# Chapter 2: Assessment of the Pacific cod (Gadus macrocephalus) stock in the Gulf of Alaska for 2013 

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## EXECUTIVE SUMMARY

## Summary of Changes in Assessment Inputs

Relative to the November 2011 assessment, the following changes have been made in the current assessment:

Changes in the Input Data

1. Fishery: 2011 and preliminary 2012 total Pacific cod catch;
2. Fishery: 2011 observer data for Pacific cod catch-at-length;
3. Fishery: seasonal- and gear-specific catch for 1991 through 2011 were updated;
4. Survey: 2011 Pacific cod age composition and mean size-at-age from the NMFS GOA bottom trawl survey

## Changes in Assessment Methodology

Several alternative model configurations were considered in this assessment. The full set of model configurations is shown below (Model A is last year's model):

| Model | A | B | 1 | 1 Q | 2 | 2 Q | 3 | 3 Q | 4 | 5 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Catchability value | 1.04 | 1.04 | 1.00 | 1.04 | 1.00 | 1.04 | 1.00 | 1.04 | 1.00 | 1.00 |
| Tail compression off? | no | yes | yes | yes | yes | yes | yes | yes | yes | yes |
| Sub-27 abundance/sizecomp data omitted? | no | no | no | no | yes | yes | no | no | yes | no |
| Sub-27 mean size at age data omitted? | no | no | no | no | yes | yes | yes | yes | yes | yes |
| 27-plus mean size at age data omitted? | no | no | no | no | no | no | yes | yes | yes | no |

The stock assessment model configurations were run with Stock Synthesis version 3.24i.

## Summary of Results

| Quantity | As estimated or specified last year for: |  | As estimated or recommended this year for: |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 2012 | 2013 | 2013 | 2014 |
| $M$ (natural mortality rate) | 0.38 | 0.38 | 0.38 | 0.38 |
| Tier | 3a | 3a | 3a | 3a |
| Projected total biomass (t) | 521,000 | 530,000 | 449,300 | 440,300 |
| Female spawning biomass (t) |  |  |  |  |
| Projected | 121,000 | 127,000 | 111,000 | 112,900 |
| $B_{100 \%}$ | 261,000 | 261,000 | 234,800 | 234,800 |
| $\mathrm{B}_{40 \%}$ | 104,000 | 104,000 | 93,900 | 93,900 |
| $B_{35 \%}$ | 91,400 | 91,400 | 82,100 | 82,100 |
| $F_{\text {OFL }}$ | 0.53 | 0.53 | 0.61 | 0.61 |
| $\operatorname{maxF}_{\text {ABC }}$ | 0.44 | 0.44 | 0.49 | 0.49 |
| $F_{\text {ABC }}$ | 0.44 | 0.44 | 0.49 | 0.49 |
| OFL (t) | 104,000 | 108,000 | 97,200 | 101,100 |
| $\operatorname{maxABC}(\mathrm{t})$ | 87,600 | 91,000 | 80,800 | 84,200 |
| $\mathrm{ABC}(\mathrm{t})$ | 87,600 | 91,000 | 80,800 | 84,200 |
| Status | As determined last year for: |  | As determined this year for: |  |
|  | 2010 | 2011 | 2011 | 2012 |
| Overfishing | no | n/a | no | n/a |
| Overfished | n/a | no | $\mathrm{n} / \mathrm{a}$ | No |
| Approaching overfished | n/a | no | $\mathrm{n} / \mathrm{a}$ | No |

## Responses to SSC and Plan Team Comments on Assessments in General

## Retrospective analysis

From the December 2011 SSC minutes: The SSC is pleased to see that many assessment authors have examined retrospective bias in the assessment and encourages the authors and Plan Teams to determine guidelines for how to best evaluate and present retrospective patterns associated with estimates of biomass and recruitment. We recommend that all assessment authors (Tier 3 and higher) bring retrospective analyses forward in next year's assessments.

From the September 2012 Plan Team minutes: The Teams recommend that authors conduct a retrospective analysis back 10 years (thus, back to 2002 for the 2012 assessments), and show the patterns for spawning biomass (both the time series of estimates and the time series of proportional changes relative to the 2012 run). This is consistent with a December 2011 NPFMC SSC request for stock assessment authors to conduct a retrospective analysis. The base model used for the retrospective analysis should be the author's recommended model, even if it differs from the accepted model from previous year.

Due to technical difficulties in model implementation, this request will be addressed next year.

## Total catch accounting

From the September 2012 Plan Team minutes: The Teams recommend that authors continue to include other removals in an appendix for 2012. Authors may apply those removals in estimating ABC and OFL; however, if this is done, results based on the approach used in the previous assessment much also be presented.

Other catch removals for GOA Pacific cod are minimal and were not applied in the estimation of 2013 and 2014 ABC and OFL.

## Responses to SSC and Plan Team Comments Specific to this Assessment

From the May 2012 Plan Team minutes: "For both the EBS and GOA, the Teams recommend that the authors attempt to explore the divergent ageing bias trends in the two regions and the impacts thereof."

This issue will be addressed in a future SAFE document.
From the May 2012 Plan Team minutes: "For both the EBS and GOA, the Teams recommend that the authors attempt to evaluate the biological basis for estimated patterns of seasonal weight at length."

This issue will be addressed in a future SAFE document.
From the September 2012 Plan Team minutes: "The Team recommended that discrepancy in likelihoods (Table 1 in assessment) be examined more closely. It appears Model C had a worse fit to the indices compared to Model A (as expected since fewer catchability parameters are involved) whereas the length and age composition data had a far better fit for Model C compared to Model A.

The Team recommended that the GOA Pacific cod author explore models with the following specifications:

1) Fix $q=1$
2) Drop sub-27 data
3) Drop mean length-at-age
4) Author's own explorations

From the October 2012 SSC minutes:"The Plan Team reviewed a suite of GOA Pacific cod models that centered on SSC, Plan Team and public comments and recommendations. The Plan Team recommended that the base model used last year be brought forward for consideration in November and that the authors explore models that consider fixed Q, drop the sub 27 size category, drop the mean length-at-age data and authors' preferred model. The SSC agrees with Plan Team recommendations and looks forward to future model developments and a more thorough documentation of the recent model improvements."

These model configurations were pursued and are presented below.

## INTRODUCTION

## General

Pacific cod (Gadus macrocephalus) is a transoceanic species, occurring at depths from shoreline to 500 m . The southern limit of the species' distribution is about $34^{\circ} \mathrm{N}$ latitude, with a northern limit of about $63^{\circ} \mathrm{N}$ latitude. Pacific cod is distributed widely over Gulf of Alaska (GOA), as well as the eastern Bering Sea (EBS) and the Aleutian Islands (AI) area. Tagging studies (e.g., Shimada and Kimura 1994) have demonstrated significant migration both within and between the EBS, AI, and GOA. Recent research indicates the existence of discrete stocks in the EBS and AI (Canino et al. 2005, Cunningham et al. 2009, Canino et al. 2010, Spies 2012). Pacific cod is not known to exhibit any special life history characteristics that would require it to be assessed or managed differently from other groundfish stocks in the GOA.

## Review of Life History

Pacific cod eggs are demersal and adhesive. Eggs hatch in about 15 to 20 days. Spawning takes place in the sublittoral-bathyal zone ( 40 to 290 m ) near bottom. Eggs sink to the bottom after fertilization and are somewhat adhesive. Optimal temperature for incubation is $3^{\circ}$ to $6^{\circ} \mathrm{C}$, optimal salinity is 13 to 23 parts per thousand (ppt), and optimal oxygen concentration is from 2 to 3 ppm to saturation. Little is known about the optimal substrate type for egg incubation.

Little is known about the distribution of Pacific cod larvae, which undergo metamorphosis at about 25 to 35 mm . Larvae are epipelagic, occurring primarily in the upper 45 m of the water column shortly after hatching, moving downward in the water column as they grow.

Juveniles occur mostly over the inner continental shelf at depths of 60 to 150 m . Adults occur in depths from the shoreline to 500 m , although occurrence in depths greater than 300 m is fairly rare. Preferred substrate is soft sediment, from mud and clay to sand. Average depth of occurrence tends to vary directly with age for at least the first few years of life. However, in the GOA trawl survey, the percentage of fish residing in waters less than 100 m tends to increase with length beyond about 90 cm . The GOA trawl survey also indicates that fish occupying depths of $200-300 \mathrm{~m}$ are typically in the $40-90 \mathrm{~cm}$ size range.

It is conceivable that mortality rates, both fishing and natural, may vary with age in Pacific cod. In particular, very young fish likely have higher natural mortality rates than older fish (note that this may not be particularly important from the perspective of single-species stock assessment, so long as these higher natural mortality rates do not occur at ages or sizes that are present in substantial numbers in the data). For example, Leslie matrix analysis of a Pacific cod stock occurring off Korea estimated the instantaneous natural mortality rate of 0 -year-olds at $910 \%$ per year (Jung et al. 2009). This may be compared to a mean estimate for age 0 Atlantic cod (Gadus morhua) in Newfoundland of $4.42 \%$ per day, with a $95 \%$ confidence interval ranging from about $3.31 \%$ to $5.03 \%$ (Gregory et al. in review); and age 0 Greenland cod (Gadus ogac) of $2.12 \%$ per day, with a $95 \%$ confidence interval ranging from about $1.56 \%$ to $2.68 \%$ (Robert Gregory and Corey Morris, pers. commun.).

Although little is known about the likelihood of age-dependent natural mortality in adult Pacific cod, it has been suggested that Atlantic cod may exhibit increasing natural mortality with age (Greer-Walker 1970).

At least one study (Ueda et al. 2006) indicates that age 2 Pacific cod may congregate more, relative to age 1 Pacific cod, in areas where trawling efficiency is reduced (e.g., areas of rough substrate), causing their selectivity to decrease. Also, Atlantic cod have been shown to dive in response to a passing vessel (Ona and Godø 1990), which may complicate attempts to estimate catchability or selectivity. It is not known whether Pacific cod undertake a similar response.

As noted above, Pacific cod are known to undertake seasonal migrations, the timing and duration of which may be variable (Savin 2008).

## FISHERY

During the two decades prior to passage of the Magnuson Fishery Conservation and Management Act (MFCMA) in 1976, the fishery for Pacific cod in the GOA was small, averaging around 3,000 t per year. Most of the catch during this period was taken by the foreign fleet, whose catches of Pacific cod were usually incidental to directed fisheries for other species. By 1976, catches had increased to $6,800 \mathrm{t}$. Catches of Pacific cod since 1991 are shown in Table 2.1. Presently, the Pacific cod stock is exploited by a multiple-gear fishery, including trawl, longline, pot, and jig components. Trawl gear took the largest share of the catch in every year but one from 1991-2002, although pot gear has taken the largest singlegear share of the catch in each year since 2003 (not counting 2012, for which data are not yet complete). Figure 2.1 shows landings by gear and season since 1977. Table 2.1 shows the catch by jurisdiction and gear type.

The history of acceptable biological catch (ABC) and total allowable catch (TAC) levels is summarized and compared with the time series of aggregate commercial catches in Table 2.2. For the first year of management under the MFCMA (1977), the catch limit for GOA Pacific cod was established at slightly less than the 1976 total reported landings. During the period 1978-1981, catch limits varied between 34,800 and $70,000 \mathrm{t}$, settling at $60,000 \mathrm{t}$ in 1982. Prior to 1981 these limits were assigned for "fishing years" rather than calendar years. In 1981 the catch limit was raised temporarily to $70,000 \mathrm{t}$ and the fishing year was extended until December 31 to allow for a smooth transition to management based on calendar years, after which the catch limit returned to $60,000 \mathrm{t}$ until 1986, when ABC began to be set on an annual basis. From 1986 (the first year in which an ABC was set) through 1996, TAC averaged about $83 \%$ of ABC and catch averaged about $81 \%$ of TAC. In 8 of those 11 years, TAC equaled ABC exactly. In 2 of those 11 years (1992 and 1996), catch exceeded TAC.

To understand the relationships between ABC, TAC, and catch for the period since 1997, it is important to understand that a substantial fishery for Pacific cod has been conducted during these years inside State of Alaska waters, mostly in the Western and Central Regulatory Areas. To accommodate the Statemanaged fishery, the Federal TAC was set well below ABC ( $15-25 \%$ lower) in each of those years. Thus, although total (Federal plus State) catch has exceeded the Federal TAC in all but three years since 1997, this is basically an artifact of the bi-jurisdictional nature of the fishery and is not evidence of overfishing. At no time since the separate State waters fishery began in 1997 has total catch exceeded ABC, and total catch has never exceeded OFL.

Changes in ABC over time are typically attributable to three factors: 1) changes in resource abundance, 2) changes in management strategy, and 3) changes in the stock assessment model. Assessments conducted prior to 1988 were based on survey biomass alone. From 1988-1993, the assessment was based on stock reduction analysis (Kimura et al. 1984). From 1994-2004, the assessment was conducted using the Stock Synthesis 1 modeling software (Methot 1986, 1990) with length-based data. The assessment was migrated to Stock Synthesis 2 in 2005 (Methot 2005), at which time age-based data began to enter the assessment. Several changes have been made to the model within the SS2 framework (renamed "Stock Synthesis," without a numeric modifier, in 2008) each year since then.

