

## 4. Assessment of the yellowfin sole stock in the Bering Sea and Aleutian Islands

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### Executive Summary

#### Summary of Changes in Assessment Inputs

##### *Changes to the input data*

- 1) 2012 fishery age composition.
- 2) 2012 survey age composition.
- 3) 2013 trawl survey biomass point estimate and standard error.
- 4) Estimate of the discarded and retained portions of the 2012 catch.
- 5) Estimate of total catch made through the end of 2013.

##### *Changes to the assessment methodology*

No changes were made to the assessment methodology. The assessment updates last year's with results and management quantities that are similar to the 2012 assessment. Yellowfin sole continue to be well-above  $B_{MSY}$  and the annual harvest remains below the ABC level. The stock is trending upwards due to a strong 2003 year class.

### Summary of Results

Quantity	As estimated or specified last year for: 2013      2014		As estimated or recommended this year for: 2014      2015	
$M$ (natural mortality rate)	0.12	0.12	0.12	0.12
Tier	1a	1a	1a	1a
Projected total (age 6+) biomass (t)	1,963,000	1,960,000	2,113,000	2,188,000
Female spawning biomass (t)				
Projected	582,300	601,000	581,100	594,800
$B_0$	966,900		989,800	
$B_{MSY}$	353,000		366,000	
$F_{OFL}$	0.112	0.112	0.123	0.123
$maxF_{ABC}$	0.105	0.105	0.113	0.113
$F_{ABC}$	0.105	0.105	0.113	0.113
OFL (t)	220,000	219,000	259,700	268,900
maxABC (t)	206,000	206,000	239,800	248,300
ABC (t)	206,000	206,000	239,800	248,300
Status	As determined last year for: 2011      2012		As determined this year for: 2012      2013	
Overfishing	No	n/a	No	n/a
Overfished	n/a	No	n/a	No
Approaching overfished	n/a	No	n/a	No

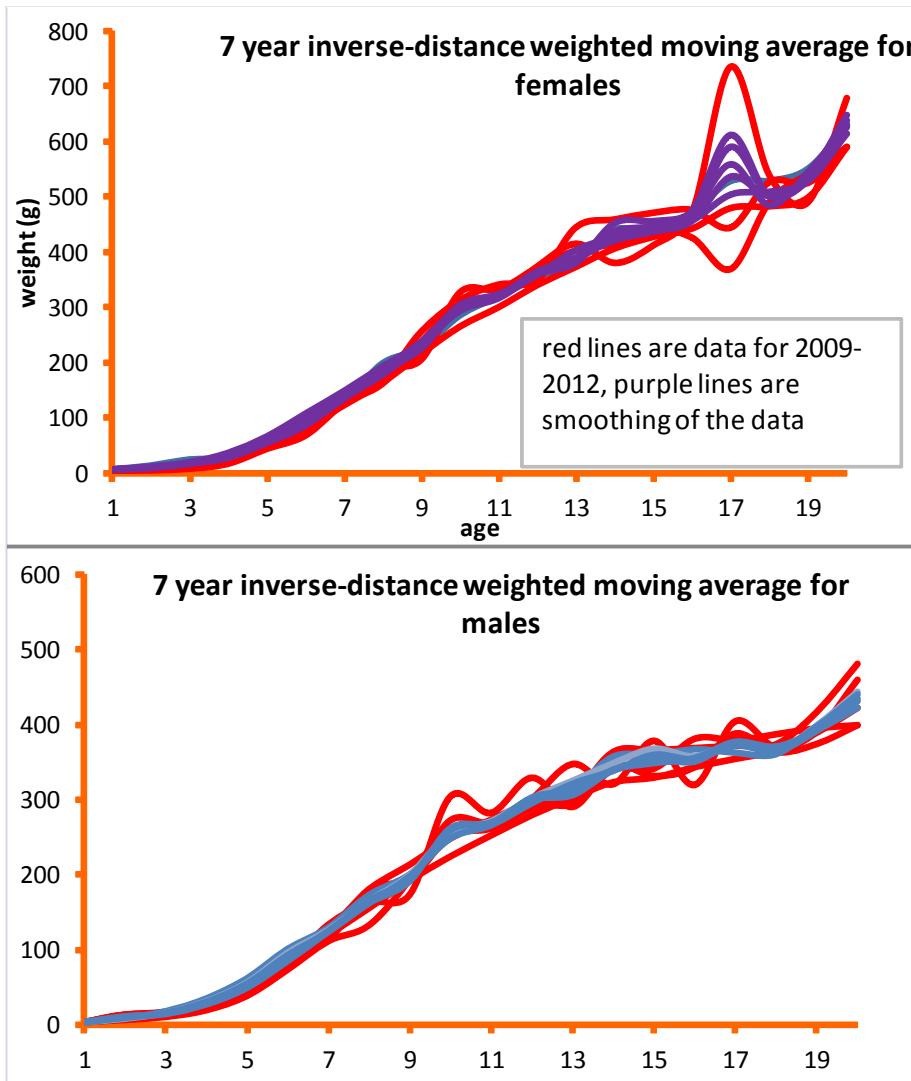
## **Responses to SSC and Plan Team Comments on Assessments in General**

No general comments.

## **Responses to SSC and Plan Team Comments specific to this Assessment**

Last year, the SSC supported the Plan Team's suggestion of examining simpler or non-parametric alternative growth models. The assessment authors indicated that an alternative growth model designed to smooth the empirical weight at age data should be implemented in next year's assessment. The SSC appreciates these efforts and looks forward to the results of this analysis.

The annual sex-specific weight-at-age data input into the accepted model were smoothed using a seven-year inverse-distance averaging weighting method. This method allows for the greatest weighting of annual data to come from the most recent years, consistent with the notion that somatic growth in yellowfin sole is temperature-dependent and that Bering Sea warm/cold years are autocorrelated. Results indicate that higher estimates of female spawning biomass, ABC and OFL come from the use of smoothed weight-at-age data. This occurs because the high wt-at-age for 17 year old females in 2009 influences many years of smoothed estimates, but only affects the 2009 wt-at-age in the empirical data set.



	<b>empirical data</b>	<b>smoothed</b>
<b>ABC</b>	240	253
<b>OFL</b>	260	271
<b>FSB</b>	581	600

## Introduction

The yellowfin sole (*Limanda aspera*) is one of the most abundant flatfish species in the eastern Bering Sea (EBS) and is the target of the largest flatfish fishery in the world. They inhabit the EBS shelf and are considered one stock. Abundance in the Aleutian Islands region is negligible.

Yellowfin sole are distributed in North American waters from off British Columbia, Canada, (approx. lat. 49° N) to the Chukchi Sea (about lat. 70° N) and south along the Asian coast to about lat. 35° N off the South Korean coast in the Sea of Japan. Adults exhibit a benthic lifestyle and occupy separate winter, spawning and summertime feeding distributions on the eastern Bering Sea shelf. From over-winter grounds near the shelf margins, adults begin a migration onto the inner shelf in April or early May each year for spawning and feeding. In recent years, the directed fishery has typically occurred from late winter through autumn (Wilderbuer et al. 1992). Yellowfin sole are managed as a single stock in the BSAI management area as there is presently no evidence of stock structure.

## Fishery

Yellowfin sole have annually been caught with bottom trawls on the Bering Sea shelf since the fishery began in 1954 and were overexploited by foreign fisheries in 1959-62 when catches averaged 404,000 t annually (Fig. 4.1, top panel). As a result of reduced stock abundance, catches declined to an annual average of 117,800 t from 1963-71 and further declined to an annual average of 50,700 t from 1972-77. The lower yield in this latter period was partially due to the discontinuation of the U.S.S.R. fishery. In the early 1980s, after the stock condition had improved, catches again increased reaching a peak of over 227,000 t in 1985.

During the 1980s, there was also a major transition in the characteristics of the fishery. Yellowfin sole were traditionally taken exclusively by foreign fisheries and these fisheries continued to dominate through 1984. However, U.S. fisheries developed rapidly during the 1980s in the form of joint ventures, and during the last half of the decade began to dominate and then take all of the catch as the foreign fisheries were phased out of the EBS. Since 1990, only domestic harvesting and processing has occurred.

The management of the yellowfin sole fishery changed significantly in 2008 with the implementation of Amendment 80 to the BSAI Fisheries Management Plan. The Amendment directly allocated fishery resources among BSAI trawl harvesters in consideration of their historic harvest patterns and future harvest needs in order to improve retention and utilization of fishery resources by the non-AFA trawl catcher/processor fleet. This was accomplished by extending the groundfish retention standards to all H&G vessels and also by providing the ability to form cooperatives within the newly formed Amendment 80 sector. In addition, Amendment 80 also mandated additional monitoring requirements which included observer coverage on all hauls, motion-compensating scales for weighing samples, flow scales to obtain accurate catch weight estimates for the entire catch, no mixing of hauls and no on-deck sorting. The partitioning of TAC and PSC (prohibited species catch) among cooperatives has significantly changed the way the annual catch has accumulated and the rate of target catch per bycatch ton (Fig 4.1, bottom panel).

Yellowfin sole are usually headed and gutted, frozen at sea, and then shipped to Asian countries for further processing (see “market profile” in the 2011 economic SAFE report for details). In 2010,

following a comprehensive assessment process, the yellowfin sole fishery was certified under the Marine Stewardship Council environmental standard for sustainable and well-managed fisheries. The certification also applies to all the major flatfish fisheries in the BSAI and GOA. The total annual catch (t) since implementation of the MFCMA in 1977 is shown in Table 4.1.

Also in 2010, federally permitted vessels using non-pelagic trawl gear whose harvest results in flatfish retained catch that is greater than any other retained fishery category were required to use modified trawl gear. The modifications required the use of elevating devices to raise the section of the trawl warps between the doors and the trawl wing tips by 2.5 inches off the seafloor. The purpose of the management action was to reduce damage of non-target animals, particularly those that form habitat structure or support other fisheries while not substantially reducing flatfish catch rates or causing gear handling problems (Rose et al. 2010).

The 1997 catch of 181,389 t was the largest since the fishery became completely domestic but was at lower levels from 1998 – 2010, averaging 94,004 t. The catch has increased the past three years (2010–2013) averaging 150,000 t. The 2012 catch totaled 147,183 t (73% of the ABC), the third highest annual catch in the past 15 years. For 2013, the catch distribution has been spread out fairly evenly by month (with the exception of July) with the majority coming from 4 BSAI management areas (509, 513, 514, 521). As of mid-October 2013, the fishing season is ongoing. In order to estimate the total 2013 catch for the stock assessment model, the average proportion of the 2008–2012 cumulative catch attained by the 37<sup>th</sup> week of the year (mid-September) was applied to the 2013 catch amount at the same time period and results in a 2013 catch estimate of 151,000 t (74% of the ABC and the highest catch since 1997). The size composition of the 2013 catch for both males and females, from observer sampling, are shown in Figure 4.2, the catch proportions by month and area are shown in Figure 4.3, and maps of the locations where yellowfin sole were caught in 2013, by month, are shown in Figure 4.4. The average age of yellowfin sole in the 2012 catch is estimated at 12.2 and 12.0 years for females and males, respectively.

The time-series of catch in Table 6.1 also includes yellowfin sole that were discarded in domestic fisheries during the period 1987 to the present. Annual discard estimates were calculated from at-sea sampling (Table 4.2). The rate of discard has ranged from a low of 2% of the total catch in 2012 to 30% in 1992. The trend has been toward fuller retention of the catch in recent years, and with the advent of the Amendment 80 harvest practices, discarding is at its lowest level since these estimates have become available. Historically, discarding primarily occurred in the yellowfin sole directed fishery, with lesser amounts in the Pacific cod, Pollock, rock sole, flathead sole, and “other flatfish” fisheries (Table 4.3).

## Data

The data used in this assessment include estimates of total catch, bottom trawl survey biomass estimates and their attendant 95% confidence intervals, catch-at-age from the fishery and population age composition estimates from the bottom trawl survey. Weight-at-age and proportion mature-at-age are also available from studies conducted during the bottom trawl surveys.

Data source	years
Fishery catch	1954-2013
Fishery age composition	1964-2012
Survey biomass and standard error, bottom temperature	1982-2013
Survey age composition	1979-2012
Annual length-at-age and weight-at-age	1979-2012
Maturity at age	Samples collected in 1992 and 1993

### *Fishery Catch and Catch-at-Age*

This assessment uses fishery catch data from 1955- 2012 (Table 4.1), including an estimate of the 2013 catch, and fishery catch-at-age (proportions) from 1964-2012 (Table 4.4, 1977-2011). The 2012 fishery age composition is primarily composed of fish older than 9 years with a large amount of 20+ fish.

### *Survey Biomass Estimates and Population Age Composition Estimates*

Indices of relative abundance available from AFSC surveys have also shown a major increase in the abundance of yellowfin sole during the late 1970s increasing from 21 kg/ha in 1975 to 51 kg/ha in 1981 (Fig. 4.2 in Bakkala and Wilderbuer 1990). These increases have also been documented through Japanese commercial pair trawl data and catch-at-age modeling in past assessments (Bakkala and Wilderbuer 1990).

Since 1981, the survey CPUEs have fluctuated widely (Fig. 4.5). Biomass estimates for yellowfin sole from the annual bottom trawl survey on the eastern Bering Sea shelf are shown in Table 4.5. The data show a doubling of survey biomass between 1975 and 1979 with a further increase to over 3.3 million t in 1981. Total survey abundance estimates fluctuated erratically from 1983 to 1990 with biomass ranging from as high as 3.5 million t in 1983 to as low as 1.9 million t in 1986. Biomass estimates since 1990 indicate an even trend at high levels of abundance for yellowfin sole, with the exception of the results from the 1999 and 2000 summer surveys, which were at lower levels. Surveys from 2001-2005 estimated an increase each year but the estimates since 2006 indicate a stable level with some annual variability. However, the 2012 estimate is a 19% decrease from 2011 and the 2013 estimate is 17% increase over 2012. Fluctuations of the magnitude shown between 1980 and 1990 and again between 1998 and 1999, 2008 and 2009 and also 2011 and 2012 are unreasonable considering the combined elements of slow growth and long life span of yellowfin sole and low to moderate exploitation rate, characteristics which should produce more gradual changes in abundance.

Variability of yellowfin sole survey abundance estimates (Fig. 4.6) is in part due to the availability of yellowfin sole to the survey area (Nichol, 1998). Yellowfin sole are known to undergo annual migrations from wintering areas off the shelf-slope break to near shore waters where they spawn throughout the spring and summer months (Nichol, 1995; Wakabayashi, 1989; Wilderbuer et al., 1992). Exploratory survey sampling in coastal waters of the eastern Bering Sea indicate that yellowfin sole concentrations can be greater in these shallower areas not covered by the standard AFSC survey than in the survey proper. Commercial bottom trawlers have commonly found high concentrations of yellowfin sole in areas such as near Togiak Bay (Low and Narita, 1990) and in more recent years from Kuskokwim Bay to just south of Nunivak Island. The coastline areas are sufficiently large enough to offer a substantial refuge for yellowfin sole from the current survey.

Over the past 15 years survey biomass estimates for yellowfin sole have shown a positive correlation with shelf bottom temperatures (Nichol, 1998); estimates have generally been lower during cold years. The 1999 survey, which was conducted in exceptionally cold waters, indicated a decline in biomass that was unrealistic. The bottom temperatures during the 2000 survey were much warmer than in 1999, and the biomass increased, but still did not approach estimates from earlier years. Average bottom temperature and biomass both increased again during the period 2001 – 2003, with the 2003 value the highest temperature and biomass observed over the 22 year time series. Given that both 1999 and 2000 surveys were conducted two weeks earlier than previous surveys, it is possible that the time difference may also have also affected the availability of yellowfin sole to the survey. If, for example, the timing of peak yellowfin sole spawning in nearshore waters corresponded to the time of the survey, a greater proportion of the population would be unavailable to the standard survey area. This trend was observed again in 2009 and 2012 when the temperatures and the bottom trawl survey point estimates were lower. Summer shelf bottom temperatures in 2012 were the 2<sup>nd</sup> coldest recorded by the survey and the time-series and resulted in a 19% decline from 2011.

We propose two possible reasons why survey biomass estimates are lower during years when bottom temperatures are low. First, catchability may be lower because yellowfin sole may be less active when cold. Less active fish may be less susceptible to herding, and escapement under the footrope of survey gear may increase if fish are less active. Secondly, bottom temperatures may influence the timing of the inshore spawning migrations of yellowfin sole and therefore affect their availability to the survey area. Because yellowfin sole spawning grounds include nearshore areas outside the survey area, availability of fish within the survey area can vary with the timing of this migration and the timing of the survey. In the case of 2012, a very cold year in the Bering Sea, it is unclear from examining survey station catches along the survey border near Kuskowkim bay if a significant portion of the biomass lies outside this border (Fig 4.7).

Yellowfin sole population numbers-at-age estimated from the annual bottom trawl surveys are shown in Table 4.6 and their occurrence in trawl survey hauls and associated collections of lengths and age structures since 1982 are shown in Table 4.7. Their total tonnage caught in the resource assessment surveys since 1982 are listed in Table 4.8 and also in an appendix table with IPHC survey catches.

#### *Length and Weight-at-Age*

Past assessments of yellowfin sole have used sex-specific, time-invariant growth based on the average length-at-age and weight-at-length relationships from the time-series of survey observations summed over all years since 1982. These weight-at-age estimates were estimated from the following relationships:

Parameters of the von Bertalanffy growth curve have been estimated for yellowfin sole, by sex, from the trawl survey database as follows:

	$L_{inf}$	K	$t_0$	n
Males	33.7	0.161	-0.111	656
Females	37.8	0.137	0.112	709

A sex-specific length-weight relationship was also calculated from the survey database using the usual power function, weight (g) = a Length(cm)<sup>b</sup>, where a and b are parameters estimated to provide the best fit to the data (Fig. 4.8).

	a	b	n
males	0.00854	3.081	2,701
females	0.0054	3.227	3,662

These estimates of weight at length were applied to the annual trawl survey estimates of population length at age averaged over all years, by sex, to calculate the weight at each age (Fig. 4.8).

Recent applications of dendrochronology (tree-ring techniques) have been used to develop biochronologies from the otolith growth increments of northern rock sole (*Lepidopsetta polyxystra*), yellowfin sole and Alaska plaice (*Pleuronectes quadrifilis*) in the eastern Bering Sea. These techniques ensure that all growth increments are assigned the correct calendar year, allowing for estimation of somatic growth by age and year for chronologies that span approximately 25 years (Matta et al. 2010). The analysis indicated that yellowfin sole somatic growth has annual variability and is positively correlated with May bottom water temperature in the Bering Sea (Fig. 4.9).

The relationship between temperature and growth was further explored by reanalyzing yellowfin sole growth by age and year. Length-weight data collected when obtaining otolith (age) samples in RACE surveys (n=7,000 from 1987, 1994 and 1999-2009) also indicate that weight at age exhibits annual variability and is highly correlated with summer bottom water temperature observations with a lag of 2-3 years for the temperature effect to be seen (shown for age 5 fish in figure 4.10). These observations were

then extended back to 1979 using survey population length-at-age estimates (since weight-at-age is a power function of the length-at-age, Clark et al. 1999, Walters and Wilderbuer 2000).

In this assessment the reanalyzed growth data were incorporated and growth was modeled as time-varying and temperature-dependent functions input into an age-structured stock assessment model and then comparing the results with the base model that uses time-invariant growth. Four growth models were developed as follows: Mean age-specific somatic body mass (here referred to as weight-at-age) is modeled as a von Bertalanfy growth function in the initial year of the stock assessment (1954) and projected forward such that the model expected mean weight at age  $j$  in year  $i$  for a given sex is constant over the projection (Model 0). In Model 1 the annual observed population mean weight-at-age (time-varying) is used in the stock assessment model. Model 2 is a fit to the data used in Model 1 by the estimation of year and age specific parameters and Model 3 estimates annual weight-at-age as a function of annual May sea surface temperature anomalies. The growth models are as follows:

Model			
0	$\hat{w}_{ij} = \hat{w}_{i-1,j-1} + g_j$	$i > 1954, j > 1$	Constant fixed growth
1	$\hat{w}_{ij} = w_{ij},$ $\hat{w}_{ij} = \bar{w}_{\cdot j}$	$1982 \leq i \leq 2009,$ $i < 1982, i > 2009$	As estimated for each age from survey data
2	$\hat{w}_{ij} = \hat{w}_{i-1,j-1} + g_j e^{\varepsilon_{ij}}$	$i > 1954, j > 1$ $\varepsilon_i \sim N(0, \sigma_g^2)$	Year-effect freely estimated on growth increment
3	$\hat{w}_{ij} = \hat{w}_{i-1,j-1} + g_j e^{\varepsilon_{ij}}$ $\varepsilon_{ij} = T_i \alpha + \delta_i$	$i > 1954, j > 1$ $\delta_i \sim N(0, \sigma_{res}^2)$	Year-effect on growth increment linked to temperature conditions

where  $w_{ij}$  represents the observed estimates of mean weights at age and year,  $g_j$  is the expected growth increment in the most recent completed year (as estimated from the a sex-specific von-Bertalanfy growth curve) and  $\varepsilon_i$  is a process error term which is modeled as to have an optional year-effect and separate age effect in model 2. For model 3 temperature anomalies are introduced for the entire period and the parameter  $\alpha$  scales them and the residual variance is computed  $\sigma_g^2$ .

For all models except 1, the negative log-likelihood function for the weight-at-age data applied was:

$$-\ln L_w = \sum_{i=1982}^{2009} \sum_{j=5}^{15} \frac{n_{ij} (\ln w_{ij} - \ln \hat{w}_{ij})^2}{2\sigma_{ij}^2}$$

### Maturity-at-age

Maturity information collected from yellowfin sole females during the 1992 and 1993 eastern Bering Sea trawl surveys is used in this assessment (Table 4.10). Nichol (1995) estimated the age of 50% maturity at 10.5 years based on the histological examination of 639 ovaries. In the case of most north Pacific flatfish species, including yellowfin sole, sexual maturity occurs well after the age of entry into the fishery.

Yellowfin sole are 90% selected to the fishery by age 11 whereas females have been found to be 61% mature at this age.

## Analytic Approach

### **Model Structure**

The abundance, mortality, recruitment and selectivity of yellowfin sole were assessed with a stock assessment model using the AD Model Builder language (Ianelli and Fournier 1998). The conceptual model is a separable catch-age analysis that uses survey estimates of biomass and age composition as auxiliary information (Fournier and Archibald 1982). The assessment model simulates the dynamics of the population and compares the expected values of the population characteristics to the characteristics observed from surveys and fishery sampling programs. This is accomplished by the simultaneous estimation of the parameters in the model using the maximum likelihood estimation procedure. The fit of the simulated values to the observable characteristics is optimized by maximizing a log(likelihood) function given some distributional assumptions about the observed data.

Since the sex-specific weight-at-age for yellowfin sole diverges after age of maturity (about age 10 for 50% of the stock) with females growing larger than males, the current assessment model is coded to accommodate the sex-specific aspects of the population dynamics of yellowfin sole. The model allows for the input of sex-specific estimates of fishery and survey age composition and weight-at-age and provides sex-specific estimates of population numbers, fishing mortality, selectivity, fishery and survey age composition and allows for the estimation of sex-specific natural mortality and catchability. The model retains the utility to fit combined sex data inputs.

The suite of parameters estimated by the model are classified by three likelihood components:

<b>Data component</b>	<b>Distributional assumption</b>
Trawl fishery catch-at-age	Multinomial
Trawl survey population age composition	Multinomial
Trawl survey biomass estimates and S.E.	Log normal

The total log likelihood is the sum of the likelihoods for each data component (Table 4.11). The likelihood components may be weighted by an emphasis factor, however, equal emphasis was placed on fitting each likelihood component in the yellowfin sole assessment except for the catch. The AD Model Builder software fits the data components using automatic differentiation (Griewank and Corliss 1991) software developed as a set of libraries (AUTODIFF C++ library). Table 4.11 also presents the key equations used to model the yellowfin sole population dynamics in the Bering Sea and Table 4.12 provides a description of the variables used in Table 4.11.

