |  |
|  | 2009 | 2010 | 2010 | 2011 |
| Overfishing | no | n/a | no | n/a |
| Overfished | n/a | no | n/a | no |
| Approaching overfished | n/a | no | n/a | no |

## Plan Team Summary Tables

| Species | Year | Biomass $^{\mathbf{1}}$ | $\mathbf{O F L}^{\mathbf{2 , 3}}$ | $\mathbf{A B C}^{\mathbf{2 , 3}}$ | TAC $^{\mathbf{2 , 3}}$ | Catch $^{\mathbf{4}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2010 | 303,140 | 59,295 | 47,422 | 10,411 | 3,842 |
| Flathead | 2011 | 297,130 | 61,412 | 49,133 | 10,587 | 2,339 |
| sole | 2012 | 292,189 | 59,380 | 47,407 |  |  |
|  | 2013 | 286,274 | 60,219 | 48,081 |  |  |

${ }^{1}$ Age 3+ biomass from the assessment model.
${ }^{2}$ http://www.fakr.noaa.gov/sustainablefisheries/specs10_11/goa_table1.pdf
${ }^{3}$ http://www.fakr.noaa.gov/sustainablefisheries/specs11_12/goa_table1.pdf
${ }^{4}$ As of Sept. 24, 2011.

| Stock/ <br> Assemblage | Area | 2011 <br> $\mathbf{O F L}^{\mathbf{1}}$ | ABC $^{\mathbf{1}}$ | TAC $^{\mathbf{1}}$ | Catch $^{\mathbf{2}}$ | $\mathbf{2 0 1 2}$ <br> $\mathbf{O F L}^{\mathbf{3}}$ | $\mathbf{A B C}^{\mathbf{3}}$ | $\mathbf{2 0 1 3}$ <br> $\mathbf{O F L}^{\mathbf{3}}$ | $\mathbf{A B C}^{\mathbf{3}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | W | -- | 17,442 | 2,000 | 341 | -- | 15,300 | -- | 15,518 |
| Flathead | C | -- | 28,104 | 5,000 | 1,998 | -- | 25,838 | -- | 26,205 |
| sole | WYAK | -- | 2,064 | 2,064 | 0 | -- | 4,558 | - | 4,623 |
|  | SEO | -- | 1,523 | 1523 | 0 | -- | 1,711 | -- | 1,735 |
|  | Total | 61,412 | 49,133 | 10,587 | 2,339 | 59,380 | 47,407 | 60,219 | 48,081 |

${ }^{1}$ http://www.fakr.noaa.gov/sustainablefisheries/specs11_12/goa_table1.pdf
${ }^{2}$ As of Sept. 24, 2011.
${ }^{3}$ Based on the assessment model.

## SSC Comments Specific to the Flathead Sole Assessments

SSC request: The SSC requested that the authors provide a graph allowing comparison of the fit of alternative models, including confidence intervals or SE's, to survey biomass to facilitate model comparison and evaluation of fits.

Author response: We developed an R function that will allow this comparison among multiple models in the future (see Figure 6.8 in the GOA rex sole chapter for an example application). However, we did not evaluate alternative models in this assessment.

SSC Comments on Assessments in General
SSC request: The SSC requested that the next round of assessments consider the possible use of ADF\&G bottom trawl survey data to expand the spatial and depth coverage.

Author response: The current assessment model does not accommodate surveys from multiple sources. We are developing a new assessment model that will incorporate surveys from multiple sources as one of its new features. When completed, this new model will allow us to explore the utility of using the ADF\&G bottom trawl survey data in future assessments.

## Introduction

Flathead sole (Hippoglossoides elassodon) are distributed from northern California, off Point Reyes, northward along the west coast of North America and throughout the GOA and the BS, the Kuril Islands, and possibly the Okhotsk Sea (Hart 1973). They occur primarily on mixed mud and sand bottoms (Norcross et al., 1997; McConnaughey and Smith, 2000) in depths < 300 m (Stark and Clausen, 1995). The flathead sole distribution overlaps with the similar-appearing Bering flounder (Hippoglossoides robustus) in the northern half of the Bering Sea and the Sea of Okhotsk (Hart, 1973), but not in the Gulf of Alaska.

Adults exhibit a benthic lifestyle and occupy separate winter spawning and summertime feeding distributions on the EBS shelf and in the GOA. From over-winter grounds near the shelf margins, adults begin a migration onto the mid and outer continental shelf in April or May each year for feeding. The spawning period may range from as early as January but is known to occur in March and April, primarily in deeper waters near the margins of the continental shelf. Eggs are large ( 2.75 to 3.75 mm ) and females have egg counts ranging from about $72,000(20 \mathrm{~cm}$ fish $)$ to almost $600,000(38 \mathrm{~cm}$ fish). Eggs hatch in 9 to 20 days depending on incubation temperatures within the range of 2.4 to $9.8^{\circ} \mathrm{C}$ and have been found in ichthyoplankton sampling on the southern portion of the BS shelf in April and May (Waldron 1981). Larvae absorb the yolk sac in 6 to 17 days, but the extent of their distribution is unknown. Nearshore sampling indicates that newly settled larvae are in the 40 to 50 mm size range (Norcross et al. 1996). Fifty percent of flathead sole females in the GOA are mature at 8.7 years, or at about 33 cm (Stark, 2004). Juveniles less than age 2 have not been found with the adult population and probably remain in shallow nearshore nursery areas.

## Fishery

Flathead sole in the Gulf of Alaska are caught in a directed fishery using bottom trawl gear. Typically 25 or fewer shore-based catcher vessels from 58-125' participate in this fishery, as do 5 catcher-processor vessels ( $90-130$ '). Fishing seasons are driven by seasonal halibut PSC apportionments, with approximately 7 months of fishing occurring between January and November. Catches of flathead sole occur only in the Western and Central management areas in the gulf (statistical areas 610 and $620+630$, respectively). Recruitment to the fishery begins at about age 3.

Historically, catches of flathead sole have exhibited decadal-scale trends (Table 8.1, Fig. 8.1). From a high of $\sim 2000 t$ in 1980, annual catches declined steadily to a low of $\sim 150 t$ in 1986 but thereupon increased steadily, reaching a high of $\sim 3100 \mathrm{t}$ in 1996. Catches subsequently declined over the next three years, reaching a low of $\sim 900 \mathrm{t}$ in 1999, followed by an increasing trend through 2010, when the catch reached its highest level ever ( $3,842 \mathrm{t}$ ). As of Sept. 24, the catch in 2011 was $2,310 \mathrm{t}$. Based on the trend in weekly cumulative catch, the total 2011 catch is projected to be 2,891 t-almost one quarter less than in 2010.

Based on observer data, the majority of the flathead sole catch in the Gulf of Alaska is taken in the Shelikof Strait and on the Albatross Bank near Kodiak Island, as well as near Unimak Island (Figure 8.2). The spatial pattern of catches has been reasonably consistent over the past three years. Most of the catch is taken in the first and second quarters of the year (Figure 8.3).

Annual catches of flathead sole have been well below TACs in recent years, although the population appears to be capable of supporting higher exploitation rates (Table 8.2a). Limits on flathead sole catches are driven by within-season closures of the directed fishery due to restrictions on halibut PSC, not by attainment of the TAC (Table 8.2b). Recognizing this, TACs have typically been set much lower than the recommended ABC. Prior to 2003, flathead sole was a Tier 5 species and ABC's were based on natural mortality rates. Following the development and adoption of an age-structured assessment model in 2003,

ABCs for flathead sole in the Gulf of Alaska almost doubled from 2002 to 2003, from ~23,000 to $42,000 \mathrm{t}$. TACs, however, increased only moderately as a result.

Flathead sole are also caught in the pursuit of other species as bycatch. They are caught in the Pacific cod, bottom pollock and other flatfish fisheries and are caught with these species in the flathead soledirected fishery. The gross retention rate for flathead sole over all fisheries has been $87 \%$ or larger since 2005, and higher than $95 \%$ since 2009 (Table 8.2a).

## Data

## Fishery Data

This assessment used fishery catches from 1984 through Sept. 24, 2011 (Table 8.1, Fig. 8.1), as well as estimates of the proportion of individuals caught by length group and sex for the years 1985-2011 from fishery observer sampling (as of Sept. 24; Tables 8.3a, b). Sample sizes for the size compositions are shown in Table 8.4a. Age composition data from the fishery is not currently used in the assessment model. Limited age data is available from observer sampling for this stock, although some ageing of observer samples has been completed.

## Survey Data

Because flathead sole are often taken incidentally in target fisheries for other species, CPUE from commercial fisheries seldom reflects trends in abundance for this species. It is therefore necessary to use fishery-independent survey data to assess the condition of this stock.

This assessment used estimates of total biomass for flathead sole in the Gulf of Alaska from triennial (1984-1999) and biennial (2001-2011) groundfish surveys conducted by the Alaska Fisheries Science Center's Resource Assessment and Conservation Engineering (RACE) division to provide an index of population abundance (Table 8.5, Figure 8.4). Although survey depth coverage has been inconsistent for depth strata $>500 \mathrm{~m}$, the fraction of the flathead sole stock occurring in these depth strata is miniscule (Table 8.6), so we have not attempted to correct the survey estimates of total biomass for missing depth strata. In addition, the 2001 survey estimate did not sample the eastern section of the Gulf. We estimated the average fraction of stock biomass occurring in the unsampled area from the 1993, 1996 and 1999 surveys ( $\sim 11 \%$ ) and assigned a corresponding availability factor of 0.9 to the 2001 survey to correct for the missing area (Table 8.5). Since 1984, survey estimates of total biomass have fluctuated about a mean of $\sim 220,000 \mathrm{t}$ with no apparent trend. Estimated total biomass was $\sim 236,000 \mathrm{t}$ in 2011, a $5 \%$ increase from the 2009 survey estimate of $\sim 225,000 t$ (the largest in the time series) but a $16 \%$ decrease over the 2007 estimate of $\sim 280,000 \mathrm{t}$, the largest in the time series.

Size and age compositions (numbers of individuals by size or age group) from the RACE GOA groundfish surveys were also incorporated into the assessment model (Tables 8.7-8). Survey size compositions were available for every survey year, while age compositions were available for all surveys except the most recent (2011). Because age compositions were calculated from age-length data using the corresponding size compositions, size compositions were de-weighted in the model likelihood for years where age composition data was also available to avoid double counting. Survey size composition data was fully weighted in the model likelihood only for years when age composition data was unavailable (2011). Sample sizes for the survey size and age compositions are given in Table 8.4b.

Data on individual growth was incorporated into the assessment using sex-specific age-length transition matrices developed previously (Table 8.9a, b). Previously-developed sex-specific weight-at-age relationships and female maturity schedules (Table 8.10) were also used in this assessment.

To summarize, the following data were incorporated in the assessment:

| Source | type | years |
| :--- | :--- | :--- |
| Fishery | catch | $1984-2011$ |
|  | size compositions | $1985-2011$ |
|  | biomass | $1984-1999$ (triennial); <br> $2001-2011 ~(b i e n n i a l) ~$ |
|  | size compositions | $1984-1999$ (triennial); <br> $2001-2011 ~(b i e n n i a l) ~$ |
|  |  | $1984-1999$ (triennial); <br> $2001-2009$ (biennial) |

## Analytic Approach

## Model structure

A number of alternative model configurations have been considered in previous assessments (Stockhausen et al., 2005, 2007, 2009). Due to time constraints, we did not explore any alternative models this year. Rather, we used the model configuration preferred by the GOA Plan Team and SSC from the last full assessment (the "base case" model in Stockhausen et al., 2009) to fit the data, estimate parameters, and calculate Tier 3 biological and management reference points.

The assessment was conducted using a split-sex, age-structured model with parameters evaluated in a maximum likelihood context. The model structure (Appendix A) was developed following Fournier and Archibald's (1982) methods, with many similarities to Methot (1990). We implemented the model using automatic differentiation software developed as a set of libraries under C++ (ADModel Builder). ADModel Builder can estimate a large number of parameters in a non-linear model using automatic differentiation software extended from Greiwank and Corliss (1991) and developed into C++ class libraries. This software provides the derivative calculations needed for finding the minimum of an objective function via a quasi-Newton function minimization routine (e.g., Press et al. 1992). It also gives simple and rapid access to these routines and provides the ability to estimate the variancecovariance matrix for all parameters of interest.

The current assessment model covers 1984-2011. Age classes included in the model run from age 3 to 20. Age at recruitment was set at 3 years in the model due to the small number of fish caught at younger ages. The oldest age class in the model, age 20, serves as a plus group; the typical maximum age of flathead sole based on otolith age determinations has been estimated at 25 years (Turnock et al., 2003a). Details of the population dynamics and estimation equations, description of variables and likelihood components are presented in Appendix A (Tables A.1, A.2, and A.3). Model parameters that are typically fixed are presented in Table A.4. A total of 81 parameters were estimated in the final model (Table A.5).

## Parameters estimated independently

Model parameters related to natural mortality, growth, weight, maturity and survey catchability were fixed in the final model (Table A.4).

## Natural mortality

As in the previous assessment (Stockhausen et al., 2009), natural mortality ( $M$ ) was fixed at $0.2 \mathrm{yr}^{-1}$ for both sexes in all age classes. This value was based on a maximum observed age for flathead sole of 22 years (Spencer et al., 1999). Although maximum observed age has increased to 31 years in the Bering

Sea, an analysis of independent estimates of natural mortality for flathead sole is not inconsistent with continued use of this value (Stockhausen, et al. 2010b).

## Growth

Individual growth was incorporated in the model using sex-specific age-length transition matrices (Table 8.9). These were identical to those used in the previous assessment (Stockhausen et al., 2009). In terms of the von Bertalanffy growth equation, $\mathrm{L}_{\text {inf }}$ was estimated at 44.37 cm for females and 37.36 cm for males (Figure 8.6 a). The length at age $2\left(L_{2}\right)$ was estimated at 10.17 cm for males and 13.25 cm for females. The growth parameter $k$ was estimated at 0.157 for females and 0.204 for males. Length at age $t$ was modeled as:

$$
L_{t}=L_{\mathrm{inf}}+\left(L_{2}-L_{\mathrm{inf}}\right) e^{-k(t-2)}
$$

## Weight at length

The weight-length relationship used for flathead sole was identical to that used in the previous assessment (Stockhausen et al., 2007): $W=0.00428 L^{3.2298}$ for both sexes combined (weight in grams and length in centimeters). Weight-at-age (Table 8.10, Figure 8.6b) was estimated using the mean length-at-age and the weight-length relationship.

## Maturity

The maturity schedule for Gulf of Alaska flathead sole was estimated using histological analysis of ovaries collected in January 1999 (Stark, 2004; Table 8.10, Figure 8.6c). A total of 180 samples were analyzed for estimation of age at maturity. Size at $50 \%$ mature was estimated to be 33.3 cm with a slope of $0.52 \mathrm{~cm}^{-1}$ from a sample of 208 fish. Age at $50 \%$ mature was 8.74 years with a slope of $0.773 \mathrm{yr}^{-1}$. Size at $50 \%$ mature was estimated at 32.0 cm for Bering Sea flathead sole (not significantly different from the GOA results), however, age at $50 \%$ mature was 9.7 due to slower growth in the Bering sea.

## Survey catchability

Based on results from the 2003 assessment (Turnock et al., 2003a), which indicated that estimating survey catchability was problematic, we fixed overall survey catchability ( $Q$ in Table A.1) in the model to a value of 1 .

## Parameters estimated conditionally

A total of 85 parameters were estimated in the model (Table A.5). These consisted primarily of parameters on the recruitment of flathead sole to the population (46 parameters total, including ones determining the initial age composition) and values related to annual fishing mortality ( 29 parameters total). The separable age-component of fishing mortality was modeled using ascending logistic functions estimated separately for males and females (4 parameters total). The same approach was also used to estimate relative age-specific survey catchability (4 parameters total).