Historically, the majority of the GOA catch has come from the Central regulatory area. To some extent the distribution of effort within the GOA is driven by regulation, as catch limits within this region have been apportioned by area throughout the history of management under the MFCMA. Changes in areaspecific allocation between years have usually been traceable to changes in biomass distributions estimated by Alaska Fisheries Science Center trawl surveys or management responses to local concerns.

Currently, the ABC allocation follows the average biomass distribution estimated by the three most recent trawl surveys, and the TAC allocation is within one percent of this distribution on an area-by-area basis. The complete history of allocation (in percentage terms) by regulatory area within the GOA is shown in Table 2.3.

The catches shown in Tables 2.1 and 2.2 include estimated discards (Table 2.4).
In addition to area allocations, GOA Pacific cod is also allocated on the basis of processor component (inshore/offshore) and season. The inshore component is allocated $90 \%$ of the TAC and the remainder is allocated to the offshore component. Within the Central and Western Regulatory Areas, $60 \%$ of each component's portion of the TAC is allocated to the A season (January 1 through June 10) and the remainder is allocated to the B season (June 11 through December 31, although the B season directed fishery does not open until September 1).

NMFS has also published the following proposed rule to implement Amendment 83 to the GOA Groundfish FMP (effective January 1, 2012):
"Amendment 83 allocates the Pacific cod TAC in the Western and Central regulatory areas of the GOA among various gear and operational sectors, and eliminates inshore and offshore allocations in these two regulatory areas. These allocations apply to both annual and seasonal limits of Pacific cod for the applicable sectors. These apportionments are discussed in detail in a subsequent section of this rule. Amendment 83 is intended to reduce competition among sectors and to support stability in the Pacific cod fishery. The final rule implementing Amendment 83 limits access to the Federal Pacific cod TAC fisheries prosecuted in State of Alaska (State) waters adjacent to the Western and Central regulatory areas in the GOA, otherwise known as parallel fisheries. Amendment 83 does not change the existing annual Pacific cod TAC allocation between the inshore and offshore processing components in the Eastern regulatory area of the GOA.
"In the Central GOA, NMFS must allocate the Pacific cod TAC between vessels using jig gear, catcher vessels (CVs) less than 50 feet ( 15.24 meters) length overall using hook-and-line gear, CVs equal to or greater than 50 feet ( 15.24 meters) length overall using hook-and-line gear, catcher/processors (C/Ps) using hook-and-line gear, CVs using trawl gear, C/Ps using trawl gear, and vessels using pot gear. In the Western GOA, NMFS must allocate the Pacific cod TAC between vessels using jig gear, CVs using hook-and-line gear, C/Ps using hook-and-line gear, CVs using trawl gear, and vessels using pot gear. Table 3 lists the proposed amounts of these seasonal allowances. For the Pacific cod sector splits and associated management measures to become effective in the GOA at the beginning of the 2012 fishing year, NMFS published a final rule ( 76 FR 74670, December 1, 2011) and will revise the final 2012 harvest specifications (76 FR 11111, March 1, 2011)."
"NMFS proposes to calculate of the 2012 and 2013 Pacific cod TAC allocations in the following manner. First, the jig sector would receive 1.5 percent of the annual Pacific cod TAC in the Western GOA and 1.0 percent of the annual Pacific cod TAC in the Central GOA, as required by proposed $\S$ $679.20(\mathrm{c})(7)$. The jig sector annual allocation would further be apportioned between the A ( 60 percent) and B (40 percent) seasons as required by § 679.20(a)(12)(i). Should the jig sector harvest 90 percent or more of its allocation in a given area during the fishing year, then this allocation would increase by one percent in the subsequent fishing year, up to six percent of the annual TAC. NMFS proposes to allocate the remainder of the annual Pacific cod TAC based on gear type, operation type, and vessel length overall in the Western and Central GOA seasonally as required by proposed § 679.20(a)(12)(A) and (B)."

The longline and trawl fisheries are also associated with a Pacific halibut mortality limit which sometimes constrains the magnitude and timing of harvests taken by these two gear types.

## DATA

This section describes data used in the current assessment model. It does not attempt to summarize all available data pertaining to Pacific cod in the GOA.

## Commercial Catch Data

## Catch Biomass

Catches for the period 1977-2012 are shown for the three main gear types in Table 2.6. This also shows gear-specific catches by "selectivity seasons," which are obtained from combinations of "catch seasons. The catch seasons are defined as January-February, March-April, May-August, September-October, and November-December. Three selectivity seasons are defined by combining catch seasons 1 and 2 into selectivity season 1 , equating catch season 3 with selectivity season 2 , and combining catch seasons 4 and 5 into selectivity season 3 . The catch seasons used were the result of a statistical analysis described in the 2010 preliminary assessment (Thompson et al. 2010), and the selectivity seasons were chosen to correspond as closely as possible to the traditional seasons used in previous assessments (given the revised catch seasons). In years for which estimates of the distribution by gear or period were unavailable, proxies based on other years' distributions were used. Catches for the years 1977-1980 may or may not include discards.

## Catch Per Unit Effort

Fishery catch per unit effort data are available by gear and season for the years 1991-2011. These are included in the models for purposes of comparison only and are not used in parameter estimation. Catch Size Composition
Fishery size compositions are presently available, by gear, for at least one gear type in every year from 1977 through the first part of 2011. Beginning with the 2010 assessment (Thompson et al. 2010), size composition data are based on $1-\mathrm{cm}$ bins ranging from 4 to 120 cm . Because displaying these data would add a large number of pages to the present document, they are not shown here but are available at: http://www.afsc.noaa.gov/REFM/Docs/2011/GOA_Pcod_fishery_sizecomp data.xlsx.

## Survey Data

## Survey Age Composition

Age compositions from each survey except 1984 are available (note that the sample size for the 1987 was very small, however). The age compositions and actual sample sizes are shown in Table 2.7.

## Survey Size Composition

For the last few assessments, the size composition data from the trawl surveys of the GOA conducted by the Alaska Fisheries Science Center have been partitioned into two length categories: fish smaller than 27 cm (the "sub-27" survey) and fish 27 cm and larger (the " 27 -plus" survey). The relative size compositions from 1984-2011 are shown for the sub-27 and the 27-plus survey in Table 2.8, using the same $1-\mathrm{cm}$ length bins defined above for the fishery catch size compositions. Columns in this table sum to the actual number of fish measured in each year.

## Mean Size at Age

Mean size-at-age data are available for all of the years in which age compositions are available. These are shown in Table 2.9; the sample sizes are shown in Table 2.10.

## Abundance Estimates

Estimates of total abundance (both in biomass and numbers of fish) obtained from the trawl surveys are shown in Table 2.11, together with their respective coefficients of variation.

The highest biomass ever observed by the survey was the 2009 estimate of $752,651 \mathrm{t}$, and the low point was the preceding (2007) estimate of $233,310 \mathrm{t}$. The 2009 biomass estimate represented a $223 \%$ increase over the 2007 estimate. The 2011 biomass estimate was down $33 \%$ from 2009 , but still $115 \%$ above the 2007 estimate.

In terms of population numbers, the record high was observed in 2009, when the population was estimated to include over 573 million fish. The 2005 estimate of 140 million fish was the low point in the time series. The 2009 abundance estimate represented a $199 \%$ increase over the 2007 estimate. The 2011 abundance estimate was down $39 \%$ from 2009, but still $81 \%$ above the 2007 estimate.

## ANALYTIC APPROACH

## Model Structure

## History of Previous Model Structures Developed Under Stock Synthesis

Beginning with the 1994 SAFE report (Thompson and Zenger 1994), a model using the Stock Synthesis 1 (SS1) assessment program (Methot 1986, 1990, 1998, 2000) and based largely on length-structured data formed the primary analytical tool used to assess the GOA Pacific cod stock.

SS1 was a program that used the parameters of a set of equations governing the assumed dynamics of the stock (the "model parameters") as surrogates for the parameters of statistical distributions from which the data were assumed to be drawn (the "distribution parameters"), and varies the model parameters systematically in the direction of increasing likelihood until a maximum is reached. The overall likelihood was the product of the likelihoods for each of the model components. In part because the overall likelihood could be a very small number, SS1 used the logarithm of the likelihood as the objective function. Each likelihood component was associated with a set of data assumed to be drawn from statistical distributions of the same general form (e.g., multinomial, lognormal, etc.). Typically, likelihood components were associated with data sets such as catch size (or age) composition, survey size (or age) composition, and survey abundance (either biomass or numbers, either relative or absolute).

SS1 permitted each data time series to be divided into multiple segments, resulting in a separate set of parameter estimates for each segment. In the base model for the GOA Pacific cod assessment, for example, possible differences in selectivity between the mostly foreign (also joint venture) and mostly domestic fisheries were accommodated by splitting the fishery size composition time series into pre-1987 and post-1986 segments during the era of SS1-based assessments.

Until 2010, each year was been partitioned into three seasons defined as January-May, June-August, and September-December (these seasonal boundaries were suggested by industry participants in the EBS fishery). Four fisheries were defined during the era of SS1-based assessments: The January-May trawl fishery, the June-December trawl fishery, the longline fishery, and the pot fishery.

Following a series of modifications from 1993 through 1997, the base model for GOA Pacific cod remained completely unchanged from 1997 through 2001. During the late 1990s, a number of attempts were made to estimate the natural mortality rate $M$ and the shelf bottom trawl survey catchability coefficient $Q$, but these were not particularly successful and the Plan Team and SSC always opted to retain the base model in which $M$ and $Q$ were fixed at traditional values of 0.37 and 1.0 , respectively.

A minor modification of the base model was suggested by the SSC in 2001, namely, that consideration be given to dividing the domestic era into pre-2000 and post-1999 segments. This modification was tested in the 2002 assessment (Thompson et al. 2002), where it was found to result in a statistically significant improvement in the model's ability to fit the data.

A major change took place in the 2005 assessment (Thompson and Dorn 2005), as the model was migrated to the newly developed Stock Synthesis 2 (SS2) program, which made use of the ADMB modeling architecture (Fournier 2005) currently used in most age-structured assessments of BSAI and GOA groundfish. The move to SS2 facilitated improved estimation of model parameters as well as statistical characterization of the uncertainty associated with parameter estimates and derived quantities such as spawning biomass. Technical details of SS2 were described by Methot (2005, 2007).

The 2006 assessment model (Thompson et al. 2006) was structured similarly to the 2005 assessment model; the primary change being external estimation of growth parameters.

A technical workshop was convened in April, 2007 to consider a wide range of issues pertaining to both the BSAI and GOA Pacific cod assessments (Thompson and Conners 2007).

The 2007 assessment model (Thompson et al. 2007b) for Pacific cod in the GOA was patterned after the model used in that year's assessment of the BSAI Pacific cod stock (Thompson et al. 2007a), with several changes as described in the assessment document. However, the 2007 assessment model was not accepted by the Plan Team or the SSC.

For the 2008 assessment, the recommended model for the GOA was based largely on the recommended model from the 2008 BSAI Pacific cod assessment. Among other things, this model used an explicit algorithm to determine which fleets (including surveys as well as fisheries) would be forced to exhibit asymptotic selectivity, and another explicit algorithm to determine which selectivity parameters would be allowed to vary periodically in "blocks" of years and to determine the appropriate block length for each such time-varying parameter. One other significant change in the recommended model from the 2008 GOA assessment, which was not shared by the BSAI assessment, was a substantial downweighting of the age composition data. This downweighting was instituted as a means of keeping the root mean squared error of the fit to the survey abundance data close to the sampling variability of those data.

The 2009 assessment (Thompson et al. 2009) featured a total of ten models reflecting a great many alternative assumptions and use or non-use of certain data, particularly age composition data. Relative to the 2008 assessment, the main changes in the model accepted by the Plan Team and SSC were as follow: 1) input standard deviations of all "dev" vectors were set iteratively by matching the standard deviations of the set of estimated "devs;" 2) the standard deviation of length at age was estimated outside the model as a linear function of mean length at age; 3) catchability for the pre-1996 trawl survey was estimated freely while catchability for the post-1993 trawl survey was fixed at the value that sets the average (weighted by numbers at length) of the product of catchability and selectivity for the $60-81 \mathrm{~cm}$ size range equal to the point estimate of 0.92 obtained by Nichol et al. (2007); 4) potential ageing bias was accounted for in the ageing error matrix by examining alternative bias values in increments of 0.1 for ages 2 and above, resulting in a positive bias of 0.4 years for these ages (age-specific bias values were also examined, but did not improve the fit significantly); 5) weighting of the age composition data was returned to its traditional level; 6) except for the parameter governing selectivity at age 0 , all parameters of the selectivity function for the post-1993 years of the 27-plus trawl survey were allowed to vary in each survey year except for the most recent; and 7) cohort-specific growth devs were estimated for all years through 2008.

Many changes were made or considered in the 2010 stock assessment model (Thompson et al. 2010). Five models were presented preliminary assessment, as requested by the Plan Teams in May, with subsequent concurrence (given two minor modifications) by the SSC in June. Following review in September and October, three of these models, or modifications thereof, were requested by the Plan Teams or SSC to be included in the final assessment. Relative to the 2009 assessment, the main changes in the model that was ultimately accepted by the Plan Team and SSC in 2010 were as follow: 1) exclude the single record (each) of fishery age composition and mean length-at-age data, 2) use a finer length bin structure than previous models, and 3) re-evaluate the existing seasonal structure used in the model and revise it as appropriate, and 4) remove cohort-specific growth rates (these were introduced for the first time in the 2009 assessment). The new length bin structure consisted of $1-\mathrm{cm}$ bins, replacing the combination of $3-\mathrm{cm}$ and $5-\mathrm{cm}$ bins used in previous assessments. The new seasonal structure consisted of five catch seasons defined as January-February, March-April, May-August, September-October, and November-December; and three selectivity seasons defined as January-April, May-August, and September-December; with spawning identified as occurring at the beginning of the second catch season (March).