Sharp increases in trawl survey abundance estimates for most species of Bering Sea flatfish between 1981 and 1982 indicate that the 83-112 trawl was more efficient for capturing these species than the 400-mesh eastern trawl used in 1975, and 1979-81. Allowing the model to tune to these early survey estimates would most likely underestimate the true pre-1982 biomass, thus exaggerating the degree to which biomass increased during that period. Although this underestimate would have little effect on the estimate of current yellowfin sole biomass, it would affect the spawner and recruitment estimates for the time-series. Hence, the pre-1982 survey biomass estimates were omitted from the analysis.

The model of yellowfin sole population dynamics was evaluated with respect to the observations of the time-series of survey and fishery age compositions and the survey biomass trend since 1982.

### **Parameters Estimated Outside the Assessment Model**

Natural mortality (M) was initially estimated by a least squares analysis where catch-at-age data were fitted to Japanese pair trawl effort data while varying the catchability coefficient (q) and M

simultaneously. The best fit to the data (the point where the residual variance was minimized) occurred at a M value of 0.12 (Bakkala and Wespestad 1984). This was also the value which provided the best fit to the observable population characteristics when M was profiled over a range of values in the stock assessment model using data up to 1992 (Wilderbuer 1992). Since then, natural mortality has been estimated as a free parameter in some of the stock assessment model runs which have been evaluated for the past five years. A natural mortality value of 0.12 is used for both sexes in the base model presented in this assessment.

Yellowfin sole maturity schedules were estimated from in-situ observations as discussed in a previous section (Table 4.10).

## Parameters Estimated Inside the Assessment Model

The parameters estimated by the model are presented below:

Fishing mortality	Selectivity	Survey catchability	Year class strength	Spawner-recruit	Total
61	248	2	100	2	413

The increase in the number of parameters estimated in this assessment compared to last year (4) can be accounted for by the input of another year of fishery data and the entry of another year class into the observed population and two more sex-specific fishery selectivity parameters.

### *Year class strengths*

The population simulation specifies the numbers-at-age in the beginning year of the simulation, the number of recruits in each subsequent year, and the survival rate for each cohort as it moves through the population over time using the population dynamics equations given in Table 4.11.

### *Selectivity*

Fishery and survey selectivity was modeled separately for males and females using the two parameter formulation of the logistic function (Table 4.11). The model was run with an asymptotic selectivity curve for the older fish in the fishery and survey, but still was allowed to estimate the shape of the logistic curve for young fish. The oldest year classes in the surveys and fisheries were truncated at 20 and allowed to accumulate into the age category 20+ years. A single selectivity curve, for both males and females, was fit for all years of survey data.

Given that there have been annual changes in management, vessel participation and most likely gear selectivity, time-varying fishing selectivity curves were estimated. A logistic equation was used to model fishery selectivity and is a function of time-varying parameters specifying the age and slope at 50% selection,  $\varphi_t$  and  $\eta_t$ , respectively. The fishing selectivity ( $S^f$ ) for age  $a$  and year  $t$  is modeled as,

$$S_{a,t}^f = \left[ 1 + e^{\eta_t(a - \varphi_t)} \right]^{-1}$$

where  $\eta_t$  and  $\varphi_t$  are time-varying and partitioned (for estimation) into parameters representing the mean and a vector of deviations (log-scale) conditioned to sum to zero. The deviations are constrained by a lognormal prior with a variance that was iteratively estimated. The process of iterating was to first set the variance to a high value (diffuse prior) of 0.5<sup>2</sup> and estimate the deviations. The next step was to compare the variability of model estimates. The variability of the model estimates were then rounded up slightly and fixed for subsequent runs. The 2013 values were fixed as the average of the 3 most recent years.

### *Fishing Mortality*

The fishing mortality rates ( $F$ ) for each age and year are calculated to approximate the catch weight by solving for  $F$  while still allowing for observation error in catch measurement. A large emphasis (300) was placed on the catch likelihood component to force the model to match the observed catch.

### *Survey Catchability*

A past assessment (Wilderbuer and Nichol 2001) first examined the relationship between estimates of survey biomass and bottom water temperature. To better understand how water temperature may affect the catchability of yellowfin sole to the survey trawl, catchability was estimated for each year in the stock assessment model as:

$$q = e^{-\alpha + \beta T}$$

where  $q$  is catchability,  $T$  is the average annual bottom water temperature anomaly at survey stations less than 100 m, and  $\alpha$  and  $\beta$  are parameters estimated by the model. The catchability equation has two parts. The  $e^{-\alpha}$  term is a constant or time-independent estimate of  $q$ . The model estimate of  $\alpha = -0.132$  indicates that  $q > 1$  suggesting that yellowfin sole are herded into the trawl path of the net which is consistent with the experimental results for other flatfish species. The second term,  $e^{\beta T}$  is a time-varying (annual)  $q$  which responds to metabolic aspects of herding or distribution (availability) which can vary annually with bottom water temperature. The result of incorporating bottom temperature to estimate annual  $q$  is shown in Figure 4.11 (for the base model).

### *Spawner-Recruit Estimation*

Annual recruitment estimates were constrained to fit a Ricker (1958) form of the stock recruitment relationship as follows:

$$R = \alpha S e^{-\beta S}$$

where  $R$  is age 1 recruitment,  $S$  is female spawning biomass ( $t$ ) the previous year, and  $\alpha$  and  $\beta$  are parameters estimated by the model. The spawner-recruit fitting is estimated in a later phase after initial estimates of survival, numbers-at-age and selectivity are obtained.

## **Results**

### **Model Evaluation**

The model evaluation for this stock assessment involved a three-step process. The first step was to evaluate the productivity of the yellowfin sole stock by an examination of which sets of years to include for spawner-recruit fitting. The second step then evaluated the growth models presented in a previous section and the third step evaluated various hypothesized states of nature by fitting natural mortality and catchability estimates in various combinations.

The SSC determined in December 2006 that yellowfin sole would be managed under the Tier 1 harvest guidelines, and therefore future harvest recommendations would be based on MSY and  $F_{MSY}$  values calculated from a spawner-recruit relationship. MSY is an equilibrium concept and its value is dependent on both the spawner-recruit estimates which are assumed to represent the equilibrium stock size-recruitment relationship and the model used to fit the estimates. In the yellowfin sole stock assessment model, a Ricker form of the stock-recruit relationship was fit to various combinations of these data and estimates of  $F_{MSY}$  and  $B_{MSY}$  were calculated, assuming that the fit to the stock-recruitment data represents the long-term productivity of the stock.

For this assessment, 2 different stock-recruitment time-series were investigated: the full time-series 1955-2006 (Model A) and the post-regime shift era, 1978-2006 (Model B) (Fig. 4.12) (see Joint Plan Team

recommendations for September 2012). Very different estimates of the long-term sustainability of the stock ( $F_{MSY}$  and  $B_{MSY}$ ) are obtained, depending on which years of stock-recruitment data were included in the fitting procedure (Table 4.13). When the entire time-series from 1955-2006 was fit, the large recruitments that occurred at low spawning stock sizes in the 1960s and early 1970s determined that the yellowfin sole stock was most productive at a smaller stock size with the result that  $F_{MSY}$  is nearly equal to  $F_{35\%}$  ( $F_{35\%} = 0.133$ ) and  $B_{MSY}$  is 318,000 (Model A). If we limit the analysis to consider only recruitments which occurred after the well-documented regime shift in 1977, a lower value of  $F_{MSY}$  is obtained (0.118) and  $B_{MSY}$  is 366,000 t. Table 4.13 indicates that the ABC values from the Model A harvest scenario for 2014 would be 44,000 t higher than Model B. Posterior distributions of  $F_{MSY}$  for these models indicate that this parameter is estimated with less uncertainty for Model A resulting in the reduced buffer between ABC and OFL relative to Model B (8% for Model B vs. 1% for Model A, Table 4.13 and Fig 4.13).

It is important for the Tier 1 calculations to identify which subset of the stock recruitment data is used. Using the full time series to fit the spawner recruit curve estimates that the stock is most productive at a small stock size. Thus MSY and  $F_{MSY}$  are relatively high values and  $B_{MSY}$  is a lower value. If the stock was productive in the past at a small stock size because of non density dependent factors (environment), then reducing the stock size to low levels could be detrimental to the long-term sustainability of the stock if the environment, and thus productivity, had changed from the earlier period. Since observations of yellowfin sole recruitment at low stock sizes are not available from multiple time periods, it is uncertain if future recruitment events at low stock conditions would be as productive as during the late 1960s-early 1970s.

Given the uncertainty of the productivity of yellowfin sole at low spawning stock sizes, and because the AFSC policy for reference point time-series selection is to use the post 1977 regime shift values unless there is a compelling reason to do otherwise, the productivity of yellowfin sole in this assessment is estimated by fitting the 1977-2006 spawner-recruit data in the model.

The second step in the model evaluation is the evaluation of the growth model for yellowfin sole. Estimates of female spawning biomass from the four growth models (described in a previous section) are shown in Figure 4.15 and indicate that all 4 models estimate the same trend with Model 1 estimates trending lower in the last six years than the others. Estimates of 2012 ABC and  $F_{ABC}$  are also similar with Model 1 values lower than the others.

	model 0	model 1	model 2	model 3
2013 FSB	742,000	581,100	652,300	Did
2014 $F_{ABC}$	0.107	0.113	0.111	Not
2014 ABC	257,600	239,800	262,300	Converge

Growth Model 1 was selected as the model of choice for this assessment since 1) It does not use time invariant growth as in Model 0 (which is not realistic and unsupported by the growth data) but instead relies on the annually collected survey population and age data to calculate annual estimates of length at age and weight-at-age, and 2) in its present formulation, Model 2 does not fit the FSB trend of model 1 in the most recent years and thus warrants more exploration before incorporation into the assessment model. Model 3 did not converge with the addition of the 2013 data and cannot be evaluated until a determination is made of what caused the convergence problem. Both Models 2 and 3 could also further benefit from using the estimated population as a covariate to model the annual growth increment due to density dependent effects. The model modifications have not been pursued for this assessment.

The third step in the model evaluation for this assessment entails the use of a single structural model to consider the uncertainty in the key parameters M and catchability. This is the Model which has been the model of choice in the past 6 assessments and operates by fixing M at 0.12 for both sexes and then

estimates q using the relationship between survey catchability and the annual average water temperature at the sea floor (from survey stations at less than 100 m). The other models used in the evaluation represented various combinations of estimating M or q as free parameters with different amounts of uncertainty in the parameter estimates (Wilderbuer et al. 2010). The results are detailed in those assessments and are not repeated here except for the following observations.

Modeling survey catchability as a nonlinear function of bottom water temperature returns a mean value of 1.104. This value is consistent with supporting evidence from experiments examining the bridle efficiency of the Bering Sea survey trawl which indicate that yellowfin sole are herded into the trawl path from an area between the wing tips of the net and the point where the bridles contact the seafloor (Somerton and Munro 2001). It is also consistent with our hypothesis of the timing of the survey relative to the temperature dependent timing of the annual spawning migration to nearshore areas which are outside of the survey area. The herding experiments suggest that the survey trawl catchability is greater than 1.0. The likelihood profile of q from the model indicated a small variance with a narrow range of likely values with a low probability of q being equal to the value of 1.0 in a past assessment (Wilderbuer and Nichol 2003).

A model that allows M to be estimated as a free parameter for males with females fixed at 0.12 provided a better fit to the sex ratio estimated from the annual trawl survey age compositions than did the base model (both sexes fixed at M = 0.12). However, since the population sex ratio annually observed at the time of the survey is a function of the timing of the annual spawning in adjacent inshore areas, it is questionable that providing the best fit to these observations is really fitting the population sex ratio better. Thus, the model configuration which utilizes the relationship between annual seafloor temperature and survey catchability with M fixed at 0.12 for both sexes is used to base the assessment of the condition of the Bering Sea yellowfin sole resource for the 2013 fishing season.

### **Time Series Results**

The 2013 trawl survey point estimate increased 17% from 2012. This resulted in higher model estimates of population numbers at age and biomass for the time-series back to the mid-1960s relative to last year's assessment. In addition, a large 2003 year class (10 years old in 2013) is present in the population at their cohort maximum. The model results indicate the stock has been in a slowly declining condition since the mid-1980s but is now in an increasing trend due to the strong 2003 year class. The estimates of female spawning biomass, total biomass and ABC are higher than those used to manage the stock in 2012. Seven of the past 8 years have had negative bottom temperature anomalies in the Bering Sea. The temperature-dependent q adjustment for 2013 was 1.104 indicating minimal herding effects in cooler water conditions.

#### *Fishing Mortality and Selectivity*

The assessment model estimates of the annual fishing mortality in terms of age-specific annual F and on fully selected ages are given in Tables 4.14 and 4.15, respectively. The full-selection F has averaged 0.08 over the period of 1978-2012 with a maximum of 0.12 in 1978 and a minimum in 2001 at 0.044. Selectivities estimated by the model (Table 4.16, Fig. 4.14) indicate that both sexes of yellowfin sole are 50% selected by the fishery at about age 9 and nearly fully selected by age 13, with annual variability.

#### *Abundance Trend*

The model estimates q at an average value of 1.104 for the period 1982-2013 which results in the model estimate of the 2013 age 2+ total biomass at 2,117,500 t (Table 4.17). Model results indicate that yellowfin sole total biomass (age 2+) was at low levels during most of the 1960s and early 1970s (700,000-800,000 t) after a period of high exploitation (Table 4.17, Fig. 4.16, bottom left panel). Sustained above average recruitment from 1967-76 combined with light exploitation resulted in a biomass increase to a peak of 2.9 million t by 1984. The population biomass has since been in a slow decline as the strong 1981 and 1983 year-classes have passed through the population with only the 1991 and 1995

year classes at levels observed during the 1970s. However, the total biomass is now increasing again due to the strong 2003 cohort. The present biomass is estimated at 78% of the peak 1984 level.

The female spawning biomass has also declined since the peak in 1994, with a 2013 estimate of 576,650 t (36% decline). The spawning biomass has been in a gradual decline for the past 19 years and is at the  $B_{40\%}$  level and 1.5 times the  $B_{MSY}$  level (Fig. 4.16). The model estimate of yellowfin sole population numbers at age for all years is shown in Table 4.18 and the resulting fit to the observed fishery and survey age compositions input into the model are shown in the Figure 4.17. The fit to the trawl survey biomass estimates are shown in Figure 4.16. Allowing  $q$  to be correlated with annual bottom temperature provides a better fit to the bottom trawl survey estimates (Fig. 4.18). Table 4.19 lists the numbers of female spawners estimated by the model for all ages and years. The estimated average age of yellowfin sole in the population is 6.1 years for males and females.

Both the trawl survey and the stock assessment model indicate that the yellowfin sole resource increased during the 1970s and early 1980s to a peak level during the mid-1980s. The yellowfin sole population biomass slowly decreased over the nineteen years since the mid-1990s as the majority of year-classes during those years were below average strength. Above-average recruitment from the strong 2003 year-class is expected to maintain the abundance of yellowfin sole at a level above  $B_{MSY}$  in the near future. The stock assessment projection model indicates an increasing trend in female spawning biomass through 2019 if the fishing mortality rate continues at the same level as the average of the past 5 years.

#### *Recruitment Trends*

The primary reason for the sustained increase in abundance of yellowfin sole during the 1970s and early 1980s was the recruitment of a series of stronger than average year classes spawned in 1967-76 (Figure 4.19 and Table 4.20). The 1981 year class was the strongest observed (and estimated) during the 47 year period analyzed and the 1983 year class was also very strong. Survey age composition estimates and the assessment model also estimate that the 1987 and 1988 year classes were average and the 1991 and 1995 year classes were above average. With the exception of these 4 year classes, recruitment from 14 of the last 18 years estimated (since the strong 1983 year-class) has been below the 48 year average, which has caused the population to gradually decline. The 1995 year-class were at the maximum of their cohort biomass in 2005 and still contribute to the mature adult reservoir of spawners. The 2003 year-class has now been observed multiple times in the age compositions and are clearly a strong year class similar to some of the hallmark recruitment mentioned above and will contribute to the reservoir of spawning fish in the coming years.

#### *Historical Exploitation Rates*

Based on results from the stock assessment model, annual average exploitation rates of yellowfin sole ranged from 3 to 8% of the total biomass since 1977, and have averaged 5% (Table 4.14). Posterior distributions of selected parameters from the preferred stock assessment model used in the assessment are shown in Figure 4.20. The values and standard deviations of some selected model parameters are listed in Table 4.21.

A within-model retrospective analysis is also included for the recommended assessment model where retrospective female spawning biomass is calculated by working backwards in time dropping data one year at a time (Fig. 4.21).

#### **Harvest Recommendations**

After decreasing since the peak value in 1984, estimates from the stock assessment model indicate the total biomass is now increasing. The estimate of age 6+ total biomass for 2014 is 2,113,000 t.

The SSC has determined that yellowfin sole qualify as a Tier 1 stock and therefore the 2014 ABC is calculated using Tier 1 methodology. In 2006 the SSC selected the 1978-2001 data set for the Tier 1 harvest recommendation. Using this approach again for the 2014 harvest (now the 1978-2006 time-series) recommendation (Model B in Table 4.13 with growth option 1), the  $F_{ABC} = F_{harmonic\ mean} = 0.113$ .

The Tier 1 harvest level is calculated as the product of the harmonic mean of  $F_{MSY}$  and the geometric mean of the 2014 biomass estimate, as follows:

$$B_{gm} = e^{\frac{\ln \hat{B} - \frac{cv^2}{2}}{2}}, \text{ where } B_{gm} \text{ is the geometric mean of the 2014 biomass estimate, } \hat{B} \text{ is the point estimate of the 2014 biomass from the stock assessment model and } cv^2 \text{ is the coefficient of variation of the point estimate (a proxy for sigma);}$$

and

$$\bar{F}_{har} = e^{\frac{\ln \hat{F}_{msy} - \frac{\ln sd^2}{2}}{2}}, \text{ where } \bar{F}_{har} \text{ is the harmonic mean, } \hat{F}_{msy} \text{ is the peak mode of the } F_{MSY} \text{ distribution and } sd^2 \text{ is the square of the standard deviation of the } F_{MSY} \text{ distribution. This calculation gives a Tier 1 ABC harvest recommendation of } 239,800 \text{ t and an OFL of } 259,700 \text{ t for 2014. This gives an 8\% (19,800 t) buffer between ABC and OFL, 2\% more than the same buffer calculated for 2013 in last year's assessment.}$$

The stock assessment analysis must also consider harvest limits, usually described as overfishing fishing mortality levels with corresponding yield amounts. Amendment 56 to the BSAI FMP sets the Tier 1 harvest limit at the  $F_{MSY}$  fishing mortality value. The overfishing fishing mortality values, ABC fishing mortality values and their corresponding yields are given as follows:

<u>Harvest level</u>	<u>F value</u>	<u>2014 Yield</u>
<b>Tier 1 <math>F_{OFL} = F_{MSY}</math></b>	<b>0.123</b>	<b>259,700 t</b>
<b>Tier 1 <math>F_{ABC} = F_{harmonic\ mean}</math></b>	<b>0.113</b>	<b>239,800 t</b>

#### *Status Determination*

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with the vector of 2013 numbers at age estimated in the assessment. This vector is then projected forward to the beginning of 2014 using the schedules of natural mortality and selectivity described in the assessment and the best available estimate of total (year-end) catch for 2013. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TAC for 2014, are as follows (“ $\max F_{ABC}$ ” refers to the maximum permissible value of  $F_{ABC}$  under Amendment 56):

*Scenario 1:* In all future years,  $F$  is set equal to  $\max F_{ABC}$ . (Rationale: Historically, TAC has been constrained by ABC, so this scenario provides a likely upper limit on future TACs.)

*Scenario 2:* In all future years,  $F$  is set equal to a constant fraction of  $\max F_{ABC}$ , where this fraction is equal to the ratio of the  $F_{ABC}$  value for 2014 recommended in the assessment to the  $\max F_{ABC}$  for 2014. (Rationale: When  $F_{ABC}$  is set at a value below  $\max F_{ABC}$ , it is often set at the value recommended in the stock assessment.)

*Scenario 3:* In all future years,  $F$  is set equal to 50% of  $\max F_{ABC}$ . (Rationale: This scenario provides a likely lower bound on  $F_{ABC}$  that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

*Scenario 4:* In all future years,  $F$  is set equal to the 2009-2013 average  $F$ . (Rationale: For some stocks, TAC can be well below ABC, and recent average  $F$  may provide a better indicator of  $F_{TAC}$  than  $F_{ABC}$ .)

*Scenario 5:* In all future years,  $F$  is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as  $B_{35\%}$ ):

*Scenario 6:* In all future years,  $F$  is set equal to  $F_{OFL}$ . (Rationale: This scenario determines whether a stock is overfished. If the stock is expected to be above its MSY level in 2012 and above its MSY level in 2025 under this scenario, then the stock is not overfished.)

*Scenario 7:* In 2014 and 2015,  $F$  is set equal to  $\max F_{ABC}$ , and in all subsequent years,  $F$  is set equal to  $F_{OFL}$ . (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2026 under this scenario, then the stock is not approaching an overfished condition.)

Simulation results shown in Table 4.22 indicate that yellowfin sole are not currently overfished and are not approaching an overfished condition. The projection of yellowfin sole female spawning biomass through 2024 is shown in Figure 4.22 and a phase plane figure of the estimated time-series of yellowfin sole female spawning biomass relative to the harvest control rule is shown in Figure 4.23.

#### *Scenario Projections and Two-Year Ahead Overfishing Level*

In addition to the seven standard harvest scenarios, Amendments 48/48 to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. The 2013 numbers at age from the stock assessment model are projected to 2014 given the 2013 catch and then a 2014 catch of 150,000 t is applied to the projected 2014 population biomass to obtain the 2015 OFL.

Tier 1 Projection		SSB	Geometric mean 6+ total biomass		
Year	Catch			ABC	OFL
2014	150,000	581,100	2,113,500	239,800	259,700
2015		594,800	2,188,100	243,300	268,900

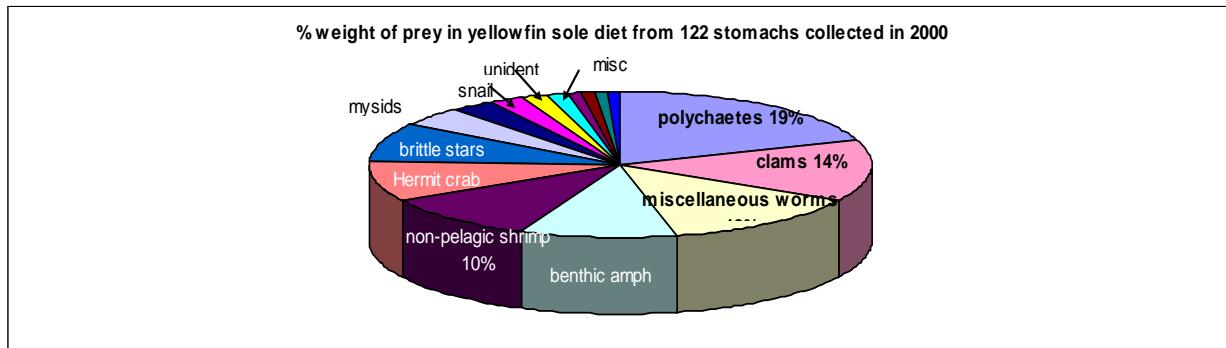
## Ecosystem Considerations

### Ecosystem Effects on the stock

#### 1) Prey availability/abundance trends

Yellowfin sole diet by life stage varies as follows: Larvae consume plankton and algae, early juveniles consume zooplankton, late juvenile stage and adults prey includes bivalves, polychaetes, amphipods,

mollusks, euphausiids, shrimps, brittle stars, sculpins and miscellaneous crustaceans. Information is not available to assess the abundance trends of the benthic infauna of the Bering Sea shelf. The original description of infaunal distribution and abundance by Haflinger (1981) resulted from sampling conducted in 1975 and 1976 and has not been re-sampled since. The large populations of flatfish which have occupied the middle shelf of the Bering Sea over the past twenty-five years for summertime feeding do not appear food-limited. These populations have fluctuated due to the variability in recruitment success which suggests that the primary infaunal food source has been at an adequate level to sustain the yellowfin sole resource.



