# 5. Assessment of Greenland Turbot in the Eastern Bering Sea and Aleutian Islands 

James N. Ianelli, Thomas K. Wilderbuer, and Dan Nichol<br>U.S. Department of Commerce<br>National Oceanic and Atmospheric Administration<br>National Marine Fisheries Service<br>Alaska Fisheries Science Center<br>7600 Sand Point Way NE, Seattle, WA 98115-6349

## Executive summary

Relative to last year's assessment, the following changes have been made in the current assessment.

## Changes to the input data

1. 2010 and 2011 catch data were updated (and added).
2. The Eastern Bering Sea (EBS) shelf survey 2011 biomass and length composition estimates were added.
3. Several additional years of NMFS bottom-trawl shelf-survey age data were completed and are included.
4. Fishery catch-at-length data were updated for longline and trawl gear from 2004-2011.

## Changes to the assessment model

The software used was Stock Synthesis (SS3) version 3.21d (windows 64-bit). Changes from previous years included refinements to estimating selectivities (additional parameters estimated), an alternative model in which male natural mortality is estimated (with female mortality fixed), and some adjustments to the length bin structure.
Changes in the assessment results
The 2011 EBS shelf trawl survey biomass increased by $11 \%$ from the 2010 estimate to $26,200 \mathrm{t}$ and is more than double the 2009 estimate. It is also the largest biomass estimate since 1998 and ranks $7^{\text {th }}$ highest from the survey since 1987 ( 25 years). In terms of population numbers, the estimate for 2011 was the highest ever recorded with 2010 estimate coming in second (since the standard survey began in 1982). The total 2011 population estimate (in numbers, all ages) from the EBS survey was over five times the average for Greenland turbot. The high numbers were almost entirely due to an increase in Greenland turbot less than about 30 cm . Results from the longline survey (which provides an index of the adult population) indicate a low but steady population trend.
Model results based on these surveys and data from longline and trawl fisheries result in an estimate of $B_{40 \%}$ equal to $21,560 t$ (female spawning biomass). The current estimate of the 2011 female spawning biomass is $51,300 \mathrm{t}$. There appears to be a favorable recruitment pattern in recent years and the 2010 survey saw the largest abundance of $1-y r$ old Greenland turbot on record. Recommended ABCs are set to the maximum permissible there are indications of strong recruitment. The corresponding maximum permissible ABC levels for Greenland turbot under Tier 3a are 9,660 and 8,029 t, for 2012 and 2013 espectively. The 2012 and 2013 overfishing levels, based on the $F_{35 \%}$ rate are $11,658 \mathrm{t}$ and $9,697 \mathrm{t}$ corresponding to a full-selection $F$ of 0.453 .

## Response to SSC comments

There were no comments specific to Greenland turbot assessment this year. Under general SAFE report comments they requested that if alternative models are available, they be arrayed so that ABC and OFL and status determinations can be selected. They also requested that authors incorporate their best estimate of total landings that will occur for the entire year.

Two model alternatives for ABC recommendations are provided in entirety. Catch for 2011 was based on the best available total-year estimate.

Summary

| Quantity | As estimated or specified last year for: 2011 <br> 2012 |  | As estimated or recommended this year for: 2012 |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |
| $M$ (natural mortality rate) | 0.112 | 0.112 |  |  |
| Tier | 3a | 3 a |  |  |
| Projected total (age 1+) biomass (t) | 74,000 | 69,700 | 76,850 | 73,910 |
| Female spawning biomass (t) |  |  |  |  |
| Projected | 45,504 | 40,407 | 47,687 | 41,441 |
| $B_{100 \%}$ | 71,048 | 71,048 | 53,900 | 53,900 |
| $B_{40 \%}$ | 28,419 | 28,419 | 21,560 | 21,560 |
| $B_{35 \%}$ | 24,867 | 24,867 | 18,870 | 18,870 |
| $F_{\text {OFL }}$ | 0.293 | 0.293 | 0.453 | 0.453 |
| $\operatorname{maxF}_{\text {ABC }}$ | 0.247 | 0.247 | 0.367 | 0.367 |
| $F_{\text {ABC }}$ | 0.247 | 0.247 | 0.367 | 0.367 |
| OFL (t) | 7,217 | 6,764 | 11,658 | 9,697 |
| maxABC ( t ) | 6,137 | 5,750 | 9,660 | 8,029 |
| ABC (t; BSAI) | 6,137 | 5,750 | 9,660 | 8,029 |
| EBSAleutian Islands | 4,590 | 4,301 | 7,226 | 6,006 |
|  | 1,547 | 1,449 | 2,434 | 2,023 |
| Status | As determined last year for: |  | As determined this year for: |  |
|  | 2009 | 2010 | 2010 | 2011 |
| Overfishing | No | n/a | No | n/a |
| Overfished | n/a | No | n/a | No |
| Approaching overfished | n/a | No | n/a | No |

## Introduction

Greenland turbot (Reinhardtius hippoglossoides) within the US 200-mile exclusive economic zone are mainly distributed in the eastern Bering Sea (EBS) and Aleutian Islands region. Juveniles are believed to spend the first 3 or 4 years of their lives on the continental shelf and then move to the continental slope (Alton et al. 1988; Sohn 2009; Fig. 5.1). Juveniles are absent in the Aleutian Islands regions, suggesting that the population in the Aleutians originates from the EBS or elsewhere. In this assessment, Greenland turbot found in the two regions are assumed to represent a single management stock. NMFS initiated a tagging study in 1997 to supplement earlier international programs. Results from conventional and archival tag return data suggest that individuals can range distances of several thousands of kilometers and spend summer periods in deep water in some years and in other years spend time on the shallower EBS shelf region.

Prior to 1985 Greenland turbot and arrowtooth flounder were managed together. Since then, the Council has recognized the need for separate management quotas given large differences in the market value between these species. Furthermore, the abundance trends for these two species are clearly distinct (e.g., Wilderbuer and Sample 1992).

The American Fisheries Society uses "Greenland halibut" as the common name for Reinhardtius hippoglossoides instead of Greenland turbot. To avoid confusion with the Pacific halibut, Hippoglossus stenolepis, common name of Greenland turbot which is also the "official" market name in the US and Canada (AFS 1991) is retained. For further background on this assessment and the methods used refer to Ianelli and Wilderbuer (1995).

## Catch history and fishery data

Catches of Greenland turbot and arrowtooth flounder were not reported separately during the 1960s. During that period, combined catches of the two species ranged from 10,000 to $58,000 \mathrm{t}$ annually and averaged $33,700 \mathrm{t}$. Beginning in the 1970s the fishery for Greenland turbot intensified with catches of this species reaching a peak from 1972 to 1976 of between $63,000 \mathrm{t}$ and $78,000 \mathrm{t}$ annually (Fig. 5.2). Catches declined after implementation of the MFCMA in 1977, but were still relatively high in 1980-83 with an annual range of 48,000 to $57,000 \mathrm{t}$ (Table 5.1). Since 1983, however, trawl harvests declined steadily to a low of $7,100 \mathrm{t}$ in 1988 before increasing slightly to $8,822 \mathrm{t}$ in 1989 and $9,619 \mathrm{t}$ in 1990. This overall decline is due mainly to catch restrictions placed on the fishery because of apparent low levels of recruitment. From 1990-1995 Council set the ABC's (and TACs) to $7,000 \mathrm{t}$ as an added conservation measure citing concerns about recruitment. Since 1996 the ABC levels have varied but averaged 7,660 t (with catch for that period averaging $4,550 \mathrm{t}$ ). The size frequency of the catch from the fisheries over time shows the how there are usually two modes present representing males at the smaller mode and females at the larger (Fig. 5.3).
Recently the trawl-caught Greenland turbot has exceeded the level of catch by longline vessels since about 2008 (Table 5.1). The shift in the proportion of catch by sector was due to changes arising from Amendment 80 passed in 2007. Amendment 80 to the BSAI Fishery Management Plan (FMP) was designed to improve retention and utilization of fishery resources. The amendment extended the American Fisheries Act (AFA) Groundfish Retention Standards to all vessels and established a limited access privilege program for the non-AFA trawl catcher/processors. This authorized the allocation of groundfish species quotas to fishing cooperatives and effectively provided better means to reduce bycatch and increase the value of targeted species.
The longline fleet generally targets pre-spawning aggregations of Greenland turbot; the fishery opens May 1 but usually occurs June-Aug in the EBS to avoid killer whale predation. Catch information prior to 1990 included only the tonnage of Greenland turbot retained by Bering Sea fishing vessels or processed onshore (as reported by PacFIN). Discard levels of Greenland turbot have typically been highest in the sablefish fisheries (at about one half of all sources of Greenland turbot discards during 1992-2002) while

Pacific cod fisheries and the "flatfish" fisheries also have contributed substantially to the discard levels (Table 5.2). About $12.6 \%$ of all Greenland turbot caught in groundfish fisheries were discarded (on average) during 2004-2010.

By gear-type and region, trawl catch was most significant in the Aleutian Islands in 2009 and 2010 whereas in the EBS there was high trawl catch in 2008 but then a switch to higher longline catches in 2009-2011 (Table 5.3). By target fishery, the gain in trawl-fishery has occurred primarily in the Greenland turbot and "arrowtooth flounder" (and for 2011, the Kamchatka flounder) fisheries in 2008 2011 (Table 5.4).

The catch data were used as presented above for both the longline and trawl fisheries. The early catches included Greenland turbot and arrowtooth flounder together. To separate them, the ratio of the two species for the years 1960-64 were assumed to be the same as the mean ratio caught by USSR vessels from 1965-69.

## Size and age composition

Fishery age composition data are becoming available for Greenland turbot. Age-determination methods have improved in the last few years and new age data are used this year. Extensive length frequency compositions have been collected by the NMFS observer program from the period 1980 to 1991. The length composition data from the trawl and longline fishery are presented in the appendix (along with the expected values from the assessment model) and absolute sample sizes for the period of the domestic fishery by sex and fishery from 1989 - 2010 are given in Table 5.5.

## Catch totals from research and other sources

Annual research catches (t, 1977-2011) from NMFS longline and trawl surveys are estimated as follows:

| Year | 1979 | 1980 | 1981 | 1982 | 1983 | 1984 | 1985 | 1986 | 1987 | 1988 | 1989 | 1990 | 1991 | 1992 | 1993 | 1994 |
| :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| NMFS Bottom |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| trawl survey <br> Longline surveys | 62.5 | 48.4 | 103 | 124 | 3 | 2 |  | 1 | 1 | 175 | 0.2 | 0.5 | 19 | 0.6 | 0.7 | 9.0 |
| Year | 1995 | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 |
| NMFS Bottom <br> trawl survey <br> Longline surveys | 1.4 | 1.5 | 1.2 | 1.4 | 1.0 | 5.1 | 1.1 | 5.3 | 1.1 | 11.0 | 0.7 | 0.76 | 0.59 | 2.0 |  |  |

An updated database for 2010 sport and research catches indicates the following for Greenland turbot:

| Source | t |
| ---: | ---: |
| 2010 Aleutian Island Bottom Trawl Survey | 0.530 |
| 2010 Bering Sea Acoustic Survey | 0.000 |
| 2010 Bering Sea Bottom Trawl Survey | 0.816 |
| 2010 Bering Sea Slope Survey | 5.21 |
| 2010 Northern Bering Sea Bottom Trawl Survey | 0.004 |
| Blue King Crab Pot | 0.056 |
| IPHC (halibut commission) | 2.989 |
| NMFS_LL survey | 0.364 |

Recent analyses examining the bycatch of Greenland turbot in directed halibut fisheries indicate an average of just over 109 t from 2001-2010 with about 49 t average since 2006 (NMFS Regional Office).

## Resource Surveys

## EBS slope and shelf bottom trawl survey

The older juveniles and adults on the slope have been surveyed every third year from 1979-1991 (also in 1981) as part of a U.S.-Japan cooperative agreement. From 1979-1985, the slope surveys were conducted by Japanese shore-based (Hokuten) trawlers chartered by the Japan Fisheries Agency. In 1988, the NOAA ship Miller Freeman was used to survey the resources on the EBS slope region. In this same year, chartered Japanese vessels performed side-by-side experiments with the Miller Freeman for calibration purposes. However, the Miller Freeman sampled a smaller area and fewer stations in 1988 than the previous years. The Miller Freeman sampled 133 stations over a depth interval of 200-800 m while during earlier slope surveys the Japanese vessels usually sampled 200-300 stations over a depth interval of 200-1000 m. In 2002, the AFSC re-established the bottom trawl survey of the upper continental slope of the eastern Bering Sea and a second survey was conducted in 2004. There was a gap in the planned biennial slope surveys when the 2006 survey was canceled but the surveys resumed in the summer of 2008 and in 2010 (Table 5.6).

The trawl slope-surveys are likely to represent under-estimates of the BSAI-wide biomass of Greenland turbot since fish are found consistently in other regions. Hence, the slope survey is treated as an index representing $75 \%$ of the stock based on earlier assessment analyses (Ianelli et al. 1993). The logic was based on average area-swept biomass estimates relative to the total compared to EBS on-shelf biomass and estimates of biomass from the Aleutian Islands (which are omitted from the model). A similar issue likely affects the distribution of Greenland turbot on the shelf region, particularly given the extent of the cold pool and warm conditions in recent years (Figs. 5.4 and 5.5). The shelf survey biomass estimates are therefore treated as a relative abundance index.

The estimated biomass of Greenland turbot in this region has fluctuated over the years. When USJapanese slope surveys were conducted in 1979, 1981, 1982 and 1985, the combined survey biomass estimates from the shelf and slope indicate a decline in EBS abundance. After 1985, the combined shelf plus slope biomass estimates (comparable since similar depths were sampled) have averaged $55,000 \mathrm{t}$, with a 2004 level of $57,500 \mathrm{t}$. The average shelf-survey biomass estimate during the last 18 years (19932011) is $24,680 \mathrm{t}$. The number of hauls and the levels of Greenland turbot sampling in the shelf surveys are presented in Table 5.7. In 2011 and 2010 the abundance estimates from the shelf surveys indicate a significant increase of Greenland turbot recruitment but also the proportion of tows with Greenland turbot present has increased (Fig. 5.6). These observations suggest that the extent of the spatial distribution has remained relatively constant prior to 2010 (with a slight increase) and that the most recent surveys have both higher densities and broader spatial distribution.

As reported in Ianelli et al (2010) the 2010 EBS slope trawl survey biomass estimate was up by $11 \%$ from the next most recent slope survey (in 2008). Most of the change in estimated biomass was due to Greenland turbot abundances observed in the $400-600 \mathrm{~m}$ depth strata (Table 5.8).

## Survey size and age composition

A time series of estimated size composition of the population was available for the shelf and slope trawl surveys and for the longline survey. The slope surveys typically sample more turbot than the shelf trawl surveys; consequently, the number of fish measured in the slope surveys is greater. The shelf survey appears to be useful for detecting some recruitment patterns that are consistent with the trends in biomass (Fig. 5.7). In the last 5 years signs of recruits (Greenland turbot less than about 40 cm ) is clear after a dearth of small fish during 2004-2006.
Survey size-at-age data was available and used for estimating growth and growth variability were previously available from 1975, 1979-1982. Gregg et al. (2006) revised age-determination methods for Greenland turbot and this year survey age composition data from 2003-2009 were included (previously only data from 1994, 1998, and 2007 surveys were available).