Following a review by the Center for Independent Experts earlier in the year that resulted in a total of 128 unique recommendations from the three reviewers, the 2011 stock assessment (Thompson et al. 2011) again considered several possible model changes. Three models were requested by the Plan Teams to be included in the final GOA assessment. The SSC concurred, and added one more model. The model that was ultimately accepted by the Team and SSC differed from the 2010 model in the following respects:

- The age corresponding to the $L 1$ parameter in the length-at-age equation was increased from 0 to 1.3333 , to correspond to the age of a 1 -year-old fish at the time of the survey, which is when the age data are collected. This change was adopted to prevent mean size at age from going negative (as sometimes happened in previous EBS Pacific cod models), and to facilitate comparison of estimated and observed length at age and variability in length at age.
- The parameters governing variability in length at age were re-tuned. This was necessitated by the change in the age corresponding to the $L 1$ parameter (above).
- A column for age 0 fish was added to the age composition and mean-size-at-age portions of the data file. Even though there are virtually no age 0 fish represented in these two portions of the data file, unless a column for age 0 is included, SS will interpret age 1 fish as being ages 0 and 1 combined, which can bias the estimates of year class strength.
- Ageing bias was estimated internally. To preserve a large value for the strength of the 1977 year class and to keep the mean recruitment from the pre-1977 environmental regime lower than the mean recruitment from the post-1976 environmental regime, ageing bias was constrained to be positive (this constraint ultimately proved to be binding only at the maximum age).

It should also be noted that, consistent with Plan Team policy adopted in 2010, quantities that were estimated iteratively in the 2009 assessment were not re-estimated in the 2010 assessment (with the exception of the parameters governing variability in length at age, for the reason listed above).

## Model Structures Considered in This Year's Assessment

The author's preferred model configuration from 2011, Model 3, was carried over for this assessment. There are two variants of this model configuration:

- 2011 model with 2012 data
- 2011 model with 2012 data and tail compression turned off

These model configurations will be referred to as the "2011 models" in subsequent text.

The following new model configurations were developed for consideration from the September 2012 Plan Team minutes and discussion:

| Model | Description |
| :--- | :--- |
| 1 | Fix q at 1.0 as most of the tuned runs were close to 1.0 ; request that the mean catchability for $60-$ <br> 81 cm be presented to contrast with experimental value of 0.916 |
| 2 | Drop sub-27 data to evaluate effect on recruitment estimates and potential interaction with other <br> data sources |
| 3 | Drop mean length-at-age, as the lack of fit was quite high as indicated by the large contribution <br> to the total likelihood. |
| 4 | as in 2) but drop 27-plus mean length-at-age <br> as in 1) but drop sub-27 mean length-at-age |

The following two model configurations were in response to request 1 in the table above; Model 1 has the same structure as Model C from the September 2012 document (Appendix attached):

- Model 1: q set to $1.0,2$ periods of catchability and selectivity for the sub-27 survey
- Model 1Q: Model 1 with q set to 1.04 (the value used in 2011)

The following two model configurations were in response to request 2 in the table above:

- Model 2: q set to 1.0 , all sub-27 survey data is omitted
- Model 2Q: Model 2 with q set to 1.04 (the value used in 2011)

The following two model configurations were in response to request 3 in the table above:

- Model 3: q set to $1.0,2$ periods of catchability and selectivity for the sub- 27 survey, all sub-27 and 27-plus survey mean length-at-age omitted
- Model 3Q: Model 3 with q set to 1.04 (the value used in 2011)

The following two model configurations were run to determine the impact of subsets of the mean length-at-age data:

- Model 4: Model 2 with the 27-plus mean length-at-age data omitted
- Model 5: Model 1 with the sub-27 mean length-at-age data omitted

The new model configurations have the "tail compression" turned off. This option is part of the assessment program and had been interacting with recruitment estimates in un-anticipated ways (as presented at the September meeting [Appendix attached]). The full set of models is shown below:

| Model | 2011 | 2011 tc | 1 | $1 Q$ | 2 | 2 Q | 3 | 3 Q | 4 | 5 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Catchability value | 1.04 | 1.04 | 1.00 | 1.04 | 1.00 | 1.04 | 1.00 | 1.04 | 1.00 | 1.00 |
| Tail compression off? | no | yes | yes | yes | yes | yes | yes | yes | yes | yes |
| Sub-27 abundance/sizecomp data |  |  |  |  |  |  |  |  |  |  |
| omitted? | no | no | no | no | yes | yes | no | no | yes | no |
| Sub-27 mean size at age data omitted? | no | no | no | no | yes | yes | yes | yes | yes | yes |
| 27-plus mean size at age data omitted? | no | no | no | no | no | no | yes | yes | yes | no |

Version 3.24i of SS was used to run all the models in this assessment.

## Parameters Estimated Outside the Assessment Model

## Natural Mortality

In the 1993 BSAI Pacific cod assessment (Thompson and Methot 1993), the natural mortality rate $M$ was estimated using SS1 at a value of 0.37. All subsequent assessments of the BSAI and GOA Pacific cod stocks (except the 1995 GOA assessment) have used this value for $M$, until the 2007 assessments, at which time the BSAI assessment adopted a value of 0.34 and the GOA assessment adopted a value of 0.38 . Both of these were accepted by the respective Plan Teams and the SSC. The new values were based on Equation 7 of Jensen (1996) and ages at $50 \%$ maturity reported by (Stark 2007; see "Maturity" subsection below). In response to a request from the SSC, the 2008 BSAI assessment included further discussion and justification for these values.

For historical completeness, other published estimates of $M$ for Pacific cod are shown below:

| Area | Author | Year | Value |
| :--- | :--- | :--- | :--- |
| Eastern Bering Sea | Low | 1974 | $0.30-0.45$ |
|  | Wespestad et al. | 1982 | 0.70 |
|  | Bakkala and Wespestad | 1985 | 0.45 |
|  | Thompson and Shimada | 1990 | 0.29 |
|  | Thompson and Methot | 1993 | 0.37 |
| Gulf of Alaska | Thompson and Zenger | 1993 | 0.27 |
|  | Thompson and Zenger | 1995 | 0.50 |
| British Columbia | Ketchen | 1964 | $0.83-0.99$ |
|  | Fournier | 1983 | 0.65 |

All of the models in this assessment set $M$ independently at the SSC-approved value of 0.38 .

## Catchability

In the 2009 assessment (Thompson et al. 2009), catchability for the post-1993 27-plus trawl survey was estimated iteratively by matching the average (weighted by numbers at length) of the product of catchability and selectivity for the $60-81 \mathrm{~cm}$ size range equal to the point estimate of 0.92 obtained by Nichol et al. (2007). The resulting value of 1.04 was retained for several of the models in the present assessment; others set catchability equal to 1.00 , per Plan Team request.

## Variability in Estimated Age

Variability in estimated age in SS is based on the standard deviation of estimated age. Weighted least squares regression has been used in the past several assessments to estimate a linear relationship between standard deviation and age. The regression was recomputed in 2011, yielding an estimated intercept of 0.023 and an estimated slope of 0.072 (i.e, the standard deviation of estimated age was modeled as 0.023 $+0.072 \times$ age , which gives a weighted $R^{2}$ of 0.88 . This regression was used for all models in the present assessment.

## Variability in Length at Age

The last few assessments have used a regression approach to estimate the parameters of the schedule of variability in length at age, based on the outside-the-model estimates of standard deviation of length at age and mean length at age from the survey age data (Thompson et al. 2009). The best fit was obtained by assuming that the standard deviation is a linear function of length at age. The regression was reestimated in 2011 after updating with the most recent data, giving an intercept of 2.248 and a slope of 0.044 . This regression was used for all models in the present assessment.

Use of this regression requires an iterative, "quasi-conditional" procedure for specifying the standard deviations of length at ages 0 and 20, because the regression is a function of length at age, and length at age is estimated conditionally (i.e., inside the model).

In the 2011 model, the age corresponding to the $L 1$ parameter in the length-at-age equation was increased from 0 to 1.3333 (to correspond to the age of a 1-year-old fish at the time of the survey, when the age data are collected). This made it necessary to re-do the iterative tuning process for this model.

## Weight at Length

Season-specific parameters governing the weight-at-length schedule were estimated in the 2010 assessment (based on data through 2008), giving the following values:

| Season: | Jan-Feb | Mar-Apr | May-Aug | Sep-Oct | Nov-Dec |
| :--- | :--- | :--- | :--- | :--- | :--- |
| $\alpha:$ | $8.799 \times 10^{-6}$ | $8.013 \times 10^{-6}$ | $1.147 \times 10^{-5}$ | $1.791 \times 10^{-5}$ | $7.196 \times 10^{-6}$ |
| $\beta:$ | 3.084 | 3.088 | 2.990 | 2.893 | 3.120 |
| Samples: | 36,566 | 29,753 | 6,950 | 9,352 | 2,957 |

The above parameters were retained for all models in the present assessment.

## Maturity

A detailed history and evaluation of parameter values used to describe the maturity schedule for BSAI Pacific cod was presented in the 2005 assessment (Thompson and Dorn 2005). A length-based maturity schedule was used for many years. The parameter values used for this schedule in the 2005 and 2006 assessments were set on the basis of a study by Stark (2007) at the following values: length at $50 \%$ maturity $=50 \mathrm{~cm}$ and slope of linearized logistic equation $=-0.222$. However, in 2007, changes in SS allowed for use of either a length-based or an age-based maturity schedule. Beginning with the 2007 assessment, the accepted model has used an age-based schedule with intercept $=4.3$ years and slope $=$ -1.963 (Stark 2007). The use of an age-based rather than a length-based schedule follows a recommendation from the maturity study's author (James Stark, Alaska Fisheries Science Center, personal communication). The age-based parameters were retained for all models in the present assessment.

## Parameters Estimated Inside the Assessment Model

Parameters estimated conditionally (i.e., within individual SS runs, based on the data and the parameters estimated independently) in all models include the von Bertalanffy growth parameters, two ageing bias parameters, log mean recruitment before and since the 1976-1977 regime shift, annual recruitment deviations, initial fishing mortality, gear-season-and-block-specific fishery selectivity parameters, survey selectivity parameters, and pre-1996 catchability for the 27-plus survey. In addition, the 2011 models and Models 1, 1Q, 3, 3Q, and 5 estimate annual deviations for catchability in the sub-27 survey. The same functional form (pattern 24 for length-based selectivity, pattern 20 for age-based selectivity) used to define the selectivity schedules in last year's assessments was used again this year. This functional form is constructed from two underlying and rescaled normal distributions, with a horizontal line segment joining the two peaks. This form uses the following six parameters (selectivity parameters are referenced by these numbers in several of the tables in this assessment):

1. Beginning of peak region (where the curve first reaches a value of 1.0)
2. Width of peak region (where the curve first departs from a value of 1.0)
3. Ascending "width" (equal to twice the variance of the underlying normal distribution)
4. Descending width
5. Initial selectivity (at minimum length/age)
6. Final selectivity (at maximum length/age)

All but the "beginning of peak region" parameter are transformed: The widths are log-transformed and the other parameters are logit-transformed.

Fishery selectivities are length-based and trawl survey selectivities are age-based in all models considered in this assessment.

Uniform prior distributions are used for all parameters, except that dev vectors are constrained by input standard deviations ("sigma"), which imply a type of joint prior distribution. These input standard deviations were determined iteratively in the 2009 assessment (Thompson et al. 2009) by matching the standard deviations of the estimated devs. The same input standard deviations were used in all models in the present assessment.

For all parameters estimated within individual SS runs, the estimator used is the mode of the logarithm of the joint posterior distribution, which is in turn calculated as the sum of the logarithms of the parameterspecific prior distributions and the logarithm of the likelihood function.

In addition to the above, the full set of year-, season-, and gear-specific fishing mortality rates are also estimated conditionally, but not in the same sense as the above parameters. The fishing mortality rates are determined exactly rather than estimated statistically because SS assumes that the input total catch data are true values rather than estimates, so the fishing mortality rates can be computed algebraically given the other parameter values and the input catch data.

## Likelihood Components

All models included likelihood components for trawl survey relative abundance, fishery and survey size composition, survey age composition, survey mean size at age, recruitment, parameter deviations, and "softbounds" (equivalent to an extremely weak prior distribution used to keep parameters from hitting bounds), and initial (equilibrium) catch. In addition, all models except 3, 3Q, and 4 included a likelihood component for survey mean size at age.

In SS, emphasis factors are specified to determine which likelihood components receive the greatest attention during the parameter estimation process. As in previous assessments, likelihood components were given an emphasis of 1.0 in the present assessment.