### *2) Predator population trends*

As juveniles, it is well-documented from studies in other parts of the world that flatfish are prey for shrimp species in near shore areas. This has not been reported for Bering Sea yellowfin sole due to a lack of juvenile sampling and collections in near shore areas, but is thought to occur. As late juveniles they have been found in stomachs of Pacific cod and Pacific halibut; mostly on small yellowfin sole ranging from 7 to 25 cm standard length..

Past, present and projected future population trends of these predator species can be found in their respective SAFE chapters in this volume and also from Annual reports compiled by the International Pacific Halibut Commission. Encounters between yellowfin sole and their predators may be limited since their distributions do not completely overlap in space and time.

### *3) Changes in habitat quality*

Changes in the physical environment which may affect yellowfin sole distribution patterns, recruitment success and migration timing patterns are catalogued in the Ecosystem Considerations Appendix of this SAFE report. Habitat quality may be enhanced during years of favorable cross-shelf advection (juvenile survival) and warmer bottom water temperatures with reduced ice cover (higher metabolism with more active feeding).

## **Fishery Effects on the ecosystem**

- 1) The yellowfin sole target fishery contribution to the total bycatch of other target species is shown for 1992-2012 in Table 4.23. The catch of non-target species from 2003-2012 is shown in Table 4.24. The yellowfin sole target fishery contribution to the total bycatch of prohibited species is shown for 2010 and 2011 in Table 13 of the Economic SAFE (Appendix C) and is summarized for 2011 as follows:

Prohibited species	Yellowfin sole fishery % of total bycatch
Halibut mortality	28.5
Herring	5.0
Red King crab	14.1
<u>C. bairdi</u>	62.9
Other Tanner crab	70.6
Salmon	<1

- 2) Relative to the predator needs in space and time, the yellowfin sole target fishery has a low selectivity for fish 7-25 cm and therefore has minimal overlap with removals from predation.
- 3) The target fishery is not perceived to have an effect on the amount of large size target fish in the population due to its history of light exploitation (6%) over the past 30 years.
- 4) Yellowfin sole fishery discards are presented in the Catch History section.
- 5) It is unknown what effect the fishery has had on yellowfin sole maturity-at-age and fecundity.
- 6) Analysis of the benthic disturbance from the yellowfin sole fishery is available in the Preliminary draft of the Essential Fish Habitat Environmental Impact Statement.

<b>Ecosystem effects on yellowfin sole</b>			
Indicator	Observation	Interpretation	Evaluation
<i>Prey availability or abundance trends</i>			
Benthic infauna	Stomach contents	Stable, data limited	Unknown
<i>Predator population trends</i>			
Fish (Pacific cod, halibut, skates)	Stable	Possible increases to yellowfin sole mortality	
<i>Changes in habitat quality</i>			
Temperature regime	Cold years yellowfin sole catchability and herding may decrease, timing of migration may be prolonged	Likely to affect surveyed stock	No concern (dealt with in model)
Winter-spring environmental conditions	Affects pre-recruit survival	Probably a number of factors	Causes natural variability
<b>Yellowfin sole effects on ecosystem</b>			
Indicator	Observation	Interpretation	Evaluation
<i>Fishery contribution to bycatch</i>			
Prohibited species	Stable, heavily monitored	Minor contribution to mortality	No concern
Forage (including herring, Atka mackerel, cod, and pollock)	Stable, heavily monitored	Bycatch levels small relative to forage biomass	No concern
HAPC biota	Low bycatch levels of (spp)	Bycatch levels small relative to HAPC biota	No concern
Marine mammals and birds	Very minor direct-take	Safe	No concern
Sensitive non-target species	Likely minor impact	Data limited, likely to be safe	No concern
<i>Fishery concentration in space and time</i>	Low exploitation rate	No concern Little detrimental effect	
<i>Fishery effects on amount of large size target fish</i>			
	Low exploitation rate	Natural fluctuation	No concern
<i>Fishery contribution to discards and offal production</i>			
	Stable trend	Improving, but data limited	Possible concern
<i>Fishery effects on age-at-maturity and fecundity</i>			
	Unknown	NA	Possible concern

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## Tables

Table 4.1--Catch (t) of yellowfin sole 1964-2013. Catch for 2013 is an estimate through the end of 2013.

Year	Foreign	Domestic		Total
		JVP	DAP	
1964	111,777			111,777
1965	53,810			53,810
1966	102,353			102,353
1967	162,228			162,228
1968	84,189			84,189
1969	167,134			167,134
1970	133,079			133,079
1971	160,399			160,399
1972	47,856			47,856
1973	78,240			78,240
1974	42,235			42,235
1975	64,690			64,690
1976	56,221			56,221
1977	58,373			58,373
1978	138,433			138,433
1979	99,019			99,019
1980	77,768	9,623		87,391
1981	81,255	16,046		97,301
1982	78,331	17,381		95,712
1983	85,874	22,511		108,385
1984	126,762	32,764		159,526
1985	100,706	126,401		227,107
1986	57,197	151,400		208,597
1987	1,811	179,613	4	181,428
1988		213,323	9,833	223,156
1989		151,501	1,664	153,165
1990		69,677	14,293	83,970
1991			115,842	115,842
1992			149,569	149,569
1993			106,101	106,101
1994			144,544	144,544
1995			124,740	124,740
1996			129,659	129,659
1997			181,389	181,389
1998			101,201	101,201
1999			67,320	67,320
2000			83,850	83,850
2001			63,395	63,395
2002			73,000	73,000
2003			74,418	74,418
2004			69,046	69,046
2005			94,383	94,383
2006			99,068	99,068
2007			121,029	121,029
2008			148,894	148,894
2009			107,528	107,528
2010			118,624	118,624
2011			151,164	151,164
2012			147,183	147,183
2013			153,000	153,000

Table 4.2 Estimates of retained and discarded (t) yellowfin sole  
 caught in Bering Sea fisheries.

<b>Year</b>	<b>Retained</b>	<b>Discarded</b>
1987	3	1
1988	7,559	2,274
1989	1,279	385
1990	10,093	4,200
1991	89,054	26,788
1992	103,989	45,580
1993	76,798	26,838
1994	107,629	36,948
1995	96,718	28,022
1996	101,324	28,334
1997	149,570	31,818
1998	80,365	20,836
1999	55,202	12,118
2000	69,788	14,062
2001	54,759	8,635
2002	62,050	10,950
2003	63,732	10,686
2004	57,378	11,668
2005	85,321	9,062
2006	90,570	8,498
2007	109,084	11,945
2008	141,253	7,659
2009	92,488	5,733
2010	113,244	5,380
2011	146,419	4,745
2012	143,737	3,446

Table 4.3. Discarded and retained catch of non-CDQ yellowfin sole, by fishery, in 2012.

Source: AKFIN.

<b>Trip Target Name</b>	<b>Discarded</b>	<b>Retained</b>
Atka Mackerel	0	0
Pollock - bottom	52	831
Pacific Cod	1,057	740
Alaska Plaice	15	588
Other Flatfish - BSAI	0	
Halibut		
Rockfish	1	0
Flathead Sole	2	50
Kamchatka flounder	1	
Pollock - midwater	439	131
Rock Sole - BSAI	356	9,201
Sablefish - BSAI		
Greenland Turbot - BSAI		
Arrowtooth Flounder	1	1
Yellowfin Sole - BSAI	3,131	130,588



<b>1981</b>	0.10	0.11	0.06	0.09	0.16	0.18	0.12	0.08	0.03	0.02	0.01
<b>1982</b>	0.05	0.14	0.11	0.09	0.18	0.17	0.10	0.05	0.03	0.01	0.00
<b>1983</b>	0.12	0.08	0.14	0.08	0.09	0.13	0.12	0.09	0.05	0.03	0.02
<b>1984</b>	0.06	0.08	0.10	0.19	0.08	0.13	0.10	0.12	0.05	0.02	0.03
<b>1985</b>	0.03	0.05	0.12	0.11	0.14	0.12	0.11	0.13	0.09	0.03	0.02
<b>1986</b>	0.06	0.06	0.07	0.17	0.13	0.11	0.08	0.06	0.07	0.05	0.12
<b>1987</b>	0.03	0.06	0.04	0.08	0.11	0.12	0.13	0.10	0.06	0.06	0.18
<b>1988</b>	0.03	0.05	0.13	0.04	0.07	0.10	0.07	0.12	0.10	0.04	0.23
<b>1989</b>	0.01	0.07	0.09	0.09	0.06	0.06	0.10	0.05	0.09	0.09	0.27
<b>1990</b>	0.05	0.02	0.26	0.07	0.11	0.07	0.03	0.03	0.02	0.17	0.18
<b>1991</b>	0.03	0.14	0.02	0.23	0.09	0.08	0.02	0.07	0.09	0.05	0.18
<b>1992</b>	0.01	0.04	0.16	0.04	0.26	0.08	0.07	0.02	0.05	0.03	0.21
<b>1993</b>	0.03	0.02	0.03	0.16	0.01	0.20	0.05	0.06	0.04	0.05	0.31
<b>1994</b>	0.04	0.05	0.05	0.06	0.27	0.00	0.17	0.02	0.07	0.04	0.22
<b>1995</b>	0.07	0.10	0.04	0.03	0.03	0.18	0.00	0.27	0.02	0.05	0.21
<b>1996</b>	0.03	0.07	0.12	0.04	0.05	0.06	0.14	0.02	0.20	0.02	0.20
<b>1997</b>	0.04	0.05	0.12	0.06	0.03	0.06	0.06	0.19	0.02	0.12	0.20
<b>1998</b>	0.04	0.05	0.12	0.06	0.03	0.06	0.06	0.19	0.02	0.12	0.20
<b>1999</b>	0.02	0.03	0.04	0.04	0.14	0.08	0.04	0.06	0.07	0.10	0.37
<b>2000</b>	0.01	0.02	0.09	0.05	0.04	0.11	0.13	0.08	0.03	0.06	0.36
<b>2001</b>	0.02	0.04	0.08	0.13	0.08	0.07	0.12	0.08	0.06	0.03	0.26
<b>2002</b>	0.02	0.04	0.06	0.07	0.11	0.08	0.06	0.08	0.08	0.04	0.38
<b>2003</b>	0.01	0.08	0.08	0.05	0.08	0.15	0.06	0.04	0.03	0.05	0.34
<b>2004</b>	0.03	0.02	0.18	0.08	0.06	0.04	0.09	0.04	0.02	0.08	0.34
<b>2005</b>	0.03	0.05	0.05	0.14	0.05	0.06	0.06	0.10	0.05	0.01	0.34
<b>2006</b>	0.12	0.10	0.06	0.07	0.25	0.04	0.01	0.02	0.06	0.02	0.18
<b>2007</b>	0.02	0.08	0.06	0.04	0.08	0.15	0.06	0.06	0.05	0.06	0.31
<b>2008</b>	0.03	0.08	0.08	0.09	0.05	0.05	0.12	0.07	0.04	0.05	0.31
<b>2009</b>	0.04	0.06	0.11	0.11	0.06	0.09	0.06	0.08	0.05	0.04	0.29
<b>2010</b>	0.06	0.05	0.08	0.07	0.07	0.06	0.07	0.07	0.10	0.05	0.30
<b>2011</b>	0.04	0.09	0.07	0.09	0.07	0.11	0.06	0.05	0.05	0.07	0.29
<b>2012</b>	0.05	0.07	0.10	0.09	0.10	0.03	0.10	0.04	0.02	0.05	0.33

Table 4.5—Yellowfin sole biomass estimates (t) from the annual Bering Sea shelf bottom trawl survey and upper and lower 95% confidence intervals.

Year	Total	Lower CI	Upper CI
1975	972,500	812,300	1,132,700
1979	1,866,500	1,586,000	2,147,100
1980	1,842,400	1,553,200	2,131,700
1981	2,394,700	2,072,900	2,716,500
1982	3,377,800	2,571,000	4,184,600
1983	3,535,300	2,958,100	4,112,400
1984	3,141,200	2,636,800	3,645,600
1985	2,443,700	1,563,400	3,324,000
1986	1,909,900	1,480,700	2,339,000
1987	2,613,100	2,051,800	3,174,400
1988	2,402,400	1,808,400	2,996,300
1989	2,316,300	1,836,700	2,795,800
1990	2,183,800	1,886,200	2,479,400
1991	2,393,300	2,116,000	2,670,700
1992	2,172,900	1,898,900	2,690,600
1993	2,465,400	2,151,500	2,779,300
1994	2,610,500	2,266,800	2,954,100
1995	2,009,700	1,724,800	2,294,600
1996	2,298,600	1,749,900	2,847,300
1997	2,163,400	1,907,900	2,418,900
1998	2,329,600	2,033,130	2,626,070
1999	1,306,470	1,118,800	1,494,150
2000	1,581,900	1,382,000	1,781,800
2001	1,863,700	1,605,000	2,122,300
2002	2,016,700	1,740,700	2,292,700
2003	2,239,600	1,822,700	2,656,600
2004	2,530,600	2,147,900	2,913,300
2005	2,823,500	2,035,800	3,499,800
2006	2,133,070	1,818,253	2,447,932
2007	2,152,738	1,775,191	2,530,285
2008	2,099,521	1,599,100	2,600,000
2009	1,739,238	1,435,188	2,043,288
2010	2,367,830	1,807,430	2,928,230
2011	2,403,021	1,926,371	2,879,671
2012	1,951,400	1,675,982	2,226,819
2013	2,279,004	1,934,134	2,623,874

Table 4.6. Yellowfin sole population numbers-at-age (millions) estimated from the annual bottom trawl surveys, 1982-2011.

year/age	Females															
	2	3	4	5	6	7	8	9	10	11	12	13	14	15	16	17+
1979	21	113	150	442	616	386	555	801	626	528	219	274	59	35	29	15
1980	1	92	342	518	800	1055	413	661	880	651	765	285	113	33	23	23
1981	0	20	195	839	692	1321	1155	261	477	744	527	311	168	55	23	45
1982	38	183	349	1211	1485	1424	1619	843	829	832	704	409	246	159	51	84
1983	0	5	59	154	751	1413	843	1065	936	753	1155	866	295	160	60	54
1984	0	53	278	264	427	745	841	1111	1080	941	541	583	480	239	174	133
1985	0	3	105	442	587	406	632	915	441	518	545	384	298	321	205	127
1986	0	8	24	219	349	666	279	574	519	377	284	318	196	250	136	259
1987	0	0	70	120	803	458	843	259	376	599	356	449	243	270	247	688
1988	0	0	7	370	71	1495	560	557	184	239	351	208	360	273	219	886
1989	0	0	14	98	718	234	1337	593	446	74	179	308	234	238	183	565
1990	0	0	70	102	325	1066	192	1257	408	482	101	72	107	78	231	605
1991	0	10	127	248	123	405	896	151	1263	213	525	63	128	87	123	807
1992	0	19	247	485	520	213	286	938	94	825	75	309	129	137	170	715
1993	0	24	100	357	634	434	269	224	1314	78	866	157	165	69	68	674
1994	0	54	95	223	518	905	555	482	284	1170	516	44	274	142	42	588
1995	0	19	153	288	181	889	627	274	135	25	634	21	561	104	80	512
1996	0	16	154	809	288	279	434	517	206	146	151	602	116	637	47	619
1997	0	18	324	502	725	256	239	506	228	114	176	184	500	44	314	533
1998	0	10	83	479	420	900	260	203	370	413	369	170	176	265	67	1167
1999	0	3	65	198	175	185	727	104	107	245	190	186	72	102	175	425
2000	0	11	54	248	208	304	444	537	189	198	237	219	65	117	145	572
2001	0	1	71	239	522	248	403	415	654	374	83	191	154	127	189	617
2002	0	16	123	170	255	778	346	290	229	457	221	91	307	116	152	805
2003	0	15	115	241	251	287	1143	225	279	286	251	103	115	170	168	943
2004	10	33	192	430	560	441	217	966	221	212	218	219	106	20	167	1020
2005	0	53	167	194	602	433	213	487	834	196	144	191	324	170	53	1332

<b>2006</b>	0	67	302	376	276	634	470	176	325	738	133	133	71	156	175	514
<b>2007</b>	0	37	515	348	376	277	504	308	124	227	504	119	137	127	105	724
<b>2008</b>	0	24	115	736	621	546	359	355	198	117	259	350	153	79	85	732
<b>2009</b>	5	38	204	204	1187	609	488	259	210	218	129	138	196	88	43	444
<b>2010</b>	0	33	328	386	438	895	554	517	329	335	155	166	135	173	99	684
<b>2011</b>	0	14	243	539	707	463	769	410	457	204	226	149	142	145	186	619
<b>2012</b>	10	50	229	394	503	293	243	752	256	334	106	156	37	150	128	547

Table 4.6.(continued)

**Males**

<b>year/age</b>	<b>2</b>	<b>3</b>	<b>4</b>	<b>5</b>	<b>6</b>	<b>7</b>	<b>8</b>	<b>9</b>	<b>10</b>	<b>11</b>	<b>12</b>	<b>13</b>	<b>14</b>	<b>15</b>	<b>16</b>	<b>17+</b>
<b>1979</b>	21	115	143	390	381	303	583	847	604	406	349	247	54	76	29	36
<b>1980</b>	20	78	306	632	853	1221	457	558	616	568	444	370	147	18	8	8
<b>1981</b>	0	50	200	1047	640	1280	858	394	372	546	534	266	66	83	55	12
<b>1982</b>	89	193	428	1780	1781	1059	1673	644	774	463	471	482	302	8	24	8
<b>1983</b>	0	1	65	183	724	1729	808	1049	676	699	722	566	425	550	77	51
<b>1984</b>	0	68	246	323	497	734	830	612	788	718	358	379	201	316	122	106
<b>1985</b>	0	41	172	419	559	263	652	527	401	451	360	224	260	157	112	65
<b>1986</b>	0	13	47	108	373	652	262	327	284	335	211	205	115	210	82	252
<b>1987</b>	0	5	41	106	838	467	673	445	328	277	210	147	106	142	185	600
<b>1988</b>	0	2	10	435	49	1163	553	443	85	187	28	177	336	189	28	599
<b>1989</b>	0	2	23	181	788	177	1306	513	357	135	50	103	54	204	35	478
<b>1990</b>	0	11	47	121	316	888	195	1144	318	263	40	65	67	24	55	389
<b>1991</b>	0	0	103	354	139	275	1046	68	1137	328	244	74	64	60	53	420
<b>1992</b>	0	0	146	445	566	262	226	812	114	907	193	213	12	12	61	607
<b>1993</b>	0	20	52	233	646	393	279	247	1096	69	842	53	53	50	0	341
<b>1994</b>	4	22	71	166	427	953	656	308	191	822	26	622	46	132	11	303
<b>1995</b>	0	0	169	120	270	667	565	94	179	75	478	13	603	49	24	418
<b>1996</b>	0	76	95	837	244	227	425	344	331	141	139	399	61	449	125	495
<b>1997</b>	0	10	214	425	798	181	184	446	245	194	214	108	514	79	264	416
<b>1998</b>	0	48	70	351	569	832	159	226	204	272	346	140	157	191	113	814

<b>1999</b>	0	5	100	142	225	243	575	146	94	309	269	75	53	28	119	425
<b>2000</b>	0	0	36	219	259	143	509	583	78	215	133	77	92	78	66	547
<b>2001</b>	0	0	87	141	652	341	375	357	562	208	87	158	65	73	140	432
<b>2002</b>	0	58	72	158	309	758	318	333	262	442	194	120	220	161	133	507
<b>2003</b>	0	24	95	178	258	251	1074	238	363	53	284	173	10	71	57	682
<b>2004</b>	4	63	114	469	447	199	395	993	263	81	195	223	103	47	249	456
<b>2005</b>	0	49	166	187	474	476	204	288	972	123	142	121	133	69	93	726
<b>2006</b>	0	101	173	348	332	505	393	288	298	384	116	155	89	39	11	590
<b>2007</b>	0	58	481	352	405	284	545	209	166	252	338	101	133	72	59	620
<b>2008</b>	0	10	99	662	462	483	344	453	225	144	185	329	63	66	35	581
<b>2009</b>	0	65	144	289	946	462	555	248	249	217	78	31	195	30	29	363
<b>2010</b>	0	78	199	418	371	1032	462	510	171	189	159	53	117	151	78	678
<b>2011</b>	1	7	150	385	482	358	792	398	224	176	77	81	136	103	157	440
<b>2012</b>	0	69	274	352	344	273	238	425	297	179	98	67	91	34	100	2

Table 4.7-Occurance of yellowfin sole in the Bering Sea trawl survey and collections of length and age structures and the number of otoliths aged from each survey.

Year	Total Hauls	Hauls w/Len	Number lengths	Hauls w/otoliths	Hauls w/ages	Number otoliths	Number ages
1982	334	246	37023	35	35	744	744
1983	353	256	33924	37	37	709	709
1984	355	271	33894	56	56	821	796
1985	357	261	33824	44	43	810	802
1986	354	249	30470	34	34	739	739
1987	357	224	31241	16	16	798	798
1988	373	254	27138	14	14	543	543
1989	374	236	29672	24	24	740	740
1990	371	251	30257	28	28	792	792
1991	372	248	27986	26	26	742	742
1992	356	229	23628	16	16	606	606
1993	375	242	26651	20	20	549	549
1994	375	269	24448	14	14	526	522
1995	376	254	22116	20	20	654	647
1996	375	247	27505	16	16	729	721
1997	376	262	26034	11	11	470	466
1998	375	310	34509	15	15	575	570
1999	373	276	28431	31	31	777	770
2000	372	255	24880	20	20	517	511
2001	375	251	26558	25	25	604	593
2002	375	246	26309	32	32	738	723
2003	376	241	27135	37	37	699	695
2004	375	251	26103	26	26	725	712
2005	373	251	24658	34	34	644	635
2006	376	246	28470	39	39	428	426
2007	376	247	24790	66	66	779	772
2008	375	238	25848	65	65	858	830
2009	376	235	22018	70	70	784	752
2010	376	228	20619	77	77	841	827
2011	376	228	21665	65	64	784	753
2012	376	242	23519	72	72	993	973
2013	376	232	23261	70		821	

Table 4.8—Total tonnage of yellowfin sole caught in resource assessment surveys in the eastern Bering Sea from 1977-2013.

<u>Year</u>	<u>Research catch (t)</u>
1977	60
1978	71
1979	147
1980	92
1981	74
1982	158
1983	254
1984	218
1985	105
1986	68
1987	92
1988	138
1989	148
1990	129
1991	118
1992	60
1993	95
1994	91
1995	95
1996	72
1997	76
1998	79
1999	61
2000	72
2001	75
2002	76
2003	78
2004	114
2005	94
2006	74
2007	74
2008	69
2009	60
2010	79
2011	77
2012	64
2013	75







Table 4.10. Female yellowfin sole proportion mature at age from Nichol (1994).