## Aleutian Islands survey

As reported in Ianelli et al (2010) the 2010 Aleutian Islands bottom trawl survey estimate was 6,800 t, well below the 1991-2010 average level of 15,800 $t$ (Table 5.6). The distribution of Greenland turbot in 2010 indicate lower abundances and variability of Greenland turbot in the survey compared to other recent surveys (Fig. 5.8). The breakdown of area specific survey biomass for the Aleutian Islands region shows that the eastern region has the highest densities and contains about $61 \%$ of the biomass, on based on average biomass from survey data from 1991-2010 (Fig. 5.9; Table 5.9). The trawl-survey area-swept data for the Aleutian Islands component of the Greenland turbot stock is not presently included in the stock assessment model.

## Longline survey

The sablefish longline survey alternates years between the Aleutian Islands and the Eastern Bering Sea slope region. This year the EBS region was covered but an unusually high number of orca depredation events occurred: 10 out of 16 stations were affected. Some investigations on how to account for these events highlight the need for more detailed analysis. For this year's assessment, the 2008 and 2010 data were assigned CVs of 0.3 whereas a CV of 0.2 was used for the earlier period.
The survey time series (through 2011) indicates that about $33 \%$ of the population along the combined slope regions survey is found within the northeast (NE) and southeast (SE) portions of the Aleutian Islands:

| Relative Population No. (RPN) |  |  | Year |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Area | 1996 | 1997 | 1998 | 1999 | 2000 | 2001 | 2002 | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| Bering 1 |  | 13,491 |  | 10,068 |  | 5,123 |  | 6,206 |  | 2,297 |  | 1,235 |  | 2,612 |  | 1,821 |
| Bering 2 |  | 27,936 |  | 33,848 |  | 24,766 |  | 24,660 |  | 15,268 |  | 13,523 |  | 21,192 |  | 12,164 |
| Bering 3 |  | 6,172 |  | 6,156 |  | 5,005 |  | 3,784 |  | 1,826 |  | 1,754 |  | 640 |  | 705 |
| Bering 4 |  | 11,729 |  | 13,072 |  | 16,082 |  | 11,965 |  | 3,717 |  | 1,561 |  | 3,406 |  | 1,494 |
| NE Aleutians | 23,133 |  | 23,121 |  | 12,987 |  | 10,942 |  | 8,551 |  | 3,031 |  | 3,155 |  | 2,033 |  |
| NW Aleutians | 7,212 |  | 7,208 |  | 4,049 |  | 3,411 |  | 2,666 |  | 945 |  | 984 |  | 634 |  |
| SE Aleutians | 2,142 |  | 1,791 |  | 1,201 |  | 1,397 |  | 936 |  | 566 |  | 297 |  | 163 |  |
| SW Aleutians | 6,775 |  | 5,665 |  | 3,800 |  | 4,420 |  | 2,962 |  | 1,789 |  | 939 |  | 517 |  |
| Bering Sea |  | 59,328 |  | 63,144 |  | 50,975 |  | 46,616 |  | 23,107 |  | 18,074 |  | 27,850 |  | 16,184 |
| Aleutians (total) | 39,262 |  | 37,784 |  | 22,037 |  | 20,170 |  | 15,115 |  | 6,331 |  | 5,374 |  | 3,347 |  |
| Combined index | 119.5 | 88.4 | 115.0 | 94.0 | 67.1 | 75.9 | 61.4 | 69.4 | 46.0 | 34.4 | 19.3 | 26.9 | 16.4 | 41.5 | 10.2 | 24.1 |

The combined time series shown above (1996-2011) was used as a relative abundance index. It was computed by taking the average RPN from 1996-2011 for both areas and computing the average proportion. The combined $R P N$ in each year ( $R P N_{t}^{c}$ ) was thus computed as:

$$
R P N_{t}^{c}=I_{t}^{A I} \frac{R P N_{t}^{A I}}{p^{A I}}+I_{t}^{E B S} \frac{R P N_{t}^{E B S}}{p^{E B S}}
$$

where $I_{t}^{A I}$ and $I_{t}^{\text {EBS }}$ are indicator function ( 0 or 1 ) depending on whether a survey occurred in either the Aleutian Islands or EBS, respectively. The average proportions (1996-2011) are given here by each area as: $p^{A I}(\sim 33 \%)$ and $p^{E B S}(\sim 67 \%)$. Note that each year data are added to this time series, the estimate of the combined index changes (slightly) in all years and that this approach assumes that the population proportion in these regions is constant. The time series of length frequency data from the longline survey extends back to the cooperative longline survey and is shown in Fig. 5.10.

## Analytic approach

## Model Structure

A version of the stock synthesis program (Methot 1990) has been used to model the eastern Bering Sea component of Greenland turbot since 1994. The software and assessment model configuration has changed over time, particularly in the past four years as newer versions have become available.

Total catch estimates used in the model were from 1960 to 2011. Model parameters were estimated by maximizing the log posterior distribution of the predicted observations given the data. Prior distributions consisting of very small penalties on recruitment deviations from a mean value were assumed. This was required to stabilize estimates of recruitment early in the time series when data were limiting. The model included two fisheries, those using fixed gear (longline and pots) and trawls, together with three surveys covering various years (Table 5.10). An archive of the software and model configuration for the final model can be found at http://www.afsc.noaa.gov/refm/docs/2011/BSAIGturbot.zip.

Some key assumptions from past models were retained—most importantly that the slope-trawl survey is treated as an absolute index representing 75\% of the Greenland turbot stock inhabiting US waters. In 2011 the NMFS survey extended into waters well north of the normal survey area but relatively minor amounts of Greenland turbot were found (Ianelli et al. 2010).

## Selectivity Patterns

Gender-specific size-based selectivity functions were estimated for each survey and fishery described below. For the trawl fishery, foreign operations occurred prior to the domestic fishery hence could be treated as a single fishery with an allowance for different size-based selectivity patterns estimated for the respective periods. Because the EBS shelf trawl surveys appear to cover only part of the range of this stock, selectivity was allowed to vary over time at roughly 5 -year periods. This increased the overall model uncertainty but reflected the uncertain nature of Greenland turbot occurrence on the EBS shelf region.

## Parameters estimated independently

## Natural mortality, length at age, length-weight relationship

The natural mortality of Greenland turbot was assumed to be 0.112 based on Cooper et al. (2007). This is also more consistent with re-analyses of age structures that suggest Greenland turbot live beyond 30 years (Gregg et al. 2006).

Parameters describing length-at-age are estimated within the model. Length at age 1 is assumed to be the same for both sexes and the variability in length at age 1 was assumed to have a $9 \%$ CV while at age 25 the CV was freely estimated. This appears to encompass the observed variability in length-at-age. As with last year, size-at-age information from the methods described by Gregg et al. (2006) were used and this information is summarized in Table 5.11.

The length-weight relationship for Greenland turbot estimated by Ianelli et al. (1993) was:

$$
\begin{aligned}
w= & 2.69 \times 10^{-6} L^{3.3092} \\
& \text { for females } \\
& \text { and } \\
w= & 6.52 \times 10^{-6} L^{3.068}
\end{aligned} \text { for males }
$$

where $L=$ length in mm , and $w=$ weight in grams.

## Maturation and fecundity

Recent studies on the fecundity of Greenland turbot indicate that estimates of fecundity at length are somewhat higher than most estimates from other studies and areas (Cooper et al., 2007). In particular, the
values were higher than that found from D'yakov's (1982) study. The data for proportion mature at size from the new study suggest a larger length at $50 \%$ maturity but data were too limited to provide revised estimates. For this analysis, a logistic maturity-at-size relationship was used with $50 \%$ of the female population mature at $60 \mathrm{~cm} ; 2 \%$ and $98 \%$ of the females are assumed to be mature at about 50 and 70 cm respectively. This is based on an approximation from D'yakov's (1982) study.

## Parameters estimated conditionally

The key parameters estimated within the model include:

- Annual age-0 recruitment estimates from 1960-2010 (parameterized as deviations scaled to the estimated mean recruitment)
- Selectivity for the 2 fisheries, and 3 surveys over different time frames, and
- Growth: 5 parameters (2 for each sex, one in common).
- Natural mortality of males (relative to females which were fixed at 0.112 ;-model 2 in this year's assessment)


## Model evaluation

Size composition data are unavailable until 1977 hence recruitment estimates information during the early period (1960s) are highly uncertain. The removal of 574,000 tons of Greenland turbot between 1972 and 1981 (compared to a total 64,300 t from 1997-2011) and the observed trends in abundance indicate that recruitment during the 1960s must have been high. Lacking information on the age (or sizes) of these fish impedes estimation of which (or how many) year classes were high. In previous assessments sensitivity to these estimates was performed. Evaluations of alternative model configurations were limited due to complexity related to selectivities, gear types, and general paucity of information specific to Greenland turbot.

Since gender-specific natural mortality rates are considered important for a number of flatfish species (e.g., Ralston 2006, Wilderbuer and Turnock 2009), we evaluated the ability to estimate male natural mortality relative to females (which remained fixed at 0.112 ).

## Results

Comparing the model with and without assuming male mortality was the same as females (at 0.112) resulted in better overall fits ( $\sim 10$ ln-likelihood units) when male mortality was freely estimated. Strictly speaking, the addition of a single parameter to improve model fit would be justified. However, the improvements come largely from the stock recruitment likelihood component (Table 5.12). While model 2 might justifiably be considered the best choice for proceeding with ABC recommendations, we feel that it would be premature-principally because the stock recruitment function in a model like this is essentially a prior on how recruitment may vary relative to spawning biomass and may be inappropriate to include as a model selection criterion. However, we provide the summary tables and projection results for Model 2 as an attachment should the Plan Teams or SSC wish to use those results. Results for model fits and other tables will all refer to Model 1 option (male and female natural mortality fixed at 0.112 ).
The model fit to the index data shows residual patterns that are autocorrelated but fall generally well within the confidence bands assumed for the data (Fig. 5.11). Such autocorrelation might be expected for species such as Greenland turbot in the BSAI region since they appear to be affected by periods of warm and cold bottom temperatures rather than real changes in abundance. The fit to the slope survey (considered to best cover the habitat for Greenland turbot and be less susceptible to temperature changes) fits well.

## Trends in Abundance

The biomass of Greenland turbot increased during the 1970s from the early 1960s level and is currently about $61 \%$ of the level expected under no fishing using average recruitment since 1977. The recent trend shows a gradual decline over the past 3 decades with the 2012 total begin-year biomass (age 1 and older) estimate to about 76,800 t (Table 5.13).

The historical fishing mortality rates (combined gears) began at high levels (but highly uncertain), decreased then peaked in recent decades in 1980 through 1983 (Table 5.13; Fig. 5.12). A comparison of this year's model result with the 2010 assessment shows Table 5.13. The current assessment results in similar abundances during the mid 1970s compared to assessments prior to 2009 (Fig. 5.13). The estimated historical numbers at age is given in Table 5.14.

## Selectivity

Estimates of selectivity (using recommended option \#24 in SS3-a "double normal" parameterization which describes the ascending and descending limbs of the curve with defined initial and final selectivity levels as parameters) provide patterns that appear reasonable over time for the shelf survey and other gear types (Figs. 5.14 and 5.15). Since the male selectivity estimates were different for these gear types, the proportions at sex between gear types over time was examined. This showed that the trawl fishery tends to catch slightly less than $50 \%$ females whereas the longline fishery catch has comprised about $70 \%$ females with a distinct trend towards 50:50 in recent years (Fig. 5.16). The slope trawl survey also shows that there is variability in the proportions female over time. In 2002, 2008, and 2010 the survey resulted in far more males than females-opposite of what was estimated for 2004 (more females than males; Fig. 5.17). This highlights the variability within the same region between years regarding observed sex ratios and also the difficulty in fitting the data-especially when sex ratio differences appear to occur at relatively small sizes (e.g., $<30 \mathrm{~cm}$ ).
Selectivity of Greenland turbot varied considerably between all of the surveys and fisheries. The shelf survey selected only small fish whereas the slope survey caught much larger fish. A similar pattern was observed between the trawl and longline fisheries with the longline fishery consistently catching larger Greenland turbot (e.g., Fig. 5.14). Note that the average selectivity estimates for the slope and shelf surveys indicate that the surveys sample intermediate size fish $(35-50 \mathrm{~cm})$ poorly. The reason for this is unclear; however, it could be related to the apparent bi-modality in the size distribution observed in the trawl fishery. The approximate age and sex-specific selectivity estimates (for 2010) from each gear type for Greenland turbot in the BSAI is given in Table 5.15. These are approximate because selectivity processes are modeled as a function of size. Similarly, approximate age-and-sex-specific weights are available and these are used in the projection model (Table 5.16).

Fit to age and size composition data
The model fit the available length-at-age data reasonably well and indicates that females grow larger than male Greenland turbot (Fig. 5.18). Size composition observations from the fisheries and surveys are matched by the model predictions reasonably well (see Attachment 5.A, Figs. 5.22-5.27). The discrepancies observed may be attributed to three issues. First, in some years, relatively few fish were measured so adjustments of the model to those data would depend on the trade-off in fitting other data, which may have had more extensive sampling. Second, unaccounted fish movement and hence changing availability affects fits to size composition data when an "average" (as opposed to annually varying) gear selectivity is used. Finally, natural mortality rate is undoubtedly variable among cohorts and years, the extent of which would affect our ability to model the age structure of the population accurately.

## Recruitment

Recruitment of young juvenile Greenland turbot appeared to have been poor for about 15 years since the early 1980s after several strong year-classes during the 1970s. Recently, and especially in 2010, there has been evidence of big increases in recruitment for Greenland turbot (Fig. 5.19).

## Maximum Sustainable Yield

Maximum sustainable yield (MSY) calculations require assumptions about the stock recruitment relationship including that the spawning stock range is covered. For Greenland turbot is is likely that the stock extends beyond the range of the area surveyed and fished. For this reason, a harvest strategy using spawning biomass per recruit as proxies for $F_{m s y}$ (e.g., $F_{35 \%}$ ) was selected in the absence of information on the stock-recruitment productivity relationship required for calculating MSY levels.

## Projections and harvest alternatives

## Amendment 56 Reference Points

The recommended harvest levels vary considerably among models depending on the assumptions made about the catchability coefficients from the slope-trawl survey (Ianelli et al. 1999). Since there are several areas of uncertainty surrounding this assessment, recommendations were based on a conservative model configuration.(assuming slope-survey catchability=0.75). The status of the projected spawning biomass in year 2012 relative to $B_{40 \%}$ would place Greenland turbot in Tier 3a of Amendment 56.

The $B_{40 \%}$ value using the mean recruitment estimated for the period 1978-2009 gives a long-term average female spawning biomass of $21,560 \mathrm{t}$. The estimated 2012 female spawning biomass is about $47,700 \mathrm{t}$, above the estimate of $B_{35 \%}(18,870 \mathrm{t})$.