## Use of Size Composition Data in Parameter Estimation

Size composition data are assumed to be drawn from a multinomial distribution specific to a particular year, gear, and season within the year. In the parameter estimation process, SS weights a given size composition observation (i.e., the size frequency distribution observed in a given year, gear, and season) according to the emphasis associated with the respective likelihood component and the sample size specified for the multinomial distribution from which the data are assumed to be drawn. In developing the model upon which SS was originally based, Fournier and Archibald (1982) suggested truncating the multinomial sample size at a value of 400 in order to compensate for contingencies which cause the sampling process to depart from the process that gives rise to the multinomial distribution. For many years, the Pacific cod assessments assumed a multinomial sample size equal to the square root of the true length sample size, rather than the true length sample size itself. Given the true length sample sizes observed in the GOA Pacific cod data, this procedure tended to give values somewhat below 400 while still providing SS with usable information regarding the appropriate effort to devote to fitting individual length samples.

Although the "square root rule" for specifying multinomial sample sizes gave reasonable values, the rule itself was largely ad hoc. In an attempt to move toward a more statistically based specification, the 2007 BSAI assessment (Thompson et al. 2007a) used the harmonic means from a bootstrap analysis of the available fishery length data from 1990-2006. The harmonic means were smaller than the actual sample sizes, but still ranged well into the thousands. A multinomial sample size in the thousands would likely overemphasize the size composition data. As a compromise, the harmonic means were rescaled proportionally in the 2007 BSAI assessment so that the average value (across all samples) was 300 . However, the question then remained of what to do about years not covered by the bootstrap analysis (2007 and pre-1990) and what to do about the survey samples. The solution adopted in the 2007 BSAI assessment was based on the consistency of the ratios between the harmonic means (the raw harmonic means, not the rescaled harmonic means) and the actual sample sizes. For the years prior to 1999, the ratio was very consistently close to 0.16 , and for the years after 1998, the ratio was very consistently close to 0.34 .

This consistency was used to specify input sample sizes for size composition data in all GOA assessments since 2007 as follows: For fishery data, the sample sizes for length compositions from years prior to 1999 were tentatively set at $16 \%$ of the actual sample size, and the sample sizes for length compositions from 2007 were tentatively set at $34 \%$ of the actual sample size. For the trawl survey, sample sizes were tentatively set at $34 \%$ of the actual sample size. Then, all sample sizes were adjusted proportionally so that the average was 300 .

## Use of Age Composition Data in Parameter Estimation

Like the size composition data, the age composition data are assumed to be drawn from a multinomial distribution specific to a particular gear, year, and season within the year. Input sample sizes for the multinomial distributions were computed by scaling the actual number of otoliths read in each year proportionally such that the average of the input sample sizes was equal to 300 .

To avoid double counting of the same data, all models ignore size composition data from each year in which survey age composition data are available.

## Use of Fishery CPUE and Survey Relative Abundance Data in Parameter Estimation

Fishery CPUE data are included in the models for comparative purposes only. Their respective catchabilities are estimated analytically, not statistically.

For the trawl surveys, each year's survey abundance datum is assumed to be drawn from a lognormal distribution specific to that year. The model's estimate of survey abundance in a given year serves as the geometric mean for that year's lognormal distribution, and the ratio of the survey abundance datum's standard error to the survey abundance datum itself serves as the distribution's coefficient of variation, which is then transformed into the "sigma" parameter for the lognormal distribution.

## Use of Recruitment Deviation "Data" in Parameter Estimation

The recruitment deviations likelihood component is different from traditional likelihoods because it does not involve "data" in the same sense that traditional likelihoods do. Instead, the log-scale recruitment deviation plays the role of the datum with mean zero and specified (or estimated) standard deviation; but, of course, the devs are parameters, not data.

## RESULTS

## Model evaluation

The 2011 model configurations and eight new model configurations were evaluated, differentiated by the data used in model fitting. The model evaluation criteria included the relative sizes of the likelihood components, and how well the model estimates fit to the 27 -plus and sub-27 survey indices, the survey age composition data, reasonable curves for fishery sand survey selectivity, and that the model estimated the variance-covariance matrix.

## Comparing and Contrasting the Models

The model configurations evaluated focused on exploring the impact of different combinations of 27-plus and sub-27 survey data on model fit. The trends in spawning biomass (Fig. 2.2) and recruitment (Fig. 2.3) were similar for all model configurations, although Models $3,3 \mathrm{Q}$, and 4 , which omitted survey mean size-at-age data, estimated consistently higher biomass than the other models. The patterns in fit to the 27-plus survey were similar for most years across all model configurations, although less so for the three most recent survey years (Fig. 2.4). The 2011 model configurations had virtually identical fits to the sub27 survey, as these model configurations estimate separate $q$ values for each survey year (Fig. 2.5); the other model configurations which included sub-27 data had virtually identical fits when estimating $q$ and selectivity for two periods.

## Evaluation Criteria

Table 2.12 lists the number of parameters, the values of the objective function components, and the "effective q" for all model configurations. The "effective q" is for comparison with the product of catchability and selectivity for the $60-81 \mathrm{~cm}$ size range equal to the point estimate of 0.916 obtained by Nichol et al. (2007); all model configurations had a value for "effective q" less than 0.916 . The 2011 models fit the sub-27 survey indices well due to time-varying q ; the other model configurations did not fit the sub- 27 survey indices as well. None of the models fit the 27 -plus survey indices well, although Model 4 fit the best of all model configurations. Model 4 fit the 27-plus survey age composition the best of all model configurations; Model 2 has the second best fit. All model configurations had reasonable fishery selectivity curves. All model configurations converged and produced variance-covariance matrices.

## Selection of Final Model

Model 2 was selected as the preferred model, as the biomass estimates were similar to estimates from other model configuration which included fitting to survey mean size-at-age data, and it had fewer parameters due to excluding the sub-27 survey data. The sub-27 survey indices are highly variable (Fig. 2.5 ), and there is considerable uncertainty associated with the probability and consistency of capture of sub- 27 cm fish in the trawl survey.

Although Model 4 had lower objective function component values than Model 2 for several categories, Model 4 estimated the mean length for age- 1 fish to be 27 cm , which is significantly higher than the expected value of 20.5 cm , and higher than the value estimated by Model 2, 24.9 cm (see table below). Therefore, Model 2 is recommended as the base model for this year's assessment.

|  | 2011 2011 model |  | Model | Model |  | Model | Model | Model | Model | Model |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | Model

$\mathrm{A}_{\text {min }}$ is age 1.33333 (age-1 fish at the time of the NMFS GOA bottom trawl survey, when the data were collected)

## Final Parameter Estimates and Associated Schedules

Table 2.13 lists the fixed and estimated parameter values for Model 2. Spawning biomass has decreased from a peak in 1980 to a low in 2008 and is increasing (Fig. 2.6). Age-0 recruits had the highest value at the beginning of the time series and has had moderate variability around 230 million since then (Fig. 2.7). The estimates of the 27-plus survey indices do not fit the data well (Fig. 2.8). There does not appear to be a strong relationship between spawning biomass and recruitment (Fig. 2.9). Fishery selectivity for all seasons and gear types varies across fisheries, seasons, and gear types (Fig. 2.10). Survey selectivity for the 27 -plus survey is variable (Fig. 2.11). The fits to the fishery catch-at-length data were reasonable in most years, with poor fits to some years in the 1980s for the Jan-Apr trawl fishery (Fig. 2.12). The fits to the 27 -plus survey length composition data captured the general patterns (Fig. 2.13). The fits to the 27plus survey age composition data captured much of the variability between survey years (Fig. 2.14). The fits to the 27-plus survey mean size-at-age data were reasonable (Fig. 2.15). Estimated length-at-age curve is shown in Fig. 2.16.

## Time Series Results

## Definitions

The biomass estimates presented here will be defined in two ways: 1) age $0+$ biomass, consisting of the biomass of all fish aged 0 years or greater in January of a given year; and 2) spawning biomass, consisting of the biomass of all spawning females in a given year. The recruitment estimates presented here will be defined as numbers of age-0 fish in a given year.

## Biomass

Table 2.14 shows the time series of GOA Pacific cod female spawning biomass for the years 1977-2012 as estimated last year and this year under Model 2. The estimated spawning biomass time series are accompanied by their respective standard deviations.

## Recruitment and Numbers at Age

Table 2.15 shows the time series of GOA Pacific cod age-0 recruits for the years 1977-2012 as estimated last year and this year under Model 2. The estimated recruitment time series are accompanied by their respective standard deviations. Table 2.5 shows the numbers-at-age for 1977-2012.

## Fishing Mortality

Table 2.6 shows the "effective" annual fishing mortality by age and year for ages 1-19 and years 19772011.

## Harvest Recommendations

## Amendment 56 Reference Points

Amendment 56 to the GOA Groundfish Fishery Management Plan (FMP) defines the "overfishing level" (OFL), the fishing mortality rate used to set OFL ( $F_{\text {OFL }}$ ), the maximum permissible ABC, and the fishing mortality rate used to set the maximum permissible ABC. The fishing mortality rate used to set ABC $\left(F_{A B C}\right)$ may be less than this maximum permissible level, but not greater. Because reliable estimates of reference points related to maximum sustainable yield (MSY) are currently not available but reliable estimates of reference points related to spawning per recruit are available, Pacific cod in the GOA have generally been managed under Tier 3 of Amendment 56 (with the exception of the current year, when the stock is being managed under Tier 5). Tier 3 uses the following reference points: $B_{40 \%}$, equal to $40 \%$ of the equilibrium spawning biomass that would be obtained in the absence of fishing; $F_{35 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $35 \%$ of the level that would be obtained in the absence of fishing; and $F_{40 \%}$, equal to the fishing mortality rate that reduces the equilibrium level of spawning per recruit to $40 \%$ of the level that would be obtained in the absence of fishing. The following formulae apply under Tier 3:

$$
\begin{aligned}
& \text { 3a) Stock status: } B / B_{40 \%}>1 \\
& F_{\text {OFL }}=F_{35 \%} \\
& F_{A B C} \leq F_{40 \%} \\
& \text { 3b) Stock status: } 0.05<B / B_{40 \%} \leq 1 \\
& F_{\text {OFL }}=F_{35 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& F_{A B C} \leq F_{40 \%} \times\left(B / B_{40 \%}-0.05\right) \times 1 / 0.95 \\
& \text { 3c) Stock status: } B / B_{40 \%} \leq 0.05 \\
& F_{\text {OFL }}=0 \\
& F_{A B C}=0
\end{aligned}
$$

Other useful biomass reference points which can be calculated using this assumption are $B_{100 \%}$ and $B_{35 \%}$, defined analogously to $B_{40 \%}$. These reference points are estimated as follows, based on Model 2:

| Reference point: | $B_{35 \%}$ | $B_{40 \%}$ | $B_{100 \%}$ |
| :--- | ---: | ---: | ---: |
| Spawning biomass: | $82,100 \mathrm{t}$ | $93,900 \mathrm{t}$ | $234,800 \mathrm{t}$ |

For a stock exploited by multiple gear types, estimation of $F_{35 \%}$ and $F_{40 \%}$ requires an assumption regarding the apportionment of fishing mortality among those gear types. For this assessment, the apportionment was based on Model 2's estimates of fishing mortality by gear for the five most recent complete years of data (2007-2011). The average fishing mortality rates for those years implied that total fishing mortality was divided among the three main gear types according to the following percentages: trawl $27 \%$, longline $25 \%$, and pot $48 \%$. This apportionment results in estimates of $F_{35 \%}$ and $F_{40 \%}$ equal to 0.61 and 0.49 , respectively.

## Specification of OFL and Maximum Permissible ABC

Spawning biomass for 2013 is estimated by Model 2 at a value of $111,000 \mathrm{t}$. This is well above the $B_{40 \%}$ value of $93,900 \mathrm{t}$, thereby placing Pacific cod in sub-tier "a" of Tier 3. Given this, Model 2 estimates OFL, maximum permissible ABC, and the associated fishing mortality rates for 2013 and 2014 as follows (2014 values are predicated on the assumption that 2013 catch will equal 2013 maximum permissible ABC):

| Units | Year | Overfishing | Maximum |
| :--- | ---: | ---: | ---: |
|  | Level | Permissible ABC |  |
| Harvest amount | 2013 | $\mathbf{9 7 , 2 0 0} \mathrm{t}$ | $\mathbf{8 0 , 8 0 0 0 0 \mathrm { T }}$ |
| Harvest amount | 2014 | $101,100 \mathrm{t}$ | $84,200 \mathrm{t}$ |
| Fishing mortality rate | 2013 | 0.61 | 0.49 |
| Fishing mortality rate | 2014 | 0.61 | 0.49 |

The age $0+$ biomass projections for 2013 and 2014 from Model 2 are 449,300 t and 440,300 t .

## ABC Recommendation

In 2005, the SSC used a two-year stair-step approach to recommend the 2006 ABC. In 2006, the GOA Plan Team and SSC recommended keeping ABC at the 2006 level for 2007. In 2007, the GOA Plan Team and SSC adopted a Tier 5 approach for setting the 2008 ABC. In 2008-2010, the GOA Plan Team and SSC recommended setting 2009 ABC at the maximum permissible level under Tier 3.

Following recent practice, this year's ABC recommendations for 2013 and 2014 are at their respective maximum permissible levels of $\mathbf{8 0 , 8 0 0 0 0} \mathrm{t}$ and $84,200 \mathrm{t}$.