Annual recruitment to the age 3 year class was parameterized in the model using one parameter for the log-scale mean recruitment and 45 parameters for the annual log-scale deviation from the mean. Recruitments were estimated back to 1967 to provide an initial age distribution for the model in its starting year (1984). In an analogous fashion, fully-recruited fishing mortality was parameterized in the model using one parameter for the log-scale mean and 28 parameters for the annual log-scale deviation from the mean.

Parameters in the model were selected based on minimizing an objective function equivalent to a negative log-likelihood function, hence the parameter estimates are maximum likelihood estimates. Components that contribute to the overall negative log likelihood include those related to observed fishery catches, fishery size compositions, survey biomass estimates, survey size compositions, survey age composition,
and recruitment deviations (Table A.3). The observed fishery catch was assumed to have a lognormal error structure, as was estimated survey biomass. The recruitment deviation parameters were incorporated directly into the overall likelihood via three components: "early" recruitment, "ordinary" recruitment and "late" recruitment (Table A.3). The "early" recruitment component incorporated deviations from 1967 to 1983, i.e. those that determined the initial model age structure and were thus uninformed by contemporaneous catch data. The "ordinary" recruitment component incorporated deviations from 1984-2008, while the "late" recruitment component incorporated deviations from 20092011. "Late" recruitments are weighted separately in the likelihood from "ordinary" recruitments because there is generally little data to constrain recruitment estimates for the final few years in the model. This partitioning does not reflect any assumptions regarding changes in productivity with time: All three components were formulated assuming a lognormal error structure. The size and age compositions were assumed to be drawn from different sex-specific multinomial distributions. If this assumption were strictly correct, then the number of individuals contributing to each composition would be the appropriate corresponding sample size. However, because fish of the same size and age tend to be found together, size and age compositions tend to be overdispersed with respect to actual multinomial distributions. Also, the use of high sample sizes can lead to numerical problems in estimating the model parameters. Previous experience indicates that using a uniform sample size of 200 for compositions with more than 200 individuals provides an adequately simple solution to the problem of assigning sample sizes. Thus, a sample size of 200 was used for all compositions used in the likelihood (all age compositions, as well as all fishery size compositions and survey size compositions from years with no corresponding age compositions).

Different weights can be assigned to each likelihood component to increase or decrease the relative degree of model fit to the data underlying the respective component; a larger weight induces a closer fit to a given likelihood component. Typically, a relatively large weight (e.g., 30) is applied to the catch component while smaller weights (e.g., 1) are applied to the survey biomass, recruitment, and size and age composition components. This reflects a belief that total catch data are reasonably well known (smaller variance) than the other types of data. For the recruitment components, larger weights applied to a component force the deviations contributing to that component closer to zero (and thus force recruitment closer to the geometric mean over the years that contribute to the component). The weights used in this assessment are given in Table 8.11.

Initial values for the estimable parameters were set as listed in Table 8.12. To test that the resulting model solution was indeed the global, rather than a local, maximum on the likelihood surface, we started the model using several different parameter sets. Most runs converged to the same (maximum) likelihood value and parameter estimates, providing evidence that the original solution was indeed the global maximum. The scores associated with different components of the maximum likelihood (e..g, fishery catch, survey biomass) are given in Table 8.13.

## Final parameter estimates

The estimated maximum likelihood parameter values are given in Table 8.14 for all model parameters.

## Model evaluation

Model estimates of fishery catch closely matched the observed values (Table 8.15 and Figure 8.7). This, however, is expected because a large weighting factor (30) was placed on the catch biomass component in the likelihood function. The model did not fit observed survey biomass values as closely as it did the catch (Table 8.15 and Figure 8.8), but model estimates of survey biomass fell outside the $95 \%$ confidence intervals of the actual surveys for only two out of eleven survey years (1984 and 2001) so the fit was deemed satisfactory.

The model did not fit the fishery size compositions as well, although its performance appeared to be reasonably good in most years (Figures 8.9 and 8.10 for females and males, respectively). Fits to the fishery size compositions were poorest when the observed size composition was dominated by a single size class and thus sharply peaked-for example, females in 1987 or males in 1990 (Figure 8.9). The smoothing inherent in using an age-length transition matrix to convert age classes to size classes in the model precludes close fits to peaked size compositions: the peak will be underestimated and the tails will be overestimated. It also appears, however, that these sharply-peaked size compositions may be associated with lower sample sizes or higher overdispersion (the size composition from 1990 was based on 149 individuals from only 3 different hauls). As with the fishery size compositions, model fits to the survey size compositions were poorest when the observed size compositions were sharply peaked (Figures 8.11 and 8.12). Also, the model consistently overestimates survey composition at small sizes ( $<20 \mathrm{~cm}$ ) for males (Figure 8.12) but not for females (Figure 8.11). This may indicate a shortcoming at small sizes of the current age/size transition matrix used in the model for males (Table 8.9b). However, this currently has little effect on the way the model fits the data because the 2011 survey size composition enters fully-weighted into the likelihood maximization-the remainder are deweighted in favor of the corresponding survey age compositions. Finally, the model also fit the survey age compositions reasonably well (Figures 8.16 and 8.17). Model fits to age and size compositions are quite similar to those obtained in the last full assessment (Stockhausen et al., 2009).

## Results

The estimated selectivity curves for the fishery and survey indicate that the fishery generally catches older flathead sole than the survey (Figures 8.15-16). For the fishery, age at $50 \%$ selection was 10.4 years for females and 10.1 for males. For the survey, the ages at $50 \%$ selection were younger: 5.4 years for females and 4.9 for males. The rates of increase in the selectivity curves at $50 \%$ selection were slightly smaller for females vs. males, but quite similar between the fishery and the survey. Examination of the marginal posterior distributions for these parameters indicates that they were well-estimated in the model (Figure 8.16).

The model also estimates other population variables of interest, such as time series of total biomass, spawning biomass, recruitment and fully-selected fishing mortality. In this assessment, total biomass is represented by age $3+$ biomass whereas spawning biomass is female spawning biomass. Model quantities such as median recruitment, median fishing mortality, total biomass, spawning biomass and recruitment all seem to be reasonably well-estimated, based on examination of MCMC posterior distributions (Figures 8.17-18). The maximum likelihood estimates of these quantities tends to be very close to the posterior modes, with median recruitment (Fig. 8.17, upper left plot) and the 2010 recruitment (Figure 8.18, lower plot) being exceptions to this. Median recruitment was estimated at 84 million individuals, while the mode was at 79 million. Median fishing mortality was estimated at $0.015 \mathrm{yr}^{-1}$. Total biomass in 2011 was estimated at $297,000 \mathrm{t}$, spawning stock biomass at $102,000 \mathrm{t}$ and recruitment at 139 million. Model estimates suggest that age 3+ biomass generally underwent a steady increase from 210,000 t in 1984 to $305,000 \mathrm{t}$ in 2009, followed by a very slight decline in the past two years to $297,000 \mathrm{t}$ (Table 8.16 and Figure 8.19). The estimated age 3+ biomass in this assessment is somewhat (10\%) lower than that estimated in previous assessments (Table 8.16, Figure 8.18). Female spawning biomass was estimated to have increased fairly rapidly from 1984 ( $53,000 \mathrm{t}$ ) until 1990 ( $86,000 \mathrm{t}$ ), then to have continued increasing at a much slower rate. The estimated 2011 spawning biomass is the largest in the time series $(102,000 \mathrm{t})$.

Model estimates of annual recruitment (age 3 numbers) ranged from a low of 174 million individuals in 1999 to a high of 362 million in 2008 (Table 8.17 and Figure 8.20). Prior to 2000, recruitment was generally below the long-term average ( 228 million), while it has generally been higher since 2000. In 2011, recruitment was estimated below the long-term average, but this is expected because of the
structure of the recruitment likelihood component. Results from the current assessment are generally similar to those estimated in previous assessments, particularly prior to 2001 (Table 8.17, Figure 8.20). Except in 2008, this year's model estimates for recruitment after 2001 are somewhat smaller than those obtained in 2009 and somewhat less variable than those obtained in 2007.

Marginal posterior distributions based on MCMC integration are shown in Figure 8.21 for $\mathrm{F}_{35 \%}$ and $\mathrm{F}_{40 \%}$, $\mathrm{B}_{35 \%}$ and $\mathrm{B}_{40 \%}$, and max ABC and OFL for 2012 (based on Tier 3a rules). Based on these distributions, all six quantities appear to be well-estimated in the model. A control rule plot based on the temporal trajectory of estimated fishing mortality and spawning biomass indicates that the GOA flathead sole stock has not been overfished nor has overfishing occurred (Figure 8.22).

## Projections and Harvest Alternatives

The reference fishing mortality rate for flathead sole is determined by the amount of reliable population information available (Amendment 56 of the Fishery Management Plan for the groundfish fishery of the Bering Sea/Aleutian Islands). Estimates of $F_{40 \%}, F_{35 \%}$, and $S P R_{40 \%}$ were obtained from a spawner-perrecruit analysis. Assuming that the average recruitment from the 1981-2008 year classes estimated in this assessment represents a reliable estimate of equilibrium recruitment, then an estimate of $B_{40 \%}$ is calculated as the product of $S P R_{40 \%}$ times the equilibrium number of recruits; this quantity is $49,899 \mathrm{t}$. The 2011 spawning stock biomass is estimated at $102,000 \mathrm{t}$. Since reliable estimates of the 2011 spawning biomass (B), $B_{40 \%}, F_{40 \%}$, and $F_{35 \%}$ exist and $B>B_{40 \%}(102,000 t>41,547 \mathrm{t})$, the flathead sole reference fishing mortality is defined in Tier 3a.

For this tier, $F_{A B C}$ is constrained to be $\leq F_{40 \%}$, and $F_{\text {OFL }}$ is defined to be $F_{35 \%}$. The values of these quantities are:

| estimated $2011 \text { SSB }$ | $=$ | 102,000 t |
| :---: | :---: | :---: |
| B $40 \%$ | = | 41,547 t |
| $F_{40 \%}$ | = | 0.450 |
| $F_{A B C}$ (max) | = | 0.450 |
| $B_{35 \%}$ | = | 36,354 t |
| $F_{35 \%}$ | = | 0.593 |
| $F_{\text {OFL }}$ | = | 0.593 |

Because the flathead sole stock has not been overfished in recent years and the stock biomass is relatively high, it is not recommended to adjust $F_{A B C}$ downward from its upper bound; thus, the year 2011 recommended ABC associated with $F_{A B C}$ of 0.450 , is $47,407 \mathrm{t}$. The fishing mortality associated with overfishing ( $F_{O F L}$ ) is 0.593 . The corresponding OFL for 2012 is $59,380 \mathrm{t}$.

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2011 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2012 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2011. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn
from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2012, are as follow (" $m a x F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2012 recommended in the assessment to the max $F_{A B C}$ for 2012. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to $50 \%$ of $\max F_{A B C}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 4: In all future years, $F$ is set equal to the 2006-2011 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

The recommended $F_{A B C}$ and the maximum $F_{A B C}$ are equivalent in this assessment, so scenarios 1 and 2 yield identical results. The 12 -year projections of the mean harvest, spawning stock biomass and fishing mortality for the five scenarios are shown in Tables 8.18-20.

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether the flathead sole stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follows (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2012, then the stock is not overfished.)

Scenario 7: In 2012 and 2013, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2024 under this scenario, then the stock is not approaching an overfished condition.)

The results of these two scenarios indicate that the stock is not overfished and is not approaching an overfished condition. With regard to assessing the current stock level, the expected stock size in the year 2012 of scenario 6 is $104,301 \mathrm{t}$, more than 2 times $B_{35 \%}(36,354 \mathrm{t})$. Thus the stock is not currently
overfished. With regard to whether the stock is approaching an overfished condition, the expected spawning stock size in the year 2024 of scenario $7(38,140 \mathrm{t})$ is greater than $B_{35 \%}$; thus, the stock is not approaching an overfished condition.

Estimating an ABC and OFL for 2013 is somewhat problematic as these values depend on the catch that will be taken in 2012. The actual catch taken in the GOA flathead sole fishery has been substantially smaller than the TAC for the past several years, but the catch has been rising steadily since 1999 until this year (Figure 8.1). The year end 2011 catch was predicted to be $2,891 \mathrm{t}$, substantially less than in 2010 ( $3,842 \mathrm{t}$; the largest catch in the time series). However, it is not known whether this year's catch is an anomaly to the generally-increasing trend, or not. As such, we assumed that a reasonable estimate of the catch to be taken in 2012 was the same as that taken in 2010. Using these catch values and the estimated population size at the start of 2011 from the model, we projected the stock ahead through 2011-2012 and calculated the ABC and OFL for 2013. The estimated ABC for 2013 is $48,081 \mathrm{t}$ while the estimated OFL is $60,219 \mathrm{t}$. Total biomass for 2013 is estimated at $286,274 \mathrm{t}$, while female spawning biomass is estimated at 105,127 .

## Area allocation of harvests

TAC's for flathead sole in the Gulf of Alaska are divided among four smaller management areas (Western, Central, West Yakutat and Southeast Outside). As previous assessments, the area-specific ABC's for flathead sole in the GOA are divided up over the four management areas by applying the fraction of the most recent survey biomass estimated for each area (relative to the total over all areas) to the 2012 and 2013 ABC's. The area-specific allocations for 2012 and 2013 are:

|  | Western | Central | Yest | Southeast |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Outside |  |  |  |  | Total |  | Area Apportionment | $32.3 \%$ | $54.5 \%$ | $9.6 \%$ | $3.6 \%$ |
| :--- | :---: | :---: | :---: | :---: | :---: |
| $100.0 \%$ |  |  |  |  |  |
| 2012 ABC $(\mathrm{t})$ | 15,300 | 25,838 | 4,558 | 1,711 | 47,407 |
| 2013 ABC $(\mathrm{t})$ | 15,518 | 26,205 | 4,623 | 1,735 | 48,081 |

## Ecosystem Considerations

## Ecosystem effects on the stock

## Prey availability/abundance trends

Based on results from an ecosystem model for the Gulf of Alaska (Aydin et al., 2007), flathead sole in the Gulf of Alaska occupy an intermediate trophic level as both juvenile and adults (Fig. 8.21). Pandalid shrimp and brittle stars were the most important prey for adult flathead sole in the Gulf of Alaska (64\% by weight in sampled stomachs; Yang and Nelson, 2000; Fig. 8.22a), while euphausids and mysids constituted the most important prey items for juvenile flathead sole (Fig. 8.22b).. Other major prey items included polychaetes, mollusks, bivalves and hermit crabs for both juveniles and adults. Commercially important species that were consumed included age-0 Tanner crab (3\%) and age-0 walleye pollock (< $0.5 \%$ by weight). Little to no information is available to assess trends in abundance for the major benthic prey species of flathead sole.

## Predator population trends

Important predators on flathead sole include arrowtooth flounder, walleye pollock, Pacific cod, and other groundfish (Fig. 8.23). Pacific cod and Pacific halibut are the major predators on adults, while arrowtooth flounder, sculpins, walleye pollock and Pacific cod are the major predators on juveniles. The flatfishdirected fishery constitutes the third-largest known source of mortality on flathead sole adults. However, the largest component of mortality on adults is unexplained.

Arrowtooth flounder are currently the most abundant groundfish in the Gulf of Alaska, and have steadily increased in abundance since the early 1970’s (Turnock et al., 2003b). The abundance of walleye pollock has declined rather steadily since the early 1990's, but recent evidence suggests the stock may be starting to increase again (Dorn et al., 2004). Pacific cod abundance in the Gulf of Alaska has been declining since 1990 (Thompson et al., 2004). Although the continued increase in abundance of arrowtooth flounder is cause for some concern, the abundance of flathead sole has actually increased in recent years. Predation by arrowtooth may be limiting the potential rate of increase of flathead sole under current conditions, but it does not appear to represent a threat to the stock.