## Specification of OFL and Maximum Permissible ABC and ABC Recommendation

In the past several years, the ABC has been set below the maximum permissible estimates. For example, in 2008 the ABC recommendation was $21 \%$ of the maximum permissible level. The rationale for these lower values have been generally due to concerns over stock structure uncertainty, lack of apparent recruitment, and modeling issues. This year a slope survey was conducted and while some areas show lower abundances (i.e., the Aleutian Islands) the signs of recruitment are the best ever seen for this stock. Hence, the arguments for keeping the ABC recommendation below the maximum permissible are less compelling. Therefore we recommend that the ABC be set to the maximum permissible.
The projected Greenland turbot maximum permissible ABC and OFL levels for 2012 and 2013 are shown below (catch for 2012 was set equal to the ABC recommendation):

|  | Catch | Maximum | Recommended |  | Female spawning |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year (for 2013projection) | permissible ABC | ABC | OFL | biomass |  |
| 2012 | $9,660 \mathrm{t}$ | $9,660 \mathrm{t}$ | $9,660 \mathrm{t}$ | $11,658 \mathrm{t}$ | $47,687 \mathrm{t}$ |
| 2013 |  | $8,029 \mathrm{t}$ | $8,029 \mathrm{t}$ | $9,697 \mathrm{t}$ | $41,441 \mathrm{t}$ |

The estimated overfishing level based on the adjusted $F_{35 \%}$ rate is $11,658 \mathrm{t}$ corresponding to a fullselection $F$ of 0.453 . The value of the Council's overfishing definition depends on the age-specific selectivity of the fishing gear, the somatic growth rate, natural mortality, and the size (or age) -specific maturation rate. As this rate depends on assumed selectivity, future yields are sensitive to relative gearspecific harvest levels. Because harvest of this resource is unallocated by gear type, the unpredictable nature of future harvests between gears is an added source of uncertainty. However, this uncertainty is considerably less than uncertainty related to treatment of survey biomass levels, i.e., factors which contribute to estimating absolute biomass (Ianelli et al. 1999). The history of stock size relative to the reference level (based on recruitments since 1977) shows that the fishing mortality has been well below the $F_{40 \%}$ level but is beginning to increase (Fig. 5.20).

## Subarea Allocation

In this assessment, the hypothesis proposed by Alton et al. (1989) regarding the stock structure of Greenland turbot in the eastern Bering Sea and Aleutian Islands regions was adopted. Briefly, spawning
is thought to occur throughout the adult range with post-larval settlement occurring on the shelf in shallow areas. The young fish on the shelf begin to migrate to the slope region at about age 4 or 5 . In our treatment, the spawning stock includes adults in the Aleutian Islands and the eastern Bering Sea. In support of this hypothesis, the length compositions from the Aleutian Islands surveys appear to have few small Greenland turbot, which suggests that these fish migrate from other areas (Ianelli et al. 1993). Historically, the catches between the Aleutian Islands and eastern Bering Sea has varied (Table 5.17).

Recent research on recruitment processes holds promise for clearer understanding (e.g., Sohn et al. (2010) and Sohn 2009). Stock structure between regions remains uncertain and therefore the policy has been to harvest the "stock" evenly by specifying region-specific ABCs. Based on eastern Bering Sea slope survey estimates and Aleutian Islands surveys, the proportions of the adult biomass in the Aleutian Islands region over the past three surveys (when both areas were covered) were $26.4 \%, 23.7 \%$, and $25.5 \%$. These average $25.2 \%$ which when applied to the BSAI ABC gives the following region-specific allocation:

|  | ABC |
| ---: | ---: |
| Aleutian Islands ABC | 2,434 |
| Eastern Bering Sea ABC | 7,226 |
| Total | 9,660 |

## Standard harvest scenarios and projections

A standard set of projections for population status under alternatives were conducted to comply with Amendment 56 of the FMP. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Protection Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).
For each scenario, the projections begin with the vector of 2011 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2012 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2011 (here assumed to be $2,858 \mathrm{t}$ ). In each subsequent year, the fishing mortality rate is prescribed based on the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1,000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.
Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2012, are as follow (" $\max F_{\text {ABC }}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):
Scenario 1: In all future years, $F$ is set equal to $\max F_{\text {ABC }}$. (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)
Scenario 2: In all future years, $F$ is set equal to the author's recommend level. Due to current conditions of strong recruitment and a projected increasing biomass, the recommendation is set equal to the maximum permissible ABC.
Scenario 3: In all future years, $F$ is set equal to the 2007-2011 average $F$. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 4: In all future years, $F$ is set equal to the $F_{75 \%}$. (Rationale: This scenario was developed by the NMFS Regional Office based on public feedback on alternatives.
Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follows (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):
Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above half of its MSY level in 2011 and above its MSY level in 2024 under this scenario, then the stock is not overfished.)

Scenario 7: In 2012 and 2013, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{O F L}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2024 under this scenario, then the stock is not approaching an overfished condition.)

Scenarios 1 through 7 were projected 13 years from 2011 (Table 5.18). Fishing at the maximum permissible rate indicate that the spawning stock will gradually drop to near the $B_{40 \%}$ by 2021 (Fig. 5.21).

Our projection model run under these conditions indicates that for Scenario 6, the Greenland turbot stock is not overfished based on the first criterion (year 2011 spawning biomass estimated at $51,278 \mathrm{t}$ relative to $0.5 B_{35 \%}=12,433 \mathrm{t}$ ). Under the guidelines, since the 2011 biomass estimate is above the $B_{35 \%}$ level (and $B_{40 \%}$ ) and the stock is not overfished.
Projections with fishing at the maximum permissible level result in an expected value of spawning biomass of $40,291 \mathrm{t}$ by 2024. These projections illustrate the impact of the recent recruitment observed in the survey. For example, under all scenarios except the no-fishing alternative, the spawning biomass is expected to decline until 2015 when the recruits in recent years mature.
Under Scenarios 6 and 7, the projected spawning biomass for Greenland turbot is not currently overfished, nor is it approaching an overfished status.

## Other Considerations

## Ecosystem considerations

Greenland turbot have undergone dramatic declines in the abundance of immature fish on the EBS shelf region compared to observations during the late 1970's. It may be that the high level of abundance during this period was unusual and the current level is typical for Greenland turbot life history pattern. Without further information on where different life-stages are currently residing, the plausibility of this scenario is speculation. Several major predators on the shelf were at relatively low stock sizes during the late 1970's (e.g., Pacific cod, Pacific halibut) and these increased to peak levels during the mid 1980's. Perhaps this shift in abundance has reduced the survival of juvenile Greenland turbot in the EBS shelf. Alternatively, the shift in recruitment patterns for Greenland turbot may be due to the documented environmental regime that occurred during the late 1970's. That is, perhaps the critical life history stages are subject to different oceanographic conditions that affect the abundance of juvenile Greenland turbot on the EBS shelf.

## Research and data gaps

A number of research and modeling issues continue to require further consideration. These include:

- An evaluation of possible differential natural mortality between males and females,
- Development of statistically based "effective sample size" values for size composition data (e.g., through boot-strapping original survey and observer data),
- Including more length-at-age information using the new methods, investigating age-specific natural mortality, and
- Evaluating the extent that Greenland turbot are affected by temperature and environmental conditions relative to survey gear.


## Summary

The pattern of total fishing mortality relative to spawning biomass suggests that the EBS Greenland turbot stock is approaching the $B_{40 \%}$ level, but that historically the fishing mortality was below the $F_{40 \%}$ level (Fig. 5.20). Management parameters of interest derived from this assessment are presented in Table 5.19.

## Acknowledgments

Chris Lunsford and Cara Rodgveller are thanked for providing the longline survey data summaries.

## References

AFS Publication, 1991. Common and Scientific Names of Fishes from the United States and Canada. American Fisheries Society Special Publication 20. C. Richard Robins, Chairman. 183 p. American Fisheries Society, 5410 Grosvenor Lane, Suite 110, Bethesda, MD 20814-2199.

Alton, M.S., R.G. Bakkala, G.E. Walters, and P.T. Munro. 1988. Greenland turbot Reinhardtius hippoglossoides of the eastern Bering Sea and Aleutian Islands region. NOAA Tech. Rep., NMFS 71, 31 p.
Beverton, R.J.H. and S.J. Holt. 1957. On the dynamics of exploited fish populations. Fish. Invest., Lond., Ser. 2, 19.

Cooper, D.W., K.P. Maslenikov, and D.R. Gunderson. 2007. Natural mortality rate, annual fecundity, and maturity at length for Greenland halibut (Reinhardtius hippoglossoides) from the northeastern Pacific Ocean. Fishery Bulletin, 105(2): 296-304.

D’yakov, Yu. P. 1982. The fecundity of the Greenland turbot, Reinhardtius hippoglossoides, (Pleuronectidae), from the Bering Sea. J. Ichthyol. [Engl. Transl. Vopr. Ikhtiol] 22(5):59-64.

Gregg, J.L., D.M. Anderl, and D.K. Kimura. 2006. Improving the precision of otolith-based age estimates for Greenland halibut (Reinhardtius hippoglossoides) with preparation methods adapted for fragile sagittae. Fish. Bull. 104:643-648 (2006).

Harrison, R.C. 1993. Data Report: 1991 Bottom trawl survey of the Aleutian Islands Area. NOAA Tech. Memo. NMFS-AFSC-12. 144p.

Healey, B.P. And J.-C. Mahé. 2006. An Assessment of Greenland Halibut (Reinhardtius hippoglossoides) in NAFO Subarea 2 and Divisions 3KLMNO. NAFO SCR Doc., No. 06/51, Ser. No. N5281.

Ianelli, J.N., T.K. Wilderbuer, and T.M. Sample. 1993. Stock assessment of Greenland turbot. In Stock assessment and fishery evaluation document for groundfish resources in the Bering Sea/Aleutian Islands region as projected for 1994. Section 4. North Pacific Fishery Management Council, Anchorage, AK.

Ianelli, J.N., T.K. Wilderbuer, and T.M. Sample. 1994. Stock assessment of Greenland turbot. In Stock assessment and fishery evaluation document for groundfish resources in the Bering Sea/Aleutian Islands region as projected for 1995. Section 4. North Pacific Fishery Management Council, Anchorage, AK.

Ianelli, J.N. and T. K. Wilderbuer. 1995. Greenland Turbot (Reinhardtius hippoglossoides) stock assessment and management in the Eastern Bering Sea. In: Proceedings of the International Symposium on North Pacific Flatfish. Alaska Sea Grant. AK-SG-95-04:407-441.
Ianelli, J.N., T.K. Wilderbuer, and T.M. Sample. 1999. Stock assessment of Greenland turbot. In Stock assessment and fishery evaluation document for groundfish resources in the Bering Sea/Aleutian Islands region as projected for 2000. Section 4. North Pacific Fishery Management Council, Anchorage, AK.

Ianelli, J.N., T.K. Wilderbuer, and D. Nichol. 2005. Stock assessment of Greenland turbot. In Stock assessment and fishery evaluation document for groundfish resources in the Bering Sea/Aleutian Islands region as projected for 2000. Section 4. North Pacific Fishery Management Council, Anchorage, AK.

Kell, L. T., I. Mosqueira, P. Grosjean, J-M. Fromentin, D. Garcia, R. Hillary, E. Jardim, S. Mardle, M.A. Pastoors, J.J. Poos, F. Scott, and R.D. Scott. 2007. FLR: an open-source framework for the evaluation and development of management strategies. - ICES Journal of Marine Science, 64: 640-646.

Kimura, D.K. 1988. Analyzing relative abundance indices with log-linear models. N. Am. Journ. Fish. Manage. 8:175-180.

Methot, R.D. 1990. Synthesis model: an adaptable framework for analysis of diverse stock assessment data. In Proceedings of the symposium on applications of stock assessment techniques to Gadids. L. Low [ed.]. Int. North Pac. Fish. Comm. Bull. 50: 259-277.

Miller, D.C.M. P.A. Shelton, B.P. Healey, W.B. Brodie, M.J. Morgan, D.S. Butterworth, R. Alpoim, D. González, F. González, C. Fernandez, J. Ianelli, J.C. Mahé, I. Mosqueira, R. Scott and A. Vazquez. 2008. Management strategy evaluation for Greenland halibut (Reinhardtius hippoglossoides) in NAFO Subarea 2 and Divisions 3LKMNO. NAFO SCR Doc. Serial. No. N5225. June 2008. 50p.

Ralston, S. 2006. An assessment of starry flounder off California, Oregon, and Washington. Pages 1-84 in Status of the Pacific Coast groundfish fishery through 2005, stock assessment and fishery evaluation: stock assessments and rebuilding analyses, Volume 2. Pacific Fishery Management Council, Portland, Oregon.

Sohn, D., L. Ciannelli, J. D. Anderson, A. Matarese, and K. M. Bailey. (2008) Early life history of Greenland halibut (Reinhardtius hippoglossoides) in the eastern Bering Sea. STAGES, Newsletter of the Early Life History Section of the American Fisheries Society. 29 (1): 1-12.
Sohn D. 2009. Ecology of Greenland Halibut (Reinhardtius hippoglossoides) during the Early Life Stages in the Eastern Bering Sea and Aleutian Islands. Master's thesis. College of Oceanic and Atmospheric Sciences, Oregon State University, Oregon, USA (June, 2009).

Sohn D., Ciannelli L., and Duffy-Anderson, J.T. 2010. Distribution and drift pathways of Greenland halibut (Reinhardtius hippoglossoides) during early life stages in the eastern Bering Sea and Aleutian Islands. Fisheries Oceanography. 19:5, 339-353.

Thompson, G.G., and M.W. Dorn. 2004. Chapter 2: Assessment of the Pacific Cod Stock in the Eastern Bering Sea and Aleutian Islands Area. North Pacific Fishery Management Council, Anchorage, AK. p. 185-302. http://www.afsc.noaa.gov/refm/docs/2004/BSAIpcod.pdf
Wilderbuer, T.K. and T.M. Sample. 1992. Stock assessment of Greenland turbot. In Stock assessment and fishery evaluation document for groundfish resources in the Bering Sea/Aleutian Islands region as projected for 1993. Section 4. North Pacific Fishery Management Council, Anchorage, AK.

Wilderbuer, T. K., and B. J. Turnock. 2009. Sex-specific natural mortality of arrowtooth flounder in Alaska: Implications of a skewed sex ratio on exploitation and management. NAJFM 29:306322.

Zenger, H.H. and M.F. Sigler. 1992. Relative abundance of Gulf of Alaska sablefish and other groundfish based on NMFS longline surveys, 1988-90. U.S. Dept. of Comm. NOAA Tech. Memo. NMFS F/NWC-216.