## Area Allocation of Harvests

For the past several years, ABC has been allocated among regulatory areas on the basis of the three most recent surveys. The previous proportions of $35 \%$ Western, $62 \%$ Central, and $3 \%$ Eastern were based on the average (across years) of the area-specific biomass estimates from the 2005-2009 surveys. If the same methodology were applied to the 2007-2011 surveys, the proportions would be $32 \%$ Western, $65 \%$ Central, and 3\% Eastern. The SSC and GOA Plan Team have requested that the simple Kalman filter approach that has been used to estimate the proportions of Pacific cod biomass in the EBS and AI since 2004 be applied to the GOA as well. Using this approach, the proportions would be $35 \%$ Western, $61 \%$ Central, and 4\% Eastern.

## Standard Harvest and Recruitment Scenarios and Projection Methodology

A standard set of projections is required for each stock managed under Tiers 1, 2, or 3 of Amendment 56. This set of projections encompasses seven harvest scenarios designed to satisfy the requirements of Amendment 56, the National Environmental Policy Act, and the Magnuson-Stevens Fishery Conservation and Management Act (MSFCMA).

For each scenario, the projections begin with an estimated vector of 2013 numbers at age. In each subsequent year, the fishing mortality rate is prescribed on the basis of the spawning biomass in that year and the respective harvest scenario. In each year, recruitment is drawn from an inverse Gaussian distribution whose parameters consist of maximum likelihood estimates determined from recruitments estimated in the assessment. Spawning biomass is computed in each year based on the time of peak spawning and the maturity and weight schedules described in the assessment. Total catch is assumed to equal the catch associated with the respective harvest scenario in all years. This projection scheme is run 1000 times to obtain distributions of possible future stock sizes, fishing mortality rates, and catches.

Five of the seven standard scenarios will be used in an Environmental Assessment prepared in conjunction with the final SAFE. These five scenarios, which are designed to provide a range of harvest alternatives that are likely to bracket the final TACs for 2013 and 2014, are as follow ("max $F_{A B C}$ " refers to the maximum permissible value of $F_{A B C}$ under Amendment 56):

Scenario 1: In all future years, $F$ is set equal to $\max F_{A B C}$. (Rationale: Historically, TAC has been constrained by ABC , so this scenario provides a likely upper limit on future TACs.)

Scenario 2: In all future years, $F$ is set equal to a constant fraction of $\max F_{A B C}$, where this fraction is equal to the ratio of the $F_{A B C}$ value for 2013 recommended in the assessment to the max $F_{A B C}$ for 2013. (Rationale: When $F_{A B C}$ is set at a value below $\max F_{A B C}$, it is often set at the value recommended in the stock assessment.)

Scenario 3: In all future years, $F$ is set equal to the 2006-2010 average F. (Rationale: For some stocks, TAC can be well below ABC, and recent average $F$ may provide a better indicator of $F_{T A C}$ than $F_{A B C}$.)

Scenario 4: In all future years, the upper bound on $F_{A B C}$ is set at $F_{60 \%}$. (Rationale: This scenario provides a likely lower bound on $F_{A B C}$ that still allows future harvest rates to be adjusted downward when stocks fall below reference levels.)

Scenario 5: In all future years, $F$ is set equal to zero. (Rationale: In extreme cases, TAC may be set at a level close to zero.)

Two other scenarios are needed to satisfy the MSFCMA's requirement to determine whether a stock is currently in an overfished condition or is approaching an overfished condition. These two scenarios are as follow (for Tier 3 stocks, the MSY level is defined as $B_{35 \%}$ ):

Scenario 6: In all future years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is overfished. If the stock is 1) above its MSY level in 2011 or 2) above $1 / 2$ of its MSY level in 2011 and expected to be above its MSY level in 2021 under this scenario, then the stock is not overfished.)

Scenario 7: In 2013 and 2014, $F$ is set equal to $\max F_{A B C}$, and in all subsequent years, $F$ is set equal to $F_{\text {OFL }}$. (Rationale: This scenario determines whether a stock is approaching an overfished condition. If the stock is expected to be above its MSY level in 2024 under this scenario, then the stock is not approaching an overfished condition.)

## Projections and Status Determination

Projections corresponding to the standard scenarios are shown for Model 2 in Table 2.18 (note that Scenarios 1 and 2 are identical in this case, because the recommended $A B C$ is equal to the maximum permissible ABC ).

In addition to the seven standard harvest scenarios, Amendments $48 / 48$ to the BSAI and GOA Groundfish Fishery Management Plans require projections of the likely OFL two years into the future. While Scenario 6 gives the best estimate of OFL for 2013, it does not provide the best estimate of OFL for 2014, because the mean 2014 catch under Scenario 6 is predicated on the 2013 catch being equal to the 2013 OFL, whereas the actual 2013 catch will likely be less than the 2013 OFL.

Under the MSFCMA, the Secretary of Commerce is required to report on the status of each U.S. fishery with respect to overfishing. This report involves the answers to three questions: 1) Is the stock being subjected to overfishing? 2) Is the stock currently overfished? 3) Is the stock approaching an overfished condition?

Is the stock being subjected to overfishing? The official catch estimate for the most recent complete year (2011) is $84,800 \mathrm{t}$. This is less than the 2011 OFL of $102,600 \mathrm{t}$. Therefore, the stock is not being subjected to overfishing.

Harvest Scenarios \#6 and \#7 are intended to permit determination of the status of a stock with respect to its minimum stock size threshold (MSST). Any stock that is below its MSST is defined to be overfished. Any stock that is expected to fall below its MSST in the next two years is defined to be approaching an overfished condition. Harvest Scenarios \#6 and \#7 are used in these determinations as follows:

Is the stock currently overfished? This depends on the stock's estimated spawning biomass in 2013:
a. If spawning biomass for 2012 is estimated to be below $1 / 2 B_{35 \%}$, the stock is below its MSST.
b. If spawning biomass for 2012 is estimated to be above $B_{35 \%}$ the stock is above its MSST.
c. If spawning biomass for 2012 is estimated to be above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the stock's status relative to MSST is determined by referring to harvest Scenario \#6 (Table 2.18). If the mean spawning biomass for 2025 is below $B_{35 \%}$, the stock is below its MSST. Otherwise, the stock is above its MSST.

Is the stock approaching an overfished condition? This is determined by referring to harvest Scenario \#7 (Table 2.18):
a. If the mean spawning biomass for 2014 is below $1 / 2 B_{35 \%}$, the stock is approaching an overfished condition.
b. If the mean spawning biomass for 2014 is above $B_{35 \%}$, the stock is not approaching an overfished condition.
c. If the mean spawning biomass for 2014 is above $1 / 2 B_{35 \%}$ but below $B_{35 \%}$, the determination depends on the mean spawning biomass for 2025. If the mean spawning biomass for 2025 is below $B_{35 \%}$, the stock is approaching an overfished condition. Otherwise, the stock is not approaching an overfished condition.

Based on the above criteria and Table 2.18, the stock is not overfished and is not approaching an overfished condition.

Biological reference points, spawning biomass, and ABC values from the current SAFE document and previous GOA Pacific cod SAFE documents for 2001-2011 are listed in Table 2.19.

## ECOSYSTEM CONSIDERATIONS

This section is largely unchanged from last year's assessment, except for the subsection on "Incidental Catch of Nontarget Species."

## Ecosystem Effects on the Stock

A primary ecosystem phenomenon affecting the Pacific cod stock seems to be the occurrence of periodic "regime shifts," in which central tendencies of key variables in the physical environment change on a scale spanning several years to a few decades (Boldt (ed.), 2005). One well-documented example of such a regime shift occurred in 1977, and shifts occurring in 1989 and 1999 have also been suggested (e.g., Hare and Mantua 2000). In the present assessment, an attempt was made to estimate the change in median recruitment of GOA Pacific cod associated with the 1977 regime shift. According to this year's model, pre-1977 median recruitment was only about $32 \%$ of post-1976 median recruitment. Establishing a link between environment and recruitment within a particular regime is more difficult. In the 2004
assessment (Thompson et al. 2004), for example, the correlations between age 1 recruits spawned since 1977 and monthly values of the Pacific Decadal Oscillation (Mantua et al. 1997) were computed and found to be very weak.

The prey and predators of Pacific cod have been described or reviewed by Albers and Anderson (1985), Livingston (1989, 1991), Lang et al. (2003), Westrheim (1996), and Yang (2004). The composition of Pacific cod prey varies to some extent by time and area. In terms of percent occurrence, some of the most important items in the diet of Pacific cod in the BSAI and GOA have been polychaetes, amphipods, and crangonid shrimp. In terms of numbers of individual organisms consumed, some of the most important dietary items have been euphausids, miscellaneous fishes, and amphipods. In terms of weight of organisms consumed, some of the most important dietary items have been walleye pollock, fishery offal, yellowfin sole, and crustaceans. Small Pacific cod feed mostly on invertebrates, while large Pacific cod are mainly piscivorous. Predators of Pacific cod include Pacific cod, halibut, salmon shark, northern fur seals, Steller sea lions, harbor porpoises, various whale species, and tufted puffin. Major trends in the most important prey or predator species could be expected to affect the dynamics of Pacific cod to some extent.

## Fishery Effects on the Ecosystem

Potentially, fisheries for Pacific cod can have effects on other species in the ecosystem through a variety of mechanisms, for example by relieving predation pressure on shared prey species (i.e., species which serve as prey for both Pacific cod and other species), by reducing prey availability for predators of Pacific cod, by altering habitat, by imposing bycatch mortality, or by "ghost fishing" caused by lost fishing gear.

## Incidental Catch of Nontarget Species

Incidental catches of nontarget species in each year 2003-2012 are shown Table 2.5. In terms of average catch over the time series, only sea stars account for more than 200 t per year.

## Steller Sea Lions

Sinclair and Zeppelin (2002) showed that Pacific cod was one of the four most important prey items of Steller sea lions in terms of frequency of occurrence averaged over years, seasons, and sites, and was especially important in winter. Pitcher (1981) and Calkins (1998) also showed Pacific cod to be an important winter prey item in the GOA and BSAI, respectively. Furthermore, the size ranges of Pacific cod harvested by the fisheries and consumed by Steller sea lions overlap, and the fishery operates to some extent in the same geographic areas used by Steller sea lion as foraging grounds (Livingston (ed.), 2002).

The Fisheries Interaction Team of the Alaska Fisheries Science Center has been engaged in research to determine the effectiveness of recent management measures designed to mitigate the impacts of the Pacific cod fisheries (among others) on Steller sea lions. Results from studies conducted in 2002-2003 were summarized by Conners et al. (2004). These studies included a tagging feasibility study, which may evolve into an ongoing research effort capable of providing information on the extent and rate to which Pacific cod move in and out of various portions of Steller sea lion critical habitat. Nearly 6,000 cod with spaghetti tags were released, of which approximately 1,000 had been returned as of September, 2003.

## Seabirds

The following is a summary of information provided by Livingston (ed., 2002): In both the BSAI and GOA, the northern fulmar (Fulmarus glacialis) comprises the majority of seabird bycatch, which occurs primarily in the longline fisheries, including the hook and line fishery for Pacific cod (Tables 2.30b and 2.30b). Shearwater (Puffinus spp.) distribution overlaps with the Pacific cod longline fishery in the Bering Sea, and with trawl fisheries in general in both the Bering Sea and GOA. Black-footed albatross (Phoebastria nigripes) is taken in much greater numbers in the GOA longline fisheries than the Bering

Sea longline fisheries, but is not taken in the trawl fisheries. The distribution of Laysan albatross (Phoebastria immutabilis) appears to overlap with the longline fisheries in the central and western Aleutians. The distribution of short-tailed albatross (Phoebastria albatrus) also overlaps with the Pacific cod longline fishery along the Aleutian chain, although the majority of the bycatch has taken place along the northern portion of the Bering Sea shelf edge (in contrast, only two takes have been recorded in the GOA). Some success has been obtained in devising measures to mitigate fishery-seabird interactions. For example, on vessels larger than 60 ft LOA, paired streamer lines of specified performance and material standards have been found to reduce seabird incidental take significantly.

## Fishery Usage of Habitat

The following is a summary of information provided by Livingston (ed., 2002): The longline and trawl fisheries for Pacific cod each comprise an important component of the combined fisheries associated with the respective gear type in each of the three major management regions (BS, AI, and GOA). Looking at each gear type in each region as a whole (i.e., aggregating across all target species) during the period 1998-2001, the total number of observed sets was as follows:

| Gear | BS | AI | GOA |
| :--- | ---: | ---: | ---: |
| Trawl | 240,347 | 43,585 | 68,436 |
| Longline | 65,286 | 13,462 | 7,139 |

In the BS, both longline and trawl effort was concentrated north of False Pass (Unimak Island) and along the shelf edge represented by the boundary of areas 513, 517 (in addition, longline effort was concentrated along the shelf edge represented by the boundary of areas 521-533). In the AI, both longline and trawl effort were dispersed over a wide area along the shelf edge. The catcher vessel longline fishery in the AI occurred primarily over mud bottoms. Longline catcher-processors in the AI tended to fish more over rocky bottoms. In the GOA, fishing effort was also dispersed over a wide area along the shelf, though pockets of trawl effort were located near Chirikof, Cape Barnabus, Cape Chiniak and Marmot Flats. The GOA longline fishery for Pacific cod generally took place over gravel, cobble, mud, sand, and rocky bottoms, in depths of 25 fathoms to 140 fathoms.

Impacts of the Pacific cod fisheries on essential fish habitat were further analyzed in an environmental impact statement by NMFS (2005).