Age	Proportion mature
1	0.00
2	0.00
3	.001
4	.004
5	.008
6	.020
7	.046
8	.104
9	.217
10	.397
11	.612
12	.790
13	.899
14	.955
15	.981
16	.992
17	.997
18	1.000
19	1.000
20	1.000

Table 4.11. Key equations used in the population dynamics model.

$$N_{t,1} = R_t = R_0 e^{\tau_t}, \quad \tau_t \sim N(0, \sigma^2_R) \quad \text{Recruitment 1956-75}$$

$$N_{t,1} = R_t = R_\gamma e^{\tau_t}, \quad \tau_t \sim N(0, \sigma^2_R) \quad \text{Recruitment 1976-96}$$

$$C_{t,a} = \frac{F_{t,a}}{Z_{t,a}} (1 - e^{-z_{t,a}}) N_{t,a}$$

Catch in year  $t$  for age  $a$  fish

$$N_{t+1,a+1} = N_{t,a} e^{-z_{t,a}}$$

Numbers of fish in year  $t+1$  at age  $a$

$$N_{t+1,A} = N_{t,A-1} e^{-z_{t,A-1}} + N_{t,A} e^{-z_{t,A}}$$

Numbers of fish in the “plus group”

$$S_t = \sum N_{t,a} W_{t,a} \phi_a$$

Spawning biomass

$$Z_{t,a} = F_{t,a} + M$$

Total mortality in year  $t$  at age  $a$

$$F_{t,a} = s_a \mu^F \exp^{\varepsilon^F_t}, \quad \varepsilon^F_t \sim N(0, \sigma^2_F)$$

Fishing mortality

$$s_a = \frac{1}{1 + (e^{-\alpha + \beta a})}$$

Age-specific fishing selectivity

$$C_t = \sum C_{t,a}$$

Total catch in numbers

$$P_{t,a} = \frac{C_{t,a}}{C_t}$$

Proportion at age in catch

$$SurB_t = q \sum N_{t,a} W_{t,a} v_a$$

Survey biomass

Table 4.11—continued.

$$qprior = \lambda \frac{0.5(\ln q_{est,t} - \ln q_{prior})^2}{\sigma_q^2} \quad \text{survey catchability prior (when estimated)}$$

$$mprior = \lambda \frac{0.5(\ln m_{est} - \ln m_{prior})^2}{\sigma_m^2} \quad \text{natural mortality prior (when estimated)}$$

$$reclike = \lambda \left( \sum_{i=1965}^{endyear} \bar{R} - R_i \right)^2 + \sum_{a=1}^{20} (\bar{R}_{init} - R_{init,a})^2 + \frac{1}{2((\sum_{i=1965}^{endyear} \bar{R} - R_i) \frac{1}{n+1})} \quad \text{recruitment likelihood}$$

$$catchlike = \lambda \sum_{i=startyear}^{endyear} (\ln C_{obs,i} - \ln C_{est,i})^2 \quad \text{catch likelihood}$$

$$surveylike = \lambda \frac{(\ln B - \ln \hat{B})^2}{2\sigma^2} \quad \text{survey likelihood}$$

$$SurvAgelike = \sum_{i,t} m_t P_{t,a} \ln \frac{\hat{P}_{t,a}}{P_{t,a}} \quad \text{survey age composition likelihood}$$

$$FishAgelike = \sum_{i,t} m_t P_{t,a} \ln \frac{\hat{P}_{t,a}}{P_{t,a}} \quad \text{fishery age composition likelihood}$$

Table 4.12. Variables used in the population dynamics model.

Variables

$R_t$	Age 1 recruitment in year $t$
$R_0$	Geometric mean value of age 1 recruitment, 1956-75
$R_\gamma$	Geometric mean value of age 1 recruitment, 1976-96
$\tau_t$	Recruitment deviation in year $t$
$N_{t,a}$	Number of fish in year $t$ at age $a$
$C_{t,a}$	Catch numbers of fish in year $t$ at age $a$
$P_{t,a}$	Proportion of the numbers of fish age $a$ in year $t$
$C_t$	Total catch numbers in year $t$
$W_{t,a}$	Mean body weight (kg) of fish age $a$ in year $t$
$\phi_a$	Proportion of mature females at age $a$
$F_{t,a}$	Instantaneous annual fishing mortality of age $a$ fish in year $t$
$M$	Instantaneous natural mortality, assumed constant over all ages and years
$Z_{t,a}$	Instantaneous total mortality for age $a$ fish in year $t$
$s_a$	Age-specific fishing gear selectivity
$\mu^F$	Median year-effect of fishing mortality
$\varepsilon_t^F$	The residual year-effect of fishing mortality
$v_a$	Age-specific survey selectivity
$\alpha$	Slope parameter in the logistic selectivity equation
$\beta$	Age at 50% selectivity parameter in the logistic selectivity equation
$\sigma_t$	Standard error of the survey biomass in year $t$

Table 4.13. Models evaluated for stock productivity in the 2013 stock assessment of yellowfin sole

	<b>Model A</b>	<b>Model B</b>
<b>Years included</b>	1955-2006	1978-2006
<b>Fmsy</b>	0.135	0.118
<b>Bmsy (t)</b>	318,000	366,000
<b>ABC (t)</b>	284,200	239,800
<b>OFL (t)</b> <b>Buffer between ABC and OFL</b>	286,100 1%	259,700 8%







Table 4.15. Model estimates of yellowfin sole full selection fishing mortality and exploitation rate (catch/total biomass).

Year	Exploitation	
	Full selection F	Rate
1964	0.27	0.14
1965	0.25	0.07
1966	0.43	0.12
1967	0.53	0.20
1968	0.33	0.11
1969	0.66	0.22
1970	0.68	0.18
1971	1.46	0.21
1972	0.25	0.06
1973	0.41	0.08
1974	0.13	0.04
1975	0.13	0.05
1976	0.12	0.03
1977	0.06	0.03
1978	0.12	0.06
1979	0.07	0.04
1980	0.08	0.03
1981	0.06	0.04
1982	0.05	0.03
1983	0.05	0.04
1984	0.08	0.05
1985	0.12	0.08
1986	0.11	0.07
1987	0.11	0.07
1988	0.14	0.08
1989	0.10	0.06
1990	0.05	0.03
1991	0.05	0.04
1992	0.08	0.06
1993	0.06	0.04
1994	0.07	0.06
1995	0.06	0.05
1996	0.07	0.05
1997	0.11	0.08
1998	0.06	0.04
1999	0.05	0.03
2000	0.06	0.04
2001	0.04	0.03
2002	0.05	0.03
2003	0.05	0.03
2004	0.04	0.03

2005	0.06	0.04
2006	0.06	0.05
2007	0.08	0.06
2008	0.10	0.07
2009	0.08	0.05
2010	0.09	0.05
2011	0.12	0.07
2012	0.13	0.06
2013		0.07









Table 4.17. Model estimates of yellowfin sole age 2+ total biomass (t) and begin-year female spawning biomass (t) from the 2012 and 2013 stock assessments.

Year	2013 Assessment		2012 Assessment	
	Female spawning biomass	Total biomass	Female spawning biomass	Total biomass
1964	128,916	779,784	127,733	771,625
1965	144,813	777,174	143,009	767,314
1966	162,027	831,076	159,627	819,804
1967	154,268	823,012	151,517	810,860
1968	133,895	746,221	130,719	733,371
1969	125,631	771,265	120,504	757,973
1970	101,100	724,316	92,773	709,264
1971	90,814	751,985	72,090	733,241
1972	91,809	784,413	47,137	745,387
1973	97,381	962,921	52,048	923,714
1974	104,688	1,147,400	65,143	1,111,290
1975	135,561	1,412,760	97,831	1,375,670
1976	181,097	1,669,720	147,292	1,629,730
1977	250,703	1,938,100	220,952	1,895,140
1978	340,085	2,192,830	313,144	2,146,370
1979	430,780	2,339,210	406,321	2,289,690
1980	534,042	2,505,400	510,239	2,452,730
1981	632,460	2,657,740	607,825	2,601,440
1982	721,023	2,787,560	695,993	2,728,410
1983	801,450	2,882,110	775,810	2,820,490
1984	860,593	2,946,590	834,221	2,882,260
1985	879,214	2,934,760	852,904	2,868,480
1986	867,261	2,853,940	840,950	2,785,530
1987	845,446	2,787,250	818,932	2,716,620
1988	812,032	2,765,630	785,310	2,692,360
1989	773,138	2,685,510	746,208	2,609,920
1990	781,876	2,676,890	754,197	2,598,860
1991	827,905	2,710,160	799,132	2,630,260
1992	863,332	2,712,710	833,483	2,631,000
1993	888,272	2,642,890	857,165	2,559,340
1994	902,563	2,601,120	870,482	2,516,060
1995	891,342	2,512,890	858,739	2,426,400
1996	869,571	2,449,370	836,645	2,360,690
1997	829,189	2,377,610	796,104	2,286,520
1998	787,211	2,249,500	753,792	2,156,230
1999	766,924	2,203,530	733,067	2,107,710
2000	758,269	2,215,810	723,588	2,116,370
2001	753,683	2,211,140	717,839	2,108,240
2002	746,861	2,210,630	710,079	2,105,270
2003	748,528	2,208,500	710,213	2,100,650

2004	745,337	2,209,820	705,755	2,100,360
2005	743,817	2,212,150	702,773	2,101,570
2006	735,119	2,188,710	692,642	2,079,070
2007	720,061	2,179,260	676,246	2,073,210
2008	688,110	2,173,370	643,154	2,070,390
2009	648,594	2,155,470	603,153	2,042,740
2010	623,607	2,198,590	578,050	2,072,760
2011	601,560	2,245,240	555,717	2,107,030
2012	581,630	2,273,960	540,236	2,117,470
2013	576,652	2,314,360		









Table 4.19—Model estimates of the number of female spawners (millions) 1964-2013.

year/age	5	6	7	8	9	10	11	12	13	14	15	16	17	18	19	20
<b>1964</b>	4.5	5.9	26.2	72.5	43.0	85.0	132.7	105.7	39.9	14.3	2.7	0.4	0.1	0.1	0.1	0.8
<b>1965</b>	2.2	4.3	15.5	40.0	102.8	53.4	89.0	116.4	81.7	28.8	10.0	1.9	0.2	0.1	0.1	0.6
<b>1966</b>	4.8	2.8	14.9	30.8	73.1	159.8	65.1	83.9	93.1	60.3	20.5	7.0	1.3	0.2	0.1	0.5
<b>1967</b>	2.4	6.1	9.5	29.5	55.9	111.8	185.0	54.8	57.2	57.5	35.7	11.9	4.0	0.7	0.1	0.3
<b>1968</b>	2.1	3.0	20.3	16.3	37.5	55.3	90.4	124.7	32.5	31.7	30.8	18.8	6.3	2.1	0.4	0.2
<b>1969</b>	3.0	2.7	10.2	40.3	30.1	60.3	72.5	91.4	99.2	22.6	20.9	19.8	12.0	4.0	1.3	0.4
<b>1970</b>	3.0	3.7	9.2	19.6	68.7	40.1	56.9	49.1	50.2	49.2	10.7	9.7	9.1	5.5	1.8	0.8
<b>1971</b>	6.3	3.9	12.8	18.3	36.2	109.8	51.1	50.7	29.9	25.1	23.0	4.9	4.4	4.1	2.5	1.2
<b>1972</b>	9.5	8.0	13.2	25.4	33.3	55.9	131.0	41.9	26.1	10.0	6.4	5.1	1.0	0.9	0.9	0.8
<b>1973</b>	9.6	12.0	26.7	23.4	37.7	42.2	59.5	116.6	32.9	19.1	7.1	4.4	3.6	0.7	0.6	1.1
<b>1974</b>	12.5	12.1	40.9	51.9	39.6	48.8	40.8	46.0	78.5	20.6	11.6	4.2	2.6	2.1	0.4	1.0
<b>1975</b>	13.8	15.9	41.3	80.7	93.8	60.9	60.8	41.5	40.8	64.8	16.4	9.1	3.3	2.0	1.6	1.1
<b>1976</b>	10.8	17.5	54.1	80.7	140.7	137.3	73.7	61.2	36.8	33.8	51.8	12.9	7.1	2.6	1.6	2.1
<b>1977</b>	7.5	13.7	59.7	106.3	144.7	214.7	171.1	75.5	54.9	30.7	27.2	41.1	10.2	5.6	2.0	2.9
<b>1978</b>	10.1	9.4	46.2	115.6	189.0	223.9	278.2	185.1	71.8	48.7	26.3	23.0	34.5	8.5	4.7	4.1
<b>1979</b>	11.9	12.8	31.9	88.8	199.8	278.4	273.6	283.0	165.6	60.0	39.3	20.9	18.2	27.2	6.7	6.9
<b>1980</b>	7.8	15.0	43.3	62.2	157.8	306.5	356.5	292.2	266.1	145.3	50.9	32.8	17.4	15.0	22.5	11.2
<b>1981</b>	9.8	9.9	50.9	84.8	112.0	245.9	396.5	381.9	274.0	232.1	122.3	42.1	27.0	14.2	12.3	27.6
<b>1982</b>	6.4	12.4	33.6	99.8	152.8	174.8	319.8	428.7	362.6	242.4	198.3	102.9	35.2	22.5	11.8	33.1
<b>1983</b>	4.1	8.1	42.2	65.7	179.5	238.7	228.8	349.5	412.6	325.6	210.4	169.5	87.4	29.8	19.0	38.0
<b>1984</b>	7.9	5.2	27.5	81.9	117.1	278.6	311.1	249.4	335.7	369.9	282.2	179.6	143.8	73.9	25.2	48.1
<b>1985</b>	5.9	10.0	17.6	53.7	146.1	179.7	355.7	330.7	233.2	292.8	311.9	234.3	148.1	118.4	60.7	60.2
<b>1986</b>	16.9	7.4	34.0	34.3	95.1	220.6	223.4	365.7	298.2	195.9	237.7	249.1	186.0	117.4	93.5	95.5
<b>1987</b>	3.1	21.4	25.2	65.9	59.5	140.7	272.1	230.0	331.6	252.4	160.3	191.5	199.4	148.6	93.5	150.5
<b>1988</b>	13.9	4.0	73.0	49.7	119.1	92.3	179.1	284.4	209.9	281.4	206.8	129.2	153.3	159.4	118.4	194.4

Table 4.19 (continued).

year/age	<b>5</b>	<b>6</b>	<b>7</b>	<b>8</b>	<b>9</b>	<b>10</b>	<b>11</b>	<b>12</b>	<b>13</b>	<b>14</b>	<b>15</b>	<b>16</b>	<b>17</b>	<b>18</b>	<b>19</b>	<b>20</b>
<b>1989</b>	4.8	17.6	13.5	143.6	88.8	179.6	113.1	180.4	251.0	172.5	223.6	161.7	100.4	118.9	123.3	241.9
<b>1990</b>	3.7	6.0	60.1	26.7	261.9	138.8	229.0	118.1	164.7	213.2	141.7	180.7	129.9	80.5	95.0	291.8
<b>1991</b>	4.9	4.6	20.7	119.1	48.8	414.1	182.8	251.2	114.0	148.3	185.7	121.4	153.9	110.4	68.2	327.9
<b>1992</b>	6.7	6.3	15.8	40.9	217.2	77.2	545.0	200.0	241.5	102.2	128.4	158.3	102.9	130.1	93.1	333.9
<b>1993</b>	6.6	8.5	21.3	31.1	73.6	336.3	98.9	578.8	186.4	209.7	85.7	106.1	129.9	84.3	106.3	348.7
<b>1994</b>	3.3	8.4	28.9	42.1	56.6	116.7	444.2	108.2	554.1	165.7	179.8	72.3	88.8	108.5	70.1	378.7
<b>1995</b>	3.6	4.1	28.5	56.7	75.2	87.0	149.2	473.3	101.5	484.8	140.2	149.7	59.8	73.3	89.3	369.6
<b>1996</b>	8.0	4.6	14.1	55.8	101.1	115.9	112.0	160.3	447.9	89.6	414.0	117.8	125.0	49.9	60.9	381.5
<b>1997</b>	4.7	10.1	15.6	27.6	100.0	156.6	149.6	120.2	151.2	393.7	76.1	346.0	97.9	103.6	41.2	365.6
<b>1998</b>	3.9	5.9	34.0	30.2	48.9	152.5	197.6	156.0	109.8	128.4	322.8	61.4	277.2	78.3	82.6	324.3
<b>1999</b>	3.9	5.0	20.0	66.4	54.1	75.8	197.4	213.3	148.2	97.3	110.0	272.2	51.5	231.9	65.3	339.3
<b>2000</b>	9.7	4.9	17.0	39.8	122.0	86.3	100.4	217.0	205.7	133.2	84.5	94.1	231.3	43.6	196.0	342.0
<b>2001</b>	4.1	12.3	16.8	33.6	72.9	193.6	113.6	109.4	207.2	183.0	114.5	71.5	79.1	194.0	36.5	449.9
<b>2002</b>	3.4	5.2	41.9	33.1	61.3	115.1	255.4	124.8	105.8	186.8	159.5	98.2	61.0	67.3	164.6	412.8
<b>2003</b>	4.2	4.3	17.9	83.1	60.7	97.3	151.9	279.7	120.1	94.9	162.1	136.2	83.4	51.6	56.8	487.6
<b>2004</b>	5.9	5.4	14.7	35.4	151.4	95.9	128.1	166.4	269.5	108.0	82.5	138.6	115.8	70.7	43.7	460.5
<b>2005</b>	3.7	7.4	18.3	29.0	64.5	239.2	126.5	140.8	160.9	243.0	94.1	70.8	118.2	98.5	60.0	427.7
<b>2006</b>	5.4	4.7	25.3	36.0	52.4	100.9	312.1	137.2	134.2	143.0	208.7	79.5	59.4	99.1	82.4	407.6
<b>2007</b>	5.6	6.8	15.8	48.4	63.1	80.4	130.4	337.6	130.9	119.4	123.0	176.8	67.0	49.9	83.0	410.5
<b>2008</b>	11.3	7.0	23.2	31.0	86.5	97.1	102.7	138.6	315.4	114.0	100.6	102.0	145.7	55.1	40.9	404.6
<b>2009</b>	5.2	14.2	23.9	45.2	54.7	130.0	121.0	106.5	126.5	268.5	93.9	81.5	82.1	117.1	44.1	357.0
<b>2010</b>	6.8	6.6	48.6	47.1	80.8	83.5	165.0	128.1	99.4	110.1	226.0	77.8	67.1	67.5	95.9	328.7
<b>2011</b>	8.2	8.6	22.5	95.6	84.7	124.5	106.1	173.9	118.5	85.7	91.8	185.4	63.4	54.6	54.8	344.5
<b>2012</b>	9.4	10.3	29.4	44.3	171.9	129.6	155.5	108.8	155.9	98.9	69.0	72.8	146.1	49.9	42.8	313.0
<b>2013</b>	11.9	11.9	35.2	58.0	80.0	266.8	164.7	161.3	98.0	129.9	79.4	54.5	57.1	114.3	38.9	277.6

Table 4.20. Model estimates of yellowfin sole age 5 recruitment (millions) from the 2012 and 2013 stock assessments.

<b>Year class</b>	<b>2013 Assessment</b>	<b>2012 Assessment</b>
<b>1964</b>	706	686
<b>1965</b>	724	727
<b>1966</b>	1,500	1,488
<b>1967</b>	2,263	2,240
<b>1968</b>	2,280	2,263
<b>1969</b>	2,986	2,955
<b>1970</b>	3,294	3,253
<b>1971</b>	2,582	2,546
<b>1972</b>	1,792	1,765
<b>1973</b>	2,412	2,375
<b>1974</b>	2,830	2,786
<b>1975</b>	1,865	1,836
<b>1976</b>	2,342	2,305
<b>1977</b>	1,529	1,503
<b>1978</b>	976	959
<b>1979</b>	1,888	1,852
<b>1980</b>	1,399	1,371
<b>1981</b>	4,033	3,947
<b>1982</b>	744	728
<b>1983</b>	3,309	3,234
<b>1984</b>	1,137	1,110
<b>1985</b>	870	848
<b>1986</b>	1,178	1,146
<b>1987</b>	1,598	1,552
<b>1988</b>	1,580	1,530
<b>1989</b>	781	755
<b>1990</b>	865	832
<b>1991</b>	1,900	1,821
<b>1992</b>	1,116	1,067
<b>1993</b>	936	889
<b>1994</b>	923	882
<b>1995</b>	2,307	2,202
<b>1996</b>	985	927
<b>1997</b>	807	776
<b>1998</b>	1,009	958
<b>1999</b>	1,400	1,334
<b>2000</b>	889	871
<b>2001</b>	1,291	1,253
<b>2002</b>	1,329	1,304
<b>2003</b>	2,684	2,663
<b>2004</b>	1,241	1,178
<b>2005</b>	1,626	1,726
<b>2006</b>	1,942	

Table 4.21—Selected parameter estimates and their standard deviation from the preferred stock assessment model.

parameter	value	std dev	parameter	value	std dev
alpha (q-temp model)	-0.10	0.04	1975 total biomass	1412800	38999
beta (q-temp model)	0.09	0.02	1976 total biomass	1669700	44874
mean_log_rec	0.83	0.10	1977 total biomass	1938100	50599
mean sel_slope_fsh (females)	1.19	0.08	1978 total biomass	2192800	55995
mean sel50_fsh (females)	8.65	0.26	1979 total biomass	2339200	60632
mean sel_slope_fsh_males	1.34	0.10	1980 total biomass	2505400	64869
mean sel50_fsh_males	8.23	0.24	1981 total biomass	2657700	68492
sel_slope_srv (females)	1.62	0.09	1982 total biomass	2787600	71680
sel50_srv (females)	5.07	0.07	1983 total biomass	2882100	74242
sel_slope_srv_males	-0.06	0.08	1984 total biomass	2946600	76404
sel50_srv_males	0.02	0.02	1985 total biomass	2934800	78006
Ricker SR logalpha	-4.15	0.54	1986 total biomass	2853900	79423
Ricker SR logbeta	-6.10	0.31	1987 total biomass	2787200	80765
Fmsy	0.21	0.10	1988 total biomass	2765600	82458
log (Fmsy)	-1.57	0.46	1989 total biomass	2685500	83609
ABC_biomass 2012	2118800	149910	1990 total biomass	2676900	85003
ABC_biomass 2013	2196000	186320	1991 total biomass	2710200	85793
msy	339.16	120.02	1992 total biomass	2712700	86190
Bmsy	366.11	65.42	1993 total biomass	2642900	86397
1954 total biomass	2314800	150920	1994 total biomass	2601100	86174
1955 total biomass	2279800	136550	1995 total biomass	2512900	85677
1956 total biomass	2229600	121360	1996 total biomass	2449400	85542
1957 total biomass	2179400	105290	1997 total biomass	2377600	85411
1958 total biomass	2141200	90224	1998 total biomass	2249500	84936
1959 total biomass	2095100	76049	1999 total biomass	2203500	85025
1960 total biomass	1920700	61494	2000 total biomass	2215800	86133
1961 total biomass	1502100	43695	2001 total biomass	2211100	87384
1962 total biomass	1035600	24388	2002 total biomass	2210600	88095
1963 total biomass	741880	14830	2003 total biomass	2208500	89309
1964 total biomass	779780	15322	2004 total biomass	2209800	90612
1965 total biomass	777170	15912	2005 total biomass	2212200	91889
1966 total biomass	831080	17081	2006 total biomass	2188700	93224
1967 total biomass	823010	17471	2007 total biomass	2179300	95457
1968 total biomass	746220	17244	2008 total biomass	2173400	98942
1969 total biomass	771260	18421	2009 total biomass	2155500	103580
1970 total biomass	724320	18929	2010 total biomass	2198600	110280
1971 total biomass	751990	20882	2011 total biomass	2245200	119580
1972 total biomass	784410	23877	2012 total biomass	2274000	132380
1973 total biomass	962920	28354	2013 total biomass	2314400	150140
1974 total biomass	1147400	33332			

Table 4.22. Projections of yellowfin sole female spawning biomass (1,000s t), catch (1,000s t) and full selection fishing mortality rate for seven future harvest scenarios.

Scenarios 1 and 2				Scenario 3			
Maximum Tier 3 ABC harvest permissible				1/2 Maximum Tier 3 ABC harvest permissible			
Female				Female			
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2013	622.106	152.986	0.11	2013	622.106	152.986	0.11
2014	632.811	160.315	0.11	2014	644.075	80.163	0.05
2015	645.048	165.083	0.11	2015	690.046	86.682	0.05
2016	672.476	173.554	0.11	2016	752.559	94.619	0.05
2017	694.340	181.600	0.11	2017	808.111	102.262	0.05
2018	718.835	184.550	0.11	2018	865.927	107.163	0.05
2019	736.988	182.338	0.11	2019	916.972	109.215	0.05
2020	737.416	176.613	0.11	2020	946.101	108.856	0.05
2021	737.509	172.343	0.11	2021	974.507	108.953	0.05
2022	718.675	167.856	0.11	2022	974.275	108.317	0.05
2023	696.676	163.730	0.11	2023	966.511	107.484	0.05
2024	675.344	160.106	0.11	2024	954.279	106.532	0.05
2025	658.346	157.230	0.11	2025	944.163	105.720	0.05
2026	649.147	155.079	0.11	2026	942.360	105.301	0.05

Scenario 4				Scenario 5			
Harvest at average F over the past 5 years				No fishing			
Female				Female			
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2013	622.106	152.986	0.11	2013	622.106	152.986	0.11
2014	635.728	139.790	0.10	2014	655.006	0	0
2015	654.634	158.858	0.10	2015	736.130	0	0
2016	685.046	167.452	0.10	2016	838.992	0	0
2017	709.403	175.585	0.10	2017	936.887	0	0
2018	736.174	178.814	0.10	2018	1039.780	0	0
2019	756.459	177.085	0.10	2019	1138.420	0	0
2020	758.556	171.896	0.10	2020	1212.850	0	0
2021	760.329	168.055	0.10	2021	1289.160	0	0
2022	742.301	163.908	0.10	2022	1326.370	0	0
2023	720.793	160.055	0.10	2023	1352.230	0	0
2024	699.565	156.631	0.10	2024	1366.950	0	0
2025	682.553	153.895	0.10	2025	1380.820	0	0
2026	673.420	152.028	0.10	2026	1403.690	0	0

Table 4.22—continued.