## Tables

Table 5.1. Catch estimates of Greenland turbot by gear type (t; including discards) and ABC and TAC values since implementation of the MFCMA.

| Year | Trawl | Longline \& Pot | Total | ABC | TAC |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1977 | 29,722 | 439 | 30,161 | 40,000 |  |
| 1978 | 39,560 | 2,629 | 42,189 | 40,000 |  |
| 1979 | 38,401 | 3,008 | 41,409 | 90,000 |  |
| 1980 | 48,689 | 3,863 | 52,552 | 76,000 |  |
| 1981 | 53,298 | 4,023 | 57,321 | 59,800 |  |
| 1982 | 52,090 | 31.8 | 52,122 | 60,000 |  |
| 1983 | 47,529 | 28.8 | 47,558 | 65,000 |  |
| 1984 | 23,107 | 12.6 | 23,120 | 47,500 |  |
| 1985 | 14,690 | 40.6 | 14,731 | 44,200 |  |
| 1986 | 9,864 | 0.4 | 9,864 | 35,000 | 33,000 |
| 1987 | 9,551 | 34 | 9,585 | 20,000 | 20,000 |
| 1988 | 6,827 | 281 | 7,108 | 14,100 | 11,200 |
| 1989 | 8,293 | 529 | 8,822 | 20,300 | 6,800 |
| 1990 | 12,119 | 577 | 12,696 | 7,000 | 7,000 |
| 1991 | 6,245 | 1,617 | 7,863 | 7,000 | 7,000 |
| 1992 | 749 | 3,003 | 3,752 | 7,000 | 7,000 |
| 1993 | 1,145 | 7,323 | 8,467 | 7,000 | 7,000 |
| 1994 | 6,426 | 3,845 | 10,272 | 7,000 | 7,000 |
| 1995 | 3,978 | 4,215 | 8,194 | 7,000 | 7,000 |
| 1996 | 1,653 | 4,902 | 6,555 | 7,000 | 7,000 |
| 1997 | 1,209 | 5,989 | 7,199 | 9,000 | 9,000 |
| 1998 | 1,830 | 7,319 | 9,149 | 15,000 | 15,000 |
| 1999 | 1,799 | 4,057 | 5,857 | 9,000 | 9,000 |
| 2000 | 1,946 | 5,027 | 6,973 | 9,300 | 9,300 |
| 2001 | 2,149 | 3,163 | 5,312 | 8,400 | 8,400 |
| 2002 | 1,033 | 2,605 | 3,638 | 8,000 | 8,000 |
| 2003 | 900 | 2,605 | 3,513 | 4,000 | 4,000 |
| 2004 | 675 | 1,544 | 2,220 | 3,500 | 3,500 |
| 2005 | 729 | 1,831 | 2,559 | 3,500 | 3,500 |
| 2006 | 360 | 1,605 | 1,965 | 2,740 | 2,740 |
| 2007 | 1,400 | 1,829 | 2,440 | 2,440 |  |
| 2008 | 809 | 1,935 | 1,741 | 2,540 | 2,540 |
| 2009 | 3,080 | 1,417 | 4,497 | 7,380 | 7,380 |
| 2010 | 2,070 | 1,975 | 4,046 | 6,120 | 6,120 |
| 2011 | 1,340 | 1,518 | $2,858 *$ | 6,140 | 5,050 |
|  |  |  |  |  |  |

[^0]Table 5.2. Estimates of BSAI discarded and retained ( t ) Greenland turbot based on NMFS estimates by "target" fishery, 1992-2011 (the "arrowtooth" fishery was combined with the Greenland turbot fishery from 2003-2011).

| Fishery: | Greenland turbot |  | Sablefish |  | Pacific cod |  | Rockfish |  | Flatfish |  | Others |  | Combined |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Retain | Discard | Retain | Discard | Retain | Discard | Retain | Discard | Retain | Discard | Retain | Discard | Retain | Discard |
| 1992 | 62 | 13 | 196 | 2,121 | 135 | 557 | 180 | 103 | 13 | 3 | 107 | 261 | 693 | 3,058 |
| 1993 | 5,685 | 332 | 235 | 880 | 160 | 108 | 572 | 87 | 19 | 185 | 10 | 194 | 6,681 | 1,786 |
| 1994 | 6,316 | 368 | 194 | 2,305 | 149 | 211 | 316 | 37 | 27 | 235 | 38 | 76 | 7,040 | 3,232 |
| 1995 | 5,093 | 327 | 157 | 1,546 | 145 | 284 | 362 | 25 | 5 | 102 | 28 | 121 | 5,790 | 2,405 |
| 1996 | 3,451 | 173 | 200 | 1,026 | 170 | 307 | 598 | 113 | 171 | 63 | 143 | 140 | 4,733 | 1,822 |
| 1997 | 4,709 | 521 | 129 | 619 | 270 | 283 | 202 | 19 | 212 | 92 | 18 | 125 | 5,540 | 1,659 |
| 1998 | 6,905 | 301 | 125 | 171 | 278 | 154 | 42 | 2 | 628 | 249 | 123 | 171 | 8,101 | 1,048 |
| 1999 | 4,009 | 227 | 179 | 120 | 180 | 50 | 25 | 2 | 600 | 269 | 134 | 61 | 5,127 | 729 |
| 2000 | 4,798 | 177 | 192 | 253 | 130 | 108 | 39 | 1 | 838 | 176 | 186 | 75 | 6,183 | 790 |
| 2001 | 2,727 | 89 | 171 | 325 | 203 | 92 | 431 | 30 | 764 | 337 | 95 | 47 | 4,391 | 920 |
| 2002 | 1,979 | 73 | 144 | 207 | 210 | 139 | 175 | 18 | 301 | 217 | 124 | 49 | 2,933 | 703 |
| 2003 | 1,842 | 95 | 98 | 534 | 165 | 95 | 198 | 5 | 114 | 176 | 79 | 55 | 2,497 | 961 |
| 2004 | 1,244 | 37 | 78 | 24 | 221 | 79 | 72 | 3 | 154 | 158 | 99 | 50 | 1,868 | 352 |
| 2005 | 1,676 | 28 | 63 | 19 | 156 | 30 | 136 | 5 | 179 | 69 | 149 | 49 | 2,359 | 200 |
| 2006 | 1,340 | 33 | 62 | 52 | 65 | 32 | 71 | 8 | 105 | 23 | 135 | 46 | 1,778 | 194 |
| 2007 | 1,225 | 28 | 59 | 71 | 128 | 91 | 36 | 13 | 59 | 39 | 198 | 50 | 1,704 | 292 |
| 2008 | 1,706 | 417 | 42 | 82 | 16 | 69 | 142 | 1 | 96 | 37 | 206 | 104 | 2,207 | 710 |
| 2009 | 3,649 | 336 | 69 | 54 | 65 | 21 | 69 | 8 | 52 | 13 | 148 | 14 | 4,053 | 445 |
| 2010 | 3,580 | 98 | 62 | 27 | 109 | 19 | 57 | 2 | 23 | 73 | 78 | 8 | 3,910 | 227 |
| 2011 | 2,330 | 14 | 41 | 7 | 142 | 7 | 20 | 1 | 577 | 39 | 64 | 10 | 3,174 | 78 |

Table 5.3. Estimates of Greenland turbot catch by gear and area based on NMFS Regional Office estimates, 2003-2011.

| Area | Gear | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| ---: | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Aleutian | Fixed | 650 | 218 | 167 | 355 | 343 | 115 | 97 | 213 | 82 |
| Islands | Trawl | 315 | 196 | 301 | 179 | 178 | 712 | 2,164 | 1,653 | 412 |
|  | Aleutian Islands Total | 965 | 414 | 439 | 468 | 534 | 521 | 827 | 2,261 | 1,866 |
|  | Fixed | 1,918 | 1,326 | 1,713 | 1,270 | 1,200 | 873 | 1,333 | 1,947 | 1,589 |
| EBS | Trawl | 575 | 479 | 427 | 183 | 280 | 1,223 | 916 | 322 | 1,170 |
|  | EBS Total | 2,493 | 1,805 | 2,120 | 2,140 | 1,452 | 1,481 | 2,095 | 2,249 | 2,272 |
|  |  | 3,458 | 2,220 | 2,608 | 1,986 | 2,002 | 2,922 | 4,510 | 4,138 | 3,252 |
| BSAI Total |  |  |  |  |  |  |  |  |  |  |

Table 5.4. Estimates of Greenland turbot catch (t) by gear and "target" fishery, 2004-2011. Source: NMFS AK Regional Office catch accounting system. Note, 2011 data are preliminary.

|  | "Target" fishery | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Longline and pot | Greenland turbot | 1,168 | 1,527 | 1,212 | 1,097 | 573 | 1,192 | 1,818 | 1,371 |
|  | Sablefish | 90 | 75 | 114 | 130 | 119 | 122 | 77 | 41 |
|  | Pacific cod | 221 | 170 | 77 | 129 | 76 | 84 | 121 | 152 |
|  | Shallow-water flatfish | 64 | 57 | 61 | 15 | 15 | 7 | 77 | 26 |
|  | Arrowtooth flounder | 0 | 2 | 140 | 16 | 0 | 9 | 53 | 0 |
|  | Others | 1 | 0 | 3 | 12 | 22 | 4 | 0 |  |
| Trawl | Greenland turbot | 61 | 24 | 0 | 2 | 205 | 1,349 | 118 | 4 |
|  | Pacific cod | 79 | 15 | 19 | 89 | 11 | 2 | 7 | 0 |
|  | Arrowtooth flounder | 53 | 154 | 21 | 3 | 1,176 | 1,435 | 1,689 | 892 |
|  | Kamchatka flounder |  |  |  |  |  |  |  | 582 |
|  | Atka mackerel | 123 | 167 | 117 | 130 | 201 | 118 | 62 | 45 |
|  | Flathead sole | 191 | 150 | 28 | 30 | 98 | 49 | 12 | 2 |
|  | Pollock | 18 | 31 | 65 | 107 | 82 | 44 | 1 | 4 |
|  | Rockfish | 74 | 139 | 74 | 47 | 143 | 73 | 59 | 22 |
|  | Other Flatfish | 51 | 34 | 1 | 12 | 11 | 4 | 1 | 0 |
|  | Rock sole | 4 | 1 | 27 | 8 | 0 | 2 | 3 | 1 |
|  | yellowfin sole | 1 | 7 | 8 | 1 | 1 | 4 | 1 | 4 |
|  | Sablefish | 12 | 7 | 0 | 0 | 6 | 0 | 12 | 6 |
|  | Others | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 5.5. Greenland turbot BSAI fishery length sample sizes by gear type and sex, 1989-2011 (Source: NMFS observer program data) and assumptions on multinomial sample sizes by gear type (columns headed SS N).

| Trawl fishery |  |  |  |  | Longline fishery |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Female | Male | \% Female | SS N | Female | Male | \% Female | SS N |
| 1989 | 1,335 | 5,002 | 21\% | 141 | 0 | 0 |  | Na |
| 1990 | 3,864 | 5,762 | 40\% | 386 | 0 | 0 |  | Na |
| 1991 | 1,851 | 1,752 | 51\% | 192 | 0 | 0 |  | Na |
| 1992 |  |  |  |  | 0 | 0 |  | Na |
| 1993 |  |  |  | 0 | 3,921 | 915 | 81\% | 392 |
| 1994 | 1,122 | 1,027 | 52\% | 102 | 503 | 150 | 77\% | 50 |
| 1995 | 217 | 355 | 38\% | 26 | 1,870 | 715 | 72\% | 187 |
| 1996 | 112 | 390 | 22\% | 11 | 941 | 442 | 68\% | 94 |
| 1997 | 0 | 0 |  | 0 | 2,393 | 1,014 | 70\% | 239 |
| 1998 | 307 | 696 | 31\% | 31 | 3,510 | 2,127 | 62\% | 351 |
| 1999 | 1,044 | 1,556 | 40\% | 104 | 7,961 | 2,835 | 74\% | 500 |
| 2000 | 724 | 1,328 | 35\% | 72 | 6,550 | 2,962 | 69\% | 500 |
| 2001 | 467 | 892 | 34\% | 47 | 4,054 | 1,550 | 72\% | 405 |
| 2002 | 186 | 433 | 30\% | 19 | 4,725 | 1,811 | 72\% | 473 |
| 2003 | 197 | 325 | 38\% | 20 | 4,608 | 2,113 | 69\% | 462 |
| 2004 | 179 | 433 | 29\% | 39 | 4,286 | 2,564 | 63\% | 435 |
| 2005 | 118 | 211 | 36\% | 21 | 4,639 | 1,902 | 71\% | 415 |
| 2006 | 15 | 76 | 16\% | 6 | 3,338 | 1,473 | 69\% | 305 |
| 2007 | 34 | 23 | 60\% | 4 | 3,816 | 2,127 | 64\% | 377 |
| 2008 | 421 | 1,572 | 21\% | 127 | 1,577 | 1,481 | 52\% | 194 |
| 2009 | 1,017 | 2,993 | 25\% | 255 | 3,486 | 2,704 | 56\% | 393 |
| 2010 | 289 | 3,560 | 8\% | 244 | 3,290 | 2,859 | 54\% | 390 |
| 2011 | 262 | 536 | 33\% | 51 | 1,043 | 641 | 62\% | 107 |

Table 5.6. Survey estimates of Greenland turbot biomass (t) for the Eastern Bering Sea shelf and slope areas and for the Aleutian Islands region, 1975-2008. Note that the shelf-survey estimates from 1985, and 1987-2008 include the northwestern strata (8 and 9) and these were the values used in the model. The Aleutian Islands surveys prior to 1990 used different operational protocols and may not compare well with subsequent surveys. The 1988 and 1991 slope estimates are from 200-800 m whereas the other slope estimates are from 200 1,000m.

| Eastern Bering Sea |  |  | Aleutian Islands |
| :---: | :---: | :---: | :---: |
| Year | Shelf | Slope | Survey |
| 1975 | 126,700 |  |  |
| 1979 | 225,600 | 123,000 |  |
| 1980 | 172,200 |  | 48,700* |
| 1981 | 86,800 | 99,600 |  |
| 1982 | 48,600 | 90,600 |  |
| 1983 | 35,100 |  | 63,800* |
| 1984 | 17,900 |  |  |
| 1985 | 7,700 | 79,200 |  |
| 1986 | 5,600 |  | 76,500* |
| 1987 | 10,600 |  |  |
| 1988 | 14,800 | 42,700 |  |
| 1989 | 8,900 |  |  |
| 1990 | 14,300 |  |  |
| 1991 | 13,000 | 40,500 | 11,925 |
| 1992 | 24,000 |  |  |
| 1993 | 30,400 |  |  |
| 1994 | 48,800 |  | 28,227 |
| 1995 | 34,800 |  |  |
| 1996 | 30,300 |  |  |
| 1997 | 29,218 |  | 28,334 |
| 1998 | 28,126 |  |  |
| 1999 | 19,797 |  |  |
| 2000 | 22,957 |  | 9,359 |
| 2001 | 25,347 |  |  |
| 2002 | 21,450 | 27,589 | 9,891 |
| 2003 | 23,685 |  |  |
| 2004 | 20,910 | 36,557 | 11,334 |
| 2005 | 21,359 |  |  |
| 2006 | 20,933 |  | 20,934 |
| 2007 | 16,726 |  |  |
| 2008 | 13,514 | 17,901 | NA |
| 2009 | 10,956 |  |  |
| 2010 | 23,415 | 19,873 | 6,795 |
| 2011 | 26,156 |  |  |

* U.S. - Japanese cooperative surveys

Table 5.7. Levels of Greenland turbot biological sampling from the EBS shelf surveys. Note that in 1982-1984, and 1986 the northwestern stations were not sampled.