## Fishery effects on ecosystem

Catches of flathead sole have been concentrated in several areas in the Gulf of Alaska over the past few years (Figure 8.2). These areas include Shelikof Straight, Portlock Bank and Davidson Bank. The ecosystem effects of this spatial concentration of fishing activity are unknown.

Prohibited species such as halibut, salmon, and crab are also taken to some extent in the flathead soledirected fishery (Table 8.21). In 2011 thus far, the overall prohibited species catch (PSC) for crab in the directed fishery was exclusively Bairdi tanner crab, with catches sometimes fluctuating by factors of 3-4 between years. The PSC for crab thus far in the 2011 directed fishery was approximately 5,000 Bairdi tanner crab, somewhat less than that caught in 2008-2010. As a fraction of the total Bairdi crab PSC, the fishery accounted for $5.8 \%$ in 2011 but less than $3 \%$ in 2008-2010-even though the absolute numbers were greater i 2008-2010. The PSC for halibut was almost $92,000 \mathrm{~kg}$ halibut-a decrease from the 2010 catch of almost 257,000 kg and similar to the 2009 and 2008 catches (approximately 98,000 and 92,000 kg , respectively). Except for 2011, these catches constituted less than $2.5 \%$ of the total halibut PSC. In 2011, the percentage was $5.7 \%$. The PSC for salmon in the directed fishery is mainly Chinook, with 498 individuals caught in 2010 and 118 in 2009. These accounted for $0.9 \%$ and $1.5 \%$ of the total salmon PSC in those years. In the previous two years (2007-8), no individuals were caught. Preliminary values for 2011 were not available at the time this document was compiled.

Bycatch of non-target species in the flathead sole fishery tends be highly variable between years, at least when expressed as a percentage of the total observed bycatch in the FMP by non-target species group (Table 8.22). In 2011, the flathead sole fishery accounted for more than $5 \%$ of the bycatch of six species groups: benthic urochordata (tunicates; 8.5\%), eelpouts (9.2\%), grenadier (6.4\%), unidentified polychaetes ( $39.2 \%$ ), sea pens and whips ( $8.6 \%$ ), and stichaeidae (pricklebacks; 12.0\%). In 2010, the fishery reportedly caught no unidentified polychaetes or grenadier, but again accounted for more than 5\% of the bycatch of benthic urochordata (14.1\%), eelpouts (11.3\%), sea pens and whips (14.0\%), and stichaeidae (13.5\%), as well as unidentified brittle stars (9.7\%), Giant grenadiers (5.1\%), greenlings (5.5\%), and pandalid shrimp (6.1\%). The fishery has had no bycatch of birds and has accounted for less than $5 \%$ of bycatch in all shark, skate, and forage fish (capelin, eulachon, sandlance) species groups over the time frame analyzed (2003-2011).

Over the past five years, the flathead sole-directed fishery caught more arrowtooth flounder than any other non-prohibited FMP species, including flathead sole (Table 8.23). Flathead sole was the second most-caught species in the directed fishery. Only small amounts of arrowtooth were retained (typically $10 \%$ ), while generally more than $90 \%$ of flathead sole was retained. Pacific cod was the third mostcaught species, with retention rates typically greater than $90 \%$.

Effects of discards and offal production on the ecosystem are unknown for the flathead sole fishery.

## Data gaps and research priorities

The AFSC's Age and Growth Program has made substantial progress in processing survey age data for flathead sole in the Gulf of Alaska. While this information has been incorporated in the current stock assessment in the form of survey age compositions, age information also enters the assessment in the form of age-length conversion matrices estimated outside the assessment model. The matrices currently used in the assessment are now several years old. One of our goals for the next assessment is to use the newly-available age data to revise growth schedules for GOA flathead and reassess these age-length conversion matrices. In addition, we anticipate incorporating such estimation directly into the assessment model, rather than performing it outside the model. This approach will also allow us to incorporate ageing error into the model structure.

Although the AFSC's Age and Growth Program has made substantial progress in processing survey age data for flathead sole in the Gulf of Alaska, the amount of fishery age data is almost nonexistent. Additional age data (both survey and fishery) should improve future stock assessments by allowing improved estimates of individual growth and age-length transition matrices, and by filling in missing years with age composition data.

Further modeling research should address the use of length-based approaches to fishery and survey selectivity in the assessment model, as well as alternative forms for the selectivity function. The utility of potential environmental predictors of recruitment (e.g., temperature) should also be investigated. We will also revisit the estimates used for natural mortality in the model.

## Summary

| Tier | 3 a |  |
| :---: | :---: | :---: |
| Reference mortality rates |  |  |
| M | 0.2 |  |
| $F_{35 \%}$ | 0.593 |  |
| $F_{40 \%}$ | 0.450 |  |
| Equilibrium female spawning biomass |  |  |
| B 100\% | 103,868 t |  |
| B $40 \%$ | 41,547 t |  |
| B $35 \%$ | 36,354 t |  |
| Fishing rates |  |  |
| $F_{\text {OFL }}$ | 0.593 |  |
| $F_{\text {ABC }}$ (maximum permissible) | 0.450 |  |
| $F_{\text {ABC }}$ (recommended) | 0.450 |  |
| Projected biomass | 2012 | 2013 |
| Age 3+ biomass (t) | 292,189 | 286,274 |
| Female spawning biomass (t) | 104,301 | 105,127 |
| Harvest limits | 2012 | 2013 |
| OFL (t) | 59,380 | 60,219 |
| ABC (maximum permissible; t) | 47,407 | 48,081 |
| ABC (recommended; t) | 47,407 | 48,081 |

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## Tables

Table 8.1. Annual catch of flathead sole in the Gulf of Alaska, from 1978 to 2011. The 2011 catch is through Sept. 24, 2011.

| year | total catch <br> (t) | Western <br> Gulf | Central <br> Gulf | West <br> Yakutat | Southeast |
| :---: | :---: | :---: | :---: | :---: | :---: |

Table 8.2a. Time series of recent reference points (ABC, OFL, TAC), total catch and retention rates for GOA flathead sole. The 2011 catch is through Sept. 24, 2011.

| Year | Author <br> ABC (t) | ABC (t) | TAC (t) | OFL (t) | Total <br> Catch (t) | \% <br> Retained |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1995 | -- | 28,790 | 9,740 | 31,557 | 2,181 |  |
| 1996 | -- | 52,270 | 9,740 | 31,557 | 3,107 |  |
| 1997 | -- | 26,110 | 9,040 | 34,010 | 2,446 |  |
| 1998 | -- | 26,110 | 9,040 | 34,010 | 1,742 |  |
| 1999 | -- | 26,010 | 9,040 | 34,010 | 900 |  |
| 2000 | -- | 26,270 | 9,060 | 34,210 | 1,547 |  |
| 2001 | -- | 26,270 | 9,060 | 34,210 | 1,911 |  |
| 2002 | 22,684 | 22,690 | 9,280 | 29,530 | 2,145 |  |
| 2003 | 41,402 | 41,390 | 11,150 | 51,560 | 2,425 | 88 |
| 2004 | 51,721 | 51,270 | 10,880 | 64,750 | 2,390 | 80 |
| 2005 | 36,247 | 45,100 | 10,390 | 56,500 | 2,530 | 87 |
| 2006 | 37,820 | 37,820 | 9,077 | 47,003 | 3,134 | 89 |
| 2007 | 39,110 | 39,110 | 9,148 | 48,658 | 3,163 | 89 |
| 2008 | 44,735 | 44,735 | 11,054 | 55,787 | 3,419 | 90 |
| 2009 | 46,464 | 46,464 | 11,181 | 57,911 | 3,658 | 96 |
| 2010 | 47,422 | 47,422 | 10,411 | 59,295 | 3,842 | 95 |
| 2011 | 49,133 | 49,133 | 10,587 | 61,412 | 2,339 | 97 |

Table 8.2b. Status of the flathead sole fishery in recent years.

| Year | Dates | Status |
| :---: | :---: | :---: |
| 2005 | Jan 20 | open |
|  | Aug 19 | halibut bycatch status |
|  | Sep 1 | open |
|  | Sep 4 | halibut bycatch status |
| 2006 | Jan 20 | open |
|  | Feb 23 | halibut bycatch status |
|  | Feb 27 | open |
|  | Jun 10 | halibut bycatch status |
|  | Jul 1 | open |
|  | Sep 1 | halibut bycatch status |
|  | Spe 6 | open |
|  | Sep 6 | halibut bycatch status |
|  | Sep 20 | open |
|  | Spe 20 | halibut bycatch status |
|  | Sep 25 | open |
|  | Sep 25 | halibut bycatch status |
|  | Oct 1 | open |
|  | Oct 8 | halibut bycatch status |
| 2007 | Jan 20 | open |
|  | Jun 4 | halibut bycatch status |
|  | Jul 1 | open |
|  | Aug 10 | halibut bycatch status |
|  | Sep 1 | open |
|  | Sep 1 | halibut bycatch status |
|  | Sep 6 | open |
|  | Sep 6 | halibut bycatch status |
|  | Sep 11 | open |
|  | Sep 11 | halibut bycatch status |
|  | Sep 21 | open |
|  | Sep 23 | halibut bycatch status |
|  | Oct 1 | open |
|  | Oct. 8 | halibut bycatch status |
|  | Oct 10 | open |
|  | Oct 15 | halibut bycatch status |
|  | Oct 22 | open |
| 2008 | Jan 20 | open |
|  | Jan 23 | A80 vessels subject to sideboard limits: halibut bycatch status |
|  | Jan 29 | A80 vessels subject to sideboard limits: open |
|  | Mar 10 | halibut bycatch status |
|  | Mar 21 | open |
|  | May 21 | halibut bycatch status |
|  | Jul 1 | open |
|  | Aug 7 | halibut bycatch status |
|  | Sep 1 | open |
|  | Sep 3 | halibut bycatch status |
|  | Sep 10 | open |
|  | Sep 11 | halibut bycatch status |
|  | Oct 1 | open |
|  | Nov 6 | halibut bycatch status |
|  | Nov 16 | open |


| Year | Dates | Status |
| :--- | :--- | :--- |
| 2009 | Jan 20 <br> Sep 2 <br> Oct 1 | open <br> halibut bycatch status <br> open |
| 2010 | Jan 20 | open |
|  | Sep 3 |  |
| Sep 11 | halibut bycatch status |  |
| open |  |  |

Table 8.3a. Annual fishery length compositions for female flathead sole. The 2011 composition is based on observer reports through Sept. 24. Fishery length compositions are normalized to 1 over both sexes.


Table 8.3b. Annual fishery length compositions for male flathead sole. The 2011 composition is based on observer reports through Sept. 24. Fishery length compositions are normalized to 1 over both sexes.


Table 8.4a. Sample sizes for the domestic fishery. Fishery age compositions are not currently used in the assessment model.

| year | Size compositions |  |  |  | Age compositions |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | hauls | total <br> indiv.s | females | males | hauls | total indiv.s | females | males |
| 1990 | 3 | 274 | 65 | 84 |  |  |  |  |
| 1991 | 48 | 4301 | 1213 | 1418 |  |  |  |  |
| 1992 | 77 | 4958 | 1650 | 2034 |  |  |  |  |
| 1993 | 55 | 4801 | 1425 | 2140 |  |  |  |  |
| 1994 | 56 | 4089 | 1175 | 930 |  |  |  |  |
| 1995 | 46 | 2818 | 1280 | 1301 |  |  |  |  |
| 1996 | 174 | 11207 | 2297 | 2330 |  |  |  |  |
| 1997 | 72 | 4827 | 1926 | 2113 |  |  |  |  |
| 1998 | 128 | 6509 | 2569 | 2896 |  |  |  |  |
| 1999 | 7 | 130 | 70 | 55 |  |  |  |  |
| 2000 | 111 | 1464 | 667 | 796 |  |  |  |  |
| 2001 | 80 | 1446 | 664 | 757 |  |  |  |  |
| 2002 | 86 | 1326 | 645 | 643 |  |  |  |  |
| 2003 | 168 | 2592 | 920 | 1609 |  |  |  |  |
| 2004 | 79 | 1590 | 816 | 765 |  |  |  |  |
| 2005 | 118 | 1838 | 882 | 947 | 46 | 159 | 77 | 82 |
| 2006 | 124 | 1872 | 835 | 990 |  |  |  |  |
| 2007 | 122 | 1830 | 840 | 985 |  |  |  |  |
| 2008 | 100 | 1628 | 815 | 798 |  |  |  |  |
| 2009 | 106 | 1860 | 986 | 851 | 43 | 180 | 91 | 85 |
| 2010 | 172 | 2852 | 1423 | 1427 |  |  |  |  |
| 2011 | 79 | 1282 | 467 | 815 |  |  |  |  |

Table 8.4 b. Sample sizes for the groundfish survey.

| year | $\begin{aligned} & \frac{\text { biomass }}{\text { total }} \\ & \text { hauls } \end{aligned}$ | hauls | Size compositions |  |  | Age compositions |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | total indiv.s | females | males | hauls | total indiv.s | females | males |
| 1984 | 929 | 264 | 25316 | 13875 | 11291 |  |  |  |  |
| 1987 | 783 | 197 | 27298 | 15931 | 11350 |  |  |  |  |
| 1990 | 708 | 286 | 24322 | 12939 | 11255 | 22 | 247 | 138 | 107 |
| 1993 | 775 | 364 | 26124 | 13592 | 12294 | 36 | 312 | 179 | 132 |
| 1996 | 807 | 417 | 21416 | 11086 | 9975 | 55 | 528 | 285 | 243 |
| 1999 | 764 | 389 | 16052 | 7941 | 8023 | 47 | 605 | 316 | 288 |
| 2001 | 489 | 245 | 11877 | 5962 | 5899 | 83 | 861 | 458 | 399 |
| 2003 | 809 | 434 | 25885 | 13279 | 12479 | 87 | 499 | 249 | 250 |
| 2005 | 839 | 413 | 23499 | 12501 | 10907 | 76 | 551 | 296 | 243 |
| 2007 | 820 | 411 | 25539 | 13563 | 11860 | 82 | 755 | 410 | 345 |
| 2009 | 823 | 454 | 21010 | 10304 | 10360 | 75 | 692 | 366 | 314 |
| 2011 | 670 | 379 | 17627 | 9290 | 8286 |  |  |  |  |

Table 8.5. Biomass estimates ( t ) by NPFMC regulatory area for GOA flathead sole from the NMFS bottom trawl surveys. Note that in 2001 the eastern GOA was not surveyed. This was accounted for in the assessment model by assuming, based on previous surveys, that availability for this year was was 0.9. The maximum depth stratum included in each survey is also noted.

| Year | Western <br> Gulf | Central <br> Gulf | West <br> Yakutat | Southeast | Total | Std. Dev | Max <br> Depth <br> (m) |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1984 | 45,100 | 158,539 | 45,694 | 9 | 249,341 | 30,355 | 1000 |
| 1987 | 33,603 | 113,483 | 30,455 | 5 | 177,546 | 18,956 | 1000 |
| 1990 | 58,740 | 161,257 | 23,019 | 40 | 243,055 | 28,877 | 500 |
| 1993 | 57,871 | 113,976 | 16,720 | 124 | 188,690 | 24,486 | 500 |
| 1996 | 66,732 | 122,730 | 12,751 | 3,308 | 205,521 | 18,430 | 500 |
| 1999 | 49,636 | 139,356 | 15,115 | 3,482 | 207,590 | 24,404 | 1000 |
| 2001 | 68,164 | 85,430 | 0 | 0 | 153,594 | 18,300 | 500 |
| 2003 | 67,055 | 170,852 | 17,154 | 2,234 | 257,294 | 19,913 | 700 |
| 2005 | 59,458 | 142,043 | 11,400 | 312 | 213,213 | 16,944 | 1000 |
| 2007 | 78,361 | 176,529 | 21,430 | 3,970 | 280,290 | 23,778 | 1000 |
| 2009 | 80,115 | 128,910 | 9,458 | 6,894 | 225,377 | 25,041 | 1000 |
| 2011 | 76,049 | 128,428 | 22,656 | 8,506 | 235,639 | 22,329 | 700 |

Table 8.6. Biomass estimates (t) by depth stratum for GOA flathead sole from the NMFS bottom trawl surveys. Note that in 2001 the eastern GOA was not surveyed.

| year | Depth range (m) |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1-100 | 101-200 | 201-300 | 301-500 | >500 |
| 1984 | 118,974 | 121,791 | 8,571 | 5 | 0 |
| 1987 | 91,482 | 75,475 | 10,553 | 36 | 0 |
| 1990 | 157,014 | 76,306 | 9,713 | 22 | -- |
| 1993 | 113,072 | 65,143 | 10,278 | 198 | -- |
| 1996 | 119,657 | 78,545 | 7,270 | 50 | -- |
| 1999 | 145,347 | 58,641 | 3,581 | 14 | 8 |
| 2001 | 93,433 | 56,133 | 4,006 | 22 | -- |
| 2003 | 146,018 | 101,421 | 9,855 | 0 | 0 |
| 2005 | 114,895 | 92,869 | 5,297 | 151 | 0 |
| 2007 | 139,806 | 130,661 | 9,823 | 0 | 0 |
| 2009 | 138,824 | 80,395 | 6,157 | 0 | 0 |
| 2011 | 149,836 | 73,718 | 12,086 | 0 | 0 |

Table 8.7. Survey size compositions for flathead sole. Numbers are in 1000's of individuals.