Scenario 6				Scenario 7			
Determination of whether yellowfin sole are currently overfished				Determination of whether the stock is approaching an overfished condition			
B35=502.900				B35=502.900			
Female				Female			
Year	spawning biomass	catch	F	Year	spawning biomass	catch	F
2013	622.106	152.986	0.11	2013	622.106	152.986	0.11
2014	628.551	189.986	0.13	2014	632.808	160.333	0.11
2015	628.630	192.315	0.13	2015	645.040	165.081	0.11
2016	644.236	199.353	0.13	2016	667.979	205.668	0.13
2017	655.500	206.128	0.13	2017	676.848	211.587	0.13
2018	670.090	207.154	0.13	2018	688.724	211.716	0.13
2019	678.967	202.400	0.13	2019	694.839	206.162	0.13
2020	671.890	194.089	0.13	2020	684.954	197.122	0.13
2021	665.014	187.788	0.13	2021	675.820	190.247	0.13
2022	642.436	181.724	0.13	2022	651.103	183.685	0.13
2023	618.181	176.376	0.13	2023	625.195	177.945	0.13
2024	596.186	170.841	0.13	2024	601.724	172.469	0.13
2025	579.741	164.842	0.13	2025	583.957	166.291	0.13
2026	571.712	160.964	0.13	2026	574.823	162.063	0.13

Table 4-23. Catch and bycatch (t) of other BSAI target species in the yellowfin sole directed fishery from 1992-2012 estimated from a combination of regional office reported catch and observer sampling of the catch.



Table 4-24. Estimated non-target species catch (t) in the yellowfin sole fishery, 2003-2013 (PSC not included).

	2003	2004	2005	2006	2007	2008	2009	2010	2011	2012	2013
Benthic urochordata	1670846	1695563	674762	520091	114427	347756	205806	155571	132867	143576	59972
Birds			0				0	0	0		
Bivalves	1543	1113	1327	343	448	1484	1300	1822	1671	692	345
Brittle star unidentified	34303	32271	28706	19961	7526	19048	5209	4082	14024	13126	1440
Capelin	3	4519	45	108	321	161	251	718	3769	2325	181
Corals Bryozoans	240	46	1232	9378	162	8309	312	504	950	677	961
Eelpouts	19044	12256	7729	4514	2344	5598	5188	5144	29320	14293	43288
Eulachon	12	278	33	115	5075	22	89	133	453	133	20
Giant Grenadier									236		
Greenlings	646	753	283	703	474	183	24	53	49	98	
Grenadier					339		358				
Gunnels					1						1
Hermit crab unidentified	87940	51999	82996	26898	35820	36606	15623	16760	15898	10097	2945
Invertebrate unidentified	556495	625561	418512	177181	40009	70401	30665	25883	65462	120896	21253
Misc crabs	14432	21524	11774	10571	27967	14095	11052	11681	20216	18002	12095
Misc crustaceans	14	186	225	2325	1402	719	1335	935	539	501	571
Misc fish	95745	91469	66164	42470	70971	66422	48913	29256	40108	86012	31512
Misc inverts (worms etc)	20	123	25	50	46	152	170	105	181	97	227
Other osmerids	4258	4292	497	634	35770	9833	849	2830	2053	4692	956
Pacific Sand lance	9	167	97	33	17	37	15	35	395	170	21
Pandalid shrimp	216	920	115	772	101	305	494	744	2273	606	1980
Polychaete unidentified	16	68	42	360	69	175	75	102	212	53	178
Scypho jellies	111900	299034	115550	46785	42346	146153	222944	152367	309001	176728	206560
Sea anemone unidentified	6087	6202	2581	4896	8791	24840	25572	20526	14668	5965	22626
Sea pens whips	9	28	164	3	12	324	185	635	20	67	39
Sea star	1939624	1865768	1606948	1308482	1456620	1831017	684867	791632	1662779	1728950	500173
Snails	118257	191064	69769	141517	95876	139765	58354	57060	74718	33708	23106
Sponge unidentified	11434	6807	12205	3118	405	6721	69506	16623	11312	12384	13331
Stichaeidae	72	32		10	784	239	10	171	384	136	147
Surf smelt					1						
urchins dollars cucumbers	2254	315	2549	845	3477	4897	7548	1278	987	754	667

Table 4.25--Yellowfin sole TAC and ABC levels, 1980- 2013

Year	TAC	ABC
1980	117,000	169,000
1981	117,000	214,500
1982	117,000	214,500
1983	117,000	214,500
1984	230,000	310,000
1985	229,900	310,000
1986	209,500	230,000
1987	187,000	187,000
1988	254,000	254,000
1989	182,675	241,000
1990	207,650	278,900
1991	135,000	250,600
1992	235,000	372,000
1993	220,000	238,000
1994	150,325	230,000
1995	190,000	277,000
1996	200,000	278,000
1997	230,000	233,000
1998	220,000	220,000
1999	207,980	212,000
2000	123,262	191,000
2001	113,000	176,000
2002	86,000	115,000
2003	83,750	114,000
2004	86,075	114,000
2005	90,686	124,000
2006	95,701	121,000
2007	136,000	225,000
2008	225,000	248,000
2009	210,000	210,000
2010	219,000	219,000
2011	196,000	239,000
2012	202,000	203,000
2013	198,000	206,000

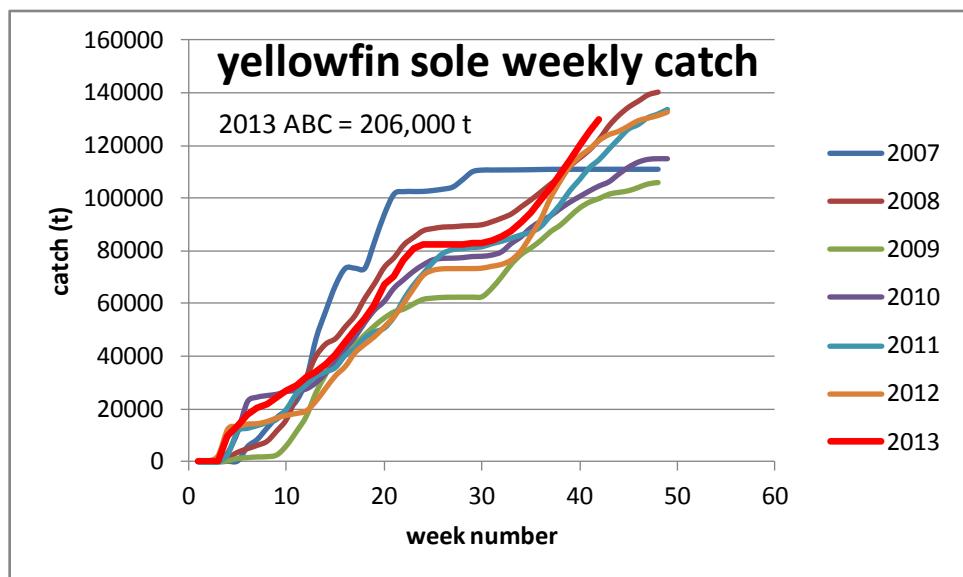
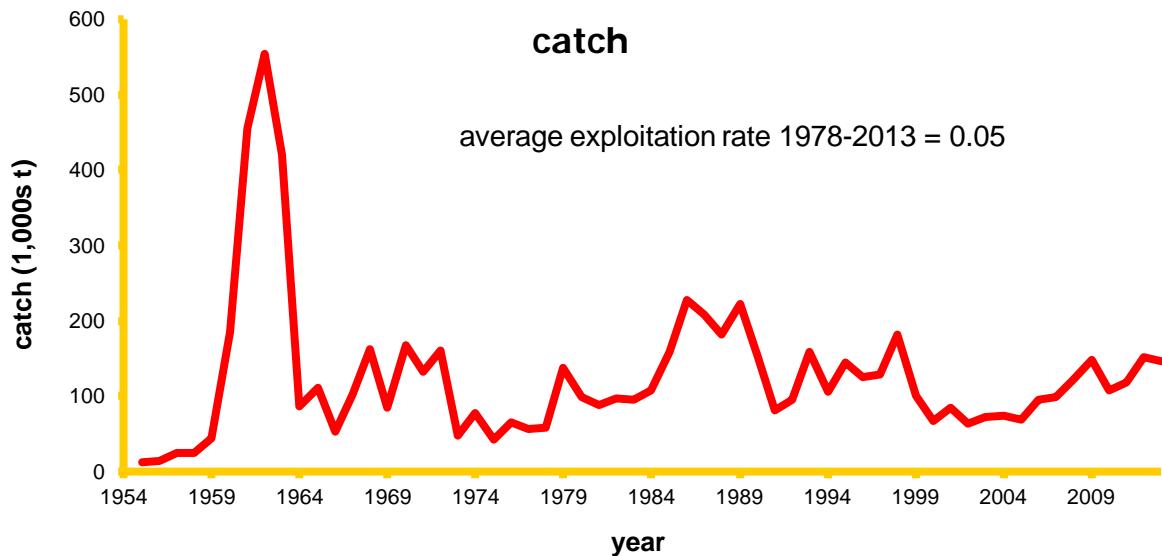


Figure 4.1—Yellowfin sole annual catch (1,000s t) in the Eastern Bering Sea from 1954-2013 (top panel) and catch by week from 2007 – September 2013 (bottom panel).

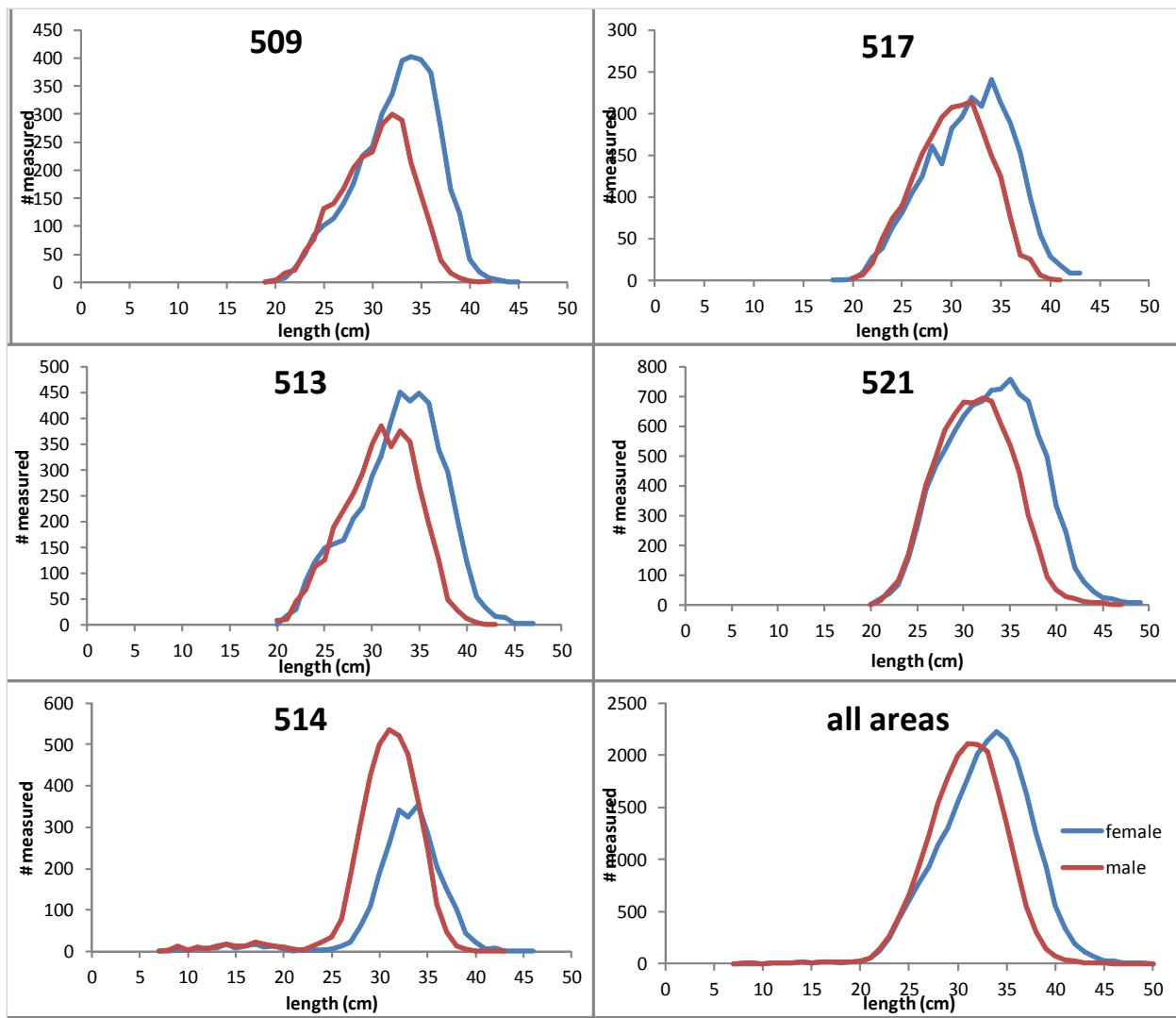
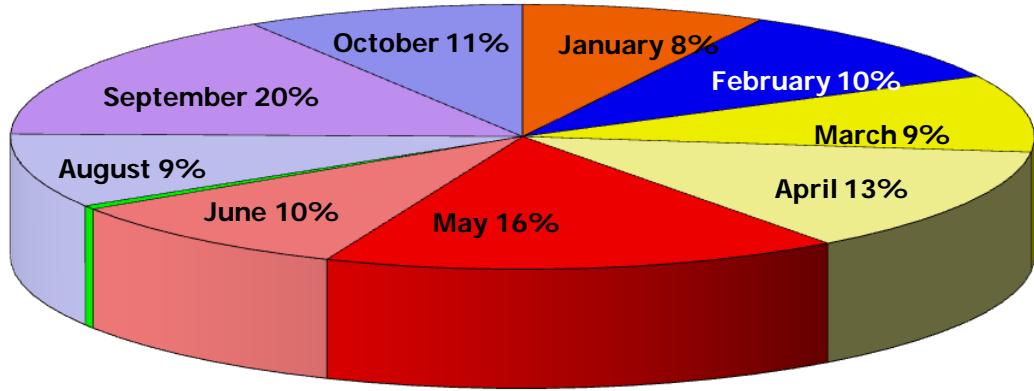


Figure 4.2--Size composition of the yellowfin sole catch in 2013 (through September), by subarea and total.

**yellowfin sole catch by month in 2013 through October**



**yellowfin sole catch by area in 2013  
(through September)**

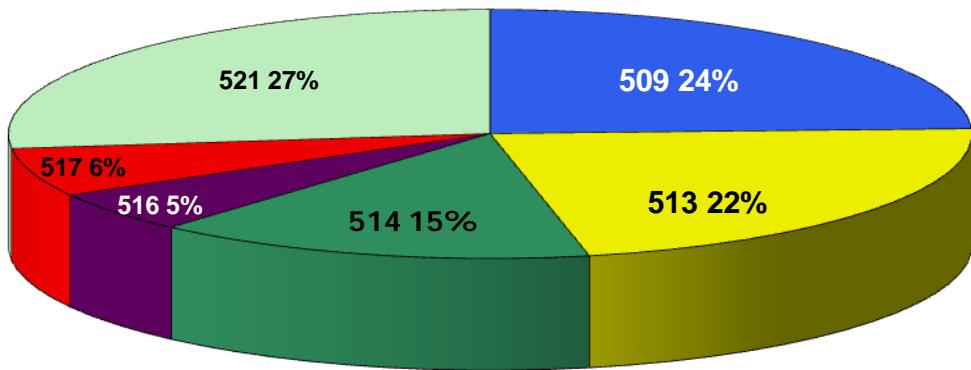


Figure 4.3     Yellowfin sole catch by month and area in the Eastern Bering Sea in 2013.

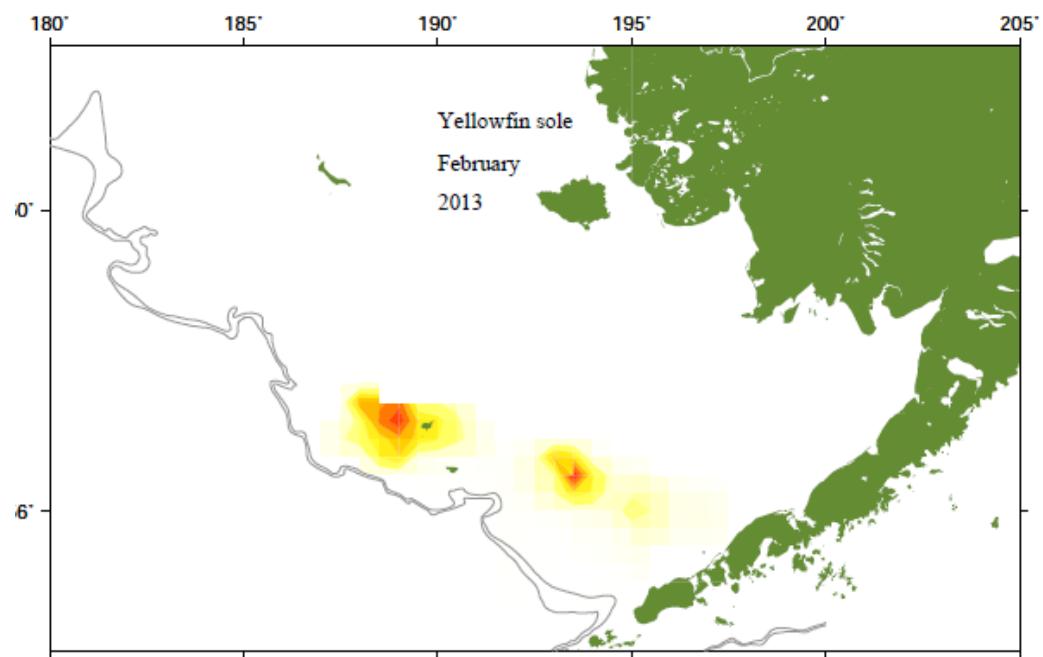
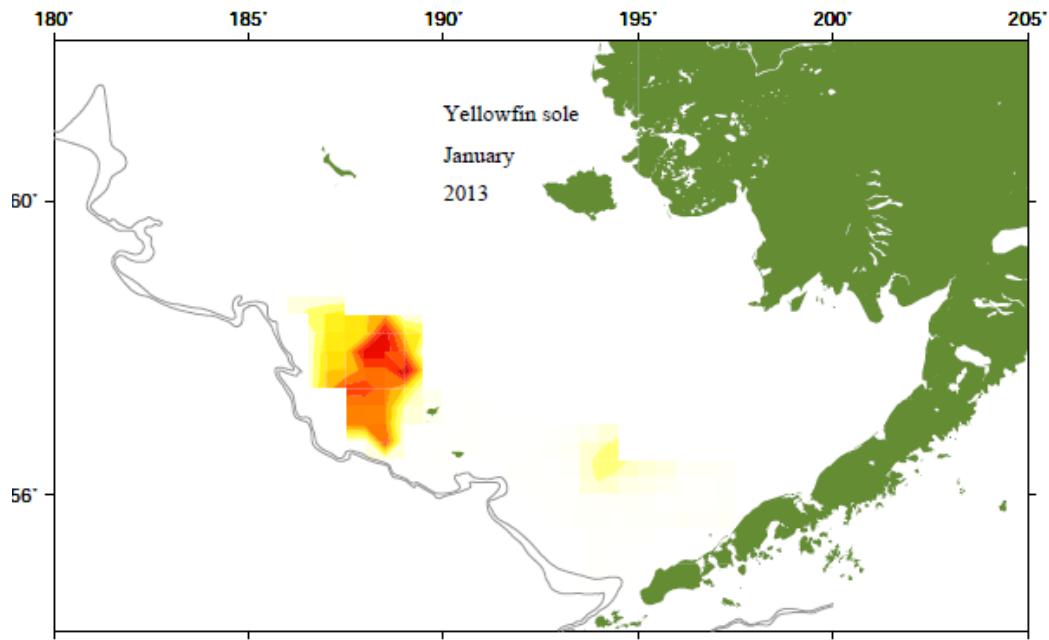


Figure 4.4—General fishery locations by month.

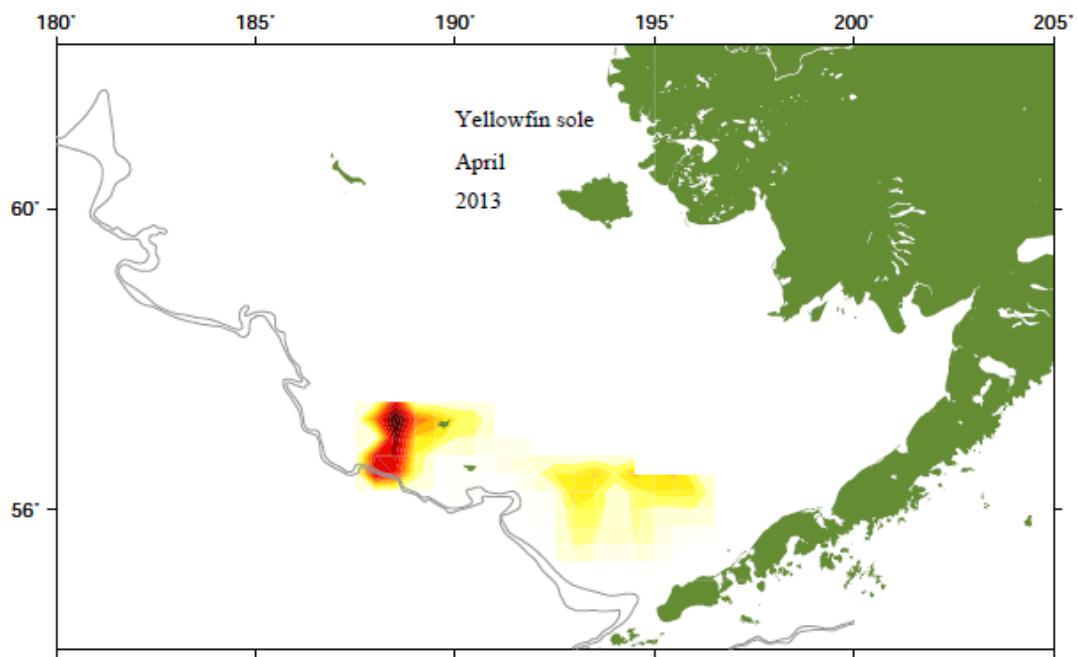
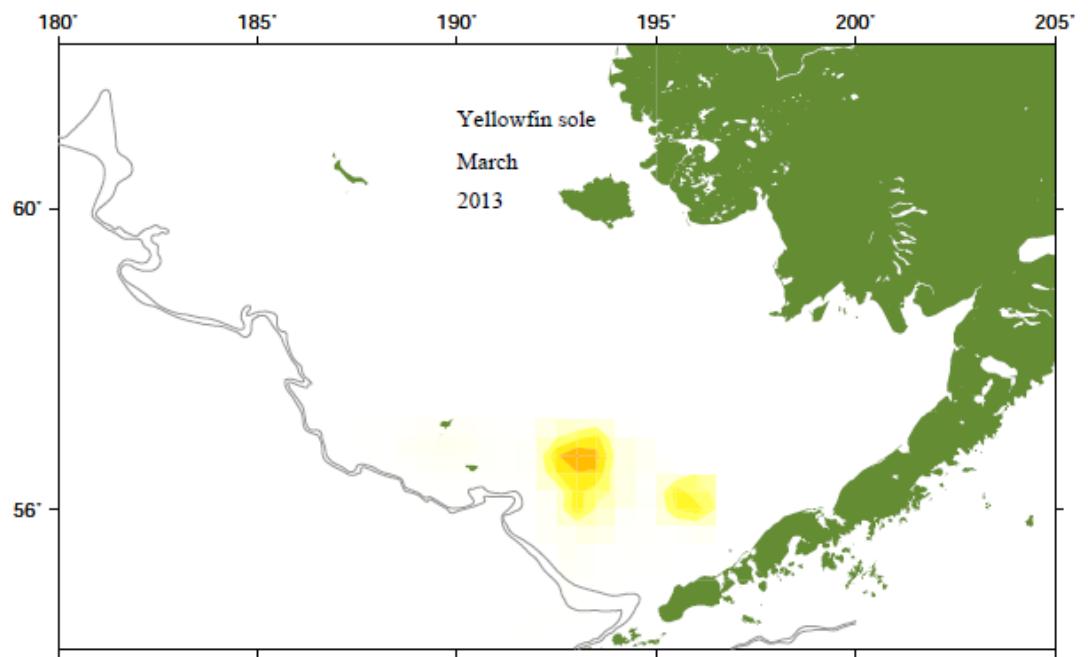


Figure 4.4—(continued).