| Year | Total <br> Hauls | Hauls w/ turbot | Length samples | Otolith sample hauls | Hauls w/age | Otolith Samples | Ages |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1982 | 334 | 41 | 1,228 | 11 | 11 | 292 | 292 |
| 1983 | 353 | 55 | 951 |  |  |  |  |
| 1984 | 355 | 27 | 536 | 20 |  | 263 |  |
| 1985 | 358 | 46 | 200 |  |  |  |  |
| 1986 | 354 | 53 | 195 |  |  |  |  |
| 1987 | 360 | 36 | 354 |  |  |  |  |
| 1988 | 373 | 58 | 414 |  |  |  |  |
| 1989 | 373 | 56 | 376 |  |  |  |  |
| 1990 | 371 | 62 | 544 |  |  |  |  |
| 1991 | 372 | 65 | 658 |  |  |  |  |
| 1992 | 356 | 64 | 616 | 5 |  | 7 |  |
| 1993 | 375 | 73 | 632 | 7 |  | 179 |  |
| 1994 | 376 | 52 | 530 | 17 |  | 196 |  |
| 1995 | 376 | 49 | 343 |  |  |  |  |
| 1996 | 375 | 75 | 450 | 8 |  | 100 |  |
| 1997 | 376 | 64 | 298 | 11 |  | 79 |  |
| 1998 | 375 | 73 | 445 | 25 | 21 | 200 | 127 |
| 1999 | 373 | 43 | 128 | 8 |  | 11 |  |
| 2000 | 372 | 57 | 248 | 34 |  | 188 |  |
| 2001 | 375 | 58 | 270 | 43 |  | 215 |  |
| 2002 | 375 | 70 | 455 | 21 |  | 71 |  |
| 2003 | 376 | 71 | 622 | 62 | 26 | 435 | 192 |
| 2004 | 375 | 64 | 606 | 45 | 45 | 290 | 280 |
| 2005 | 373 | 61 | 441 | 56 | 55 | 293 | 277 |
| 2006 | 376 | 56 | 427 | 49 | 48 | 262 | 239 |
| 2007 | 376 | 83 | 499 | 68 | 68 | 334 | 311 |
| 2008 | 375 | 78 | 406 | 59 | 59 | 245 | 235 |
| 2009 | 376 | 103 | 856 | 72 | 32 | 351 | 344 |
| 2010 | 376 | 144 | 3,199 | 70 |  | 362 |  |
| 2011 | 376 | 155 | 4,381 | 61 |  | 427 |  |

Table 5.8. Eastern Bering Sea slope survey estimates of Greenland turbot biomass ( t ), 2002, 2004, 2008, and 2010 by depth category.

| Depth (m) | $\mathbf{2 0 0 2}$ | $\mathbf{2 0 0 4}$ | $\mathbf{2 0 0 8}$ | $\mathbf{2 0 1 0}$ |
| ---: | ---: | ---: | ---: | ---: |
| $200-400$ | 4,081 | 2,889 | 4,553 | 1,166 |
| $\mathbf{4 0 0 - 6 0 0}$ | $\mathbf{1 4 , 1 7 4}$ | $\mathbf{2 5 , 3 6 0}$ | $\mathbf{6 , 7 0 7}$ | $\mathbf{1 0 , 3 5 2}$ |
| $600-800$ | 4,709 | 5,303 | 4,373 | 5,235 |
| $800-1000$ | 2,189 | 1,800 | 1,487 | 2,041 |
| $1000-1200$ | 1,959 | 1,206 | 781 | 1,079 |
| Total | 27,113 | 36,557 | 17,901 | 19,873 |

Table 5.9. Time series of Aleutian Islands survey sub-regions estimates of Greenland turbot biomass (t), 1980-2010.

| Year | Western Aleutian | Central Aleutian | Eastern Aleutian | Southern Bering Sea | Total |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1980 | 0 | 799 | 2,720 | 79 | 3,598 |
| 1983 | 525 | 2,357 | 5,747 | 1,094 | 9,722 |
| 1986 | 1,747 | 2,495 | 19,580 | 7,937 | 31,759 |
| 1991 | 2,195 | 3,280 | 4,607 | 1,803 | 11,885 |
| 1994 | 2,401 | 4,007 | 15,862 | 5,966 | 28,235 |
| 1997 | 2,137 | 3,130 | 22,708 | 359 | 28,334 |
| 2000 | 839 | 2,351 | 5,703 | 467 | 9,359 |
| 2002 | 793 | 1,658 | 6,996 | 444 | 9,891 |
| 2004 | 2,588 | 2,947 | 1,937 | 15,564 | 3,234 |
| 11,333 |  |  |  |  |  |
| 2006 | 1,973 | 1,544 | 3,698 | 1,282 | 20,934 |
| 2010 | 1,749 | 2,607 | 9,735 | 482 | 6,795 |
| Avg. since 1991 |  |  | 1,755 | 15,846 |  |

Table 5.10. Data sets used in the stock synthesis (SS3) model for Greenland Turbot in the EBS. All size and age data are specified by sex.

| Data Component | Years of data |
| :--- | :--- |
| Survey size at age data | $1994,1998,2003-2009$ |
| Shelf survey: size composition and biomass estimates | $1979-2011$ |
| Shelf survey age composition data | $1998,2003-2009$ |
| Slope survey: size composition and biomass estimates | $1979,1981,1982,1985,1988,1991$, |
|  | $2002,2004,2008,2010$ |
| Longline survey: size composition and abundance index | $1996-2011$ |
| Total fishery catch data | $1960-2011$ |
| Trawl fishery size composition | $1977-87,1989-91,1993-2011$ |
| Longline fishery size composition | $1977,1979-85,1992-2011$ |

Table 5.11. Summary of the shelf-survey length-at-age information used for this BSAI Greenland turbot assessment and sample size by age and by year.

| Mean length (cm) |  |  | Sample size |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Age | Female | Male | Female | Male |  |  |  |
| 1 | 14.28 | 13.26 | 19 | 19 |  |  |  |
| 2 | 22.70 | 22.90 | 33 | 34 |  |  |  |
| 3 | 30.65 | 29.85 | 29 | 35 |  |  |  |
| 4 | 34.22 | 35.15 | 31 | 36 |  |  |  |
| 5 | 38.95 | 40.08 | 30 | 33 |  |  |  |
| 6 | 44.31 | 43.91 | 23 | 29 |  |  |  |
| 7 | 49.46 | 46.49 | 28 | 26 |  | Sampl |  |
| 8 | 52.55 | 48.54 | 25 | 27 | Year | Female | Male |
| 9 | 56.45 | 54.12 | 19 | 28 | 1998 | 75 | 53 |
| 10 | 64.36 | 58.04 | 23 | 20 | 2003 | 107 | 60 |
| 11 | 69.85 | 59.82 | 22 | 12 | 2004 | 105 | 52 |
| 12 | 76.30 | 69.88 | 20 | 11 | 2005 | + 78 | 67 |
| 13 | 75.05 | 67.21 | 23 | 10 | 2006 | 68 | 51 |
| 14 | 81.84 | 75.50 | 19 | 4 | 2007 | 72 | 63 |
| 15 | 77.41 | 73.41 | 17 | 12 | 2008 | 69 | 46 |
| 16 | 82.15 | 73.30 | 18 | 10 |  |  |  |
| 17 | 83.62 | 74.51 | 24 | 12 |  | 69 | 53 |
| 18 | 88.93 | 75.50 | 21 | 8 |  |  |  |
| 19 | 88.27 | 76.41 | 23 | 10 |  |  |  |
| 20 | 86.41 | 79.00 | 28 | 9 |  |  |  |
| 21 | 88.36 | 73.66 | 26 | 4 |  |  |  |
| 22 | 90.22 | 77.00 | 12 | 6 |  |  |  |
| 23 | 87.77 |  | 20 |  |  |  |  |
| 24 | 87.95 | 75.85 | 15 | 13 |  |  |  |
| 25 | 86.99 | 75.15 | 21 | 9 |  |  |  |
| 26 | 93.49 | 77.51 | 16 | 5 |  |  |  |
| 27 | 87.11 | 73.50 | 11 | 4 |  |  |  |
| 28 | 91.21 | 78.00 | 12 | 4 |  |  |  |
| 29 | 90.99 | 79.60 | 10 | 5 |  |  |  |
| 30 | 89.08 | 78.50 | 25 | 10 |  |  |  |

Table 5.12. Negative log likelihood components of the Greenland turbot comparing models 1 (M fixed) and model 2 (male mortality freely estimated). Right-most two columns represent the difference from the "best fitting" for that component-shading indicates a worse fit)

| -ln likelihood component | M Fixed <br> Model 1 | M Est <br> Model 2 | M Fixed <br> Model 1 | M Est <br> Model 2 |
| ---: | ---: | ---: | ---: | ---: |
| Age composition | 210.1 | 208.6 | 1.5 | 0.0 |
| Length composition | $2,175.8$ | $2,179.3$ | 0.0 | 3.4 |
| Recruitment | 113.0 | 102.7 | 10.3 | 0.0 |
| Size at age | 833.5 | 826.9 | 6.6 | 0.0 |
| Survey indices | -26.4 | -21.4 | 0.0 | 5.0 |
| Total | $3,306.0$ | $3,296.0$ | 10.0 | 0.0 |

Table 5.13. Fishing mortality rate (maximum F over ages, combined sex and fisheries), harvest rate (measured as 1-SPR), spawning and total biomass (compared with the 2010 assessment) for BSAI Greenland turbot, 1960-2011.

| Year | Total Fishing Mortality | 1-SPR | Female Spawning Biomass |  | Total Age 1+ Biomass |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 2010 | Current | 2010 | Current |
|  |  |  | Assessment | Assessment | Assessment | Assessment |
| 1960 | 0.17 | 0.485 | 258,392 | 118,843 | 474,088 | 220,366 |
| 1961 | 0.31 | 0.725 | 249,139 | 113,381 | 448,382 | 194,345 |
| 1962 | 0.43 | 0.859 | 233,857 | 102,327 | 410,807 | 157,311 |
| 1963 | 0.31 | 0.817 | 217,227 | 85,826 | 382,489 | 129,006 |
| 1964 | 0.33 | 0.846 | 207,912 | 74,723 | 387,637 | 133,337 |
| 1965 | 0.07 | 0.404 | 199,234 | 66,321 | 405,268 | 150,022 |
| 1966 | 0.07 | 0.352 | 201,923 | 74,009 | 454,564 | 196,743 |
| 1967 | 0.09 | 0.415 | 214,194 | 92,187 | 513,301 | 253,541 |
| 1968 | 0.10 | 0.441 | 236,920 | 118,828 | 572,267 | 311,592 |
| 1969 | 0.09 | 0.384 | 267,433 | 150,934 | 628,292 | 367,441 |
| 1970 | 0.06 | 0.261 | 303,476 | 187,110 | 687,282 | 426,328 |
| 1971 | 0.10 | 0.413 | 344,265 | 227,388 | 754,980 | 495,239 |
| 1972 | 0.18 | 0.586 | 378,122 | 263,160 | 800,335 | 544,065 |
| 1973 | 0.14 | 0.526 | 398,008 | 288,527 | 802,344 | 551,959 |
| 1974 | 0.18 | 0.599 | 419,604 | 315,374 | 804,777 | 557,812 |
| 1975 | 0.16 | 0.577 | 433,239 | 333,284 | 779,701 | 535,080 |
| 1976 | 0.16 | 0.587 | 445,162 | 344,128 | 755,006 | 510,665 |
| 1977 | 0.09 | 0.392 | 448,813 | 342,858 | 726,749 | 484,585 |
| 1978 | 0.12 | 0.507 | 455,179 | 344,621 | 726,596 | 489,050 |
| 1979 | 0.13 | 0.516 | 448,328 | 336,226 | 710,828 | 483,658 |
| 1980 | 0.17 | 0.601 | 438,277 | 327,888 | 692,314 | 482,781 |
| 1981 | 0.19 | 0.639 | 421,551 | 316,764 | 658,565 | 472,334 |
| 1982 | 0.18 | 0.615 | 401,701 | 305,743 | 614,943 | 454,325 |
| 1983 | 0.16 | 0.596 | 383,432 | 300,647 | 569,952 | 434,625 |
| 1984 | 0.08 | 0.381 | 364,461 | 296,896 | 523,769 | 410,716 |
| 1985 | 0.05 | 0.268 | 354,272 | 300,123 | 496,953 | 401,782 |
| 1986 | 0.04 | 0.191 | 344,867 | 302,051 | 474,170 | 393,764 |
| 1987 | 0.04 | 0.189 | 334,294 | 300,666 | 452,610 | 384,540 |
| 1988 | 0.03 | 0.151 | 320,364 | 294,188 | 428,782 | 371,096 |
| 1989 | 0.04 | 0.192 | 304,839 | 284,481 | 405,651 | 356,651 |
| 1990 | 0.07 | 0.278 | 286,540 | 271,102 | 379,923 | 338,026 |
| 1991 | 0.04 | 0.203 | 265,271 | 254,463 | 349,900 | 314,078 |
| 1992 | 0.01 | 0.109 | 246,392 | 238,886 | 325,430 | 294,602 |
| 1993 | 0.03 | 0.229 | 229,402 | 224,266 | 305,148 | 279,179 |
| 1994 | 0.07 | 0.268 | 209,171 | 206,306 | 280,792 | 259,257 |
| 1995 | 0.05 | 0.246 | 190,557 | 189,773 | 255,222 | 237,313 |
| 1996 | 0.04 | 0.222 | 173,884 | 174,729 | 232,722 | 217,654 |
| 1997 | 0.05 | 0.256 | 158,867 | 160,564 | 212,765 | 199,870 |
| 1998 | 0.07 | 0.322 | 143,845 | 145,984 | 193,194 | 181,958 |
| 1999 | 0.05 | 0.246 | 128,066 | 130,564 | 172,775 | 162,914 |
| 2000 | 0.07 | 0.302 | 115,466 | 118,109 | 156,575 | 147,730 |
| 2001 | 0.07 | 0.273 | 102,728 | 105,568 | 140,403 | 132,537 |
| 2002 | 0.05 | 0.227 | 92,402 | 95,251 | 127,252 | 120,324 |
| 2003 | 0.05 | 0.236 | 83,723 | 86,466 | 117,110 | 111,554 |
| 2004 | 0.03 | 0.176 | 75,762 | 78,554 | 108,454 | 104,569 |
| 2005 | 0.04 | 0.211 | 69,472 | 72,690 | 102,061 | 99,974 |
| 2006 | 0.03 | 0.182 | 63,755 | 67,853 | 95,884 | 95,501 |
| 2007 | 0.03 | 0.172 | 59,422 | 64,537 | 90,773 | 91,825 |
| 2008 | 0.06 | 0.199 | 56,110 | 61,791 | 86,111 | 88,135 |
| 2009 | 0.10 | 0.301 | 53,151 | 59,031 | 81,298 | 83,822 |
| 2010 | 0.08 | 0.309 | 49,176 | 55,288 | 76,979 | 78,586 |
| 2011 | 0.08 | 0.254 |  | 51,278 |  | 75,026 |