## a) Females.

| year | Length bin cutpoints (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
| 1984 | 567 | 3,098 | 3,337 | 7,306 | 14,170 | 20,489 | 29,800 | 45,645 | 63,475 | 76,302 | 69,592 | 48,288 | 28,087 | 17,406 | 10,172 |
| 1987 | 883 | 2,085 | 2,838 | 4,499 | 6,056 | 12,189 | 17,593 | 25,161 | 36,188 | 50,279 | 54,302 | 45,327 | 26,573 | 14,468 | 8,169 |
| 1990 | 1,269 | 3,347 | 6,036 | 6,002 | 9,283 | 15,446 | 19,887 | 24,583 | 37,464 | 46,874 | 55,347 | 60,532 | 52,045 | 30,967 | 12,912 |
| 1993 | 2,584 | 5,010 | 7,744 | 12,730 | 13,216 | 15,423 | 20,372 | 21,820 | 25,208 | 31,170 | 41,270 | 43,396 | 36,634 | 23,356 | 11,688 |
| 1996 | 3,360 | 6,318 | 10,043 | 14,294 | 16,104 | 19,497 | 21,345 | 25,059 | 29,741 | 34,375 | 37,894 | 40,168 | 33,867 | 23,395 | 12,837 |
| 1999 | 2,895 | 2,940 | 5,795 | 8,718 | 10,041 | 16,134 | 18,845 | 21,287 | 25,158 | 28,741 | 31,886 | 35,669 | 31,739 | 27,829 | 17,274 |
| 2001 | 2,777 | 4,699 | 5,728 | 8,070 | 9,822 | 7,348 | 9,242 | 12,441 | 17,973 | 20,460 | 29,033 | 26,925 | 24,106 | 18,520 | 11,972 |
| 2003 | 3,066 | 6,647 | 10,771 | 15,271 | 20,514 | 25,006 | 23,932 | 24,520 | 28,685 | 35,373 | 46,891 | 47,205 | 42,591 | 35,524 | 21,997 |
| 2005 | 4,988 | 7,391 | 10,305 | 14,894 | 20,011 | 22,229 | 27,086 | 30,483 | 33,432 | 38,116 | 37,285 | 35,590 | 34,358 | 24,141 | 14,563 |
| 2007 | 2,429 | 6,105 | 10,258 | 20,784 | 19,669 | 18,962 | 23,767 | 25,095 | 35,366 | 40,489 | 50,426 | 51,280 | 44,436 | 33,157 | 20,019 |
| 2009 | 4,488 | 3,880 | 7,286 | 10,748 | 14,502 | 16,873 | 24,109 | 28,080 | 35,058 | 28,175 | 38,788 | 43,340 | 39,670 | 24,950 | 13,921 |
| 2011 | 4,073 | 5,359 | 7,286 | 10,426 | 14,248 | 23,161 | 27,878 | 27,508 | 33,530 | 41,161 | 43,376 | 40,572 | 33,238 | 24,696 | 15,377 |

b) Males.

| year | Length bin cutpoints (cm) |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 |
| 1984 | 958 | 2,651 | 3,872 | 10,794 | 19,758 | 34,522 | 54,303 | 81,720 | 76,269 | 40,785 | 19,368 | 10,317 | 5,446 | 1,990 | 541 |
| 1987 | 1,257 | 2,623 | 3,876 | 5,448 | 6,445 | 15,112 | 26,507 | 48,973 | 53,591 | 33,389 | 14,119 | 7,290 | 3,550 | 1,731 | 480 |
| 1990 | 1,061 | 4,055 | 5,883 | 8,099 | 11,657 | 19,990 | 29,710 | 45,839 | 65,958 | 73,288 | 42,626 | 12,664 | 3,977 | 850 | 386 |
| 1993 | 2,205 | 5,315 | 9,757 | 12,897 | 16,987 | 23,213 | 29,095 | 39,372 | 50,735 | 54,631 | 36,488 | 12,636 | 5,513 | 2,599 | 343 |
| 1996 | 4,039 | 6,250 | 9,608 | 14,129 | 18,421 | 22,021 | 27,807 | 37,472 | 49,772 | 52,356 | 41,352 | 17,459 | 5,026 | 1,607 | 464 |
| 1999 | 2,484 | 4,313 | 7,246 | 11,893 | 17,227 | 21,067 | 30,364 | 42,405 | 59,243 | 60,992 | 49,672 | 24,469 | 7,967 | 1,647 | 1,492 |
| 2001 | 2,519 | 5,015 | 7,128 | 8,810 | 10,981 | 13,831 | 17,031 | 27,453 | 37,617 | 39,651 | 36,558 | 19,205 | 6,125 | 2,013 | 432 |
| 2003 | 4,634 | 6,574 | 11,065 | 17,329 | 24,994 | 31,230 | 36,233 | 41,029 | 54,997 | 57,972 | 53,126 | 33,017 | 14,061 | 4,857 | 990 |
| 2005 | 4,727 | 7,283 | 12,201 | 15,830 | 23,301 | 33,863 | 45,026 | 49,439 | 52,297 | 49,895 | 37,689 | 24,343 | 9,653 | 2,244 | 543 |
| 2007 | 4,193 | 6,756 | 13,904 | 23,942 | 25,572 | 25,987 | 33,841 | 43,613 | 53,834 | 57,061 | 48,581 | 34,411 | 14,458 | 5,925 | 1,711 |
| 2009 | 3,558 | 4,350 | 8,914 | 16,447 | 23,573 | 31,578 | 41,192 | 46,564 | 52,118 | 56,359 | 47,310 | 28,938 | 11,366 | 3,234 | 745 |
| 2011 | 3,572 | 3,931 | 7,738 | 14,131 | 26,323 | 34,825 | 41,670 | 46,065 | 55,727 | 55,017 | 43,498 | 24,163 | 9,663 | 3,714 | 1,518 |

Table 8.8. Survey age compositions for flathead sole. Numbers are in 1000’s of individuals.

## a) Females.

| year | Age bins |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| 1984 | 2,554 | 5,732 | 6,984 | 36,300 | 74,155 | 104,300 | 74,810 | 47,661 | 24,199 | 24,848 | 4,627 | 2,992 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 5,347 | 5,317 | 22,314 | 39,700 | 22,879 | 41,042 | 42,977 | 12,137 | 29,626 | 31,845 | 39,978 | 45,470 | 16,957 | 16,633 | 3,215 | 9,072 | 0 | 1,850 |
| 1993 | 12,043 | 23,746 | 18,705 | 18,484 | 22,728 | 23,396 | 24,017 | 49,392 | 25,997 | 22,142 | 19,556 | 17,817 | 4,674 | 10,333 | 10,345 | 5,432 | 758 | 9,712 |
| 1996 | 14,353 | 40,180 | 36,747 | 26,716 | 45,246 | 32,697 | 20,360 | 22,297 | 24,929 | 16,811 | 17,244 | 14,740 | 6,557 | 12,507 | 2,794 | 4,049 | 803 | 1,423 |
| 1999 | 6,932 | 13,068 | 23,768 | 20,784 | 21,354 | 37,554 | 32,344 | 30,160 | 22,069 | 19,428 | 20,935 | 9,843 | 11,778 | 7,231 | 6,379 | 7,171 | 1,791 | 4,052 |
| 2001 | 11,791 | 19,800 | 14,024 | 15,168 | 28,225 | 28,937 | 16,633 | 21,531 | 19,368 | 7,060 | 11,883 | 6,196 | 4,593 | 4,184 | 4,939 | 2,684 | 1,522 | 5,470 |
| 2003 | 27,825 | 39,592 | 50,233 | 52,481 | 13,806 | 37,912 | 43,306 | 50,772 | 16,791 | 14,290 | 10,785 | 24,386 | 3,205 | 2,332 | 382 | 4,405 | 4,587 | 3,712 |
| 2005 | 21,097 | 46,779 | 48,192 | 56,383 | 33,181 | 23,400 | 32,891 | 24,245 | 16,342 | 14,216 | 9,983 | 10,575 | 9,960 | 4,152 | 5,346 | 2,470 | 1,387 | 7,602 |
| 2007 | 19,430 | 40,782 | 39,391 | 31,794 | 27,647 | 56,539 | 51,838 | 24,195 | 28,469 | 23,276 | 12,987 | 28,467 | 12,684 | 6,019 | 12,587 | 4,513 | 1,923 | 8,673 |
| 2009 | 13,273 | 36,231 | 29,081 | 50,468 | 37,877 | 21,776 | 22,589 | 34,692 | 30,904 | 16,305 | 18,961 | 8,881 | 4,319 | 6,387 | 5,712 | 2,045 | 882 | 3,118 |

b) Males.

| year | Age bins |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 |
| 1984 | 0 | 10,403 | 29,061 | 41,741 | 48,344 | 96,634 | 61,205 | 16,899 | 21,343 | 9,159 | 1,421 | 4,745 | 2,773 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 7,392 | 11,683 | 29,810 | 33,483 | 28,696 | 36,881 | 38,549 | 35,774 | 13,969 | 12,160 | 26,810 | 24,797 | 17,194 | 4,452 | 0 | 3,360 | 0 | 716 |
| 1993 | 8,357 | 31,800 | 26,579 | 38,034 | 21,547 | 48,187 | 19,371 | 15,327 | 30,706 | 12,449 | 14,387 | 13,195 | 6,993 | 5,117 | 1,902 | 0 | 68 | 7,770 |
| 1996 | 16,381 | 37,078 | 30,360 | 23,837 | 44,421 | 34,830 | 34,399 | 31,534 | 16,454 | 9,247 | 6,710 | 6,140 | 6,892 | 3,200 | 2,905 | 232 | 1,202 | 345 |
| 1999 | 9,794 | 16,733 | 31,988 | 29,956 | 21,892 | 61,304 | 44,990 | 33,109 | 26,041 | 22,030 | 10,088 | 13,624 | 4,753 | 1,572 | 7,129 | 3,766 | 1,975 | 1,340 |
| 2001 | 11,844 | 27,780 | 18,472 | 26,731 | 36,430 | 23,497 | 15,761 | 16,711 | 11,266 | 13,890 | 8,441 | 8,339 | 2,636 | 2,857 | 2,444 | 1,522 | 1,689 | 3,273 |
| 2003 | 32,103 | 53,090 | 64,911 | 68,289 | 28,709 | 16,977 | 39,693 | 21,243 | 18,447 | 5,498 | 10,919 | 3,074 | 3,654 | 1,189 | 3,116 | 3,308 | 4,701 | 4,686 |
| 2005 | 29,361 | 48,735 | 39,610 | 56,586 | 60,672 | 38,238 | 22,515 | 14,721 | 15,575 | 3,836 | 14,354 | 10,745 | 1,379 | 6,296 | 1,724 | 2,006 | 2,560 | 336 |
| 2007 | 22,275 | 40,202 | 42,624 | 50,947 | 38,778 | 53,202 | 51,104 | 11,552 | 9,996 | 12,399 | 8,488 | 20,030 | 5,047 | 8,022 | 7,170 | 4,121 | 0 | 6,497 |
| 2009 | 17,074 | 53,552 | 42,754 | 64,591 | 46,282 | 26,472 | 26,736 | 26,812 | 8,453 | 20,983 | 10,507 | 8,842 | 6,307 | 3,325 | 1,617 | 3,289 | 174 | 6,332 |

Table 8.9a. Age-length transition matrices for female flathead sole. Values at a row/column combination correspond to the fraction of individuals at the age indicated by the row that fall into the length group indicated by the column.

| age | length cut 14 | 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.154 | 0.396 | 0.348 | 0.094 | 0.008 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.004 | 0.038 | 0.175 | 0.348 | 0.302 | 0.113 | 0.018 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0.002 | 0.018 | 0.09 | 0.235 | 0.321 | 0.229 | 0.086 | 0.017 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0 | 0.002 | 0.013 | 0.063 | 0.174 | 0.281 | 0.265 | 0.146 | 0.047 | 0.009 | 0.001 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0 | 0.002 | 0.013 | 0.056 | 0.15 | 0.251 | 0.263 | 0.173 | 0.071 | 0.018 | 0.003 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0 | 0.003 | 0.016 | 0.061 | 0.148 | 0.237 | 0.251 | 0.175 | 0.08 | 0.024 | 0.005 | 0.001 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.023 | 0.074 | 0.159 | 0.235 | 0.236 | 0.162 | 0.075 | 0.024 | 0.005 | 0.001 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.035 | 0.095 | 0.179 | 0.236 | 0.217 | 0.139 | 0.063 | 0.02 | 0.004 | 0.001 |
| 11 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.004 | 0.017 | 0.055 | 0.125 | 0.202 | 0.233 | 0.19 | 0.111 | 0.046 | 0.013 | 0.003 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.032 | 0.085 | 0.161 | 0.221 | 0.218 | 0.155 | 0.079 | 0.029 | 0.008 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.019 | 0.057 | 0.125 | 0.196 | 0.225 | 0.188 | 0.114 | 0.05 | 0.016 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.012 | 0.04 | 0.096 | 0.169 | 0.219 | 0.208 | 0.144 | 0.073 | 0.027 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.008 | 0.028 | 0.074 | 0.144 | 0.206 | 0.217 | 0.169 | 0.096 | 0.04 |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.02 | 0.058 | 0.122 | 0.191 | 0.22 | 0.187 | 0.117 | 0.054 |
| 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.004 | 0.015 | 0.046 | 0.104 | 0.175 | 0.218 | 0.2 | 0.136 | 0.068 |
| 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.011 | 0.037 | 0.09 | 0.161 | 0.213 | 0.209 | 0.151 | 0.081 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.009 | 0.03 | 0.078 | 0.148 | 0.207 | 0.215 | 0.164 | 0.093 |
| 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.025 | 0.069 | 0.137 | 0.201 | 0.218 | 0.175 | 0.103 |

Table 8.9b. Age-length transition matrices for male flathead sole. Values at a row/column combination correspond to the fraction of individuals at the age indicated by the row that fall into the length group indicated by the column.