## DATA GAPS AND RESEARCH PRIORITIES

Understanding of the above ecosystem considerations would be improved if future research were directed toward closing certain data gaps. Such research would have several foci, including the following: 1) ecology of the Pacific cod stock, including spatial dynamics, trophic and other interspecific relationships, and the relationship between climate and recruitment; 2) behavior of the Pacific cod fishery, including spatial dynamics; 3) determinants of trawl survey catchability and selectivity; 4) age determination; 5) ecology of species taken as bycatch in the Pacific cod fisheries, including estimation of biomass, carrying capacity, and resilience; and 6) ecology of species that interact with Pacific cod, including estimation of biomass, carrying capacity, and resilience.

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## Tables

Table 2.1. Catch ( t ) for 1991 through 2012 by jurisdiction and gear type (as of 3 November 2012)

|  | Federal |  |  |  |  | State |  |  |  |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Trawl | Longline | Pot | Other | Subtotal | Longline | Pot | Other | Subtotal | Total |
| 1991 | 58,093 | 7,656 | 10,464 | 115 | 76,328 | 0 | 0 | 0 | 0 | 76,328 |
| 1992 | 54,593 | 15,675 | 10,154 | 325 | 80,747 | 0 | 0 | 0 | 0 | 80,747 |
| 1993 | 37,806 | 8,963 | 9,708 | 11 | 56,488 | 0 | 0 | 0 | 0 | 56,488 |
| 1994 | 31,447 | 6,778 | 9,161 | 100 | 47,485 | 0 | 0 | 0 | 0 | 47,485 |
| 1995 | 41,875 | 10,978 | 16,055 | 77 | 68,985 | 0 | 0 | 0 | 0 | 68,985 |
| 1996 | 45,991 | 10,196 | 12,040 | 53 | 68,280 | 0 | 0 | 0 | 0 | 68,280 |
| 1997 | 48,406 | 10,978 | 9,065 | 26 | 68,476 | 0 | 7,224 | 1,319 | 8,542 | 77,018 |
| 1998 | 41,570 | 10,012 | 10,510 | 29 | 62,121 | 0 | 9,088 | 1,316 | 10,404 | 72,525 |
| 1999 | 37,167 | 12,363 | 19,015 | 70 | 68,614 | 0 | 12,075 | 1,096 | 13,171 | 81,785 |
| 2000 | 25,443 | 11,660 | 17,351 | 54 | 54,508 | 0 | 10,388 | 1,643 | 12,031 | 66,560 |
| 2001 | 24,383 | 9,910 | 7,171 | 155 | 41,619 | 0 | 7,836 | 2,084 | 9,920 | 51,542 |
| 2002 | 19,810 | 14,666 | 7,694 | 176 | 42,345 | 0 | 10,423 | 1,714 | 12,137 | 54,483 |
| 2003 | 18,885 | 9,585 | 12,740 | 161 | 41,371 | 0 | 7,966 | 3,242 | 11,207 | 52,579 |
| 2004 | 17,513 | 10,380 | 14,965 | 400 | 43,258 | 0 | 10,602 | 2,765 | 13,367 | 56,625 |
| 2005 | 14,549 | 5,758 | 14,749 | 203 | 35,260 | 0 | 9,653 | 2,673 | 12,326 | 47,585 |
| 2006 | 13,131 | 10,274 | 14,795 | 118 | 38,319 | 0 | 8,890 | 646 | 9,536 | 47,854 |
| 2007 | 14,774 | 11,677 | 13,515 | 40 | 40,006 | 0 | 10,885 | 574 | 11,459 | 51,465 |
| 2008 | 20,293 | 12,358 | 11,230 | 62 | 43,943 | 0 | 13,438 | 1,568 | 15,006 | 58,949 |
| 2009 | 13,981 | 14,050 | 11,576 | 194 | 39,801 | 325 | 10,293 | 2,497 | 13,116 | 52,917 |
| 2010 | 21,791 | 16,674 | 20,114 | 423 | 59,003 | 375 | 14,604 | 4,090 | 19,069 | 78,072 |
| 2011 | 16,364 | 16,508 | 29,228 | 721 | 62,821 | 725 | 16,669 | 4,625 | 22,018 | 84,839 |
| 2012 | 19,704 | 13,017 | 19,023 | 722 | 52,467 | 628 | 15,912 | 4,594 | 21,134 | 73,601 |

Table 2.2 History of Pacific cod catch ( t , includes catch from State waters), Federal TAC (does not include State guideline harvest level), ABC, and OFL. ABC was not used in management of GOA groundfish prior to 1986. Catch for 2012 is current through 3 November. The values in the column labeled "TAC" correspond to "optimum yield" for the years 19801986, "target quota" for the year 1987, and true TAC for the years 1988-2009. The ABC value listed for 1987 is the upper bound of the range. Source: NPFMC staff.

| Year | Catch | TAC | ABC | OFL |
| :---: | :---: | :---: | :---: | :---: |
| 1980 | 35,345 | 60,000 | - | - |
| 1981 | 36,131 | 70,000 | - | - |
| 1982 | 29,465 | 60,000 | - | - |
| 1983 | 36,540 | 60,000 | - | - |
| 1984 | 23,898 | 60,000 | - | - |
| 1985 | 14,428 | 60,000 | 136,000 | - |
| 1986 | 25,012 | 75,000 | 125,000 | - |
| 1987 | 32,939 | 50,000 | 185,000* | - |
| 1988 | 33,802 | 80,000 | 99,000 | - |
| 1989 | 43,293 | 71,200 | 71,200 | - |
| 1990 | 72,517 | 90,000 | 90,000 | - |
| 1991 | 76,328 | 77,900 | 77,900 | - |
| 1992 | 80,747 | 63,500 | 63,500 | 87,600 |
| 1993 | 56,488 | 56,700 | 56,700 | 78,100 |
| 1994 | 47,485 | 50,400 | 50,400 | 71,100 |
| 1995 | 68,985 | 69,200 | 69,200 | 126,000 |
| 1996 | 68,280 | 65,000 | 65,000 | 88,000 |
| 1997 | 77,018 | 69,115 | 81,500 | 180,000 |
| 1998 | 72,525 | 66,060 | 77,900 | 141,000 |
| 1999 | 81,785 | 67,835 | 84,400 | 134,000 |
| 2000 | 66,560 | 59,800 | 76,400 | 102,000 |
| 2001 | 51,542 | 52,110 | 67,800 | 91,200 |
| 2002 | 54,483 | 44,230 | 57,600 | 77,100 |
| 2003 | 52,579 | 40,540 | 52,800 | 70,100 |
| 2004 | 56,625 | 48,033 | 62,810 | 102,000 |
| 2005 | 47,585 | 44,433 | 58,100 | 86,200 |
| 2006 | 47,854 | 52,264 | 68,859 | 95,500 |
| 2007 | 51,462 | 52,264 | 68,859 | 97,600 |
| 2008 | 58,964 | 50,269 | 64,493 | 88,660 |
| 2009 | 52,920 | 41,807 | 55,300 | 66,000 |
| 2010 | 78,071 | 59,563 | 79,100 | 94,100 |
| 2011 | 84,839 | 65,100 | 86,800 | 102,600 |
| 2012 | 73,601 | 65,700 | 87,600 | 104,000 |

Table 2.3. History of GOA Pacific cod allocations by regulatory area (in percent)

|  |  | Regulatory area |  |
| ---: | ---: | ---: | ---: |
| Year(s) | Western | Central | Eastern |
| $1977-1985$ | 28 | 56 | 16 |
| 1986 | 40 | 44 | 16 |
| 1987 | 27 | 56 | 17 |
| $1988-1989$ | 19 | 73 | 8 |
| 1990 | 33 | 66 | 1 |
| 1991 | 33 | 62 | 5 |
| 1992 | 37 | 61 | 2 |
| $1993-1994$ | 33 | 62 | 5 |
| $1995-1996$ | 29 | 66 | 5 |
| $1997-1999$ | 35 | 63 | 2 |
| $2000-2001$ | 36 | 57 | 7 |
| 2002 | 39 | 55 | 6 |
| 2002 | 38 | 56 | 6 |
| 2003 | 39 | 55 | 6 |
| 2003 | 38 | 56 | 6 |
| 2004 | 36 | 57 | 7 |
| 2004 | 35.3 | 56.5 | 8.2 |
| 2005 | 36 | 57 | 7 |
| 2005 | 35.3 | 56.5 | 8.2 |
| 2006 | 39 | 55 | 6 |
| 2006 | 38.54 | 54.35 | 7.11 |
| 2007 | 39 | 55 | 6 |
| 2007 | 38.54 | 54.35 | 7.11 |
| 2008 | 39 | 57 | 4 |
| 2008 | 38.69 | 56.55 | 4.76 |
| 2009 | 39 | 57 | 4 |
| 2009 | 38.69 | 56.55 | 4.76 |
| 2010 | 35 | 62 | 3 |
| 2010 | 34.86 | 61.75 | 3.39 |
| 2011 | 35 | 62 | 3 |
| 2011 | 35 | 62 | 3 |
| $2012(\mathrm{ABC)}$ | 28.03 | 56.94 | 2.63 |
| $2012(\mathrm{TAC)}$ | 21.02 | 42.71 | 1.97 |
|  |  |  |  |

Table 2.4 Estimated retained-and discarded GOA Pacific cod from federal waters (source: NMFS/NOAA/AKFIN)

| Year | Discarded | Retained | Grand Total |
| ---: | ---: | ---: | ---: |
| 1991 | 1,429 | 74,899 | 76,328 |
| 1992 | 3,873 | 76,199 | 80,073 |
| 1993 | 5,844 | 49,865 | 55,709 |
| 1994 | 3,109 | 43,540 | 46,649 |
| 1995 | 3,525 | 64,560 | 68,085 |
| 1996 | 7,534 | 60,530 | 68,064 |
| 1997 | 4,783 | 63,057 | 67,840 |
| 1998 | 1,709 | 59,811 | 61,520 |
| 1999 | 1,617 | 66,311 | 67,928 |
| 2000 | 1,362 | 52,904 | 54,266 |
| 2001 | 1,901 | 39,632 | 41,533 |
| 2002 | 3,713 | 38,594 | 42,307 |
| 2003 | 2,413 | 50,046 | 52,459 |
| 2004 | 1,266 | 55,305 | 56,571 |
| 2005 | 1,040 | 46,499 | 47,539 |
| 2006 | 1,831 | 45,960 | 47,791 |
| 2007 | 1,449 | 49,805 | 51,253 |
| 2008 | 3,310 | 55,375 | 58,685 |
| 2009 | 3,842 | 48,709 | 52,551 |
| 2010 | 2,831 | 74,789 | 77,620 |
| 2011 | 1,806 | 82,317 | 84,124 |
| 2012 | 964 | 72,111 | 73,075 |

Table 2.5 - Incidental catch ( t ) of non-target species groups by GOA Pacific cod fisheries, 2003-2012 (as of 28 September 2012)

| Species/group | 2003 | 2004 | 2005 | 2006 | 2007 | 2008 | 2009 | 2010 | 2011 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Benthic urochordata | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.6 | 3.0 | 0.0 | 0.2 |
| Birds | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Bivalves | 1.3 | 0.3 | 1.3 | 2.1 | 1.2 | 1.7 | 4.2 | 2.7 | 6.1 |
| Brittle star unidentified | 0.0 | 0.1 | 0.2 | 0.1 | 0.3 | 0.1 | 0.0 | 0.1 | 2.1 |
| Capelin | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Corals Bryozoans | 0.0 | 0.1 | 0.0 | 0.1 | 0.2 | 0.0 | 1.7 | 0.0 | 0.7 |
| Dark Rockfish | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.3 | 2.7 | 12.3 | 2.4 |
| Eelpouts | 0.3 | 0.0 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 | 0.1 | 0.0 |
| Eulachon | 0.0 | 0.0 | 0.0 | 2.4 | 0.0 | 0.1 | 0.0 | 0.6 | 0.0 |
| Giant Grenadier | 0.0 | 0.0 | 0.0 | 21.9 | 81.5 | 31.0 | 51.3 | 140.9 | 60.4 |
| Greenlings | 3.1 | 6.2 | 1.5 | 3.7 | 0.8 | 7.1 | 1.3 | 0.8 | 0.8 |
| Grenadier | 5.2 | 0.4 | 0.0 | 0.6 | 0.0 | 66.0 | 6.6 | 0.0 | 8.2 |
| Gunnels | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Hermit crab unidentified | 1.0 | 0.1 | 0.4 | 0.5 | 1.7 | 2.9 | 3.9 | 2.1 | 0.8 |
| Invertebrate unidentified | 0.5 | 3.7 | 0.0 | 12.6 | 1.6 | 1.3 | 0.1 | 1.1 | 8.9 |
| Misc crabs | 1.0 | 0.3 | 1.7 | 0.7 | 6.6 | 2.4 | 1.5 | 3.4 | 2.4 |
| Misc crustaceans | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |  |
| Misc fish | 85.6 | 136.5 | 152.5 | 176.0 | 539.4 | 210.5 | 99.0 | 87.6 | 134.0 |
| Misc inverts (worms etc) | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| Other osmerids | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 |
| Pacific Sand lance | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 | 0.0 |
| Pandalid shrimp | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.2 | 0.0 |
| Scypho jellies | 8.7 | 1.5 | 1.1 | 4.6 | 0.1 | 0.4 | 0.2 | 10.8 | 0.8 |
| Sea anemone unidentified | 0.9 | 1.5 | 0.7 | 0.3 | 5.1 | 6.0 | 6.6 | 7.3 | 8.8 |
| Sea pens whips | 0.1 | 0.0 | 0.0 | 3.2 | 1.0 | 0.0 | 3.3 | 3.1 | 1.4 |
| Sea star | 468.4 | 1009.3 | 937.7 | 703.5 | 299.0 | 316.5 | 471.9 | 869.4 | 717.0 |
| Snails | 5.0 | 0.6 | 4.8 | 2.9 | 0.8 | 0.9 | 2.5 | 0.7 | 1.3 |
| Sponge unidentified | 0.2 | 0.6 | 1.0 | 1.2 | 0.0 | 1.1 | 1.6 | 0.4 | 0.5 |
| Stichaeidae | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 1.8 | 0.0 | 0.0 |
| Surf smelt | 0.0 | 0.0 | 0.4 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 |
| urchins dollars cucumbers | 0.8 | 0.5 | 1.1 | 1.0 | 3.2 | 0.5 | 1.3 | 0.5 | 2.2 |