Table 5.14. Estimated beginning of year numbers of Greenland turbot by age and sex (millions).
Females

| Yr | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | $19 \quad 20+$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 87.8 | 6.1 | 61.3 | 17.0 | 32.0 | 2.4 | 1.6 | 4.0 | 23.0 | 18.6 | 15.0 | 12.21 | 10.3 | 9.1 | 8.1 | 7.3 | 6.4 | 5.4 | 0.5 | 0.43 .2 |
| 1978 | 26.8 | 78.5 | 5.5 | 54.6 | . | 28.2 | 2.1 | 1.4 | 3.5 | 19.9 | 16.0 | 12.91 | 10.5 | 8.9 | 7.9 | 7.1 | 6.3 | 5.5 | 4.7 | 0.53 .2 |
| 1979 | 12.2 | 3.9 | 70.0 | 4.9 | 48.3 | , | 24.4 | 1.8 | 1.2 | 3.0 | 16.8 | . | 11.0 | 9.0 | 7.5 | 6.7 | 6.0 | 5.4 | 4.7 | 4.03 .2 |
| 1980 | 15.3 |  | 21.4 | 62.3 | 4.3 | 2.3 | 15 | 20.9 | 1.5 | 1.0 | 2.5 | , | 11.5 | 9.3 | 7.6 | 6.4 | 5.7 | 2 | 4.6 | 4.16 .1 |
| 1981 | 1.4 | 3.7 | 9.7 | 19. | 5.0 | 3.8 | 6.2 | 9.7 | 17.5 | 1.3 | 0.8 | 2.11 | 11.9 | 9.6 | 7.8 | 6.4 | 5.4 | 4.8 | 4.3 | 3.98 .6 |
| 1982 | 2.8 | 1.3 | 12.2 | 8.6 | 16.7 | 47.7 | 3.2 | 30.5 | 8.1 | 4.5 | 1.1 | 0.7 | 1.7 | 9.8 | 8.0 | 6.5 | 5.3 | 4.5 | 4.1 | 3.610 .6 |
| 1983 | 1.2 | 2.5 | 1.1 | 10.8 | 7.6 | . 5 | 40.7 | 2.7 | 25.4 | 6.7 | 12.0 | 0.9 | 0.6 | 1.4 | 8.3 | 6.7 | 5.5 | 4.5 | 3.8 | 3.412 .1 |
| 1984 | 3.2 | 1.1 | 2.3 | 1.0 | 9.5 | 6.6 | 12.4 | 34.4 | 2.3 | 21.3 | 5.61 | 10.1 | 0.7 | 0.5 | 1.2 | 7.0 | 5.7 | 4.6 | 3.8 | 3.313 .3 |
| 1985 | 11.1 | 2.9 | 1.0 | 2.0 | 0.9 | 8.4 | 5.8 | 0.8 | 29.8 | 2.0 | 18.4 | 4.9 | 8.7 | 0.6 | 0.4 | 1.1 | 6.1 | 5.0 | 4.0 | 3.314 .5 |
| 1986 | 2.5 | 0.0 | 2.6 | 0.9 | 1.8 | 0.8 | 7.4 | 5. | 9.5 | 26.2 | 1.7 | 16.1 | 4.3 | 7.7 | 0.6 | 0.4 | 0.9 | 5.4 | 4.4 | 3.615 .7 |
| 1987 | 3.6 | 2.2 | 8.9 | 2.3 | 0.8 | 1.6 | 0.7 | 6.6 | 4.5 | 8.4 | 3.1 | 1.5 | 14.3 | 3.8 | 6.8 | 0.5 | 0.3 | 0.8 | 4.7 | 3.917 .1 |
| 1988 | 2.8 | 3.2 | 2.0 | 8.0 | 2.0 | 0.7 | 1.4 | 0.6 | 5.8 | 4.0 | 7.4 | 20.4 | 1.3 | 12.6 | 3.3 | 6.0 | 0.4 | 0.3 | 0.7 | 4.218 .5 |
| 1989 | 12.7 | 2.5 | 2.9 | 1.8 | 7.1 | 1.8 | 0.6 | 1.3 | 0.5 | 5.1 | 3.5 | 6.5 | 18.0 | 1.2 | 11.2 | 2.9 | 5.3 | 0.4 | 0.3 | 0.620 .2 |
| 1990 | 2.5 | 11.3 | 2.2 | 2.6 | 1.6 | 6.3 | 1.6 | 0.5 | 1.1 | 0.5 | 4.5 | 3.1 | 5.8 | 15.9 | 1.0 | 9.9 | 2.6 | 4.7 | 0.3 | 0.218 .5 |
| 1991 | 0.9 | 2.2 | 10.1 | 2.0 | 2.3 | 1.4 | 5.6 | 1.4 | 0.5 | 1.0 | 0.4 | 4.0 | 2.7 | 5.1 | 14.0 | 0.9 | 8.7 | 2.3 | 4.1 | 0.316 .5 |
| 1992 | 0.8 | 0.8 | 2.0 | 9.1 | 1.8 | 2.1 | 1.3 | 4.9 | 1.2 | 0.4 | 0.9 | 0.4 | 3.5 | 2.4 | 4.5 | 12.3 | 0.8 | 7.7 | 2.0 | 3.714 .8 |
| 1993 | 0.5 | 0.7 | 0.7 | 1.8 | 8.1 | 1.6 | 1.8 | 1.1 | 4.4 | 1.1 | 0.4 | 0.8 | 0.3 | 3.1 | 2.1 | 4.0 | 10.9 | 0.7 | 6.8 | 1.816 .4 |
| 1994 | 0.7 | 0.5 | 0.6 | 0.7 | 1.6 | 7.2 | 1.4 | 1.6 | 1.0 | 3.9 | 1.0 | 0.3 | 0.7 | 0.3 | 2.7 | 1.9 | 3.5 | 9.6 | 0.6 | 5.915 .9 |
| 1995 | 3.0 | 0.6 | 0.4 | 0.6 | 0.6 | 1.4 | 6.5 | 1.3 | 1.4 | 0.9 | 3.4 | 0.9 | 0.3 | 0.6 | 0.3 | 2.4 | 1.6 | 3.0 | 8.4 | 0.619 .2 |
| 1996 | 2.2 | 2.7 | 0.6 | 0.4 | 0.5 | 0.5 | 1.3 | 5.8 | 1.1 | 1.3 | 0.8 | 3.0 | 0.8 | 0.3 | 0.5 | 0.2 | 2.1 | 1.4 | 2.7 | 7.417 .3 |
| 1997 | 1.3 | 2.0 | 2.4 | 0.5 | 0.3 | 0.5 | 0.5 | 1.1 | 5.1 | 1.0 | 1.1 | 0.7 | 2.7 | 0.7 | 0.2 | 0.5 | 0.2 | 1.8 | 1.3 | 2.321 .6 |
| 1998 | 2.1 | 1.2 | 1.8 | 2.2 | 0.5 | 0.3 | 0.4 | 0.4 | 1.0 | 4.5 | 0.9 | 1.0 | 0.6 | 2.3 | 0.6 | 0.2 | 0.4 | 0.2 | 1.6 | 1.120 .9 |
| 1999 | 7.6 | 1.8 | 1.0 | 1.6 | 1.9 | 0.4 | 0.3 | 0.4 | 0.4 | 0.9 | 4.0 | 0.8 | 0.9 | 0.5 | 2.0 | 0.5 | 0.2 | 0.3 | 0.1 | 1.419 .1 |
| 2000 | 6.8 | 6.8 | 1.7 | 0.9 | 1.4 | 1.7 | 0.4 | 0.2 | 0.3 | 0.3 | 0.8 | 3.5 | 0.7 | 0.8 | 0.5 | 1.8 | 0.4 | 0.1 | 0.3 | 0.117 .9 |
| 2001 | 7.8 | 6.1 | 6.1 | 1.5 | 0.8 | 1.3 | 1.5 | 0.3 | 0.2 | 0.3 | 0.3 | 0.7 | 3.1 | 0.6 | 0.7 | 0.4 | 1.5 | 0.4 | 0.1 | 0.315 .7 |
| 2002 | 1.0 | 7.0 | 5.4 | 5.4 | 1.3 | 0.7 | 1.1 | 1.4 | 0.3 | 0.2 | 0.2 | 0.3 | 0.6 | 2.7 | 0.5 | 0.6 | 0.3 | 1.3 | 0.3 | 0.114 .0 |
| 2003 | 0.4 | 0.9 | 6.3 | 4.9 | 4.8 | 1.2 | 0.7 | 1.0 | 1.2 | 0.3 | 0.2 | 0.2 | 0.2 | 0.5 | 2.4 | 0.5 | 0.5 | 0.3 | 1.2 | 0.312 .4 |
| 2004 | 0.4 | 0.3 | 0.8 | 5.6 | 4.3 | 4.3 | 1.1 | 0.6 | 0.9 | 1.1 | 0.2 | 0.1 | 0.2 | 0.2 | 0.5 | 2.1 | 0.4 | 0.4 | 0.3 | 1.011 .2 |
| 2005 | 0.6 | 0.3 | 0.3 | 0.7 | 5.0 | 3.9 | 3.9 | 0.9 | 0.5 | 0.8 | 1.0 | 0.2 | 0.1 | 0.2 | 0.2 | 0.4 | 1.8 | 0.4 | 0.4 | 0.210 .8 |
| 2006 | 4.8 | 0.5 | 0.3 | 0.3 | 0.6 | 4.5 | 3.5 | 3.4 | 0.8 | 0.5 | 0.7 | 0.8 | 0.2 | 0.1 | 0.1 | 0.2 | 0.4 | 1.6 | 0.3 | 0.39 .7 |
| 2007 | 5.7 | 4.3 | 0.5 | 0.3 | 0.2 | 0.6 | 4.0 | 3.1 | 3.1 | 0.7 | 0.4 | 0.6 | 0.7 | 0.2 | 0.1 | 0.1 | 0.1 | 0.3 | 1.4 | 0.38 .9 |
| 2008 | 8.0 | 5.1 | 3.8 | 0.4 | 0.2 | 0.2 | 0.5 | 3.6 | 2.8 | 2.7 | 0.7 | 0.4 | 0.5 | 0.7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.3 | 1.38 .1 |
| 2009 | 8.7 | 7.2 | 4.5 | 3.4 | 0.4 | 0.2 | 0.2 | 0.5 | 3.2 | 2.4 | 2.4 | 0.6 | 0.3 | 0.5 | 0.6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.28 .3 |
| 2010 | 41.6 | 7.8 | 6.4 | 4.0 | 3.0 | 0.3 | 0.2 | 0.2 | 0.4 | 2.8 | 2.1 | 2.1 | 0.5 | 0.3 | 0.4 | 0.5 | 0.1 | 0.1 | 0.1 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 2011 |  | 37.2 | 7.0 | 5.7 | 3.6 | 2.7 | 0.3 | 0.2 | 0.1 | 0.4 | 2.4 | 1.9 | 1.8 | 0.4 | 0.2 | 0.4 | 0.4 | 0.1 | 0.1 | 0.10 .1 |

Table 5.14 (cont'd). Estimated beginning of year numbers of Greenland turbot by age and sex (millions).