| age | $\begin{array}{\|c\|} \hline \text { length cutp } \\ 14 \end{array}$ | (cm) 16 | 18 | 20 | 22 | 24 | 26 | 28 | 30 | 32 | 34 | 36 | 38 | 40 | 42 | 44 | 46 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | 0.706 | 0.265 | 0.029 | 0.001 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 | 0.038 | 0.205 | 0.405 | 0.279 | 0.067 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0.001 | 0.015 | 0.095 | 0.274 | 0.353 | 0.204 | 0.052 | 0.006 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 6 | 0 | 0.001 | 0.011 | 0.067 | 0.205 | 0.323 | 0.26 | 0.108 | 0.023 | 0.002 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 7 | 0 | 0 | 0.001 | 0.013 | 0.066 | 0.188 | 0.298 | 0.263 | 0.13 | 0.036 | 0.005 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 | 0 | 0 | 0 | 0.003 | 0.019 | 0.081 | 0.199 | 0.288 | 0.245 | 0.122 | 0.036 | 0.006 | 0.001 | 0 | 0 | 0 | 0 |
| 9 | 0 | 0 | 0 | 0.001 | 0.006 | 0.033 | 0.112 | 0.227 | 0.281 | 0.212 | 0.097 | 0.027 | 0.005 | 0 | 0 | 0 | 0 |
| 10 | 0 | 0 | 0 | 0 | 0.002 | 0.014 | 0.06 | 0.158 | 0.257 | 0.261 | 0.165 | 0.065 | 0.016 | 0.002 | 0 | 0 | 0 |
| 11 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.033 | 0.106 | 0.213 | 0.271 | 0.217 | 0.11 | 0.035 | 0.007 | 0.001 | 0 | 0 |
| 12 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.019 | 0.071 | 0.17 | 0.258 | 0.248 | 0.153 | 0.06 | 0.015 | 0.002 | 0 | 0 |
| 13 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.012 | 0.049 | 0.135 | 0.236 | 0.263 | 0.188 | 0.086 | 0.025 | 0.005 | 0.001 | 0 |
| 14 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.007 | 0.035 | 0.108 | 0.212 | 0.267 | 0.214 | 0.11 | 0.036 | 0.008 | 0.001 | 0 |
| 15 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.005 | 0.026 | 0.088 | 0.191 | 0.264 | 0.234 | 0.132 | 0.047 | 0.011 | 0.002 | 0 |
| 16 | 0 | 0 | 0 | 0 | 0 | 0 | 0.003 | 0.019 | 0.072 | 0.172 | 0.259 | 0.248 | 0.151 | 0.058 | 0.014 | 0.002 | 0 |
| 17 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.015 | 0.061 | 0.156 | 0.253 | 0.258 | 0.167 | 0.068 | 0.017 | 0.003 | 0 |
| 18 | 0 | 0 | 0 | 0 | 0 | 0 | 0.002 | 0.012 | 0.052 | 0.143 | 0.246 | 0.266 | 0.18 | 0.076 | 0.02 | 0.003 | 0 |
| 19 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.009 | 0.045 | 0.132 | 0.241 | 0.272 | 0.191 | 0.083 | 0.022 | 0.004 | 0 |
| 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0.001 | 0.008 | 0.039 | 0.122 | 0.235 | 0.278 | 0.201 | 0.089 | 0.024 | 0.004 | 0 |

Table 8.10. Age-specific schedules for flathead sole in the Gulf of Alaska. Maturity ogive is based on Stark (2004).

|  | Length (cm) |  | Weight (kg) |  | Maturity |
| :---: | ---: | ---: | ---: | ---: | ---: |
| Age | Males | Females | Males | Females | ogive |
| 3 | 15.2 | 17.8 | 0.03 | 0.05 | 0.0117 |
| 4 | 19.3 | 21.6 | 0.06 | 0.09 | 0.0251 |
| 5 | 22.6 | 24.9 | 0.10 | 0.14 | 0.0527 |
| 6 | 25.3 | 27.8 | 0.15 | 0.20 | 0.1076 |
| 7 | 27.6 | 30.2 | 0.20 | 0.26 | 0.2072 |
| 8 | 29.4 | 32.2 | 0.24 | 0.32 | 0.3615 |
| 9 | 30.8 | 34.0 | 0.28 | 0.38 | 0.5508 |
| 10 | 32.0 | 35.5 | 0.32 | 0.44 | 0.7265 |
| 11 | 33.0 | 36.8 | 0.35 | 0.49 | 0.8520 |
| 12 | 33.8 | 37.9 | 0.37 | 0.54 | 0.9257 |
| 13 | 34.5 | 38.8 | 0.40 | 0.58 | 0.9643 |
| 14 | 35.0 | 39.6 | 0.42 | 0.62 | 0.9832 |
| 15 | 35.4 | 40.3 | 0.43 | 0.66 | 0.9922 |
| 16 | 35.8 | 40.9 | 0.45 | 0.69 | 0.9964 |
| 17 | 36.1 | 41.4 | 0.46 | 0.72 | 0.9983 |
| 18 | 36.3 | 41.8 | 0.47 | 0.74 | 0.9992 |
| 19 | 36.5 | 42.2 | 0.48 | 0.76 | 0.9996 |
| 20 | 36.7 | 42.5 | 0.48 | 0.83 | 0.9998 |

Table 8.11. Likelihood multiplier settings.

| Fishery <br> length |  | biomass | Survey <br> length <br> compositions | age <br> compositions | early | ordinary |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | late | latment |
| :---: |
| 30 |

Table 8.12. Initial parameter values.

| Recruitment |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\overline{\operatorname{lm} R_{0}}$ |  |  |  |  |  |  |  |  |  |  |
| $\tau_{t}$ |  |  |  |  | 1967-2011: |  | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 |  |  |  |  |  |  |  |  |  |
| Fishing mortality |  |  |  |  |  |  |  |  |  |  |
| $\overline{\ln F} \quad 0$ |  |  |  |  |  |  |  |  |  |  |
| $\varepsilon_{t}$ |  | 1984-2011: |  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
|  | 0.0000 |  |  |  |  |  |  |  |  |  |
| Fishery Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.4 | 0.4 |  |  |  |  |  |  |  |  |
| $\mathrm{A}_{50}$ | 5 | 5 |  |  |  |  |  |  |  |  |
| scale par. | -- | -- |  |  |  |  |  |  |  |  |
| Survey Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.8 | 0.4 |  |  |  |  |  |  |  |  |
| $\mathrm{A}_{50}$ | 4 | 4 |  |  |  |  |  |  |  |  |
| scale par. | -- | -- |  |  |  |  |  |  |  |  |

Table 8.13. Log-likelihood components for the assessment model.

| likelihood <br> component | Base model |
| :--- | ---: |
| ordinary recruitment <br> 'late" recruitment <br> "early" recruitment <br> fishery catch <br> fishery size <br> composition <br> lurvey biomass <br> survey size <br> composition <br> survey age <br> composition | -16.95 |

Table 8.14. Final parameter estimates.

| Recruitment |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $\overline{\operatorname{lm} R_{0}}$ | 18.24449 |  |  |  |  |  |  |  |  |  |
| $\tau_{t}$ |  |  |  |  | 1967-2011: |  | -1.7027 | -0.8710 | -0.9444 | -1.0158 |
|  | -1.0749 | -1.1220 | -1.1564 | -1.1778 | -0.4903 | -0.8957 | -0.0161 | 0.4192 | 0.9251 | 0.7041 |
|  | 0.1579 | 0.2212 | 0.1807 | 0.1912 | 0.2059 | 0.0519 | 0.1357 | 0.3594 | 0.0905 | 0.1971 |
|  | 0.2515 | 0.4367 | 0.2083 | 0.3709 | 0.1787 | 0.2896 | 0.3149 | 0.0675 | 0.0392 | 0.4505 |
|  | 0.5796 | 0.6423 | 0.3292 | 0.3761 | 0.4638 | 0.6889 | 0.3681 | 0.7684 | 0.1948 | -0.2016 |
|  | -0.1898 |  |  |  |  |  |  |  |  |  |
| Fishing mortality |  |  |  |  |  |  |  |  |  |  |
| $\overline{\ln F}$ | -4.177899 |  |  |  |  |  |  |  |  |  |
| $\varepsilon_{t}$ |  | 1984-2011: |  | -0.2605 | -0.9900 | -1.9391 | -2.0771 | -0.9868 | -0.6984 | -0.0804 |
|  | 0.0742 | 0.2344 | 0.3850 | 0.4731 | 0.3067 | 0.6473 | 0.4136 | 0.0764 | -0.5787 | -0.0705 |
|  | 0.1199 | 0.2269 | 0.3440 | 0.3341 | 0.3831 | 0.5884 | 0.5916 | 0.6474 | 0.6867 | 0.7198 |
|  | 0.4292 |  |  |  |  |  |  |  |  |  |
| Fishery Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.8110 | 0.9033 |  |  |  |  |  |  |  |  |
| $\mathrm{A}_{50}$ | 10.42 | 10.12 |  |  |  |  |  |  |  |  |
| scale par. | -- | -- |  |  |  |  |  |  |  |  |
| Survey Selectivity |  |  |  |  |  |  |  |  |  |  |
|  | females | males |  |  |  |  |  |  |  |  |
| slope | 0.7660 | 0.8697 |  |  |  |  |  |  |  |  |
| $\mathrm{A}_{50}$ | 5.29 | 4.89 |  |  |  |  |  |  |  |  |
| scale par. | -- | -- |  |  |  |  |  |  |  |  |

Table 8.15. Estimated catch and survey biomass from the assessment model. Estimate uncertainty (standard deviation) based on model hessian.

| year | catch (t) |  | survey biomass (t) |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: |
|  | estimated | std dev | observed | estimated | std dev | observed |
| 1984 | 557 | 71 | 549 | 167,440 | 9,914 | 249,341 |
| 1985 | 331 | 42 | 320 | 180,910 | 9,958 |  |
| 1986 | 157 | 20 | 147 | 190,710 | 9,897 |  |
| 1987 | 162 | 21 | 151 | 197,680 | 9,769 | 177,546 |
| 1988 | 539 | 69 | 520 | 202,540 | 9,590 |  |
| 1989 | 769 | 98 | 747 | 205,510 | 9,375 |  |
| 1990 | 1,476 | 189 | 1,447 | 207,560 | 9,144 | 243,055 |
| 1991 | 1,746 | 224 | 1,717 | 208,580 | 8,903 |  |
| 1992 | 2,054 | 263 | 2,034 | 209,460 | 8,664 |  |
| 1993 | 2,378 | 305 | 2,366 | 210,260 | 8,438 | 188,690 |
| 1994 | 2,580 | 330 | 2,580 | 211,550 | 8,245 |  |
| 1995 | 2,179 | 278 | 2,181 | 213,100 | 8,077 |  |
| 1996 | 3,059 | 390 | 3,107 | 215,220 | 7,925 | 205,521 |
| 1997 | 2,419 | 308 | 2,446 | 216,120 | 7,782 |  |
| 1998 | 1,740 | 221 | 1,742 | 217,080 | 7,640 |  |
| 1999 | 919 | 117 | 900 | 218,030 | 7,503 | 207,590 |
| 2000 | 1,554 | 198 | 1,547 | 219,160 | 7,375 |  |
| 2001 | 1,902 | 241 | 1,911 | 197,550 | 6,547 | 170,745 |
| 2002 | 2,128 | 270 | 2,145 | 220,770 | 7,232 |  |
| 2003 | 2,393 | 304 | 2,425 | 223,820 | 7,290 | 257,294 |
| 2004 | 2,364 | 300 | 2,390 | 228,490 | 7,504 |  |
| 2005 | 2,477 | 313 | 2,530 | 233,850 | 7,878 | 213,221 |
| 2006 | 3,039 | 384 | 3,134 | 238,960 | 8,390 |  |
| 2007 | 3,076 | 389 | 3,163 | 243,190 | 9,026 | 280,290 |
| 2008 | 3,319 | 420 | 3,419 | 248,060 | 9,844 |  |
| 2009 | 3,538 | 447 | 3,658 | 252,770 | 10,886 | 225,377 |
| 2010 | 3,734 | 473 | 3,842 | 256,310 | 12,172 |  |
| 2011 | 2,847 | 361 | 2,339 | 257,190 | 13,666 | 235,639 |

Table 8.16a. Age 3+ population biomass estimated in this year's model, compared with estimates from previous assessments. Estimate uncertainty (standard deviation) based on model hessian.

| year | Age 3+ Biomass (1000's t) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2011 Assessment mean std dev |  | 2009 Assessment mean std dev |  | 2007 Assessment mean std dev |  | 2005 Assessment mean std dev |  |
| 1984 | 210 | 11 | 222 | 11 | 244 | 13 | 248 | 13 |
| 1985 | 221 | 11 | 233 | 11 | 254 | 13 | 256 | 13 |
| 1986 | 229 | 11 | 241 | 11 | 262 | 13 | 263 | 13 |
| 1987 | 234 | 11 | 247 | 11 | 266 | 12 | 266 | 12 |
| 1988 | 240 | 10 | 253 | 11 | 271 | 12 | 270 | 12 |
| 1989 | 243 | 10 | 256 | 11 | 274 | 12 | 271 | 12 |
| 1990 | 245 | 10 | 258 | 11 | 275 | 12 | 271 | 11 |
| 1991 | 247 | 10 | 259 | 11 | 276 | 11 | 271 | 11 |
| 1992 | 249 | 10 | 262 | 10 | 280 | 11 | 274 | 11 |
| 1993 | 251 | 10 | 264 | 10 | 281 | 11 | 273 | 11 |
| 1994 | 253 | 9 | 266 | 10 | 282 | 11 | 272 | 11 |
| 1995 | 255 | 9 | 267 | 10 | 283 | 11 | 272 | 11 |
| 1996 | 256 | 9 | 269 | 10 | 284 | 11 | 272 | 11 |
| 1997 | 257 | 9 | 269 | 10 | 283 | 11 | 269 | 11 |
| 1998 | 257 | 9 | 267 | 10 | 279 | 11 | 265 | 11 |
| 1999 | 256 | 9 | 263 | 10 | 274 | 11 | 258 | 11 |
| 2000 | 258 | 9 | 263 | 10 | 275 | 11 | 259 | 11 |
| 2001 | 262 | 9 | 266 | 10 | 279 | 12 | 262 | 12 |
| 2002 | 268 | 10 | 273 | 11 | 288 | 13 | 269 | 14 |
| 2003 | 273 | 10 | 281 | 12 | 297 | 14 | 280 | 16 |
| 2004 | 278 | 11 | 291 | 13 | 302 | 16 | 286 | 18 |
| 2005 | 282 | 11 | 302 | 15 | 308 | 18 | 292 | 20 |
| 2006 | 290 | 12 | 314 | 17 | 320 | 21 |  |  |
| 2007 | 294 | 14 | 325 | 19 | 322 | 24 |  |  |
| 2008 | 302 | 15 | 330 | 22 |  |  |  |  |
| 2009 | 305 | 17 | 330 | 24 |  |  |  |  |
| 2010 | 303 | 18 |  |  |  |  |  |  |
| 2011 | 297 | 20 |  |  |  |  |  |  |

Table 8.16b. Female spawning biomass estimated in this year's model, compared with estimates from previous assessments. Estimate uncertainty (standard deviation) based on model hessian.