Table 2.6 Catch ( t ) of Pacific cod by year, gear, and season for the years 1991-2012 as configured in the stock assessment models (as of 3 November 2012); values for 2012 season 5 were estimated given the average fraction of catch in season 5 for $2000-2011$ ( 0.030931 ) and the average fraction of each gear type in season 5 for $2000-2011$ ( $0.110294,0.315936$, 0.57377 , for trawl, longline, and pot, respectively).

| Year | Trawl |  |  | Longline |  |  |  | Pot |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Jan-Apr | May-Aug | Sep-Dec Total | Jan-Apr | May-Aug | Sep-Dec | Total | Jan-Apr M | May-Aug | Sep-Dec | Total |
| 1991 | 55,862 | 778 | 1,493 58,133 | 7,052 | 540 | 72 | 7,664 | 9,413 | 183 | 934 | 10,530 |
| 1992 | 51,479 | 1,828 | 1,500 54,807 | 12,545 | 966 | 2,243 | 15,754 | 9,698 | 19 | 470 | 10,187 |
| 1993 | 33,637 | 2,625 | 1,551 37,813 | 7,999 | 784 | 181 | 8,964 | 9,384 | 326 | 0 | 9,710 |
| 1994 | 29,150 | 1,433 | 877 31,460 | 6,431 | 299 | 52 | 6,782 | 8,714 | 33 | 496 | 9,243 |
| 1995 | 38,198 | 1,117 | 2,597 41,912 | 10,553 | 214 | 227 | 10,994 | 15,410 | 76 | 592 | 16,078 |
| 1996 | 40,506 | 4,023 | 1,494 46,023 | 9,885 | 215 | 106 | 10,206 | 12,025 | 27 | 0 | 12,052 |
| 1997 | 40,407 | 1,970 | 6,044 48,421 | 10,213 | 390 | 379 | 10,982 | 13,411 | 2,356 | 1,848 | 17,615 |
| 1998 | 34,372 | 4,014 | 3,200 41,586 | 9,307 | 444 | 264 | 10,015 | 17,652 | 2,137 | 1,136 | 20,925 |
| 1999 | 30,122 | 1,520 | 5,550 37,192 | 11,808 | 403 | 158 | 12,369 | 22,793 | 6,859 | 2,572 | 32,224 |
| 2000 | 21,579 | 3,148 | 750 25,477 | 11,401 | 170 | 107 | 11,678 | 25,768 | 2,938 | 699 | 29,405 |
| 2001 | 14,522 | 2,753 | 7,228 24,503 | 9,644 | 135 | 142 | 9,921 | 12,275 | 2,885 | 1,958 | 17,118 |
| 2002 | 14,466 | 4,069 | 1,309 19,844 | 11,410 | 161 | 3,159 | 14,730 | 13,049 | 2,288 | 4,573 | 19,910 |
| 2003 | 10,796 | 3,780 | 5,271 19,847 | 8,932 | 579 | 765 | 10,276 | 19,399 | 0 | 3,057 | 22,456 |
| 2004 | 9,221 | 2,429 | 6,400 18,050 | 8,259 | 268 | 2,046 | 10,573 | 23,334 | 276 | 4,392 | 28,002 |
| 2005 | 9,658 | 2,131 | 3,159 14,948 | 3,838 | 174 | 1,875 | 5,887 | 21,361 | 250 | 5,139 | 26,750 |
| 2006 | 10,028 | 2,081 | 1,332 13,441 | 6,156 | 251 | 3,948 | 10,355 | 21,417 | 261 | 2,381 | 24,059 |
| 2007 | 9,613 | 2,356 | 3,127 15,096 | 7,094 | 401 | 4,262 | 11,757 | 20,066 | 546 | 3,998 | 24,610 |
| 2008 | 11,156 | 4,107 | 6,118 21,381 | 9,312 | 641 | 2,618 | 12,571 | 20,394 | 0 | 4,600 | 24,994 |
| 2009 | 6,876 | 4,613 | 3,880 15,369 | 9,598 | 1,374 | 3,954 | 14,926 | 19,026 | 0 | 3,596 | 22,622 |
| 2010 | 11,009 | 5,110 | 7,728 23,847 | 11,672 | 776 | 5,130 | 17,578 | 31,009 | 1 | 5,638 | 36,648 |
| 2011 | 9,571 | 1,939 | 5,734 17,244 | 10,248 | 1,229 | 6,301 | 17,778 | 36,953 | 102 | 12,764 | 49,819 |
| 2012 | 17,131 | 2,685 | 2,566 22,382 | 12,178 | 560 | 2,378 | 15,116 | 30,068 | 138 | 8,247 | 38,453 |

Table 2.7 Age compositions observed by the sub-27 and 27-plus GOA bottom trawl survey, 19872011. Nact = actual sample size (these get rescaled so that the average across the combined sub-27 and 27-plus age compositions equals 300). The record for 1987 is shaded to indicate that these data are ignored in the fitting process due to very low sample size.

| Year Nact | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |  |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 28 | 0.000 | 0.921 | 0.078 | 0.002 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 20 | 0.000 | 0.995 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 110 | 0.000 | 0.981 | 0.020 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 100 | 0.000 | 0.951 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 98 | 0.000 | 0.971 | 0.029 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 125 | 0.000 | 0.919 | 0.081 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 57 | 0.000 | 0.895 | 0.105 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 65 | 0.000 | 0.870 | 0.130 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 93 | 0.000 | 0.997 | 0.004 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 83 | 0.000 | 0.937 | 0.053 | 0.010 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 66 | 0.000 | 0.981 | 0.019 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Year | Nact | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| 1987 | 110 | 0.000 | 0.006 | 0.248 | 0.253 | 0.251 | 0.157 | 0.055 | 0.019 | 0.009 | 0.002 | 0.001 | 0.000 | 0.000 |
| 1990 | 473 | 0.000 | 0.002 | 0.078 | 0.261 | 0.253 | 0.200 | 0.120 | 0.049 | 0.025 | 0.008 | 0.002 | 0.000 | 0.000 |
| 1993 | 750 | 0.000 | 0.004 | 0.102 | 0.242 | 0.288 | 0.202 | 0.112 | 0.030 | 0.016 | 0.004 | 0.001 | 0.000 | 0.000 |
| 1996 | 671 | 0.000 | 0.002 | 0.064 | 0.180 | 0.216 | 0.222 | 0.201 | 0.093 | 0.016 | 0.005 | 0.001 | 0.001 | 0.000 |
| 1999 | 584 | 0.000 | 0.001 | 0.052 | 0.173 | 0.239 | 0.278 | 0.161 | 0.058 | 0.026 | 0.009 | 0.002 | 0.001 | 0.000 |
| 2001 | 626 | 0.000 | 0.013 | 0.115 | 0.251 | 0.223 | 0.168 | 0.131 | 0.066 | 0.023 | 0.007 | 0.003 | 0.000 | 0.001 |
| 2003 | 654 | 0.000 | 0.001 | 0.032 | 0.188 | 0.275 | 0.285 | 0.133 | 0.052 | 0.027 | 0.004 | 0.001 | 0.001 | 0.001 |
| 2005 | 471 | 0.000 | 0.000 | 0.075 | 0.125 | 0.224 | 0.289 | 0.170 | 0.045 | 0.034 | 0.019 | 0.012 | 0.003 | 0.003 |
| 2007 | 378 | 0.000 | 0.018 | 0.279 | 0.295 | 0.156 | 0.110 | 0.039 | 0.023 | 0.014 | 0.027 | 0.022 | 0.002 | 0.014 |
| 2009 | 463 | 0.000 | 0.000 | 0.100 | 0.337 | 0.316 | 0.174 | 0.052 | 0.011 | 0.007 | 0.002 | 0.001 | 0.000 | 0.000 |
| 2011 | 753 | 0.000 | 0.001 | 0.106 | 0.415 | 0.291 | 0.148 | 0.034 | 0.005 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 |


| Table 2.8 - Relative sub-27 and 27-plus size composition from the 1984-2011 bottom trawl surveys (in bins of 4 to $120+\mathrm{cm}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 |
| N | 53 | 39 | 22 | 78 | 89 | 58 | 68 | 22 | 37 | 125 | 103 | 34 |
| 4 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 | 0 | 0 | 0 |
| 6 | 7 | 0 | 9 | 26 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| 7 | 65 | 0 | 27 | 14 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |
| 8 | 144 | 14 | 27 | 3 | 2 | 0 | 0 | 0 | 1 | 0 | 4 | 0 |
| 9 | 168 | 30 | 61 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 10 | 94 | 56 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 11 | 39 | 30 | 0 | 0 | 0 | 2 | 5 | 0 | 0 | 2 | 12 | 1 |
| 12 | 6 | 20 | 1 | 1 | 1 | 5 | 9 | 0 | 1 | 2 | 46 | 1 |
| 13 | 3 | 2 | 0 | 7 | 4 | 3 | 15 | 0 | 2 | 4 | 179 | 6 |
| 14 | 1 | 1 | 0 | 11 | 10 | 21 | 39 | 5 | 6 | 23 | 420 | 4 |
| 15 | 0 | 4 | 0 | 28 | 17 | 22 | 42 | 3 | 24 | 80 | 391 | 7 |
| 16 | 7 | 11 | 2 | 95 | 31 | 37 | 35 | 14 | 30 | 159 | 221 | 29 |
| 17 | 2 | 22 | 4 | 159 | 49 | 58 | 39 | 11 | 40 | 228 | 70 | 38 |
| 18 | 6 | 36 | 4 | 149 | 86 | 72 | 67 | 12 | 47 | 282 | 30 | 42 |
| 19 | 10 | 39 | 12 | 121 | 98 | 86 | 78 | 32 | 62 | 292 | 22 | 38 |
| 20 | 12 | 59 | 21 | 71 | 176 | 104 | 68 | 32 | 53 | 219 | 22 | 50 |
| 21 | 15 | 38 | 25 | 60 | 135 | 89 | 87 | 51 | 67 | 173 | 20 | 57 |
| 22 | 24 | 51 | 21 | 74 | 161 | 106 | 69 | 46 | 65 | 115 | 22 | 34 |
| 23 | 20 | 48 | 34 | 79 | 184 | 89 | 103 | 39 | 41 | 83 | 23 | 47 |
| 24 | 45 | 58 | 28 | 103 | 173 | 60 | 122 | 25 | 43 | 70 | 14 | 61 |
| 25 | 42 | 37 | 19 | 95 | 113 | 51 | 112 | 33 | 27 | 57 | 12 | 55 |
| 26 | 66 | 15 | 7 | 62 | 77 | 55 | 120 | 28 | 31 | 65 | 7 | 24 |
| Year | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 |
| N | 1126 | 1287 | 752 | 1083 | 736 | 528 | 390 | 596 | 426 | 490 | 1090 | 771 |
| 27 | 100 | 28 | 45 | 74 | 138 | 26 | 73 | 27 | 36 | 110 | 37 | 22 |
| 28 | 138 | 37 | 19 | 65 | 106 | 23 | 85 | 19 | 23 | 83 | 48 | 19 |
| 29 | 186 | 64 | 14 | 53 | 48 | 21 | 68 | 20 | 11 | 89 | 82 | 17 |
| 30 | 141 | 130 | 23 | 60 | 11 | 38 | 72 | 16 | 25 | 87 | 101 | 21 |
| 31 | 149 | 184 | 52 | 88 | 39 | 36 | 63 | 21 | 28 | 66 | 126 | 35 |
| 32 | 172 | 222 | 45 | 115 | 60 | 61 | 48 | 28 | 50 | 82 | 124 | 95 |
| 33 | 199 | 230 | 64 | 135 | 87 | 87 | 63 | 38 | 52 | 83 | 173 | 78 |
| 34 | 171 | 373 | 78 | 143 | 78 | 105 | 52 | 53 | 70 | 93 | 215 | 139 |
| 35 | 229 | 524 | 91 | 170 | 69 | 103 | 65 | 69 | 68 | 107 | 324 | 233 |
| 36 | 164 | 507 | 71 | 155 | 106 | 110 | 80 | 84 | 88 | 135 | 357 | 240 |
| 37 | 176 | 543 | 84 | 169 | 137 | 113 | 98 | 83 | 72 | 166 | 429 | 276 |
| 38 | 219 | 600 | 133 | 192 | 138 | 110 | 128 | 129 | 84 | 197 | 359 | 183 |
| 39 | 212 | 601 | 137 | 219 | 174 | 123 | 165 | 147 | 61 | 210 | 403 | 281 |
| 40 | 375 | 637 | 138 | 259 | 180 | 112 | 119 | 185 | 64 | 265 | 386 | 206 |
| 41 | 403 | 582 | 157 | 328 | 204 | 146 | 163 | 164 | 80 | 257 | 358 | 391 |
| 42 | 339 | 409 | 207 | 393 | 208 | 159 | 149 | 216 | 79 | 295 | 349 | 429 |
| 43 | 413 | 388 | 142 | 396 | 178 | 201 | 141 | 243 | 76 | 315 | 362 | 539 |
| 44 | 393 | 455 | 224 | 495 | 170 | 191 | 139 | 294 | 64 | 245 | 364 | 512 |
| 45 | 423 | 249 | 224 | 435 | 191 | 203 | 152 | 281 | 78 | 252 | 370 | 562 |
| 46 | 475 | 272 | 232 | 410 | 168 | 211 | 125 | 258 | 86 | 235 | 375 | 487 |