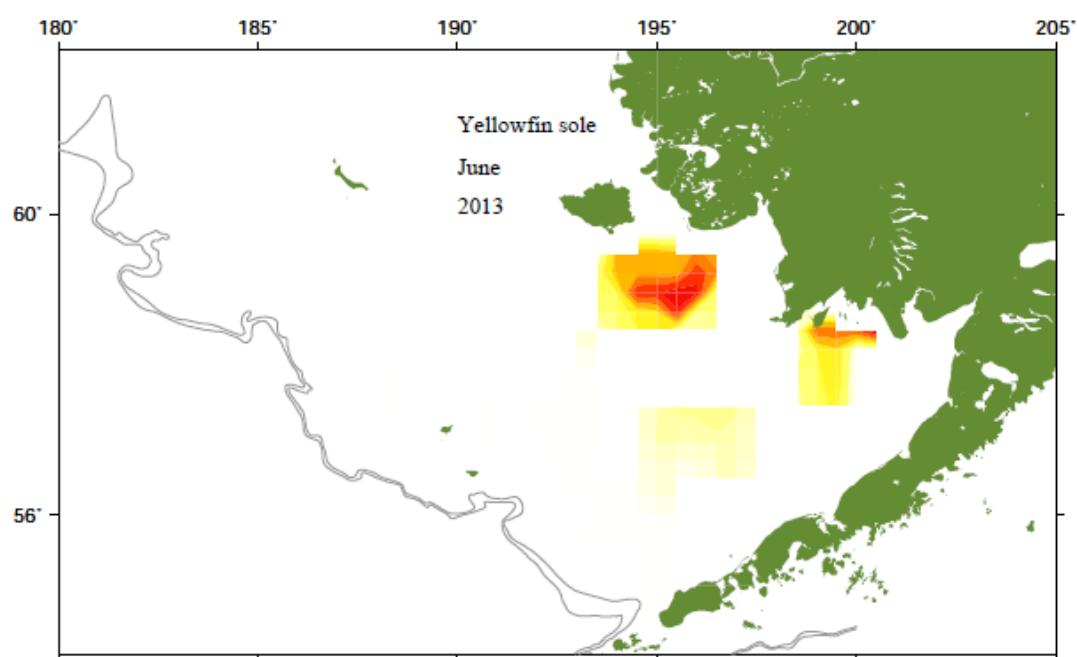
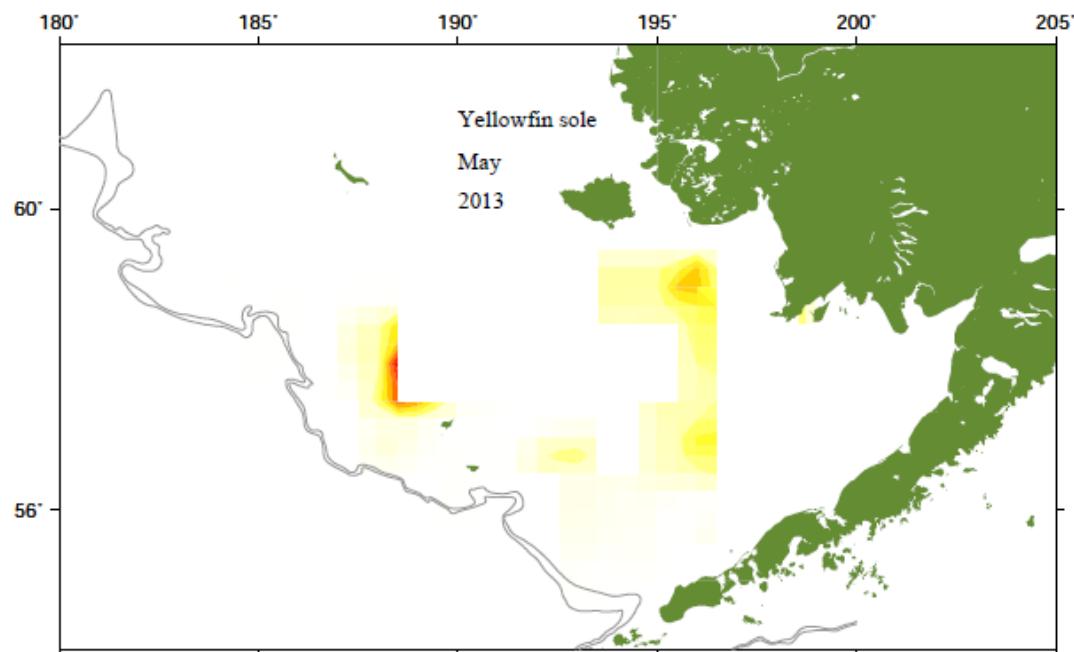


Figure 4.4—(continued).

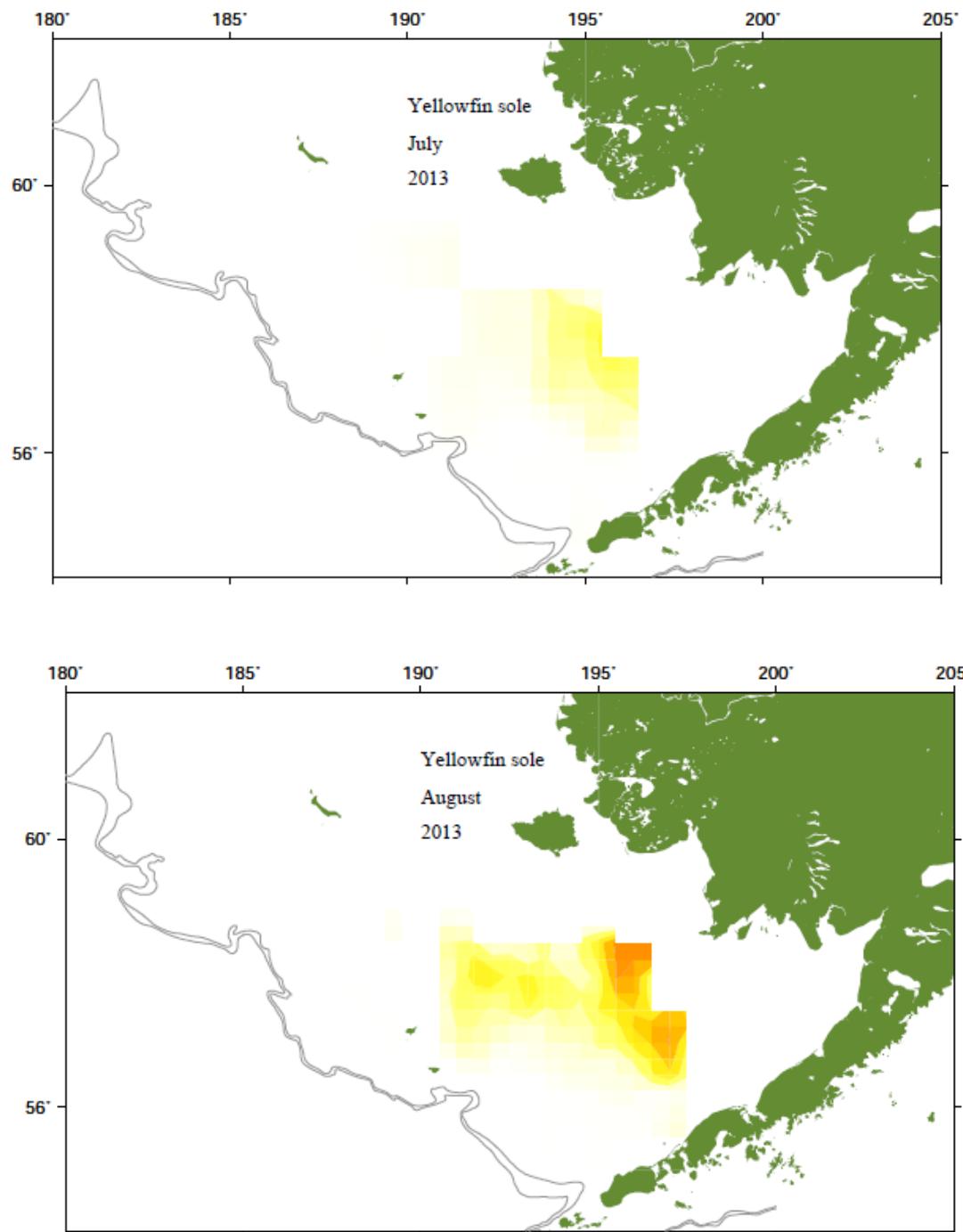


Figure 4.4—(continued).

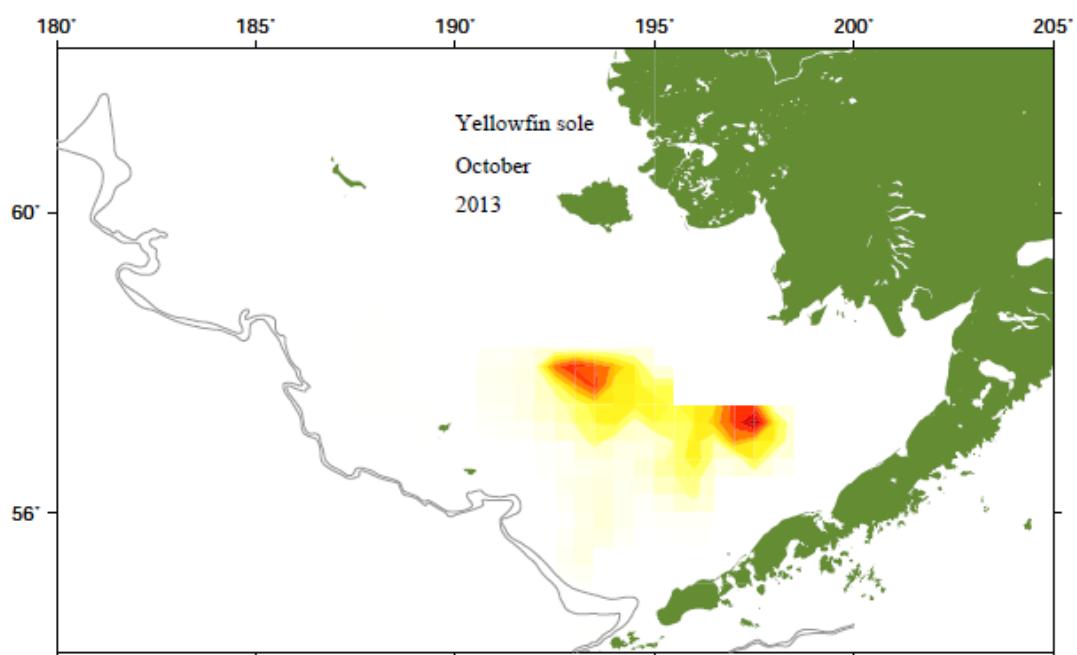
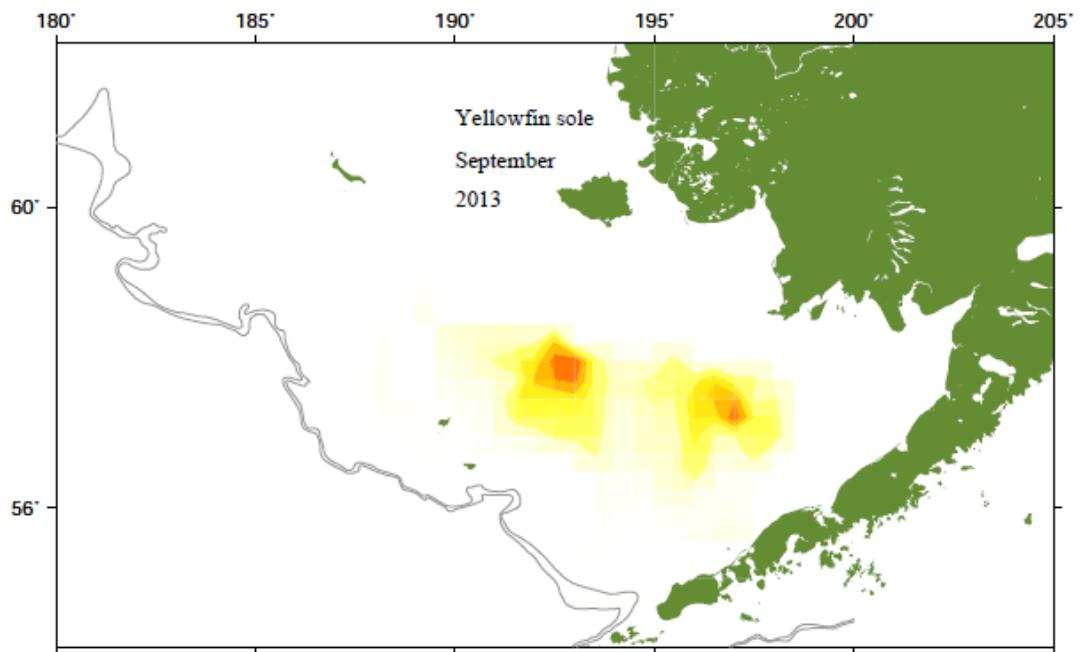


Figure 4.4— (continued).

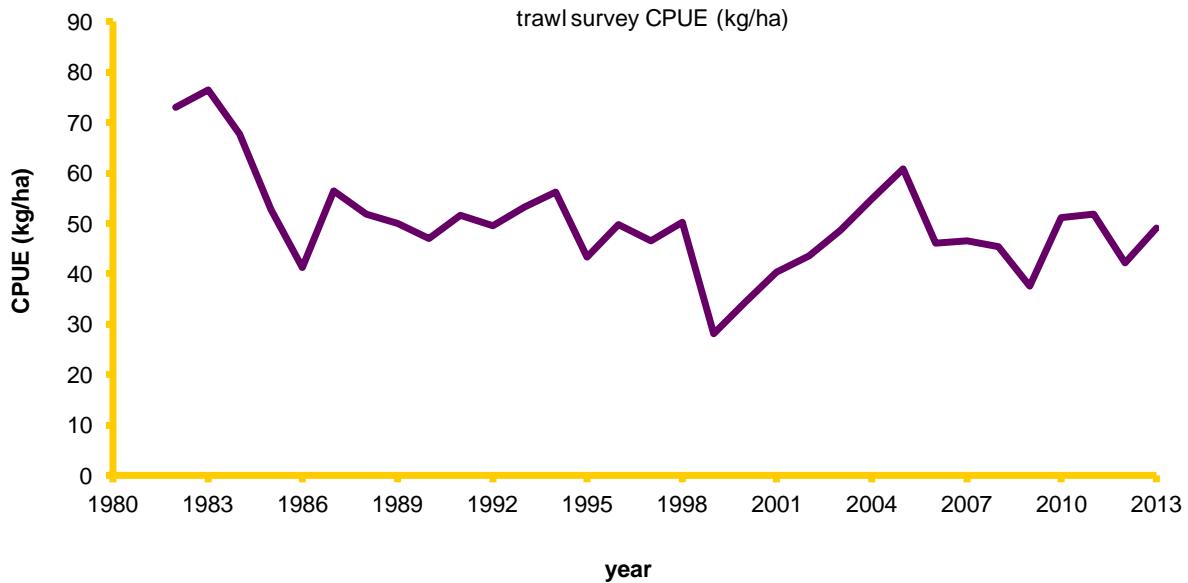


Figure 4.5.--Yellowfn sole CPUE (catch per unit effort in kg/ha) from the annual Bering Sea shelf trawl surveys, 1982-2013.

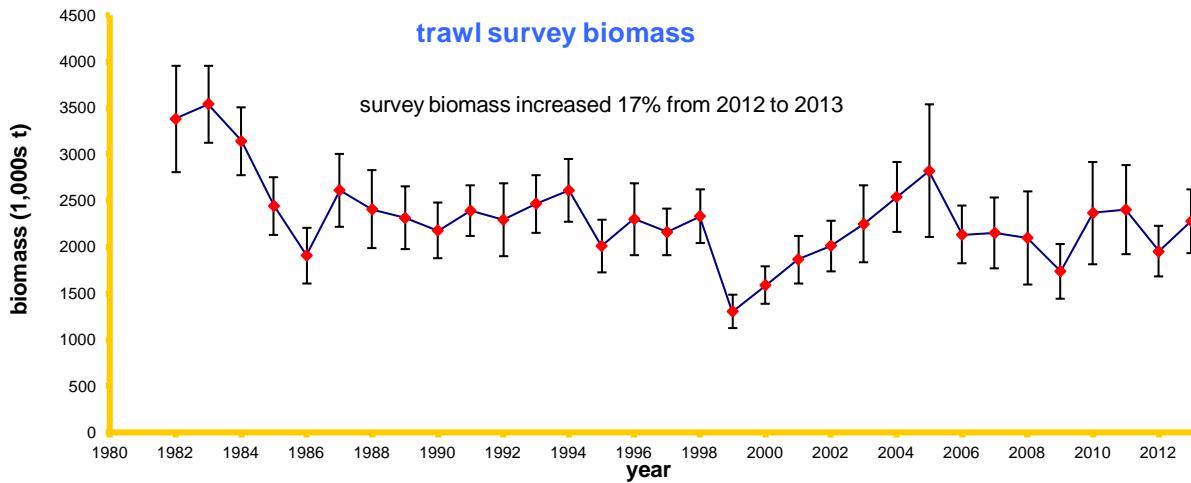


Figure 4.6.--Annual bottom trawl survey biomass point-estimates and 95% confidence intervals for yellowfin sole, 1982-2013.

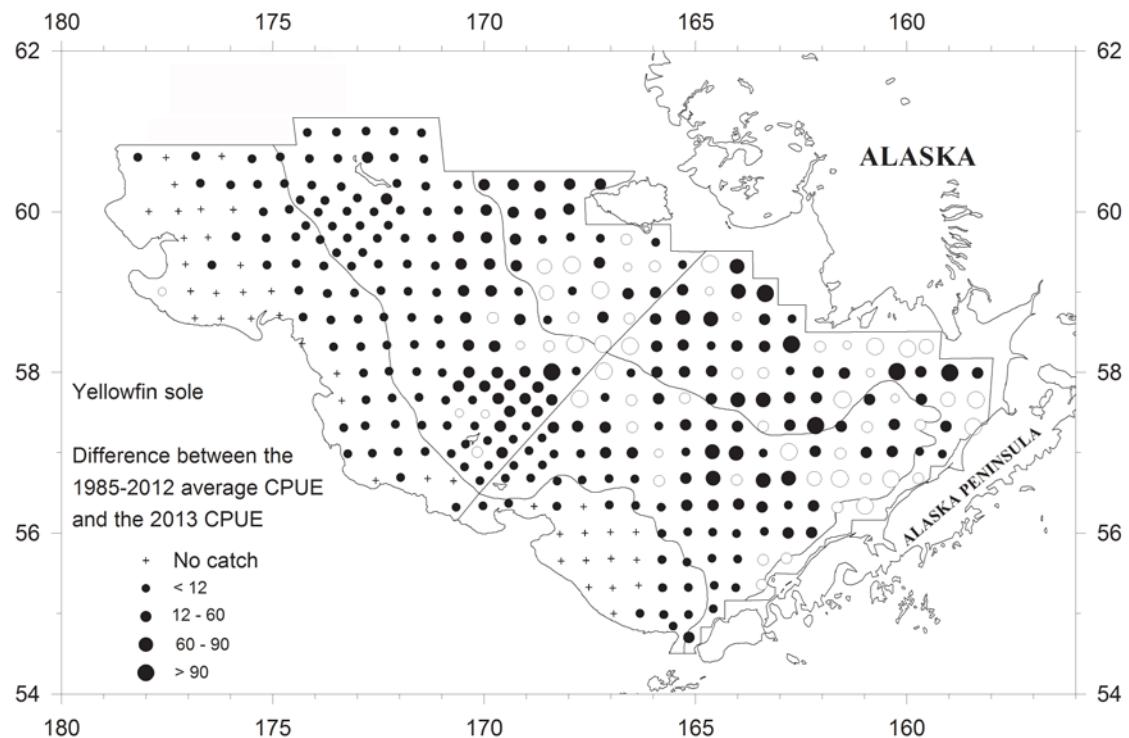


Figure 4.7.--Difference between the 1985-2012 average trawl survey CPUE for yellowfin sole and the 2013 survey CPUE. Open circles indicate that the magnitude of the catch was greater in 2013 than the long-term average, closed circles indicate the catch was greater in the long-term average than in 2013.

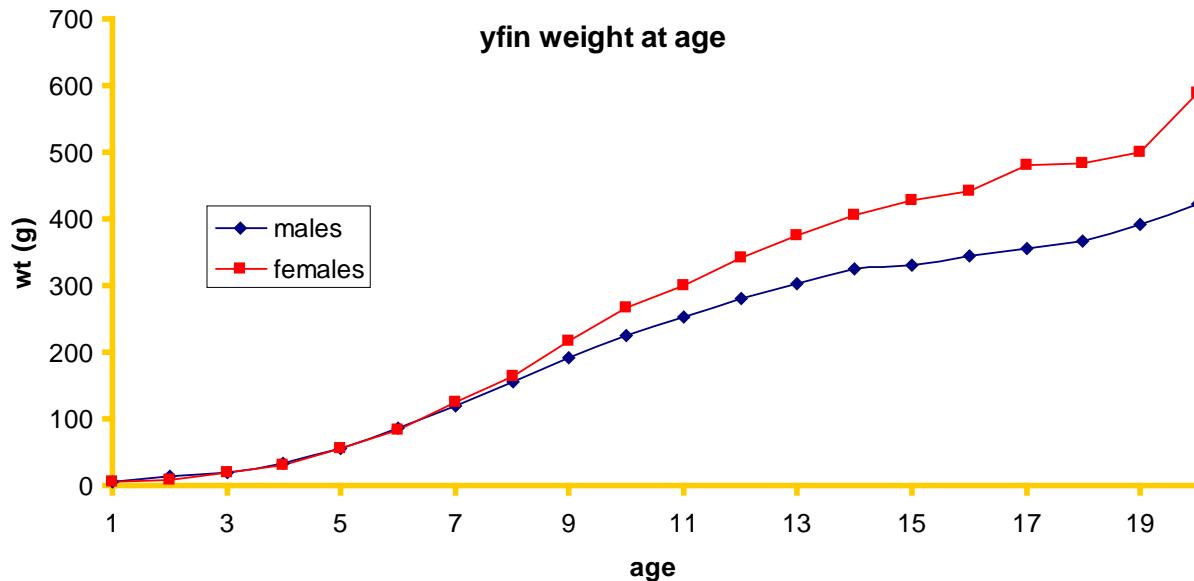


Figure 4.8--Estimates of average yellowfin sole weight-at-age (g) from trawl survey observations.