| year | Female Spawning Stock Biomass (1000's t) |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2011 Ass mean | ment std dev | $\begin{aligned} & 2009 \text { Ass } \\ & \text { mean } \end{aligned}$ | sment <br> std dev | 2007 Ass mean | sment std dev | 2005 Ass mean | sment std dev |
| 1984 | 49 | 3 | 53 | 4 | 61 | 4 | 65 | 4 |
| 1985 | 59 | 4 | 63 | 4 | 73 | 4 | 76 | 4 |
| 1986 | 68 | 4 | 73 | 4 | 83 | 5 | 85 | 5 |
| 1987 | 76 | 4 | 80 | 4 | 90 | 5 | 91 | 5 |
| 1988 | 81 | 4 | 85 | 4 | 94 | 5 | 95 | 5 |
| 1989 | 84 | 4 | 88 | 4 | 96 | 5 | 97 | 5 |
| 1990 | 86 | 4 | 90 | 4 | 97 | 5 | 97 | 5 |
| 1991 | 86 | 4 | 91 | 4 | 97 | 5 | 97 | 4 |
| 1992 | 86 | 4 | 91 | 4 | 97 | 4 | 97 | 4 |
| 1993 | 86 | 4 | 91 | 4 | 97 | 4 | 96 | 4 |
| 1994 | 86 | 4 | 91 | 4 | 97 | 4 | 95 | 4 |
| 1995 | 86 | 4 | 91 | 4 | 97 | 4 | 95 | 4 |
| 1996 | 87 | 4 | 92 | 4 | 98 | 4 | 95 | 4 |
| 1997 | 87 | 3 | 92 | 4 | 98 | 4 | 95 | 4 |
| 1998 | 88 | 3 | 93 | 4 | 99 | 4 | 95 | 4 |
| 1999 | 89 | 3 | 94 | 4 | 99 | 4 | 95 | 4 |
| 2000 | 90 | 3 | 95 | 4 | 100 | 4 | 95 | 4 |
| 2001 | 91 | 3 | 95 | 4 | 100 | 4 | 94 | 4 |
| 2002 | 91 | 3 | 95 | 4 | 99 | 4 | 93 | 4 |
| 2003 | 91 | 3 | 94 | 4 | 98 | 4 | 92 | 4 |
| 2004 | 91 | 3 | 93 | 4 | 97 | 4 | 91 | 4 |
| 2005 | 92 | 3 | 93 | 4 | 98 | 4 | 91 | 5 |
| 2006 | 93 | 4 | 95 | 4 | 100 | 5 |  |  |
| 2007 | 95 | 4 | 98 | 5 | 103 | 5 |  |  |
| 2008 | 97 | 4 | 102 | 5 |  |  |  |  |
| 2009 | 99 | 4 | 106 | 6 |  |  |  |  |
| 2010 | 101 | 5 |  |  |  |  |  |  |
| 2011 | 102 | 5 |  |  |  |  |  |  |

Table 8.17. Age 3 recruitment values as estimated in this year's model, compared with estimates from previous assessments. Estimate uncertainty (standard deviation) based on model hessian.

| Year | 2011 Assessment |  | 2009 Assessment |  | 2007 Assessment |  | 2005 Assessment |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Mean | Std Dev | Mean | Std Dev | Mean | Std Dev | Mean | Std Dev |
| 1984 | 203 | 29 | 215 | 31 | 165 | 36 | 163 | 35 |
| 1985 | 206 | 27 | 216 | 29 | 247 | 43 | 241 | 42 |
| 1986 | 177 | 24 | 185 | 26 | 239 | 39 | 233 | 38 |
| 1987 | 192 | 25 | 199 | 27 | 180 | 32 | 175 | 32 |
| 1988 | 240 | 29 | 255 | 31 | 269 | 39 | 259 | 38 |
| 1989 | 184 | 25 | 196 | 27 | 211 | 34 | 201 | 33 |
| 1990 | 204 | 25 | 212 | 27 | 224 | 34 | 212 | 33 |
| 1991 | 216 | 26 | 220 | 29 | 238 | 36 | 222 | 34 |
| 1992 | 260 | 27 | 281 | 31 | 326 | 42 | 305 | 40 |
| 1993 | 207 | 24 | 211 | 26 | 188 | 33 | 175 | 31 |
| 1994 | 243 | 26 | 252 | 30 | 215 | 38 | 200 | 36 |
| 1995 | 201 | 23 | 215 | 27 | 272 | 42 | 253 | 39 |
| 1996 | 224 | 24 | 223 | 28 | 228 | 38 | 211 | 36 |
| 1997 | 230 | 25 | 211 | 28 | 212 | 39 | 193 | 36 |
| 1998 | 179 | 22 | 169 | 25 | 154 | 34 | 140 | 31 |
| 1999 | 174 | 22 | 162 | 24 | 133 | 32 | 121 | 29 |
| 2000 | 263 | 28 | 247 | 32 | 351 | 54 | 320 | 52 |
| 2001 | 299 | 31 | 332 | 38 | 349 | 57 | 327 | 57 |
| 2002 | 319 | 34 | 368 | 44 | 366 | 69 | 359 | 73 |
| 2003 | 233 | 29 | 286 | 41 | 337 | 75 | 352 | 86 |
| 2004 | 244 | 33 | 340 | 53 | 167 | 80 | 192 | 96 |
| 2005 | 267 | 37 | 323 | 57 | 302 | 114 | 242 | 105 |
| 2006 | 334 | 49 | 365 | 78 | 447 | 174 |  |  |
| 2007 | 242 | 45 | 311 | 90 | 148 | 113 |  |  |
| 2008 | 362 | 69 | 214 | 122 |  |  |  |  |
| 2009 | 204 | 64 | 174 | 98 |  |  |  |  |
| 2010 | 137 | 86 |  |  |  |  |  |  |
| 2011 | 139 | 77 |  |  |  |  |  |  |

Table 8.18. Projected catch (t) for the seven projection scenarios.

|  |  | Catch (t) |  |  |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | :---: |
| year | scenario 1 scenario 2 | scenario 3 scenario 4 | scenario $\mathbf{5}$ | scenario 6 scenario 7 |  |  |  |  |
| 2011 | 2,891 | 2,891 | 2,891 | 2,891 | 2,891 | 2,891 | 2,891 |  |
| 2012 | 47,407 | 47,407 | 25,718 | 3,508 | 0 | 59,380 | 47,407 |  |
| 2013 | 36,167 | 36,167 | 22,773 | 3,566 | 0 | 41,437 | 36,167 |  |
| 2014 | 29,073 | 29,073 | 20,507 | 3,611 | 0 | 31,370 | 36,568 |  |
| 2015 | 24,241 | 24,241 | 18,598 | 3,621 | 0 | 25,204 | 28,127 |  |
| 2016 | 20,522 | 20,522 | 16,801 | 3,576 | 0 | 20,508 | 22,430 |  |
| 2017 | 17,570 | 17,570 | 15,093 | 3,477 | 0 | 15,707 | 16,855 |  |
| 2018 | 15,154 | 15,154 | 13,700 | 3,356 | 0 | 14,215 | 14,601 |  |
| 2019 | 14,394 | 14,394 | 12,832 | 3,255 | 0 | 14,462 | 14,578 |  |
| 2020 | 14,649 | 14,649 | 12,450 | 3,191 | 0 | 15,351 | 15,366 |  |
| 2021 | 15,114 | 15,114 | 12,355 | 3,161 | 0 | 16,107 | 16,092 |  |
| 2022 | 15,459 | 15,459 | 12,374 | 3,149 | 0 | 16,497 | 16,481 |  |
| 2023 | 15,646 | 15,646 | 12,419 | 3,147 | 0 | 16,620 | 16,609 |  |
| 2024 | 15,721 | 15,721 | 12,456 | 3,146 | 0 | 16,623 | 16,617 |  |

Table 8.19. Female spawning biomass (t) for the seven projection scenarios. The values of $B_{40 \%}$ and $B_{35 \%}$ are $41,547 \mathrm{t}$ and $36,354 \mathrm{t}$, respectively.

| Female spawning biomass (t) |  |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year | scenario 1 | scenario 2 | scenario 3 | scenario 4 | scenario 5 | scenario 6 scenario 7 |  |
| 2011 | 102,433 | 102,433 | 102,433 | 102,433 | 102,433 | 102,433 | 102,433 |
| 2012 | 104,301 | 104,301 | 104,301 | 104,301 | 104,301 | 104,301 | 104,301 |
| 2013 | 82,277 | 82,277 | 93,589 | 105,304 | 107,165 | 76,100 | 82,277 |
| 2014 | 67,620 | 67,620 | 84,618 | 105,193 | 108,749 | 59,471 | 67,620 |
| 2015 | 56,943 | 56,943 | 76,434 | 103,581 | 108,651 | 48,568 | 53,132 |
| 2016 | 48,816 | 48,816 | 68,962 | 100,798 | 107,189 | 40,910 | 43,404 |
| 2017 | 43,214 | 43,214 | 62,879 | 97,798 | 105,301 | 36,234 | 37,401 |
| 2018 | 40,306 | 40,306 | 58,790 | 95,379 | 103,765 | 35,098 | 35,500 |
| 2019 | 39,805 | 39,805 | 56,645 | 93,837 | 102,886 | 35,735 | 35,851 |
| 2020 | 40,392 | 40,392 | 55,819 | 92,952 | 102,472 | 36,828 | 36,839 |
| 2021 | 41,105 | 41,105 | 55,666 | 92,533 | 102,402 | 37,639 | 37,622 |
| 2022 | 41,596 | 41,596 | 55,755 | 92,363 | 102,502 | 38,029 | 38,010 |
| 2023 | 41,853 | 41,853 | 55,894 | 92,341 | 102,705 | 38,142 | 38,131 |
| 2024 | 41,966 | 41,966 | 56,005 | 92,299 | 102,826 | 38,146 | 38,140 |

Table 8.20. Fishing mortality for the seven projection scenarios.

|  | Fishing mortality |  |  |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| year | scenario 1 scenario 2 | scenario 3 scenario 4 scenario 5 scenario 6 scenario 7 |  |  |  |  |  |
| 2011 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 | 0.02 |
| 2012 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.59 | 0.45 |
| 2013 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.59 | 0.45 |
| 2014 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.59 | 0.59 |
| 2015 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.59 | 0.59 |
| 2016 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.58 | 0.59 |
| 2017 | 0.45 | 0.45 | 0.23 | 0.03 | 0.00 | 0.51 | 0.53 |
| 2018 | 0.43 | 0.43 | 0.23 | 0.03 | 0.00 | 0.50 | 0.50 |
| 2019 | 0.43 | 0.43 | 0.23 | 0.03 | 0.00 | 0.51 | 0.51 |
| 2020 | 0.43 | 0.43 | 0.23 | 0.03 | 0.00 | 0.52 | 0.52 |
| 2021 | 0.43 | 0.43 | 0.23 | 0.03 | 0.00 | 0.53 | 0.53 |
| 2022 | 0.44 | 0.44 | 0.23 | 0.03 | 0.00 | 0.54 | 0.54 |
| 2023 | 0.44 | 0.44 | 0.23 | 0.03 | 0.00 | 0.54 | 0.54 |
| 2024 | 0.44 | 0.44 | 0.23 | 0.03 | 0.00 | 0.54 | 0.54 |

Table 8.21. Prohibited species catches (PSC) in the flathead sole target fishery.
a) Crab PSC.

| year | PSC in target fishery (\#) |  |  |  |  | fraction of total PSC |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | King Crab |  |  | Tanner Crab |  | King Crab |  |  | Tanner Crab |  |
|  | Blue | Golden | Red | Bairdi | Opilio | Blue | Golden | Red | Bairdi | Opilio |
| 2003 | 0 | 533 | 0 | 17,330 | 174 | 0.0\% | 64.6\% | 0.0\% | 11.7\% | 11.9\% |
| 2004 | 0 | 0 | 0 | 7,275 | 0 | 0.0\% | 0.0\% | 0.0\% | 10.1\% | 0.0\% |
| 2005 | 0 | 0 | 0 | 32,471 | 0 | 0.0\% | 0.0\% | 0.0\% | 13.9\% | 0.0\% |
| 2006 | 0 | 0 | 0 | 25,884 | 0 | 0.0\% | 0.0\% | 0.0\% | 6.3\% | 0.0\% |
| 2007 | 0 | 0 | 0 | 254 | 0 | 0.0\% | 0.0\% | 0.0\% | 0.1\% | 0.0\% |
| 2008 | 0 | 0 | 0 | 7,077 | 272 | 0.0\% | 0.0\% | 0.0\% | 1.9\% | 13.6\% |
| 2009 | 0 | 0 | 0 | 7,688 | 0 | 0.0\% | 0.0\% | 0.0\% | 3.0\% | 0.0\% |
| 2010 | 0 | 0 | 0 | 6,498 | 0 | 0.0\% | 0.0\% | 0.0\% | 2.5\% | 0.0\% |
| 2011 | 0 | 0 | 0 | 5,220 | 0 | 0.0\% | 0.0\% | 0.0\% | 5.8\% | 0.0\% |

b) Halibut PSC.

| Year | directed fishery <br> halibut PSC $(\mathbf{k g})$ | \% total halibut <br> PSC |
| :---: | :---: | :---: |
| 2003 | 203,807 | $3.7 \%$ |
| 2004 | 101,755 | $1.7 \%$ |
| 2005 | 52,798 | $1.1 \%$ |
| 2006 | 36,528 | $0.6 \%$ |
| 2007 | 27,029 | $0.6 \%$ |
| 2008 | 91,959 | $1.5 \%$ |
| 2009 | 97,728 | $2.1 \%$ |
| 2010 | 256,572 | $5.7 \%$ |
| 2011 | 91,931 | $2.1 \%$ |

c) Salmon PSC. (Salmon PSC for 2011 was unavailable at time of document preparation.)

| Year | Chinook <br> fraction of <br> total |  | Non-Chinook |  |
| :---: | :---: | :---: | :---: | :---: |
| PSC (\#) | fraction of <br> total |  |  |  |
| 2003 | 612 | $3.9 \%$ | 19 | $0.2 \%$ |
| 2004 | 1,389 | $7.8 \%$ | 90 | $1.6 \%$ |
| 2005 | 16 | $0.1 \%$ | 0 | $0.0 \%$ |
| 2006 | 56 | $0.3 \%$ | 0 | $0.0 \%$ |
| 2007 | 0 | $0.0 \%$ | 0 | $0.0 \%$ |
| 2008 | 0 | $0.0 \%$ | 0 | $0.0 \%$ |
| 2009 | 118 | $1.5 \%$ | 0 | $0.0 \%$ |
| 2010 | 498 | $0.9 \%$ | 0 | $0.0 \%$ |

Table 8.22. Catch of nontarget species in the flathead sole target fishery, expressed as the fraction of species catch by all fisheries in the FMP.