## Males

| Yr | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20+ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 87.8 | 6.0 | 58.1 | 15.7 | 28.2 | 2.0 | 1.2 | 2.6 | 12.0 | 7.6 | 4.7 | 2.9 | 1.9 | 1.3 | 0.9 | 0.6 | 0.4 | 0.2 | 0.0 | 0.0 | 0.0 |
| 1978 | 26.8 | 76.5 | 5.2 | 50.4 | 13.5 | 23.7 | 1.6 | 0.9 | 2.0 | 9.0 | 5.6 | 3.4 | 2.1 | 1.3 | 0.9 | 0.6 | 0.4 | 0.3 | 0.2 | 0.0 | 0.0 |
| 1979 | 12.2 | 23.3 | 66.4 | 4.5 | 43.0 | 11.1 | 18.9 | 1.2 | 0.7 | 1.4 | 6.1 | 3.7 | 2.2 | 1.3 | 0.8 | 0.6 | 0.4 | 0.3 | 0.2 | 0.1 | 0.0 |
| 1980 | 15.3 | 10.6 | 20.3 | 57.5 | 3.8 | 35.6 | 8.9 | 14.3 | 0.9 | 0.5 | 0.9 | 4.0 | 2.4 | 1.4 | 0.8 | 0.5 | 0.3 | 0.2 | 0.2 | 0.1 | 0.1 |
| 1981 | 1.4 | 13.3 | 9.2 | 17.5 | 48.7 | 3.1 | 27.5 | 6.4 | 9.8 | 0.6 | 0.3 | 0.6 | 2.4 | 1.4 | 0.8 | 0.5 | 0.3 | 0.2 | 0.1 | 0.1 | 0.1 |
| 1982 | 2.8 | 1.2 | 11.6 | 7.9 | 14.8 | 39.3 | 2.4 | 19.5 | 4.3 | 6.2 | 0.4 | 0.2 | 0.3 | 1.3 | 0.8 | 0.4 | 0.3 | 0.2 | 0.1 | 0.1 | 0.1 |
| 1983 | 1.2 | 2.5 | 1.1 | 10.0 | 6.7 | 12.0 | 30.0 | 1.7 | 13.1 | 2.7 | 3.8 | 0.2 | 0.1 | 0.2 | 0.7 | 0.4 | 0.2 | 0.1 | 0.1 | 0.1 | 0.1 |
| 1984 | 3.2 | 1.1 | 2.1 | 0.9 | 8.4 | 5.5 | 9.2 | 21.8 | 1.2 | 8.6 | 1.7 | 2.3 | 0.1 | 0.1 | 0.1 | 0.4 | 0.2 | 0.1 | 0.1 | 0.0 | 0.1 |
| 1985 | 11.1 | 2.8 | 0.9 | 1.9 | 0.8 | 7.1 | 4.5 | 7.4 | 17.0 | 0.9 | 6.4 | 1.3 | 1.7 | 0.1 | 0.0 | 0.1 | 0.3 | 0.2 | 0.1 | 0.1 | 0.1 |
| 1986 | 2.5 | 9.7 | 2.4 | 0.8 | 1.6 | 0.7 | 6.0 | 3.7 | 6.0 | 13.6 | 0.7 | 5.0 | 1.0 | 1.3 | 0.1 | 0.0 | 0.1 | 0.2 | 0.1 | 0.1 | 0.1 |
| 1987 | 3.6 | 2.2 | 8.5 | 2.1 | 0.7 | 1.4 | 0.6 | 5.0 | 3.1 | 4.9 | 11.1 | 0.6 | 4.1 | 0.8 | 1.0 | 0.1 | 0.0 | 0.0 | 0.2 | 0.1 | 0.1 |
| 1988 | 2.8 | 3.2 | 1.9 | 7.4 | 1.8 | 0.6 | 1.2 | 0.5 | 4.2 | 2.5 | 4.0 | 9.0 | 0.5 | 3.3 | 0.6 | 0.8 | 0.0 | 0.0 | 0.0 | 0.1 | 0.2 |
| 1989 | 12.7 | 2.4 | 2.8 | 1.7 | 6.4 | 1.6 | 0.5 | 1.0 | 0.4 | 3.5 | 2.1 | 3.3 | 7.4 | 0.4 | 2.7 | 0.5 | 0.7 | 0.0 | 0.0 | 0.0 | 0.3 |
| 1990 | 2.5 | 11.1 | 2.1 | 2.4 | 1.4 | 5.5 | 1.3 | 0.4 | 0.8 | 0.3 | 2.9 | 1.7 | 2.7 | 6.0 | 0.3 | 2.2 | 0.4 | 0.6 | 0.0 | 0.0 | 0.2 |
| 1991 | 0.9 | 2.2 | 9.6 | 1.9 | 2.1 | 1.2 | 4.6 | 1.1 | 0.4 | 0.7 | 0.3 | 2.2 | 1.3 | 2.1 | 4.6 | 0.2 | 1.7 | 0.3 | 0.4 | 0.0 | 0.2 |
| 1992 | 0.8 | 0.8 | 1.9 | 8.4 | 1.6 | 1.8 | 1.0 | 3.9 | 0.9 | 0.3 | 0.5 | 0.2 | 1.8 | 1.1 | 1.7 | 3.7 | 0.2 | 1.3 | 0.3 | 0.3 | 0.2 |
| 1993 | 0.5 | 0.7 | 0.7 | 1.7 | 7.3 | 1.4 | 1.6 | 0.9 | 3.4 | 0.8 | 0.3 | 0.5 | 0.2 | 1.6 | 0.9 | 1.4 | 3.2 | 0.2 | 1.1 | 0.2 | 0.4 |
| 1994 | 0.7 | 0.5 | 0.6 | 0.6 | 1.4 | 6.3 | 1.2 | 1.3 | 0.8 | 2.9 | 0.7 | 0.2 | 0.4 | 0.2 | 1.3 | 0.8 | 1.2 | 2.6 | 0.1 | 0.9 | 0.5 |
| 1995 | 3.0 | 0.6 | 0.4 | 0.5 | 0.5 | 1.3 | 5.5 | 1.1 | 1.1 | 0.7 | 2.4 | 0.5 | 0.2 | 0.3 | 0.1 | 1.0 | 0.6 | 0.9 | 2.0 | 0.1 | 1.1 |
| 1996 | 2.2 | 2.6 | 0.5 | 0.3 | 0.5 | 0.5 | 1.1 | 4.8 | 0.9 | 1.0 | 0.5 | 1.9 | 0.4 | 0.1 | 0.2 | 0.1 | 0.8 | 0.5 | 0.7 | 1.6 | 1.0 |
| 1997 | 1.3 | 1.9 | 2.3 | 0.5 | 0.3 | 0.4 | 0.4 | 0.9 | 4.1 | 0.8 | 0.8 | 0.5 | 1.6 | 0.4 | 0.1 | 0.2 | 0.1 | 0.7 | 0.4 | 0.6 | 2.1 |
| 1998 | 2.1 | 1.1 | 1.7 | 2.0 | 0.4 | 0.3 | 0.3 | 0.3 | 0.8 | 3.5 | 0.7 | 0.7 | 0.4 | 1.3 | 0.3 | 0.1 | 0.2 | 0.1 | 0.5 | 0.3 | 2.2 |
| 1999 | 7.6 | 1.8 | 1.0 | 1.4 | 1.7 | 0.4 | 0.2 | 0.3 | 0.3 | 0.7 | 3.0 | 0.5 | 0.6 | 0.3 | 1.1 | 0.2 | 0.1 | 0.1 | 0.1 | 0.4 | 1.9 |
| 2000 | 6.8 | 6.6 | 1.6 | 0.9 | 1.3 | 1.5 | 0.3 | 0.2 | 0.3 | 0.3 | 0.6 | 2.5 | 0.4 | 0.5 | 0.2 | 0.9 | 0.2 | 0.1 | 0.1 | 0.0 | 1.9 |
| 2001 | 7.8 | 5.9 | 5.7 | 1.4 | 0.7 | 1.1 | 1.3 | 0.3 | 0.2 | 0.2 | 0.2 | 0.5 | 2.0 | 0.4 | 0.4 | 0.2 | 0.7 | 0.2 | 0.0 | 0.1 | 1.5 |
| 2002 | 1.0 | 6.8 | 5.2 | 5.0 | 1.2 | 0.6 | 1.0 | 1.1 | 0.2 | 0.1 | 0.2 | 0.2 | 0.4 | 1.6 | 0.3 | 0.3 | 0.2 | 0.5 | 0.1 | 0.0 | 1.2 |
| 2003 | 0.4 | 0.9 | 5.9 | 4.5 | 4.4 | 1.0 | 0.6 | 0.8 | 1.0 | 0.2 | 0.1 | 0.2 | 0.1 | 0.3 | 1.3 | 0.2 | 0.2 | 0.1 | 0.4 | 0.1 | 1.0 |
| 2004 | 0.4 | 0.3 | 0.8 | 5.2 | 3.9 | 3.8 | 0.9 | 0.5 | 0.7 | 0.8 | 0.2 | 0.1 | 0.1 | 0.1 | 0.3 | 1.1 | 0.2 | 0.2 | 0.1 | 0.4 | 0.9 |
| 2005 | 0.6 | 0.3 | 0.3 | 0.7 | 4.5 | 3.4 | 3.3 | 0.8 | 0.4 | 0.6 | 0.7 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.9 | 0.2 | 0.2 | 0.1 | 1.0 |
| 2006 | 4.8 | 0.5 | 0.3 | 0.2 | 0.6 | 3.9 | 3.0 | 2.9 | 0.7 | 0.4 | 0.5 | 0.6 | 0.1 | 0.1 | 0.1 | 0.1 | 0.2 | 0.7 | 0.1 | 0.1 | 0.9 |
| 2007 | 5.7 | 4.2 | 0.5 | 0.2 | 0.2 | 0.5 | 3.4 | 2.6 | 2.5 | 0.6 | 0.3 | 0.4 | 0.5 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.6 | 0.1 | 0.9 |
| 2008 | 8.0 | 4.9 | 3.6 | 0.4 | 0.2 | 0.2 | 0.4 | 3.0 | 2.2 | 2.1 | 0.5 | 0.3 | 0.4 | 0.4 | 0.1 | 0.0 | 0.1 | 0.1 | 0.1 | 0.5 | 0.8 |
| 2009 | 8.7 | 7.0 | 4.3 | 3.1 | 0.3 | 0.2 | 0.2 | 0.4 | 2.6 | 1.9 | 1.8 | 0.4 | 0.2 | 0.3 | 0.3 | 0.1 | 0.0 | 0.0 | 0.0 | 0.1 | 1.0 |
| 2010 | 41.6 | 7.6 | 6.1 | 3.7 | 2.7 | 0.3 | 0.2 | 0.1 | 0.3 | 2.1 | 1.5 | 1.4 | 0.3 | 0.2 | 0.2 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.8 |
| 2011 | 6.8 | 36.3 | 6.6 | 5.3 | 3.3 | 2.4 | 0.3 | 0.1 | 0.1 | 0.3 | 1.8 | 1.2 | 1.1 | 0.2 | 0.1 | 0.2 | 0.2 | 0.0 | 0.0 | 0.0 | 0.7 |

Table 5.15. Age-equivalent sex-specific selectivity estimates (as estimated for 2011) from each gear type for Greenland turbot in the BSAI. Note that selectivity processes are modeled as a function of size and that some selectivities-at-length are allowed to vary over time.

| Age | Trawl Fishery Female | Male | Longline fishery Female | Male |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 0.009 | 0.012 | 0.004 | 0.005 |
| 2 | 0.008 | 0.012 | 0.004 | 0.006 |
| 3 | 0.006 | 0.012 | 0.004 | 0.007 |
| 4 | 0.005 | 0.012 | 0.004 | 0.008 |
| 5 | 0.006 | 0.016 | 0.009 | 0.012 |
| 6 | 0.021 | 0.046 | 0.040 | 0.038 |
| 7 | 0.058 | 0.140 | 0.119 | 0.110 |
| 8 | 0.105 | 0.293 | 0.225 | 0.220 |
| 9 | 0.140 | 0.462 | 0.330 | 0.345 |
| 10 | 0.158 | 0.613 | 0.415 | 0.463 |
| 11 | 0.162 | 0.731 | 0.480 | 0.565 |
| 12 | 0.156 | 0.816 | 0.526 | 0.648 |
| 13 | 0.146 | 0.875 | 0.558 | 0.716 |
| 14 | 0.135 | 0.914 | 0.581 | 0.769 |
| 15 | 0.123 | 0.939 | 0.597 | 0.812 |
| 16 | 0.113 | 0.956 | 0.608 | 0.847 |
| 17 | 0.103 | 0.967 | 0.616 | 0.876 |
| 18 | 0.094 | 0.975 | 0.622 | 0.899 |
| 19 | 0.087 | 0.980 | 0.626 | 0.919 |
| 20 | 0.080 | 0.984 | 0.629 | 0.936 |
| 21 | 0.075 | 0.988 | 0.631 | 0.951 |
| 22 | 0.070 | 0.992 | 0.633 | 0.964 |
| 23 | 0.065 | 0.996 | 0.634 | 0.975 |
| 24 | 0.061 | 1.000 | 0.635 | 0.986 |
| 25 | 0.059 | 0.998 | 0.636 | 0.992 |
| 26 | 0.057 | 0.990 | 0.636 | 0.994 |
| 27 | 0.056 | 0.983 | 0.636 | 0.996 |
| 28 | 0.055 | 0.977 | 0.636 | 0.997 |
| 29 | 0.054 | 0.972 | 0.636 | 0.998 |
| 30 | 0.052 | 0.964 | 0.636 | 1.000 |

Table 5.16. Age and sex-specific mean length and weights-at-age estimates for BSAI Greenland turbot.

|  | Mid-year length (cm) |  | Mid-year weight (kg) |  |
| ---: | ---: | ---: | ---: | ---: |
| Age | Females | Males | Females | Males |
| 1 | 12.36 | 12.36 | 0.012 | 0.012 |
| 2 | 22.06 | 21.98 | 0.084 | 0.083 |
| 3 | 30.84 | 30.41 | 0.257 | 0.245 |
| 4 | 38.50 | 37.51 | 0.539 | 0.494 |
| 5 | 45.17 | 43.50 | 0.916 | 0.808 |
| 6 | 50.98 | 48.54 | 1.369 | 1.162 |
| 7 | 56.04 | 52.78 | 1.874 | 1.534 |
| 8 | 60.45 | 56.36 | 2.407 | 1.904 |
| 9 | 64.29 | 59.37 | 2.949 | 2.259 |
| 10 | 67.63 | 61.90 | 3.485 | 2.591 |
| 11 | 70.55 | 64.04 | 4.001 | 2.895 |
| 12 | 73.09 | 65.84 | 4.490 | 3.167 |
| 13 | 75.30 | 67.35 | 4.945 | 3.408 |
| 14 | 77.23 | 68.63 | 5.364 | 3.619 |
| 15 | 78.91 | 69.71 | 5.743 | 3.802 |
| 16 | 80.37 | 70.61 | 6.085 | 3.960 |
| 17 | 81.65 | 71.38 | 6.391 | 4.094 |
| 18 | 82.76 | 72.02 | 6.664 | 4.208 |
| 19 | 83.72 | 72.56 | 6.905 | 4.305 |
| 20 | 84.57 | 73.02 | 7.118 | 4.386 |
| 21 | 85.30 | 73.40 | 7.306 | 4.453 |
| 22 | 85.94 | 73.72 | 7.472 | 4.509 |
| 23 | 86.50 | 74.00 | 7.618 | 4.555 |
| 24 | 86.98 | 74.22 | 7.746 | 4.593 |
| 25 | 87.41 | 74.42 | 7.861 | 4.628 |
| 26 | 87.77 | 74.58 | 7.964 | 4.662 |
| 27 | 88.10 | 74.72 | 8.053 | 4.691 |
| 28 | 88.38 | 74.83 | 4.715 |  |
| 29 | 88.62 | 74.93 | 8.131 | 4.736 |
| 30 | 89.04 | 75.08 | 8.199 | 4.767 |
|  |  |  | 8.317 |  |

Table 5.17. Estimated total Greenland turbot harvest by area, 1977-2011. Values for 2011 are through Nov. ${ }^{\text {nd }}, 2011$ and are preliminary.

| Year | EBS | Aleutians | Year | EBS | Aleutians |
| ---: | ---: | ---: | ---: | ---: | ---: |
| 1977 | 27,708 | 2,453 | 1994 | 3,875 | 7,141 |
| 1978 | 37,423 | 4,766 | 1995 | 4,499 | 5,855 |
| 1979 | 34,998 | 6,411 | 1996 | 4,258 | 4,844 |
| 1980 | 48,856 | 3,697 | 1997 | 5,730 | 6,435 |
| 1981 | 52,921 | 4,400 | 1998 | 7,839 | 8,329 |
| 1982 | 45,805 | 6,317 | 1999 | 5,179 | 5,391 |
| 1983 | 43,443 | 4,115 | 2000 | 5,667 | 5,888 |
| 1984 | 21,317 | 1,803 | 2001 | 4,102 | 4,252 |
| 1985 | 14,698 | 33 | 2002 | 3,011 | 3,153 |
| 1986 | 7,710 | 2,154 | 2003 | 2,493 | 965 |
| 1987 | 6,519 | 3,066 | 2004 | 1,805 | 414 |
| 1988 | 6,064 | 1,044 | 2005 | 2,140 | 468 |
| 1989 | 4,061 | 4,761 | 2006 | 1,452 | 534 |
| 1990 | 7,702 | 2,494 | 2007 | 1,481 | 521 |
| 1991 | 3,781 | 4,397 | 2008 | 2,095 | 827 |
| 1992 | 1,767 | 2,462 | 2009 | 2,249 | 2,261 |
| 1993 | 4,878 | 6,330 | 2010 | 2,272 | 1,866 |
|  |  |  | 2011 | 2,759 | 493 |

Table 5.18. Mean spawning biomass, F, and yield projections for Greenland turbot, 2011-2024. The full-selection fishing mortality rates ( $F$ 's) between longline and trawl gears were assumed to be 50:50 (whereas the catch from 2008-2011 has been around 40:60). The values for $B_{40 \%}$ and $B_{35 \%}$ are 21,560 and 18,870 tons, respectively.