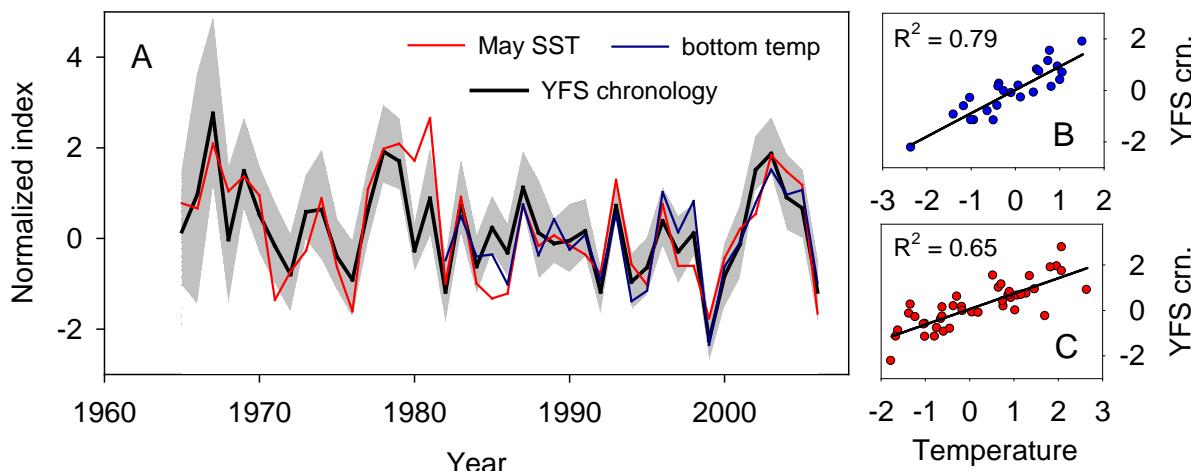


Figure 4.9--Master chronology for yellowfin sole and time series of mean summer bottom temperature and May sea surface temperature for the southeastern Bering Sea (Panel A). All data re normalized to a mean of 0 and standard deviation of 1. Correlations of chronologies with bottom temperature and sea surface temperature are shown in panels B and C, respectively. From Matta et al. 2010.

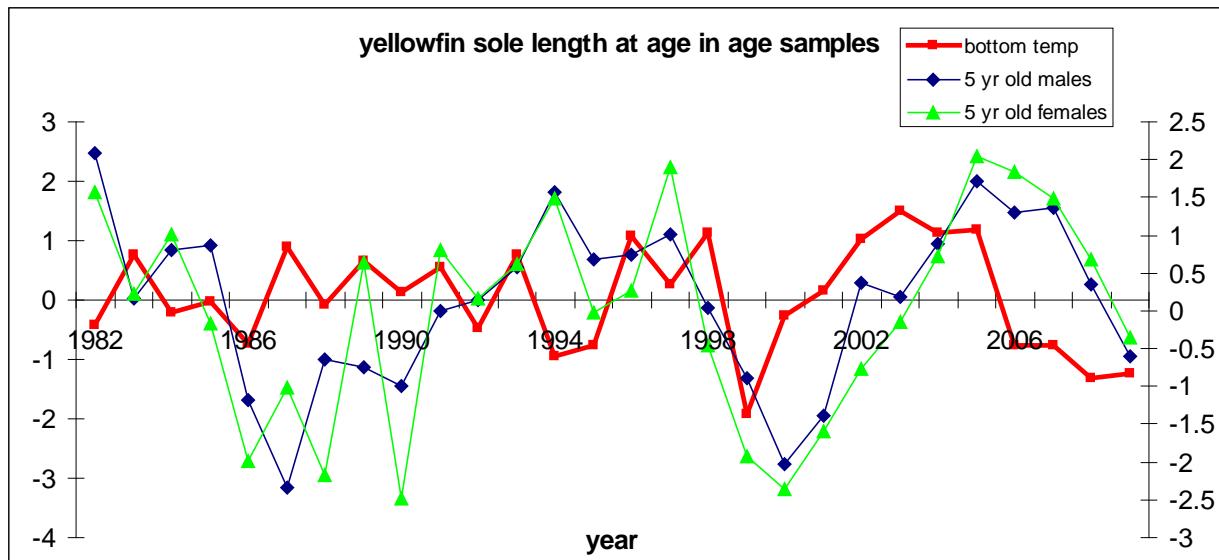


Figure 4.10—Yellowfin sole length-at-age anomalies, for males and females, and bottom temperature anomalies. Correspondence in these residuals is apparent with a 2–3 year lag effect from the mid-1990s to 2009. Late 1980s and early 1990s pattern may be a density-dependent response in growth from the large 1981 and 1983 year-classes.

#### temperature-catchability model result

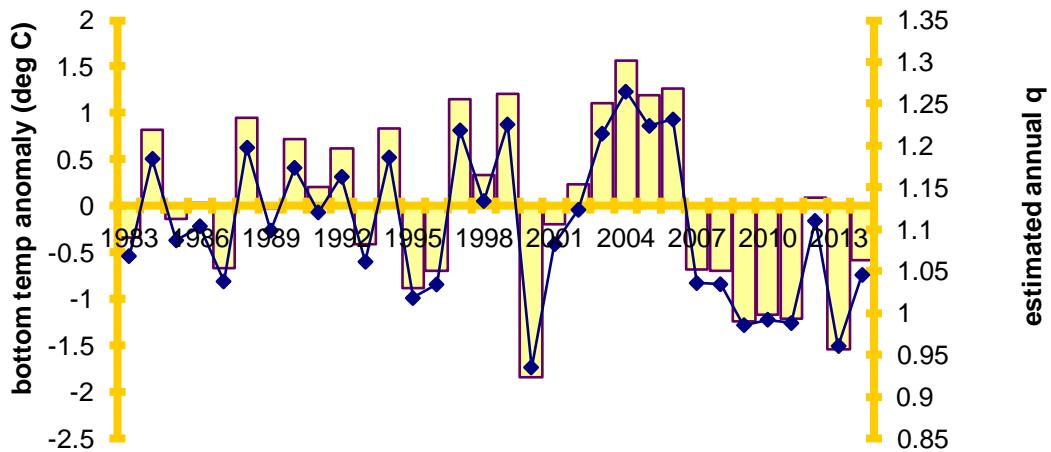


Figure 4.11--Average bottom water temperature from stations less than or equal to 100 m in the Bering Sea trawl survey (bars) and the stock assessment model estimate of  $q$  for each year 1982–2013.

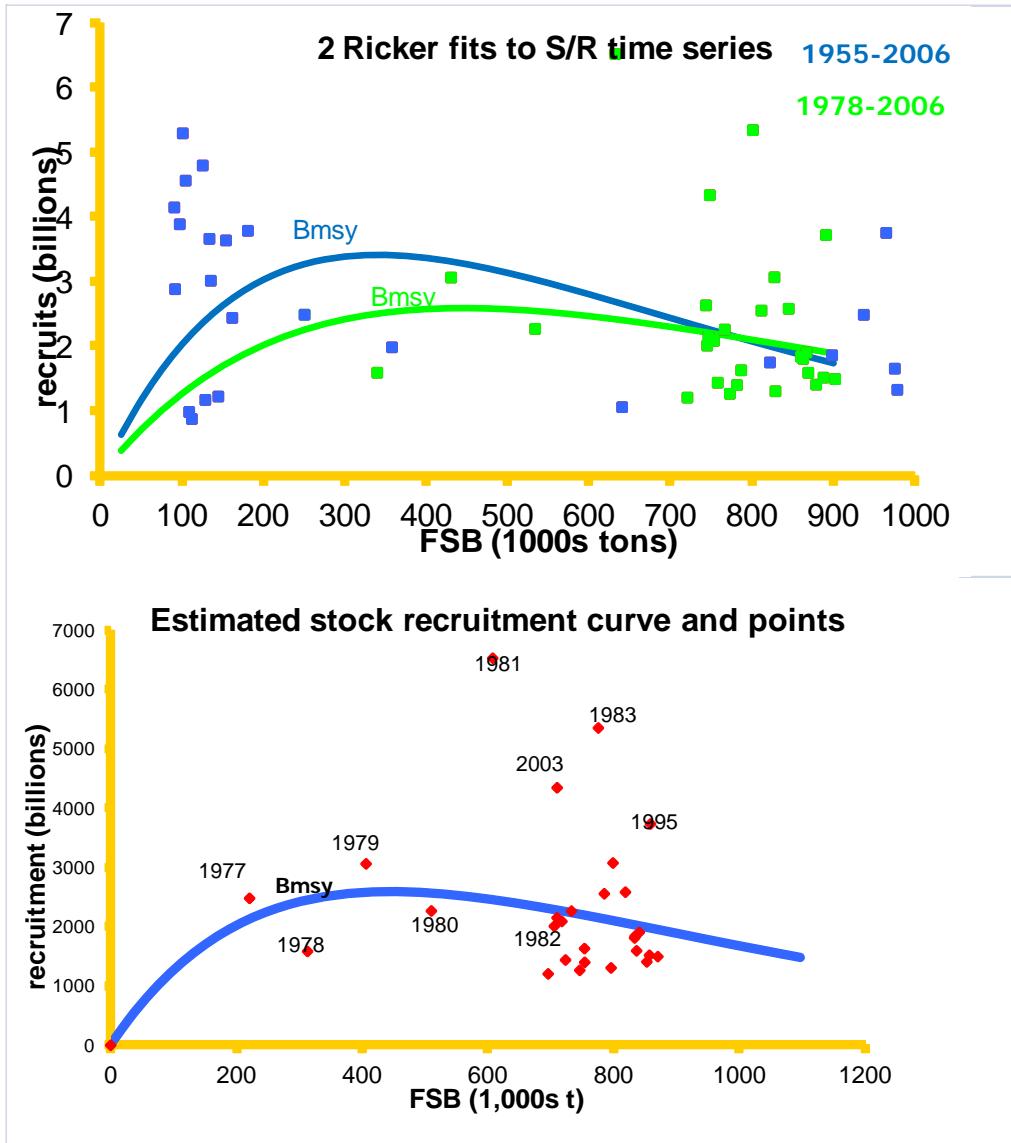


Figure 4.12-Fit of the Ricker (1958) stock recruitment model to two distinct stock recruitment time-series data sets (top panel), and the fit to the assessment preferred model (model B, lower panel).

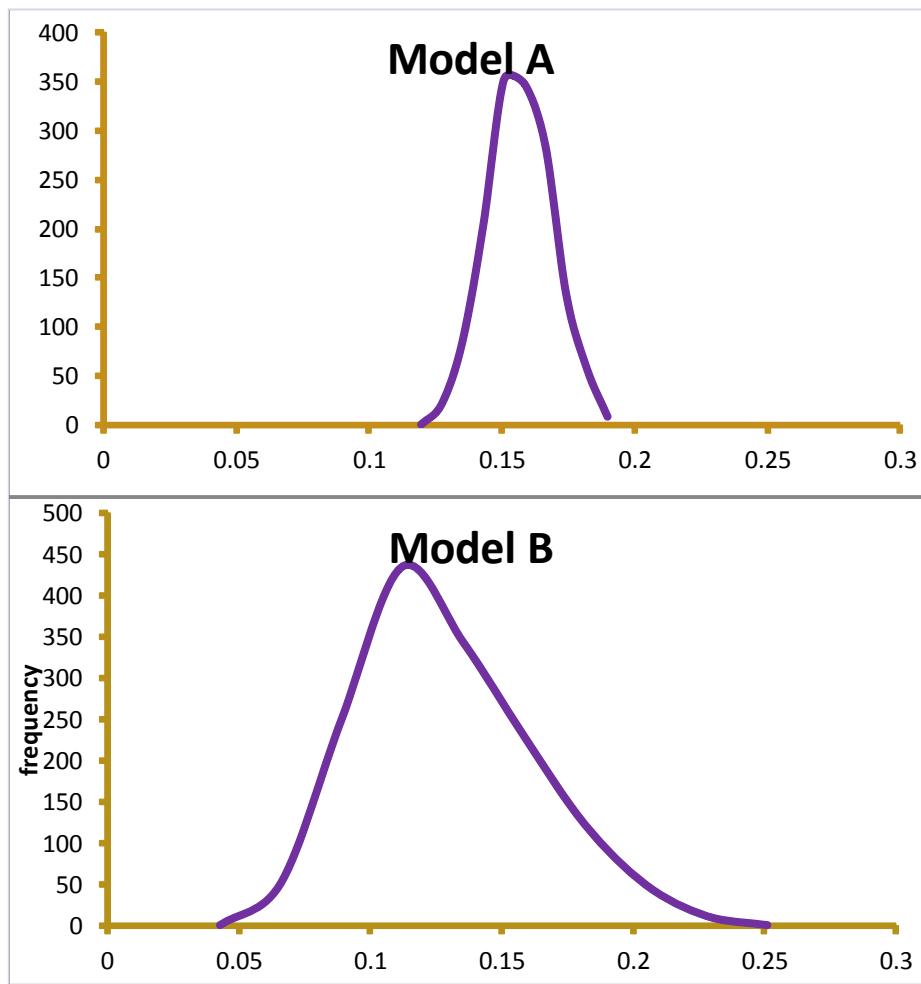


Figure 4.13--Posterior distributions of  $F_{m\text{sy}}$  for three models considered in the stock productivity analysis.

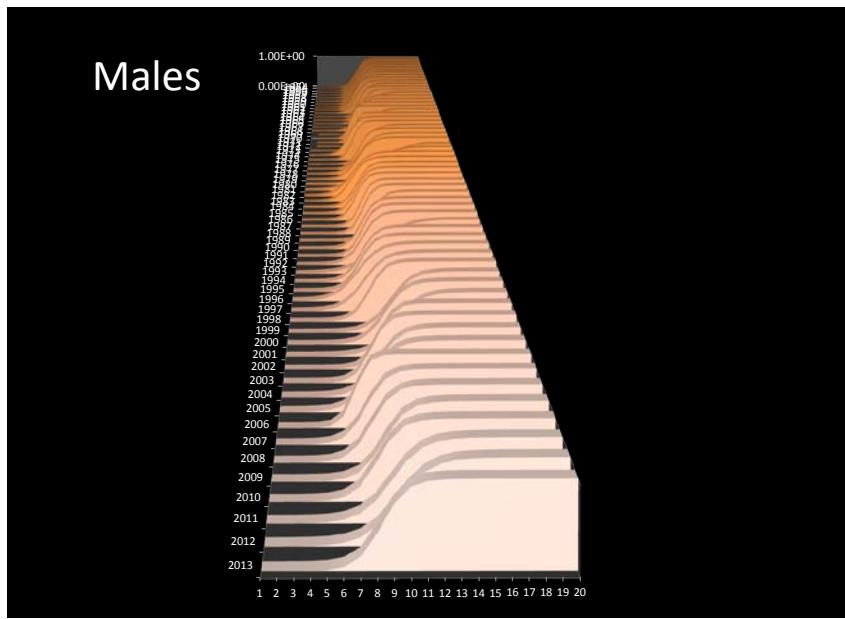


Figure 4.14a--Estimated male fishery selectivity by age and year.

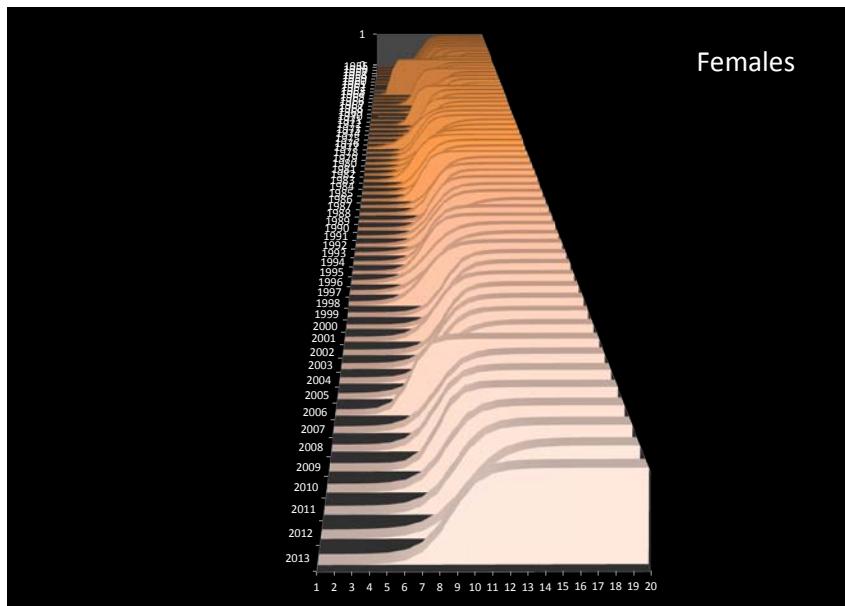


Figure 4.14b.--Estimated female fishery selectivity by age and year.

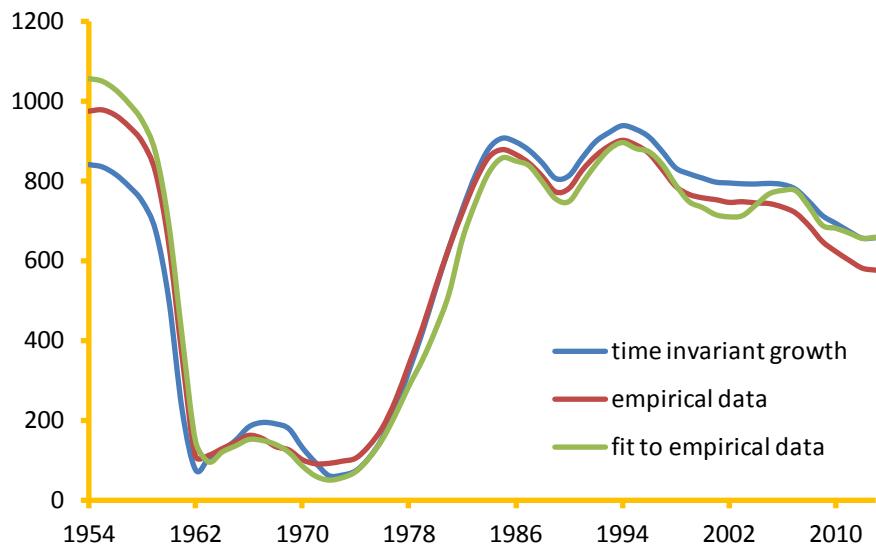


Figure 4.15—Estimates of female spawning biomass (1978-2013) from the 3 growth models considered in the assessment (Model 3, did not converge and is not shown).

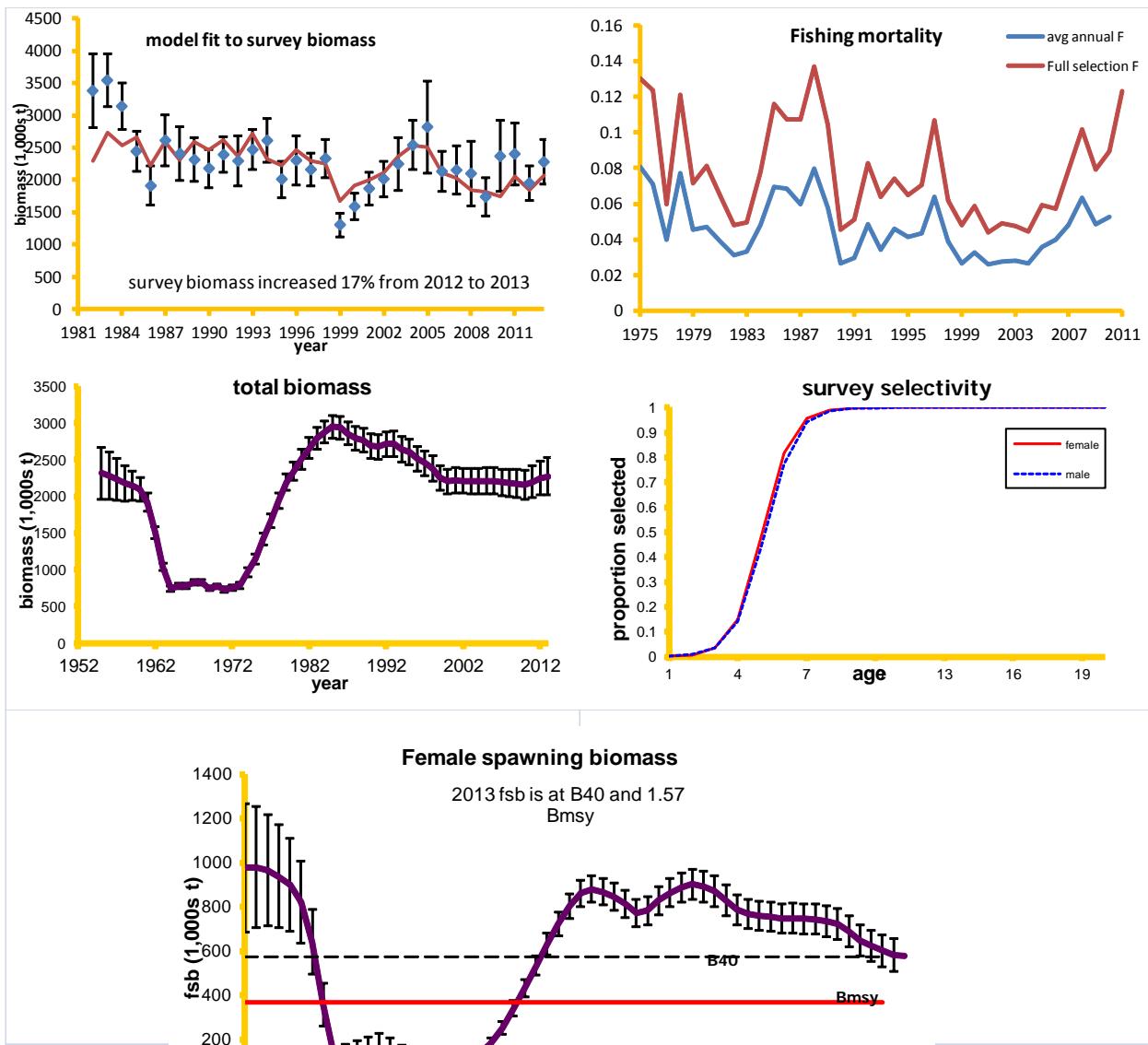


Figure 4.16.

Model fit to the survey biomass estimates (top left panel), model estimate of the full selection fishing mortality rate throughout the time-series (top right panel), model estimate of total biomass (middle left panel), the model estimate of survey selectivity (middle right panel) and the estimate of female spawning biomass (bottom left panel).

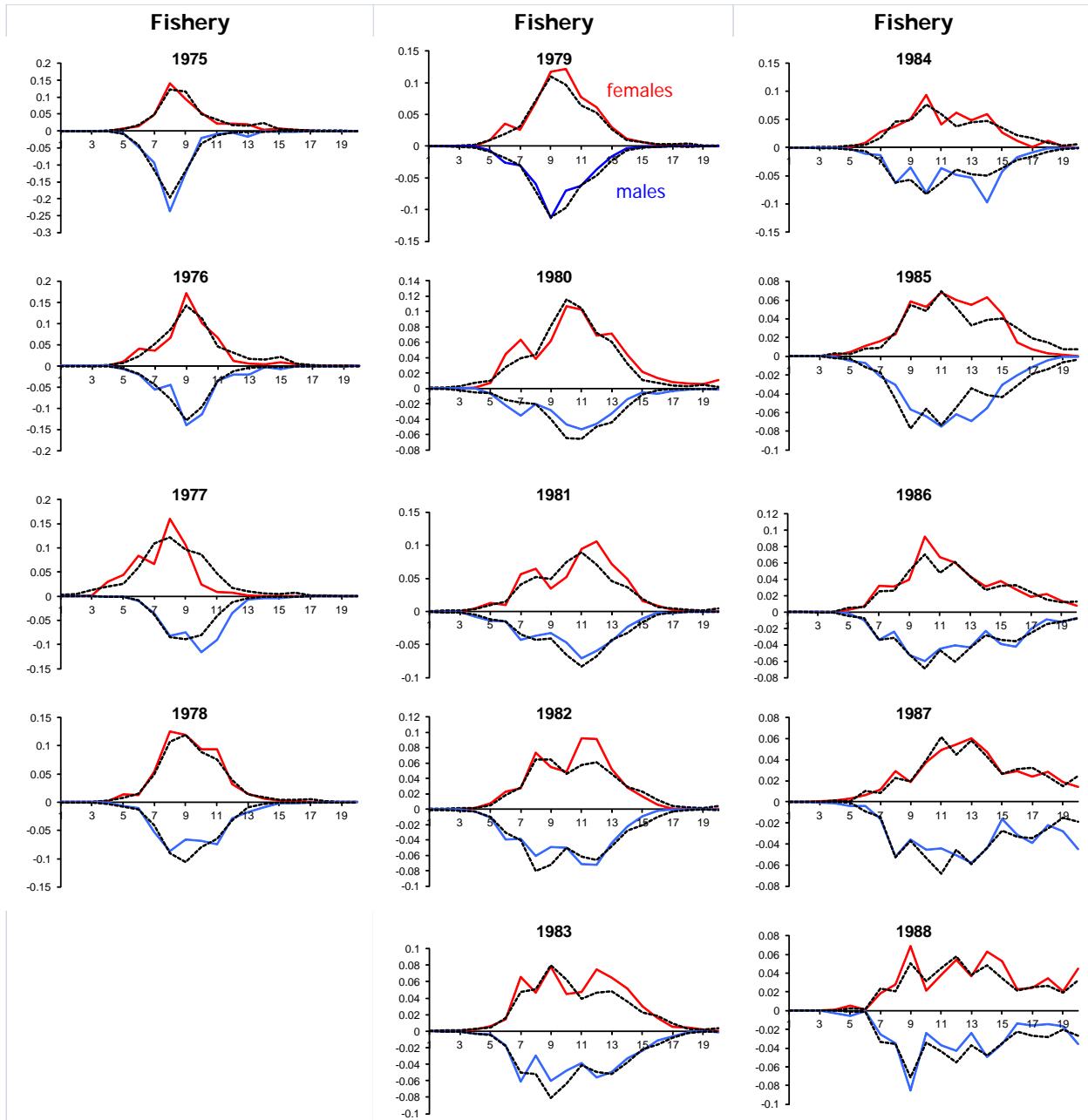


Figure 4.17. Stock assessment model fit to the time-series of fishery and survey age composition, by sex.