| Nontarget Species Group | 2011 | 2010 | 2009 | 2008 | $\begin{aligned} & \text { Year } \\ & 2007 \end{aligned}$ | 2006 | 2005 | 2004 | 2003 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Benthic urochordata | 8.5\% | 14.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.8\% | 0.0\% |
| Birds | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Bivalves | 3.0\% | 4.2\% | 2.5\% | 2.0\% | 1.7\% | 2.8\% | 6.4\% | 5.6\% | 4.2\% |
| Brittle star unidentified | 2.4\% | 9.7\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.7\% |
| Capelin | 0.0\% | 0.0\% | 0.0\% | 0.0\% | -- | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Corals Bryozoans | 0.0\% | 6.9\% | 0.0\% | 1.3\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.8\% |
| Dark Rockfish | 0.0\% | 0.0\% | 9.3\% | 0.0\% | -- | -- | -- | -- | -- |
| Eelpouts | 9.2\% | 11.3\% | 0.0\% | 6.0\% | 7.1\% | 4.9\% | 6.9\% | 6.3\% | 9.3\% |
| Eulachon | 1.8\% | 3.9\% | 2.6\% | 2.1\% | 0.0\% | 1.2\% | 2.9\% | 0.4\% | 1.4\% |
| Giant Grenadier | 0.0\% | 5.1\% | 4.4\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Greenlings | 2.4\% | 5.5\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.5\% | 0.3\% |
| Grenadier | 6.4\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.3\% | 5.8\% | 3.3\% | 4.5\% |
| Gunnels | -- | -- | -- | 2.8\% | -- | 100.0\% | -- | -- | 0.0\% |
| Hermit crab unidentified | 2.7\% | 3.6\% | 0.0\% | 1.8\% | 0.0\% | 0.0\% | 0.1\% | 0.6\% | 4.4\% |
| Invertebrate unidentified | 2.6\% | 0.0\% | 4.8\% | 2.1\% | 0.0\% | 1.6\% | 0.0\% | 2.6\% | 2.4\% |
| Lanternfishes (myctophidae) | -- | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Large Sculpins | 1.0\% | 2.0\% | 1.6\% | 0.6\% | 0.8\% | 1.0\% | 2.7\% | 3.0\% | 1.9\% |
| Misc crabs | 2.1\% | 4.2\% | 3.1\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 3.6\% | 0.9\% |
| Misc crustaceans | 0.0\% | 0.0\% | 0.0\% | -- | -- | -- | 0.0\% | 0.0\% | 0.0\% |
| Misc deep fish | -- | -- | -- | 0.0\% | -- | -- | -- | -- | -- |
| Misc fish | 1.1\% | 2.5\% | 1.5\% | 0.9\% | 0.7\% | 0.5\% | 2.0\% | 1.7\% | 2.3\% |
| Misc inverts (worms etc) | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | -- | -- |
| Octopus | 0.0\% | 1.5\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.1\% | 0.2\% |
| Other osmerids | 2.6\% | 0.0\% | 0.0\% | 0.6\% | 0.0\% | 2.7\% | 0.0\% | 0.0\% | 3.1\% |
| Other Sculpins | 2.1\% | 4.7\% | 2.3\% | 0.9\% | 0.9\% | 2.0\% | 4.2\% | 0.4\% | 3.3\% |
| Pacific Sand lance | 0.0\% | -- | 0.0\% | 0.0\% | -- | 0.0\% | 0.0\% | 0.0\% | 0.0\% |
| Pandalid shrimp | 2.1\% | 6.1\% | 2.9\% | 1.6\% | 0.9\% | 2.2\% | 3.7\% | 0.5\% | 4.8\% |
| Polychaete unidentified | 39.2\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | -- | 51.1\% | -- | 0.0\% |
| Scypho jellies | 0.0\% | 2.3\% | 3.1\% | 0.0\% | 0.0\% | 1.0\% | 0.0\% | 1.0\% | 0.8\% |
| Sea anemone unidentified | 2.2\% | 3.2\% | 2.2\% | 0.0\% | 0.0\% | 2.0\% | 0.0\% | 1.6\% | 3.8\% |
| Sea pens whips | 8.6\% | 14.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 7.2\% | 0.0\% |
| Sea star | 1.3\% | 3.1\% | 2.1\% | 1.0\% | 1.3\% | 0.8\% | 2.9\% | 2.5\% | 2.8\% |
| Shark, Other | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 1.3\% | 4.3\% | 3.9\% | 1.2\% |
| Shark, Pacific sleeper | 0.0\% | 1.6\% | 0.2\% | 0.0\% | 1.6\% | 1.0\% | 1.1\% | 0.8\% | 2.8\% |
| Shark, salmon | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.9\% | 0.0\% |
| Shark, spiny dogfish | 0.1\% | 2.0\% | 0.1\% | 0.0\% | 0.0\% | 0.1\% | 0.7\% | 0.0\% | 4.1\% |
| Skate, Alaska | 0.7\% | 0.1\% | -- | -- | -- | -- | -- | -- | -- |
| Skate, Aleutian | 0.2\% | -- | -- | -- | -- | -- | -- | -- | -- |
| Skate, Big | 1.8\% | 4.1\% | 2.6\% | 1.9\% | 0.7\% | 2.6\% | 2.5\% | 1.3\% | -- |
| Skate, Longnose | 1.8\% | 4.0\% | 2.0\% | 1.4\% | 1.2\% | 1.3\% | 1.3\% | 1.8\% | 1.1\% |
| Skate, Other | 1.7\% | 4.2\% | 2.2\% | 1.5\% | 1.8\% | 1.9\% | 4.2\% | 2.9\% | 4.4\% |
| Skate, Whiteblotched | 0.0\% | -- | -- | -- | -- | -- | -- | -- | -- |
| Snails | 2.0\% | 4.0\% | 2.8\% | 1.4\% | 0.8\% | 1.4\% | 4.3\% | 0.9\% | 4.7\% |
| Sponge unidentified | 2.1\% | 3.8\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 3.1\% | 4.9\% |
| Squid | 0.1\% | 0.1\% | 0.2\% | 0.0\% | 0.0\% | 0.1\% | 0.0\% | 0.1\% | 2.7\% |
| Stichaeidae | 12.0\% | 13.5\% | 6.8\% | 0.6\% | 0.0\% | 8.3\% | 20.8\% | 7.5\% | 19.0\% |
| Surf smelt | -- | -- | -- | 0.0\% | -- | -- | 0.0\% | 0.0\% | -- |
| urchins dollars cucumbers | 2.0\% | 3.9\% | 0.0\% | 0.0\% | 0.0\% | 0.0\% | 4.0\% | 0.9\% | 2.9\% |

Table 8.23. Catch of non-prohibited species in the flathead sole target fishery. The species accounting for the two largest totals are highlighted.

| Species | 2011 |  | 2010 |  | 2009 |  | 2008 |  | 2007 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ | total (t) | $\begin{gathered} \% \\ \text { retained } \end{gathered}$ |
| all sculpins, sharks, squid, octopus | 6 | 78\% | 22 | 20\% | 9 | 57\% | 14 | 74\% | 35 | 0\% |
| arrowtooth flounder | 779 | 7\% | 2650 | 6\% | 1337 | 10\% | 801 | 21\% | 723 | 10\% |
| Atka mackerel | 18 | 99\% | 10 | 98\% | 17 | 99\% | 3 | 98\% | 36 | 71\% |
| big skate | 39 | 94\% | 112 | 92\% | 53 | 96\% | 66 | 84\% | 23 | 99\% |
| Dover sole and turbot | 1 | 100\% | 45 | 48\% | 18 | 8\% | 4 | 98\% | 1 | 0\% |
| flathead sole | 367 | 97\% | 1242 | 96\% | 696 | 98\% | 572 | 92\% | 423 | 90\% |
| longnose skate | 12 | 95\% | 30 | 97\% | 24 | 66\% | 11 | 81\% | 13 | 19\% |
| northern rockfish | 1 | 89\% | 6 | 53\% | 1 | 89\% | 0 | 100\% | 2 | 0\% |
| other rockfish complex | 0 | -- | 2 | 4\% | 0 | -- | 2 | 53\% | 0 | 99\% |
| pacific cod | 108 | 94\% | 297 | 81\% | 279 | 97\% | 125 | 84\% | 131 | 90\% |
| pelagic shelf rockfish | 1 | 82\% | 9 | 72\% | 4 | 94\% | 2 | 100\% | 2 | 0\% |
| pollock | 57 | 94\% | 319 | 46\% | 135 | 81\% | 45 | 97\% | 27 | 99\% |
| POP | 2 | 6\% | 74 | 7\% | 2 | 5\% | 2 | 2\% | 11 | 13\% |
| rex sole | 77 | 86\% | 397 | 96\% | 184 | 94\% | 86 | 98\% | 110 | 98\% |
| rougheye | 2 | 16\% | 15 | 94\% | 3 | 44\% | 0 | 42\% | 0 | 100\% |
| sablefish | 8 | 98\% | 13 | 98\% | 19 | 77\% | 1 | 61\% | 4 | 100\% |
| shallow water flatfish | 56 | 97\% | 122 | 98\% | 95 | 98\% | 41 | 98\% | 26 | 95\% |
| shortraker | 2 | 97\% | 1 | 78\% | 3 | 98\% | 0 | -- | 0 | -- |
| thornyhead | 5 | 100\% | 13 | 76\% | 8 | 100\% | 0 | 100\% | 7 | 100\% |
| unidentified skate | 9 | 52\% | 19 | 13\% | 13 | 49\% | 5 | 28\% | 20 | 64\% |

## Figures



Figure 8.1. Fishery catches for GOA flathead sole, 1984-2011 (as of Sept. 24, 2011).


Figure 8.2. Spatial patterns of fishery catches for GOA flathead sole, 2009-2011.


Figure 8.3. Spatial patterns of fishery catches for GOA flathead sole by quarter for 2010-2011 (through Sept. 24).


Figure 8.4. GOA survey biomass for flathead sole. Error bars represent 95\% lognormal confidence intervals. The GOA survey did not include the eastern gulf in 2001. The value shown here for 2001 has been corrected to account for this (see text).


Figure 8.5. Spatial patterns of CPUE for flathead sole in the GOA groundfish surveys for 2007, 2009 and 2011.
a) Length-at-age.

b) Weight -at-age.

c) Maturity-at-age (females).


Figure 8.6. Age-specific schedules for GOA flathead sole: females solid line, males dotted line.


Figure 8.7. Estimated and observed annual catches for GOA flathead sole for the assessment model. Estimated catch $=$ solid line, observed catch $=$ dotted line with circles.


Fig. 8.8. Estimated and observed survey biomass for GOA flathead sole for the assessment model. Estimated survey biomass = triangles, observed survey biomass = circles (error bars are approximate lognormal 95\% confidence intervals).


Figure 8.9. Fits to female GOA flathead sole fishery size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Length (cm)
Figure 8.10. Fits to male GOA flathead sole fishery size composition data for the assessment model. Dashed lines represent the model estimates, solid lines represent the data.


Figure 8.11. Model fits to the female GOA flathead sole survey size composition data. Dashed lines represent the model estimates, solid lines represent the data.


Figure 8.12. Model fits to the male GOA flathead sole survey size composition data. Dashed lines represent the model estimates, solid lines represent the data.


Figure 8.13. Model fits to the female survey GOA flathead sole age composition data. Dashed lines represent the model estimates, solid lines represent the data.


Figure 8.14. Model fits to the male survey GOA flathead sole age composition data. Dashed lines represent the model estimates, solid lines represent the data.


Figure 8.15. Estimated survey and fishery selectivities. Survey selectivities are plotted in red with a dotted line, fishery selectivities are plotted in black with asolid line. Male selectivity functions are plotted with a triangle symbol, female selectivity functions are plotted without a symbol.


Figure 8.16. Marginal posterior distributions based on MCMC integration for parameters related to fishery and survey age selectivity functions. "a $\mathrm{a}_{50}$ " denotes the parameter for the age at which the logistic selectivity function is $50 \%$. " $\beta$ " is related to the slope of the selectivity function at age $=a_{50}$. The maximum likelihood estimate for each parameter is indicated by the appropriately-colored vertical line.


Figure 8.17. Marginal posterior distributions based on MCMC integration for: median recruitment (upper left), median fishing mortality (upper right), total (age 3+) biomass in 2011 (lower left), spawning biomass in 2011 (lower middle), and recruitment in 2011 (lower right). The maximum likelihood estimate for each quantity is indicated by the vertical line.


Figure 8.18. Time series plots of estimated total (age 3+) biomass and spawning biomass (upper graph) and recruitment (lower graph). $95 \%$ credibility intervals based on marginal posterior distributions from MCMC integration for parameters related to fishery (top row) and survey (bottom row) age selectivity functions. The solid lines indicate time series of maximum likelihood estimates.


Figure 8.19. Upper: : Comparison of total (age 3+) biomass estimates from the current assessment with results from the 2009, 2007, and 2005 assessments. Lower: Comparison of spawning biomass estimates from the current assessment with results from the 2009, 2007, and 2005 assessments..


Figure 8.20. Upper: Estimated age 3 recruitments with approximate 95\% lognormal confidence intervals estimated from the model hessian. Horizontal line is mean recruitment. Lower: Comparison of recruitment estimates with results from the 2009, 2007, and 2005 assessments.


Figure 8.21. Marginal posterior distributions based on MCMC integration for several managementrelated quantities estimated in the model: $\mathrm{F}_{40 \%}$ and $\mathrm{F}_{35 \%}$, (upper left), $\mathrm{B}_{40 \%}$, and $\mathrm{B}_{35 \%}$ (upper right), and 2012 ABC, OFL (lower left). The value for each quantity associated with the maximum likelihood solution is indicated by the appropriately-colored vertical line.


Figure 8.22. Control rule plot of estimated fishing mortality versus estimated female spawning biomass for GOA flathead sole. $F_{O F L}=$ upper dashed line, $F_{\max A B C}=$ lower dashed line.


Figure 8.23a. Gulf of Alaska food web from the GOA ecosystem model (Aydin et al., 2007) highlighting adult flathead sole links to predators (blue boxes and lines) and prey (green boxes and lines). Box size reflects relative standing stock biomass.


Figure 8.23b. Gulf of Alaska food web from the GOA ecosystem model (Aydin et al., 2007) highlighting juveile flathead sole links to predators (blue boxes and lines) and prey (green boxes and lines). Box size reflects relative standing stock biomass.


Figure 8.24a. Diet composition for Gulf of Alaska adult flathead sole from the GOA ecosystem model (Aydin et al., 2007).


Figure 8.24b. Diet composition for Gulf of Alaska juvenile flathead sole from the GOA ecosystem model (Aydin et al., 2007).


Figure 8.25a. Decomposition of natural mortality for Gulf of Alaska adult flathead sole from the GOA ecosystem model (Aydin et al., 2007).

GOA FH. Sole_Juv mortality


Figure 8.25b. Decomposition of natural mortality for Gulf of Alaska juvenile flathead sole from the GOA ecosystem model (Aydin et al., 2007).

Chapter 8 Appendix A: Model Equations
Table A.1. List of quantities and their definitions as used in the model.

| Quantity | Definition |
| :---: | :---: |
| T | number of years in the model. |
| A | number of age classes (18). |
| L | number of length classes (18). |
| $T_{\text {min }}$ | model start year (1984). |
| $T_{\text {max }}$ | assessment year (2011). |
| $t$ | time index. |
| $a$ | age index ( $1 \leq a \leq A ; a=1$ corresponds to age at recruitment). |
| $x$ | sex index ( $1 \leq x \leq 2 ; 1=$ female, $2=$ male $)$. |
| l | length index ( $1 \leq l \leq L$; $l=1$ corresponds to minimum length class). |
| $\left\{t^{S}\right\}$ | set of years for which survey biomass data is available. |
| $\left\{t^{\text {F, }}\right\}$ | set of years for which fishery age composition data is available. |
| $\left\{t^{F, L}\right\}$ | set of years for which fishery length composition data is available. |
| $\left\{t^{S, A}\right\}$ | set of years for which survey age composition data is available. |
| $\left\{t^{S, L}\right\}$ | set of years for which survey length composition data is available. |
| $L^{\chi}{ }_{l, a}$ | elements of length-age conversion matrix (proportion of sex $x$ fish in age class $a$ that are in length class $l$ ). (fixed) |
| $w_{x, a}$ | mean body weight (kg) of sex $x$ fish in age group $a$. (fixed) |
| $\phi_{a}$ | proportion of females mature at age $a$. (fixed) |
| $\overline{\ln R_{0}}$ | mean value of log-transformed recruitment. (estimable) |
| $\tau_{t}$ | recruitment deviation in year $t$. (estimable) |
| $M_{x}$ | instantaneous natural mortality rate. (fixed) |
| $\overline{\ln F}$ | mean value of log-transformed fishing mortality. (estimable) |
| $\varepsilon_{t}$ | deviations in fishing mortality rate in year $t$. (estimable) |
| $R_{t}$ | recruitment in year $t$. |
| $N_{t, x, a}$ | number of fish of $\operatorname{sex} x$ and age class $a$ in year $t$. |
| $C_{t, x, a}$ | catch (number) of fish of sex $x$ and age class $a$ in year $t$. |
| $p^{F, A}{ }_{t, x, a}$ | proportion of the total catch in year $t$ that is sex $x$ and in age class $a$. |
| $p^{F, L}{ }_{t, x, l}$ | proportion of the total catch in year $t$ that is sex $x$ and in length class $l$. |
| $p^{S, A}{ }_{t, x, a}$ | proportion of the survey biomass in year $t$ that is sex $x$ and in age group a. |
| $p^{s, L}{ }_{t, x, l}$ | proportion of the survey biomass in year $t$ that is sex $x$ and in age group a. |
| $C_{t}$ | total catch (yield) in tons in year $t$. |
| $F_{t, x, a}$ | instantaneous fishing mortality rate for sex $x$ and age group $a$ in year $t$. |
| $Z_{t, x, a}$ | instantaneous total mortality for sex $x$ and age group $a$ in year $t$. |
| $s^{F U}{ }_{\chi, a}$ | unnormalized fishery selectivity for sex $x$ and age group $a$. |
| $s^{S U}{ }_{\chi, a}$ | unnormalized survey selectivity for sex $x$ and age group $a$. |
| $s^{F N}{ }_{x, a}$ | normalized fishery selectivity for sex $x$ and age group $a$. |
| $s^{S N}{ }_{x, a}$ | normalized survey selectivity for sex $x$ and age group $a$. |