| Catch | Max $F_{\text {ABC }}$ | $F_{\text {ABC }}$ | 5-year avg. | $F_{75 \%}$ | No Fishing | Scenario 6 | Scenario 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 |
| 2012 | 9,660 | 9,660 | 1,624 | 2,409 | 0 | 11,658 | 9,660 |
| 2013 | 8,029 | 8,029 | 1,570 | 2,296 | 0 | 9,316 | 8,029 |
| 2014 | 7,086 | 7,086 | 1,564 | 2,259 | 0 | 7,993 | 8,571 |
| 2015 | 6,887 | 6,887 | 1,635 | 2,342 | 0 | 7,676 | 8,092 |
| 2016 | 7,721 | 7,721 | 1,853 | 2,644 | 0 | 8,648 | 8,950 |
| 2017 | 9,706 | 9,706 | 2,263 | 3,232 | 0 | 11,001 | 11,224 |
| 2018 | 11,933 | 11,933 | 2,757 | 3,939 | 0 | 13,556 | 13,724 |
| 2019 | 13,288 | 13,288 | 3,181 | 4,528 | 0 | 14,933 | 15,060 |
| 2020 | 13,434 | 13,434 | 3,458 | 4,890 | 0 | 14,793 | 14,890 |
| 2021 | 12,690 | 12,690 | 3,592 | 5,033 | 0 | 13,635 | 13,708 |
| 2022 | 11,527 | 11,527 | 3,615 | 5,012 | 0 | 12,092 | 12,147 |
| 2023 | 10,297 | 10,297 | 3,564 | 4,888 | 0 | 10,590 | 10,632 |
| 2024 | 9,195 | 9,195 | 3,470 | 4,709 | 0 | 9,329 | 9,360 |
| Fishing M. | Max FABC | FABC | 5 -year avg. | F75\% | No Fishing | Scenario 6 | Scenario 7 |
| 2011 | 0.108 | 0.108 | 0.108 | 0.108 | 0.108 | 0.108 | 0.108 |
| 2012 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.367 |
| 2013 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.367 |
| 2014 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2015 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2016 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2017 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2018 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2019 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2020 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2021 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2022 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2023 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| 2024 | 0.367 | 0.367 | 0.057 | 0.085 | 0.000 | 0.453 | 0.453 |
| Spawning biomass | Max FABC | FABC | 5 -year avg. | F75\% | No Fishing | Scenario 6 | Scenario 7 |
| 2011 | 51,278 | 51,278 | 51,278 | 51,278 | 51,278 | 51,278 | 51,278 |
| 2012 | 47,687 | 47,687 | 47,687 | 47,687 | 47,687 | 47,687 | 47,687 |
| 2013 | 41,441 | 41,441 | 45,790 | 45,375 | 46,642 | 40,325 | 41,441 |
| 2014 | 37,478 | 37,478 | 45,168 | 44,399 | 46,769 | 35,634 | 37,478 |
| 2015 | 36,528 | 36,528 | 46,800 | 45,727 | 49,067 | 34,221 | 35,699 |
| 2016 | 39,314 | 39,314 | 51,693 | 50,347 | 54,573 | 36,698 | 37,879 |
| 2017 | 44,229 | 44,229 | 58,571 | 56,959 | 62,061 | 41,347 | 42,290 |
| 2018 | 48,552 | 48,552 | 65,139 | 63,229 | 69,311 | 45,330 | 46,083 |
| 2019 | 50,761 | 50,761 | 70,172 | 67,902 | 75,166 | 47,060 | 47,659 |
| 2020 | 50,729 | 50,729 | 73,428 | 70,739 | 79,376 | 46,470 | 46,945 |
| 2021 | 49,039 | 49,039 | 75,125 | 71,987 | 82,110 | 44,256 | 44,630 |
| 2022 | 46,389 | 46,389 | 75,590 | 72,008 | 83,623 | 41,205 | 41,499 |
| 2023 | 43,346 | 43,346 | 75,150 | 71,156 | 84,183 | 37,924 | 38,154 |
| 2024 | 40,291 | 40,291 | 74,080 | 69,725 | 84,021 | 34,787 | 34,966 |

Table 5.19. Summary management values based on this assessment. Note that the fishing mortality rates assume 50:50 contribution from longline gear and trawl gear.

| Management parameter | Value |
| :--- | ---: |
| (natural mortality) | $0.112 \mathrm{yr}^{-1}$ |
| Amendment 56 Tier (in 2012) | 3 a |
| Approximate age at full recruitment | 10 years |
| $F_{35 \%}\left(F_{\text {OFL }}\right)$ | 0.453 |
| $F_{40 \%}$ | 0.367 |
| $B_{100 \%}$ | $53,900 \mathrm{t}$ |
| $B_{40 \%}$ | $21,560 \mathrm{t}$ |
| $B_{35 \%}$ | $18,870 \mathrm{t}$ |
| 2011 female spawning biomass | $51,278 \mathrm{t}$ |
| 2012 female spawning biomass | $47,687 \mathrm{t}$ |
| 2013 female spawning biomass | $41,441 \mathrm{t}$ |
| 2011 total (age 1+) biomass | $75,026 \mathrm{t}$ |
| 2012 total (age 1+) biomass | $76,845 \mathrm{t}$ |
| 2013 total (age 1+) biomass | $73,913 \mathrm{t}$ |
| $F_{\text {ABC }}=F_{40 \%}($ max permissible) | 0.367 |
| 2012 Maximum permissible ABC | $9,660 \mathrm{t}$ |
| 2013 Maximum permissible ABC | $8,029 \mathrm{t}$ |
| $F_{\text {ABC }}=F_{40 \%}$ | 0.367 |
| Recommended ABC: 2012 | $9,660 \mathrm{t}$ |
|  | $8,029 \mathrm{t}$ |
| $F_{\text {overishing }}=F_{35 \%}$ | 0.453 |
| 2012 Greenland turbot OFL | $11,658 \mathrm{t}$ |
| 2013 Greenland turbot OFL | $9,697 \mathrm{t}$ |
|  |  |

* assuming catch $2012=9,660 t$


## Figures

(a)

(b)


Figure 5.1. Schematic representation of Greenland halibut distribution and connectivity from larvae to settled juveniles. (a) Horizontally changed distribution through different life history stages (Blue circle: slope spawning ground, Green circle: shelf nursery ground of pelagic juveniles, Red circle: settlement ground). Blue arrows: possible larval transport routes from slope to shelf. (b) Vertically changed distribution as they develop. Source: Sohn (2009).


Figure 5.2. Trawl and longline catches of Greenland turbot in the combined EBS/AI area, 19602011.

# All observer length frequency data (sexes combined) 



Figure 5.3. Combined fishery catch length frequency of Greenland turbot in the BSAI area, 19892010 based on NMFS observer program data, sexes combined (based on raw observations)..


Figure 5.4. Greenland turbot relative CPUE by number based on NMFS EBS bottom trawl surveys, 2004-2011. Also shown are bottom temperature contours.


Figure 5.5. Greenland turbot relative CPUE by weight based on NMFS EBS bottom trawl surveys, 2004-2011. Also shown are bottom temperature contours.



Figure 5.6. Survey biomass estimates of Greenland turbot from the EBS shelf trawl survey (top; 1982-2011) and the proportion of tows that caught at least one Greenland turbot (bottom; 1987-2011).


Figure 5.7. Abundance-at-length (cm) for Greenland turbot observed from the summer NMFS shelf trawl surveys, 1985-2011 (sexes combined, all strata except for 1986 where only strata 16 were sampled). Note that the 2010 and 2011 are truncated in the main figure-the inset shows the same figure with the vertical scale magnified by an order of magnitude.


Figure 5.8. Greenland turbot catch per unit effort (relative values by weight, vertical bars) from the Aleutian Islands region bottom trawl survey, 2002-2010.


Figure 5.9. Greenland turbot relative biomass from the Aleutian Islands surveys by region, 19912010.


Figure 5.10. Longline survey Greenland turbot proportions at length over time (sexes combined) as used in the model.
Population index


Figure 5.11. Model fits to the EBS shelf trawl survey (top left), the EBS slope trawl survey (top right) and longline survey (bottom) indices for Greenland turbot in the EBS/AI region.


Figure 5.12. Estimated total age-specific fishing mortality rate (gears and sexes combined) for BSAI Greenland turbot, 1977-2011.


Figure 5.13 Current assessment estimates of total age 1+ biomass for Greenland turbot in the BSAI region, 1965-2012 compared to previous assessments.


Figure 5.14. Estimated selectivity-at-age patterns for EBS Greenland turbot fisheries showing differences in sex-specific availability in the current year (some earlier years had different selectivity estimates).


Figure 5.15. Average size-specific estimates of selectivity patterns for EBS Greenland turbot surveys showing differences in sex-specific availability in the current year (some earlier years had different selectivity estimates).


Figure 5.16. Observed Greenland turbot sex ratio over time from the BSAI region trawl and longline fisheries.


Figure 5.17. EBS slope trawl survey estimates of Greenland turbot sex ratio.


Age
Figure 5.18. Estimated growth (length at age) of Greenland turbot by sex (female on left, males on right) in the EBS/AI region as predicted by the model and compared to the available age data using the methods of Gregg et al (2006).


Figure 5.19. Estimated recruitment at age 0 (thousands) for Greenland turbot in the EBS/AI region, 1977-2011.


Figure 5.20. Ratio of historical $F / F_{m s y}$ versus female spawning biomass relative to $B_{m s y}$ for BSAI Greenland turbot, 1960-2011. Note that the proxies for $F_{m s y}$ and $B_{m s y}$ are $F_{35 \%}$ and $B_{35 \%}$, respectively.


Figure 5.21. Stochastic trajectory of Greenland turbot female spawning biomass (top) and catch (bottom) for the maximum allowable fishing mortality rate under Amendment 56/56, Tier 3. The dotted lines represent the $90 \%$ confidence limits. Horizontal lines with marks are the values associated with $B_{40 \%}$ and $F_{40 \%}$ while the thick horizontal line is the expected value under constant $F_{O F L}$ rate ( $F_{35 \%}$ ).

## Attachment 5.A Fits to composition data



Figure 5.22. Greenland turbot model fit to longline survey length frequency data (sexes combined). Lines are model predictions, points are data.


## Length (cm)

Figure 5.23. Greenland turbot model fit to EBS slope trawl survey length frequency data. The left set are females, while the right set are males. Lines are model predictions, points are data; females on left and males on right.


Figure 5.24. Greenland turbot model fit to EBS shelf trawl survey female length frequency data. Lines are model predictions, points are data.


Figure 5.25. Greenland turbot model fit to EBS shelf trawl survey male length frequency data. Lines are model predictions, points are data


Figure 5.26. Greenland turbot model fit to EBS longline fishery length frequency data (combined sexes). Lines are model predictions, points are data; females on left and males on right.


Figure 5.27. Greenland turbot model fit to EBS trawl fishery length frequency data. The left set are females, while the right set are males. Lines are model predictions, points are data; females on left and males on right.

## Attachment 5.B Table for Model 2 (estimate differential male natural mortality)

The following tables are provided to allow alternative evaluation for ABC/OFL and status determinations should Model 2 be selected.

Model 2 (Alternative) summary table

| Quantity | As estimated or specified last year for: |  | As estimated or recommended this year for: |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 2011 | 2012 | 2012 | 2013 |
| $M$ (natural mortality rate) | 0.112 | 0.112 | 0.112 (0.208) | 0.112 (0.208) |
| Tier | 3a | 3a | 3a | 3а |
| Projected total (age $1+$ ) biomass (t) | 74,000 | 69,700 | 112,961 | 102,406 |
| Female spawning biomass (t) |  |  |  |  |
| Projected | 49,198 | 45,504 | 85,796 | 80,298 |
| $\mathrm{B}_{100 \%}$ | 71,048 | 71,048 | 74,873 | 74,873 |
| $B_{40 \%}$ | 28,419 | 28,419 | 29,949 | 29,949 |
| $B_{35 \%}$ | 24,867 | 24,867 | 26,205 | 26,205 |
| $F_{\text {OFL }}$ | 0.293 | 0.293 | 0.688 | 0.688 |
| $\operatorname{maxF}_{\text {ABC }}$ | 0.293 | 0.293 | 0.554 | 0.554 |
| $F_{\text {ABC }}$ | 0.247 | 0.247 | NA | NA |
| OFL (t) | 7,217 | 7,217 | 17,960 | 14,486 |
| maxABC (t) | 6,137 | 6,137 | 14,772 | 11,905 |
| ABC (t; BSAI) | 6,137 | 6,137 | 14,772 | 11,905 |
| EBS portion | 4,590 | 4,590 | 11,050 | 8,905 |
| C AI portion | 4,301 | 4,301 | 3,722 | 3,000 |
|  | As determine | ar for: | As determine | this year for: |
| Status | 2009 | 2010 | 2010 | 2011 |
| Overfishing | No | n/a | No | n/a |
| Overfished | n/a | No | n/a | No |
| Approaching overfished | n/a | No | n/a | No |

Table 5B.1. Mean spawning biomass, F, and yield projections for Greenland turbot, 2011-2024 for model 2 (optionally available) under the harvest scenarios. The full-selection fishing mortality rates ( $F$ 's) between longline and trawl gears were assumed to be 50:50

| Catch | Max FABC | FABC | 5 -year avg. | F75\% | No Fishing | Scenario 6 | Scenario 7 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 2011 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 | 3,252 |
| 2012 | 14,772 | 14,772 | 1,650 | 3,559 | 0 | 17,960 | 14,772 |
| 2013 | 11,905 | 11,905 | 1,553 | 3,278 | 0 | 13,910 | 11,905 |
| 2014 | 10,065 | 10,065 | 1,493 | 3,091 | 0 | 11,414 | 12,266 |
| 2015 | 9,199 | 9,199 | 1,486 | 3,033 | 0 | 10,268 | 10,870 |
| 2016 | 9,497 | 9,497 | 1,575 | 3,192 | 0 | 10,610 | 11,043 |
| 2017 | 10,951 | 10,951 | 1,781 | 3,609 | 0 | 12,366 | 12,682 |
| 2018 | 12,633 | 12,633 | 2,030 | 4,114 | 0 | 14,332 | 14,567 |
| 2019 | 13,570 | 13,570 | 2,233 | 4,508 | 0 | 15,288 | 15,464 |
| 2020 | 13,509 | 13,509 | 2,351 | 4,710 | 0 | 14,984 | 15,117 |
| 2021 | 12,737 | 12,737 | 2,390 | 4,739 | 0 | 13,854 | 13,954 |
| 2022 | 11,641 | 11,641 | 2,371 | 4,644 | 0 | 12,419 | 12,494 |
| 2023 | 10,504 | 10,504 | 2,315 | 4,478 | 0 | 11,023 | 11,080 |
| 2024 | 9,479 | 9,479 | 2,239 | 4,279 | 0 | 9,827 | 9,869 |
| Fishing M. | Max FABC | FABC | 5 -year avg. | F75\% | No Fishing | Scenario 6 | Scenario 7 |
| 2011 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 | 0.105 |
| 2012 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.554 |
| 2013 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.554 |
| 2014 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2015 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2016 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2017 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2018 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2019 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2020 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2021 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2022 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2023 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| 2024 | 0.554 | 0.554 | 0.057 | 0.125 | 0.000 | 0.688 | 0.688 |
| Spawning biomass | Max FABC | FABC | 5-year avg. | F75\% | No Fishing | Scenario 6 | Scenario 7 |
| 2011 | 85,796 | 85,796 | 85,796 | 85,796 | 85,796 | 85,796 | 85,796 |
| 2012 | 80,298 | 80,298 | 80,298 | 80,298 | 80,298 | 80,298 | 80,298 |
| 2013 | 68,528 | 68,528 | 76,942 | 75,738 | 77,977 | 66,433 | 68,528 |
| 2014 | 60,599 | 60,599 | 75,377 | 73,152 | 77,319 | 57,199 | 60,599 |
| 2015 | 57,433 | 57,433 | 76,999 | 73,907 | 79,738 | 53,268 | 55,933 |
| 2016 | 59,954 | 59,954 | 83,230 | 79,385 | 86,686 | 55,348 | 57,431 |
| 2017 | 65,825 | 65,825 | 92,298 | 87,754 | 96,435 | 60,905 | 62,528 |
| 2018 | 71,177 | 71,177 | 101,023 | 95,747 | 105,877 | 65,875 | 67,138 |
| 2019 | 73,806 | 73,806 | 107,716 | 101,600 | 113,388 | 67,937 | 68,917 |
| 2020 | 73,413 | 73,413 | 112,009 | 104,945 | 118,604 | 66,850 | 67,607 |
| 2021 | 70,715 | 70,715 | 114,186 | 106,109 | 121,776 | 63,480 | 64,063 |
| 2022 | 66,649 | 66,649 | 114,675 | 105,589 | 123,273 | 58,895 | 59,342 |
| 2023 | 62,016 | 62,016 | 113,904 | 103,873 | 123,473 | 53,961 | 54,303 |
| 2024 | 57,369 | 57,369 | 112,239 | 101,369 | 122,701 | 49,236 | 49,497 |

(This page intentionally left blank)


[^0]:    *Catch estimated as of October 2011