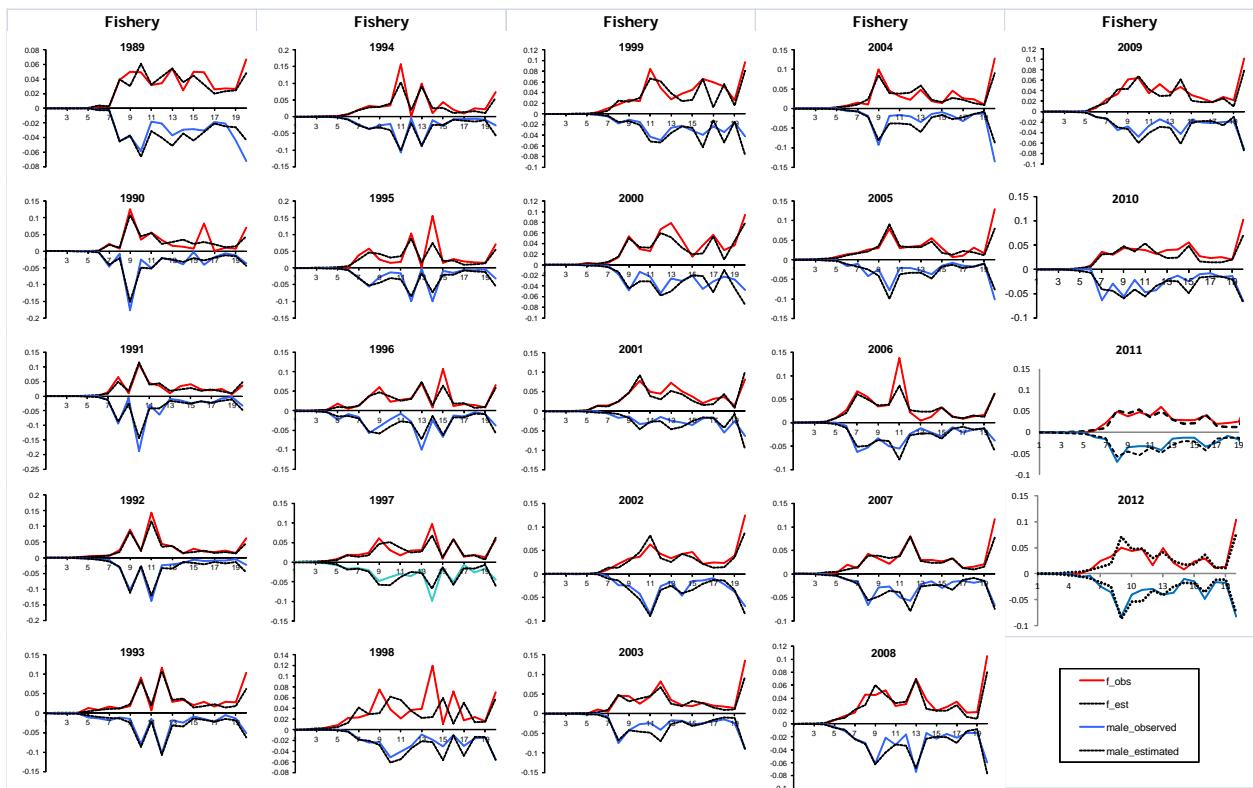


Figure 4.17 (continued).

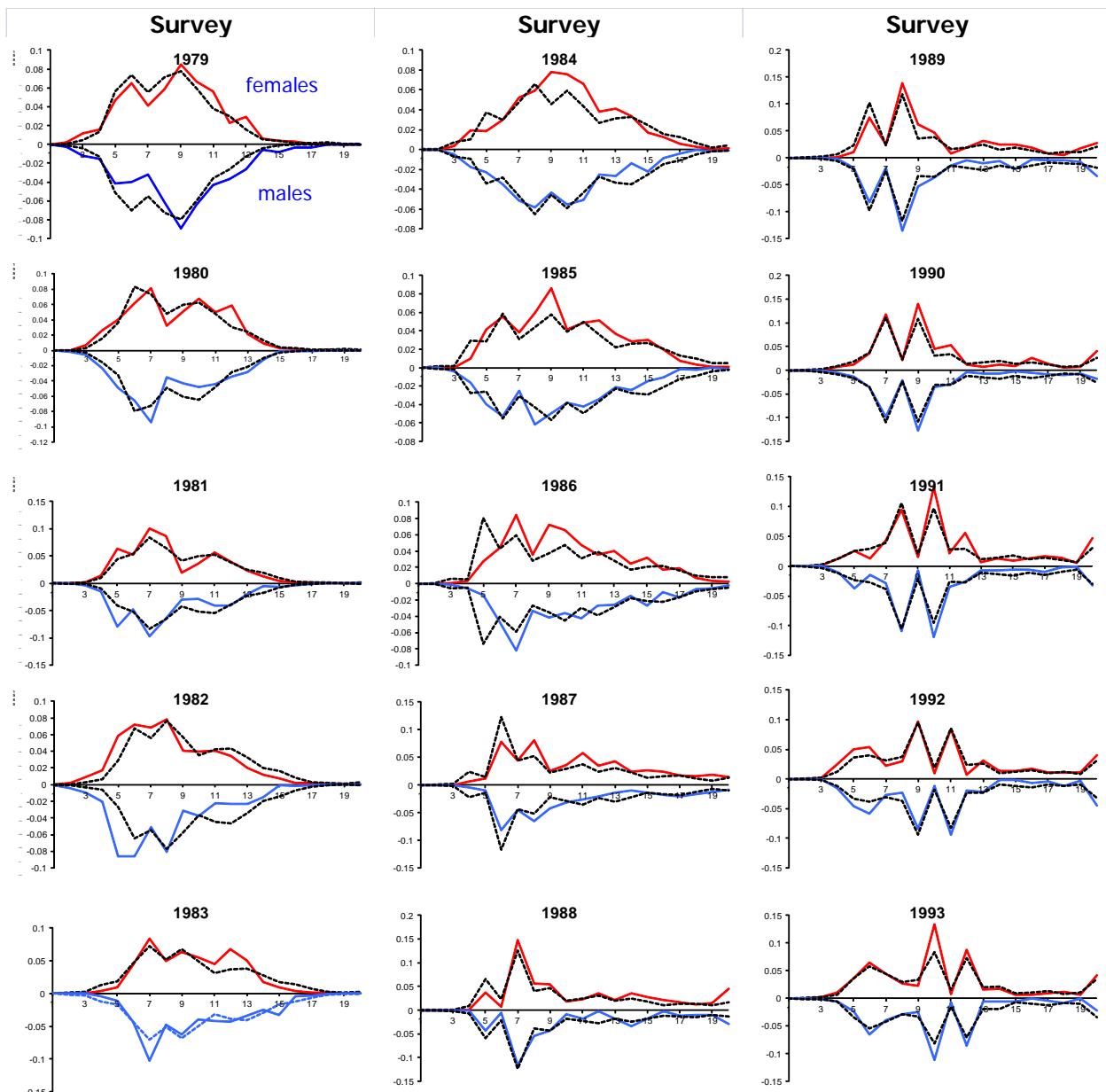


Figure 4.17 (continued).

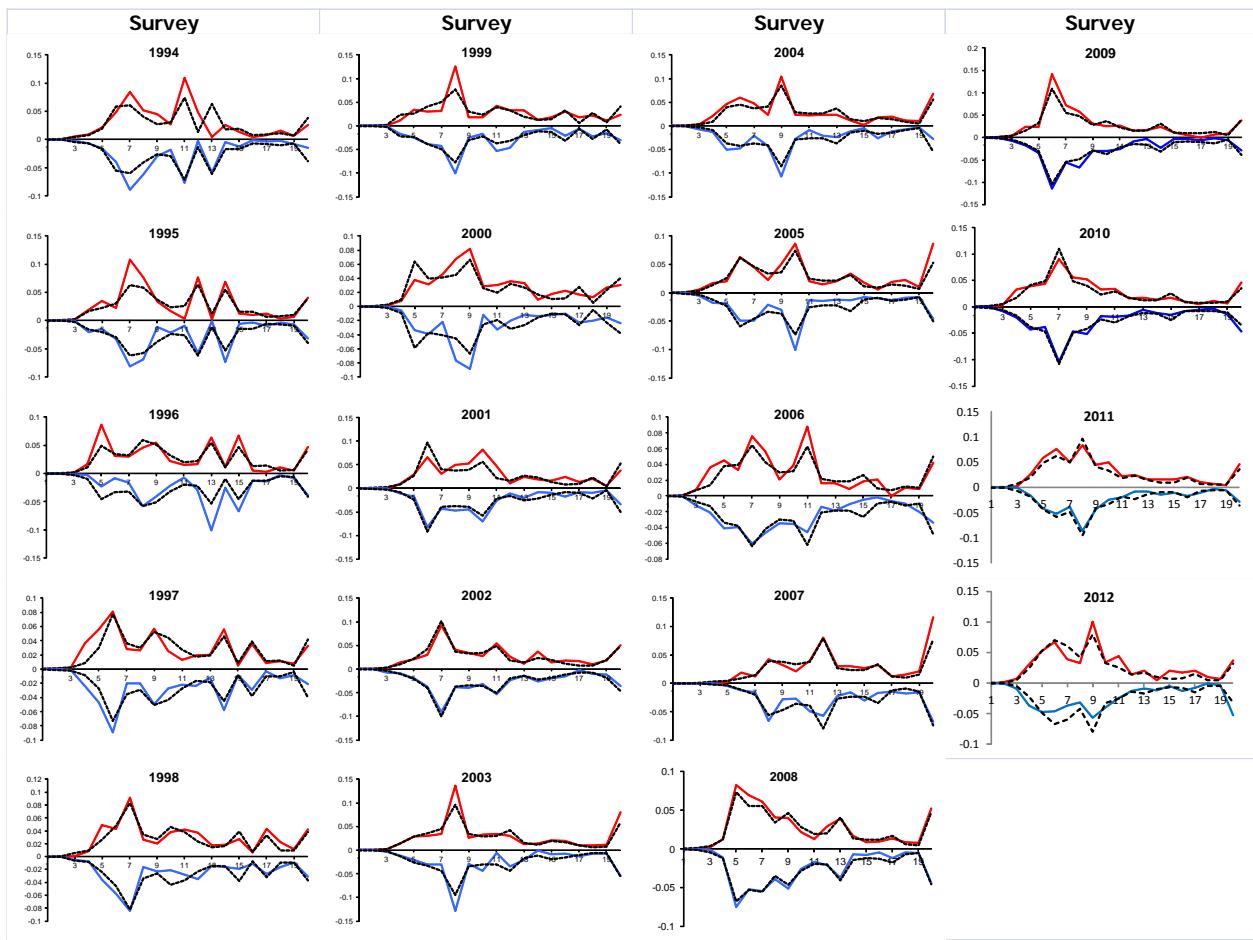


Figure 4.17 (continued).

**fit to survey with fixed q vs temp\_q**

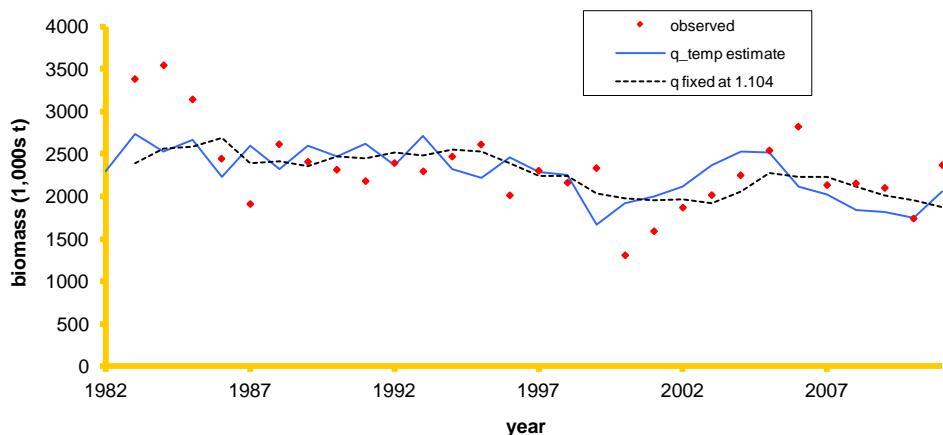


Figure 4.18.--Comparison of the fit to the survey biomass using a fixed  $q$  and the  $q$ -bottom temperature relationship.



Figure 4.19--Year class strength of age 5 yellowfin sole estimated by the stock assessment model. The dotted line is the average of the estimates from 56 years of recruitment.

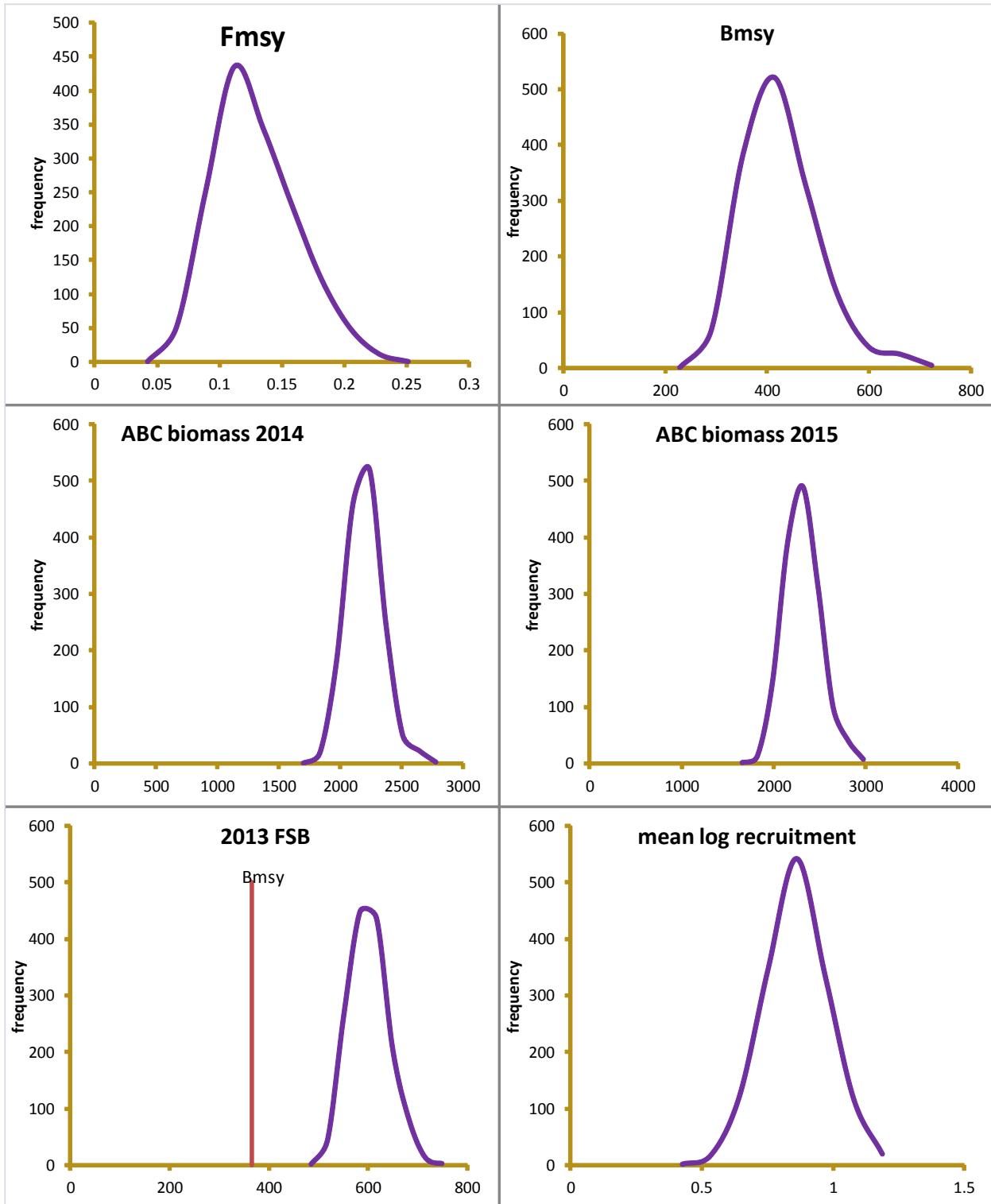


Figure 4.20.--Posterior distributions of some important parameters estimated by the preferred stock assessment model (from mcmc integration).

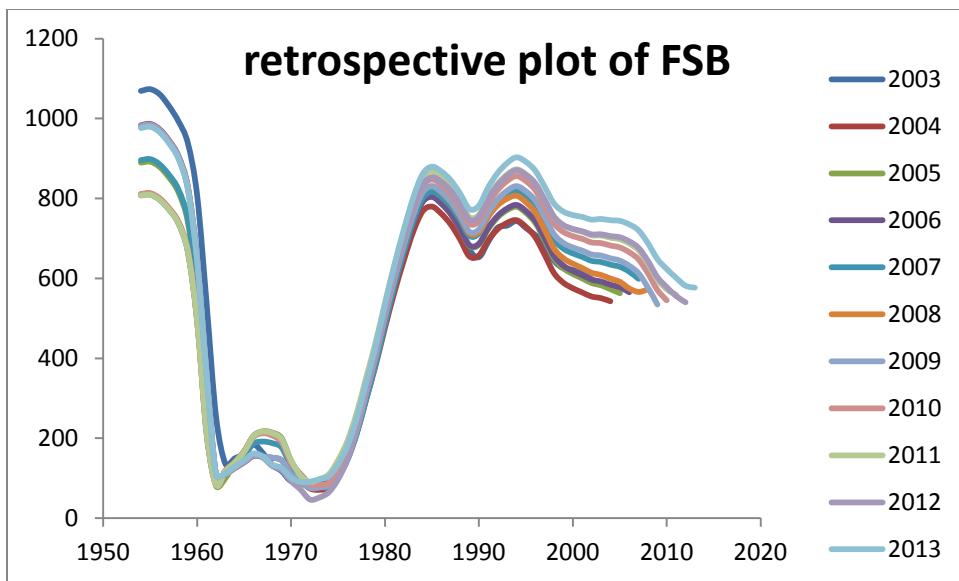


Figure 4.21—Retrospective plot of yellowfin sole female spawning biomass (1,000s t), 2003-2013, from the recommended assessment model.

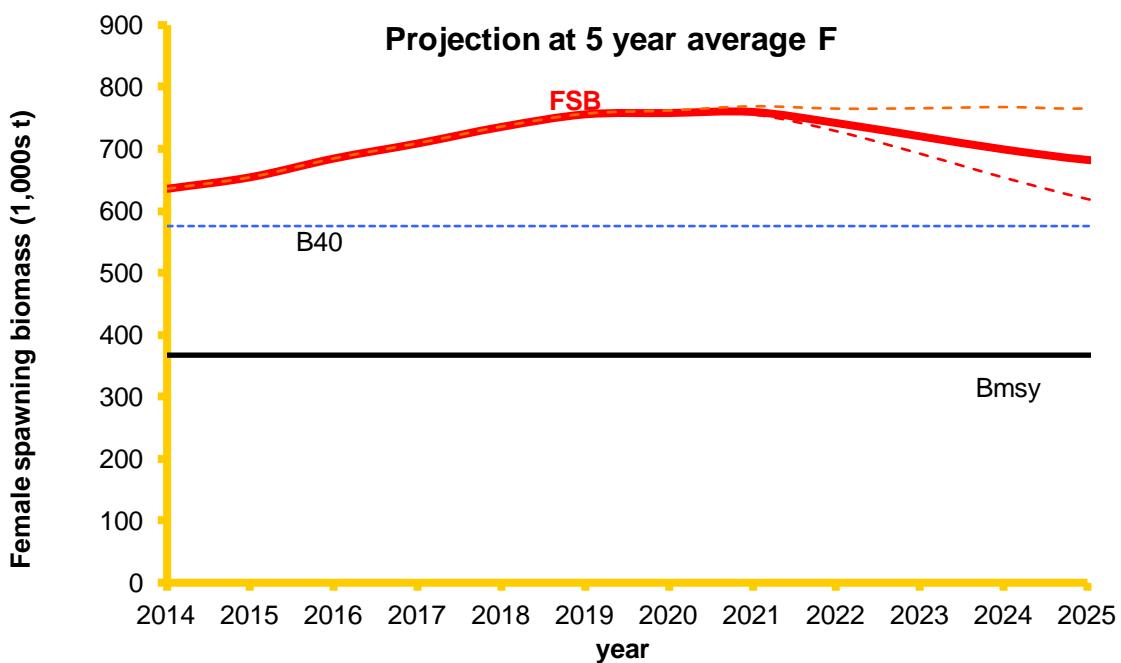


Figure 4.22--Projection of yellowfin sole female spawning biomass (1,000s t) at the average full-selection F from the past 5 years (0.104) through 2025 with  $B_{40\%}$  and  $B_{msy}$  levels indicated.

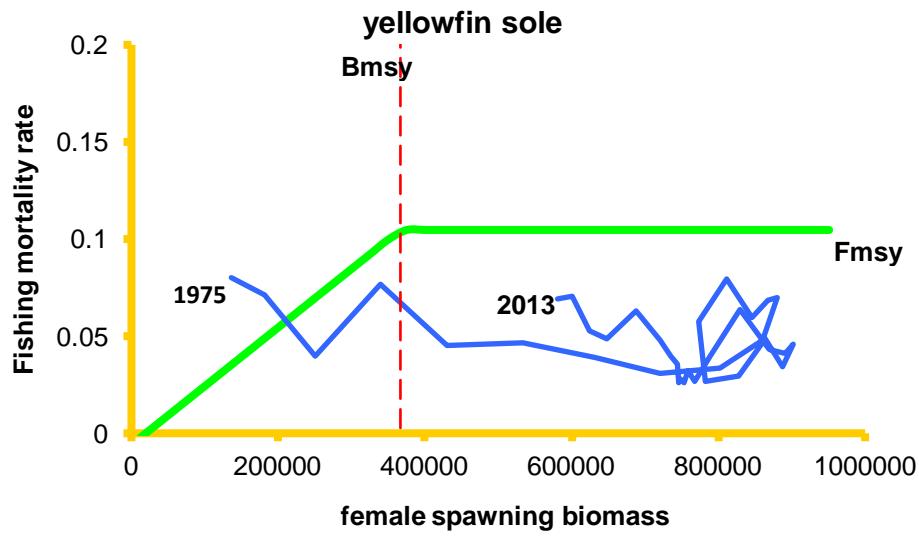


Figure 4.23.--Phase plane figure of the time-series of yellowfin sole female spawning biomass relative to the harvest control rule with 1975 and 2013 indicated.

## Appendix

IPHC research catch of yellowfin sole		
	number	weight (kg)
2007	707	502
2008	0	0
2009	0	0
2010	898	741