Table A.2. Model equations describing the model populations dynamics.

| Equation | Description |
| :---: | :---: |
| $\tau_{t} \sim N\left(0, \sigma_{R}^{2}\right)$ | Random deviate associated with recruitment. |
| $N_{t, x, 1}=R_{t}=\exp \left(\overline{\ln R_{0}}+\tau_{t}\right)$ | Recruitment (assumed equal for males and females). |
| $N_{t+1, x, a+1}=N_{t, x, a} e^{-Z_{t, x, a}}$ | Numbers at age. |
| $N_{t+1, \chi, A}=N_{t, \chi, A-1} e^{-Z_{t, x, A-1}}+N_{t, x, A} e^{-Z_{t, x, A}}$ | Numbers in "plus" group. |
| $C_{t, x, a}=\frac{F_{t, x, a}}{Z_{t, x, a}}\left(1-e^{-Z_{t, x, a}}\right) N_{t, x, a}$ | Catch at age (in numbers caught). |
| $C_{t}=\sum_{x=1}^{2} \sum_{a=1}^{A} w_{x, a} C_{t, x, a}$ | Total catch in tons (i.e., yield). |
| $F S B_{t}=\sum_{a=1}^{A} w_{1, a} \phi_{a} N_{t, 1, a}$ | Female spawning biomass. |
| $Z_{t, x, a}=F_{t, x, a}+M$ | Total mortality. |
| $F_{t, x, a}=s_{x, a}^{F} \cdot \exp \left(\overline{\overline{\ln } F}+\varepsilon_{t}\right)$ | Fishing mortality. |
| $\varepsilon_{t} \sim N\left(0, \sigma_{F}^{2}\right)$ | Random deviate associated with fishing mortality. |
| $s_{x, a}^{F U}=\frac{1}{1+e^{\left(-b_{x}^{F}\left(a g e-{ }_{50} A_{x}^{F}\right)\right)}}$ | Unnormalized fishery selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x, a}^{S U}=\frac{1}{1+e^{\left(-b_{x}^{S}\left(a g e-50 A_{x}^{\left.\left.A_{x}^{S}\right)\right)}\right.\right.}}$ | Unnormalized survey selectivity- 2 parameter ascending logistic - separate for males and females. |
| $s_{x, a}^{F N}=\exp \left(r_{x}^{F}\right) \frac{s_{x, a}^{F U}}{\max \left\{s_{1, a}^{F U}\right\}}$ | Normalized fishery selectivity. $r^{F}{ }_{1} \equiv 0$. |
| $s_{x, a}^{S N}=\exp \left(r_{x}^{S}\right) \frac{s_{x, a}^{S U}}{\max \left\{s_{1, a}^{S U}\right\}}$ | Normalized survey selectivity. $r^{\text {S }} \equiv 0$. |
| $N^{S}{ }_{t, x, a}=Q s_{x, a}^{s} N_{t, x, a}$ | Survey numbers for sex $x$, age $a$ at time $t$. |
| $S B_{t}=\sum_{x=1}^{2} \sum_{a=1}^{A} w_{x, a} N^{S}{ }_{t, x, a}$ | Total survey biomass. |
| $p_{t, x, a}^{F, A}=C_{t, x, a} / \sum_{x=1}^{2} \sum_{a=1}^{A} C_{t, x, a}$ | Proportion at age in the catch. |
| $p_{t, x, l}^{F, L}=\sum_{a=1}^{A} L_{l, a}^{X} \cdot p_{t, x, a}^{F, A}$ | Proportion at length in the catch. |
| $p_{t, x, a}^{S, A}=N^{S}{ }_{t, x, a} / \sum_{x=1}^{2} \sum_{a=1}^{A} N^{S}{ }_{t, x, a}$ | Proportion at age in the survey. |
| $p_{t, x, l}^{S, L}=\sum_{a=1}^{A} L_{l, a}^{x} \cdot p_{t, x, a}^{S, A}$ | Proportion at length in the survey. |

Table A.3. Likelihood components.

| Component | Description |
| :---: | :---: |
| $\sum_{t=1}^{T}\left[\log \left(C_{t}^{\text {obs }}\right)-\log \left(C_{t}\right)\right]^{2}$ | Catch; assumes a lognormal distribution. |
| $\sum_{t \in\left\{\mathbb{t}^{F, A}\right\}} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{F, A, o b s} \cdot \log \left(p_{t, x, a}^{F, A}\right)-\text { offset }$ | Fishery age composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{\mathbb{t}^{F, L}\right.} \sum_{x=1}^{2} \sum_{l=1}^{L} n_{t}^{\text {samp }} \cdot p_{t, x, l}^{F, L, o b s} \cdot \log \left(p_{t, x, l}^{F, L}\right)-\text { offset }$ | Fishery length composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{t^{E, A},\right.} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{S, A, o b s} \cdot \log \left(p_{t, x, a}^{S, A}\right)-\text { offset }$ | Survey age composition; assumes a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\sum_{t \in\left\{\mathbb{I}^{E}, L\right.} \sum_{x=1}^{2} \sum_{l=1}^{L} n_{t}^{\text {samp }} \cdot p_{t, x, l}^{s, L, o b s} \cdot \log \left(p_{t, x, l}^{s, L}\right)-\text { offset }$ | Survey length composition; uses a multinomial distribution. Observed sample size is $n_{t}^{\text {samp }}$. |
| $\text { offset } \left.=\sum_{t} \sum_{x=1}^{2} \sum_{a=1}^{A} n_{t}^{\text {samp }} \cdot p_{t, x, a}^{o b s} \cdot \log \left(p_{t, x, a}^{o b s}\right)\right)$ | The offset constants for age composition components are calculated from the observed proportions and the sample sizes. A similar formula is used for length composition component offsets. |
| $\sum_{t \in\left\{t^{s}\right\}}\left[\frac{\log \left[\frac{S B_{t}^{\text {obs }}}{S B_{t}}\right]}{\sqrt{2} \cdot \text { s.d. }\left(\log \left(S B_{t}^{\text {obs }}\right)\right)}\right]^{2}$ | Survey biomass; assumes a lognormal distribution. |
| $\sum_{t=T_{\text {min }}}^{T_{\text {max }}-3}\left(\tau_{t}\right)^{2}$ | Recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a log scale. |
| $\sum_{t=T_{\max }-2}^{T_{\text {max }}}\left(\tau_{t}\right)^{2}$ | "Late" recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a log scale. |
| $\sum_{t=T_{\min }-A+1}^{T_{\min }-1}\left(\tau_{t}\right)^{2}$ | "Early" recruitment; assumes a lognormal distribution, since $\tau_{t}$ is on a log scale. <br> Determines age composition at starting year of model. |

Table A.4. Parameters fixed in the model.

| Parameter | Description |
| :--- | :--- |
| $M_{x}=0.2$ | sex-specific natural mortality rate. |
| $Q=1.0$ | survey catchability. |
| $L^{x}{ }_{l, a}$ | sex-specific length-at-age conversion matrix. |
| $w_{x, a}$ | sex-specific weight-at-age. |
| $\phi_{a}$ | proportion of females mature at age $a$. |

Table A.5. Parameters estimated in the model. A total of 83 parameters were estimated.

| Parameter | Subscript <br> range | Total no. of <br> Parameters | Description |
| :--- | :---: | :---: | :--- |
| $\ln \left(R_{0}\right)$ | NA | 1 | natural log of the geometric mean <br> value of age 3 recruitment. |
| $\tau_{t}$ | $T_{\min }-A+1 \leq t \leq T_{\max }$ | 45 | log-scale recruitment deviation in <br> year $t$. |
| $\ln \left(f_{0}\right)$ | NA | 1 | natural log of the geometric mean <br> value of fishing mortality. |
| $\varepsilon_{t}$ | $T_{\min } \leq t \leq T_{\max }$ | 28 | log-scale deviations in fishing <br> mortality rate in year $t$. |
| $r^{F}{ }_{2}$ | NA | 1 | scaling from female to male fishery <br> selectivity (log-scale). |
| $b^{F},{ }_{x 0} \mathrm{~A}^{F}{ }_{x}$ | $1 \leq x \leq 2$ | 4 | sex-specific selectivity parameters <br> (slope and age at $50 \%$ selected) for <br> the fishery. |
| $r^{S}{ }_{2}$ | $S=1$ | 1 | scaling from female to male survey <br> selectivity (log-scale). |
| $b_{x}^{S},{ }_{50} \mathrm{~A}_{x}^{S}$ | $1 \leq x \leq 2$ | 4 | sex-specific selectivity parameters <br> (slope and age at $50 \%$ selected) for <br> the survey. |

## Chapter 8 Appendix B: Supplemental Catch Data

In order to comply with the Annual Catch Limit (ACL) requirements, two new datasets have been generated to help estimate total catch and removals from NMFS stocks in Alaska.

The first dataset, non-commercial removals, estimates total removals that do not occur during directed groundfish fishing activities (Table 8B.1). This includes removals incurred during research, subsistence, personal use, recreational, and exempted fishing permit activities, but does not include removals taken in fisheries other than those managed under the groundfish FMP. These estimates represent additional sources of removals to the existing Catch Accounting System estimates. For the GOA flathead sole stock, these estimates (currently available only for 2010) can be compared to research removals that have occurred in conjunction with the Gulf of Alaska Groundfish Surveys (Table 8B.2). Compared with the 2010 ABC ( $47,422 \mathrm{t}$ ), these non-commercial catches are miniscule ( $<0.2 \% \mathrm{ABC}$ ) and do not present a risk to the GOA flathead sole stock.

The second dataset, the Halibut Fishery Incidental Catch Estimation (HFICE), is an estimate of the incidental catch of groundfish in the halibut IFQ fishery in Alaska, which is currently unobserved. To estimate removals in the halibut fishery, methods were developed by the HFICE working group and approved by the Gulf of Alaska and Bering Sea/Aleutian Islands Plan Teams and the Scientific and Statistical Committee of the North Pacific Fishery Management Council. A detailed description of the methods is available in Tribuzio et al. (2011).

These estimates are for total catch of groundfish species in the halibut IFQ fishery and do not distinguish between "retained" or "discarded" catch. These estimates should be considered a separate time series from the current CAS estimates of total catch. Because of potential overlaps HFICE removals should not be added to the CAS produced catch estimates. The overlap will apply when groundfish are retained or discarded during an IFQ halibut trip. IFQ halibut landings that also include landed groundfish are recorded as retained in eLandings and a discard amount for all groundfish is estimated for such landings in CAS. Discard amounts for groundfish are not currently estimated for IFQ halibut landings that do not also include landed groundfish. For example, catch information for a trip that includes both landed IFQ halibut and sablefish would contain the total amount of sablefish landed (reported in eLandings) and an estimate of discard based on at-sea observer information. Further, because a groundfish species was landed during the trip, catch accounting would also estimate discard for all groundfish species based on available observer information and following methods described in Cahalan et al. (2010). The HFICE method estimates all groundfish caught during a halibut IFQ trip and thus is an estimate of groundfish caught whether landed or discarded. This prevents simply adding the CAS total with the HFICE estimate because it would be analogous to counting both retained and discarded groundfish species twice. Further, there are situations where the HFICE estimate includes groundfish caught in State waters and this would need to be considered with respect to ACLs (e.g. Chatham Strait sablefish fisheries), although the extent to which this occurs for flathead sole is unknown. Therefore, the HFICE estimates should be considered preliminary estimates for what is caught in the IFQ halibut fishery. Improved estimates of groundfish catch in the halibut fishery may become available following restructuring of the Observer Program in 2013.

The HFICE estimates of flathead sole catch by the halibut fishery in the Gulf of Alaska are miniscule compared with recent ABC's for the GOA stock (Table 8B.3). Based on these values, the risk to the stock from the halibut IFQ fishery is nil.

## References:

Cahalan J., J. Mondragon., and J. Gasper. 2010. Catch Sampling and Estimation in the Federal Groundfish Fisheries off Alaska. NOAA Technical Memorandum NMFS-AFSC-205. 42 p.
Tribuzio, CA, S Gaichas, J Gasper, H Gilroy, T Kong, O Ormseth, J Cahalan, J DiCosimo, M Furuness, H Shen, K Green. 2011. Methods for the estimation of non-target species catch in the unobserved halibut IFQ fleet. August Plan Team document. Presented to the Joint Plan Teams of the North Pacific Fishery Management Council.

## Tables

Table 8B.1. Non-commercial use catches of flathead sole in the Gulf of Alaska for 2010. Noncommercial use includes catches for research, recreation, subsistence, personal use and exempted fishing permits. The ABC for 2010 was $47,422 \mathrm{t}$.

| Source | Flathead Sole (t) |
| :--- | :---: |
| 2010 Shelikof Acoustic Survey | 0.0 |
| 2010 Shumigans Acoustic Survey | 0.2 |
| IPHC | 0.0 |
| large-mesh trawl | 87.6 |
| NMFS long line survey | 0.1 |
| Scallop dredge | 0.0 |
| small-mesh trawl | 12.0 |
| Structure of GOA Forage Fish Communities | 0.0 |
| Western GOA Pollock Acoustic Cooperative Survey | 0.0 |
| Grand Total | $\mathbf{1 0 0 . 0}$ |

Table 8B.2. Research catches of flathead sole from the Gulf of Alaska Groundfish Surveys. The ABC for 2011 was $49,133 \mathrm{t}$.

| year | Research Catch $(\mathrm{t})$ |
| ---: | ---: |
| 1984 | 31.33 |
| 1987 | 34.89 |
| 1990 | 22.40 |
| 1993 | 21.78 |
| 1996 | 14.03 |
| 1999 | 11.54 |
| 2001 | 9.90 |
| 2003 | 20.49 |
| 2005 | 15.58 |
| 2007 | 19.96 |
| 2009 | 16.47 |
| 2011 | 13.65 |

Table 8B.3. HFICE estimated catches of flathead sole in the Gulf of Alaska by the halibut fishery. The ABC for the GOA flathead sole fishery is also listed for each year. The ABC for 2011 was 49,133 t.

|  | Flathead sole (t) |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Western | Central | West Yakutat | Southeast | Total | ABC $(\mathrm{t})$ |
| 2001 | 0.00 | 0.00 | 0.00 | 0.07 | 0.07 | 26,270 |
| 2002 | 0.00 | 0.00 | 0.00 | 0.00 | 0.00 | 22,690 |
| 2003 | 0.00 | 0.00 | 0.00 | 0.53 | 0.53 | 41,390 |
| 2004 | 0.00 | 0.22 | 0.00 | 0.00 | 0.22 | 51,270 |
| 2005 | 0.34 | 0.16 | 0.36 | 1.15 | 2.02 | 45,100 |
| 2006 | 0.00 | 0.13 | 0.00 | 0.63 | 0.75 | 37,820 |
| 2007 | 0.02 | 0.25 | 0.00 | 2.97 | 3.24 | 39,110 |
| 2008 | 0.47 | 0.08 | 0.00 | 0.09 | 0.84 | 44,735 |
| 2009 | 0.02 | 0.01 | 0.14 | 0.07 | 0.51 | 46,464 |
| 2010 |  | 0.00 | 0.02 | 0.05 | 47,422 |  |

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