| Table 2.8 - Relative sub-27 and 27-plus size composition from the 1984-2011 bottom trawl surveys (in bins of 4 to $120+\mathrm{cm}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 |
| N | 53 | 39 | 22 | 78 | 89 | 58 | 68 | 22 | 37 | 125 | 103 | 34 |
| 47 | 448 | 301 | 325 | 474 | 172 | 197 | 138 | 236 | 74 | 163 | 343 | 436 |
| 48 | 491 | 442 | 442 | 335 | 150 | 230 | 143 | 262 | 98 | 203 | 413 | 395 |
| 49 | 629 | 453 | 495 | 357 | 207 | 196 | 178 | 244 | 90 | 169 | 384 | 384 |
| 50 | 719 | 547 | 396 | 434 | 222 | 233 | 128 | 261 | 118 | 174 | 417 | 358 |
| 51 | 829 | 694 | 366 | 411 | 201 | 176 | 170 | 226 | 116 | 131 | 491 | 308 |
| 52 | 704 | 694 | 370 | 447 | 279 | 217 | 140 | 273 | 162 | 130 | 523 | 283 |
| 53 | 689 | 709 | 345 | 590 | 297 | 217 | 129 | 265 | 185 | 104 | 503 | 327 |
| 54 | 702 | 661 | 417 | 672 | 258 | 238 | 173 | 294 | 275 | 136 | 637 | 369 |
| 55 | 691 | 706 | 324 | 696 | 339 | 245 | 145 | 336 | 248 | 144 | 529 | 373 |
| 56 | 550 | 611 | 505 | 587 | 301 | 244 | 144 | 282 | 311 | 133 | 734 | 349 |
| 57 | 609 | 521 | 390 | 631 | 349 | 296 | 135 | 328 | 308 | 179 | 587 | 368 |
| 58 | 556 | 588 | 391 | 663 | 369 | 297 | 150 | 314 | 357 | 189 | 671 | 328 |
| 59 | 473 | 517 | 405 | 523 | 314 | 288 | 141 | 365 | 318 | 190 | 605 | 320 |
| 60 | 454 | 536 | 390 | 563 | 386 | 261 | 148 | 301 | 359 | 164 | 556 | 273 |
| 61 | 376 | 426 | 433 | 484 | 443 | 260 | 143 | 300 | 252 | 178 | 438 | 250 |
| 62 | 270 | 413 | 312 | 396 | 458 | 232 | 130 | 271 | 252 | 160 | 406 | 254 |
| 63 | 339 | 391 | 281 | 351 | 383 | 197 | 185 | 236 | 214 | 163 | 372 | 216 |
| 64 | 222 | 428 | 289 | 325 | 340 | 191 | 146 | 201 | 195 | 143 | 410 | 217 |
| 65 | 213 | 343 | 275 | 319 | 439 | 183 | 153 | 161 | 120 | 116 | 238 | 143 |
| 66 | 167 | 274 | 240 | 362 | 387 | 163 | 106 | 181 | 120 | 127 | 272 | 129 |
| 67 | 141 | 275 | 181 | 312 | 327 | 142 | 112 | 134 | 87 | 109 | 231 | 110 |
| 68 | 134 | 207 | 147 | 278 | 262 | 130 | 88 | 121 | 94 | 83 | 161 | 91 |
| 69 | 133 | 198 | 133 | 216 | 271 | 130 | 66 | 103 | 69 | 93 | 123 | 59 |
| 70 | 82 | 142 | 173 | 193 | 236 | 107 | 94 | 96 | 62 | 79 | 131 | 54 |
| 71 | 55 | 140 | 119 | 137 | 160 | 81 | 59 | 74 | 52 | 59 | 98 | 44 |
| 72 | 74 | 126 | 122 | 150 | 112 | 82 | 61 | 83 | 40 | 60 | 75 | 45 |
| 73 | 81 | 85 | 91 | 123 | 93 | 69 | 35 | 69 | 34 | 42 | 72 | 27 |
| 74 | 42 | 102 | 79 | 96 | 131 | 54 | 49 | 51 | 29 | 44 | 44 | 14 |
| 75 | 64 | 73 | 51 | 64 | 59 | 36 | 36 | 46 | 27 | 51 | 65 | 17 |
| 76 | 66 | 65 | 50 | 72 | 65 | 35 | 35 | 32 | 33 | 21 | 45 | 17 |
| 77 | 60 | 34 | 42 | 59 | 47 | 34 | 30 | 36 | 20 | 31 | 36 | 11 |
| 78 | 31 | 53 | 37 | 56 | 50 | 32 | 17 | 36 | 16 | 20 | 23 | 9 |
| 79 | 84 | 47 | 36 | 46 | 52 | 22 | 13 | 39 | 9 | 10 | 18 | 6 |
| 80 | 95 | 26 | 32 | 28 | 25 | 10 | 20 | 30 | 17 | 11 | 15 | 13 |
| 81 | 41 | 20 | 29 | 27 | 27 | 10 | 12 | 28 | 12 | 10 | 11 | 4 |
| 82 | 46 | 14 | 29 | 25 | 23 | 18 | 5 | 25 | 10 | 6 | 18 | 4 |
| 83 | 25 | 22 | 23 | 14 | 15 | 12 | 4 | 14 | 11 | 7 | 6 | 2 |
| 84 | 30 | 14 | 30 | 16 | 14 | 5 | 10 | 16 | 23 | 3 | 5 | 5 |
| 85 | 27 | 23 | 15 | 15 | 12 | 11 | 2 | 12 | 11 | 2 | 5 | 1 |
| 86 | 20 | 13 | 14 | 12 | 15 | 3 | 5 | 9 | 4 | 2 | 4 | 2 |
| 87 | 22 | 7 | 10 | 10 | 9 | 3 | 4 | 8 | 11 | 4 | 4 | 2 |
| 88 | 17 | 6 | 14 | 8 | 7 | 2 | 3 | 4 | 11 | 1 | 2 | 2 |
| 89 | 11 | 8 | 21 | 15 | 7 | 8 | 4 | 3 | 9 | 3 | 2 | 1 |
| 90 | 14 | 2 | 11 | 4 | 8 | 1 | 2 | 4 | 26 | 2 | 1 | 1 |
| 91 | 6 | 2 | 10 | 4 | 13 | 2 | 2 | 6 | 17 | 1 | 0 | 1 |


| Table 2.8 - Relative sub-27 and 27-plus size composition from the 1984-2011 bottom trawl surveys (in bins of 4 to $120+\mathrm{cm}$ ) |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | 1984 | 1987 | 1990 | 1993 | 1996 | 1999 | 2001 | 2003 | 2005 | 2007 | 2009 | 2011 |
| N | 53 | 39 | 22 | 78 | 89 | 58 | 68 | 22 | 37 | 125 | 103 | 34 |
| 92 | 9 | 4 | 4 | 4 | 8 | 0 | 2 | 4 | 17 | 1 | 2 | 0 |
| 93 | 3 | 2 | 3 | 6 | 6 | 4 | 7 | 1 | 19 | 3 | 1 | 0 |
| 94 | 2 | 4 | 8 | 4 | 1 | 0 | 2 | 0 | 10 | 1 | 24 | 0 |
| 95 | 7 | 2 | 2 | 5 | 7 | 1 | 2 | 1 | 22 | 3 | 1 | 0 |
| 96 | 1 | 4 | 5 | 4 | 3 | 0 | 2 | 1 | 20 | 1 | 1 | 1 |
| 97 | 1 | 0 | 3 | 3 | 5 | 2 | 2 | 1 | 19 | 1 | 1 | 0 |
| 98 | 2 | 1 | 3 | 3 | 2 | 2 | 0 | 1 | 20 | 1 | 0 | 0 |
| 99 | 0 | 2 | 1 | 2 | 1 | 4 | 0 | 0 | 18 | 0 | 1 | 0 |
| 100 | 1 | 1 | 1 | 4 | 4 | 3 | 2 | 0 | 10 | 1 | 1 | 0 |
| 101 | 1 | 2 | 3 | 3 | 1 | 0 | 2 | 0 | 13 | 1 | 1 | 0 |
| 102 | 0 | 1 | 2 | 2 | 7 | 1 | 1 | 0 | 4 | 1 | 1 | 0 |
| 103 | 0 | 0 | 1 | 1 | 2 | 0 | 0 | 0 | 8 | 2 | 0 | 0 |
| 104 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 11 | 1 | 0 | 0 |
| 105 | 0 | 1 | 1 | 1 | 2 | 0 | 0 | 0 | 5 | 0 | 1 | 0 |
| 106 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 |
| 107 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 108 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| 109 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 110 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |
| 111 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 112 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 113 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 114 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 115 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 116 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 117 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 118 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 119 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 120+ | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Table 2.9 - Mean size-at-age (in cm) observed by the sub-27 and 27-plus GOA bottom trawl survey, 1987-2011

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 0.000 | 20.016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.000 | 21.835 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1993 | 0.000 | 20.384 | 25.652 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1996 | 0.000 | 20.440 | 25.366 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1999 | 0.000 | 20.571 | 26.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2001 | 0.000 | 21.141 | 25.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2003 | 0.000 | 21.131 | 25.041 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2005 | 0.000 | 18.941 | 24.493 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2007 | 0.000 | 17.383 | 26.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2009 | 0.000 | 19.794 | 24.898 | 25.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2011 | 0.000 | 20.829 | 25.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| 1987 | 0.000 | 0.000 | 34.251 | 43.215 | 52.832 | 59.235 | 64.794 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 1990 | 0.000 | 27.262 | 35.068 | 45.917 | 53.472 | 59.940 | 65.134 | 70.773 | 77.170 | 83.949 | 89.101 | 98.223 | 102.518 |
| 1993 | 0.000 | 27.547 | 34.306 | 44.040 | 52.123 | 58.893 | 65.611 | 70.367 | 74.692 | 87.551 | 94.429 | 97.411 | 0.000 |
| 1996 | 0.000 | 27.101 | 32.319 | 41.564 | 52.395 | 59.236 | 64.132 | 68.530 | 75.524 | 82.825 | 93.850 | 97.313 | 85.989 |
| 1999 | 0.000 | 27.361 | 32.955 | 41.050 | 48.717 | 58.167 | 64.406 | 71.194 | 71.791 | 77.824 | 80.160 | 83.688 | 0.000 |
| 2001 | 0.000 | 27.444 | 32.840 | 42.651 | 52.148 | 58.807 | 65.611 | 70.623 | 74.937 | 84.301 | 86.745 | 85.000 | 78.723 |
| 2003 | 0.000 | 29.298 | 32.645 | 43.834 | 48.972 | 57.854 | 64.947 | 71.741 | 75.490 | 84.096 | 83.477 | 75.670 | 75.965 |
| 2005 | 0.000 | 0.000 | 33.353 | 41.202 | 51.274 | 57.144 | 62.322 | 68.165 | 78.232 | 90.879 | 95.862 | 95.153 | 91.745 |
| 2007 | 0.000 | 27.470 | 35.212 | 43.362 | 55.483 | 59.665 | 63.519 | 70.055 | 69.838 | 98.805 | 103.660 | 92.826 | 0.000 |
| 2009 | 0.000 | 27.000 | 33.708 | 44.697 | 55.494 | 61.956 | 65.694 | 74.054 | 74.209 | 84.884 | 92.512 | 0.000 | 0.000 |
| 2011 | 0.000 | 27.000 | 35.708 | 44.863 | 53.947 | 62.018 | 65.501 | 75.620 | 83.818 | 0.000 | 93.530 | 0.000 | 106.283 |

Table 2.10 - Sample sizes of fish for the mean size-at-age observed by the sub-27 and 27-plus GOA bottom trawl survey, 1987-2011

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1987 | 0 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1993 | 0 | 108 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1996 | 0 | 92 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1999 | 0 | 95 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2001 | 0 | 113 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2003 | 0 | 52 | 5 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2005 | 0 | 50 | 15 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2007 | 0 | 92 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2009 | 0 | 77 | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2011 | 0 | 65 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | $12+$ |
| 1987 | 0 | 0 | 20 | 56 | 22 | 11 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 1990 | 0 | 3 | 50 | 95 | 81 | 78 | 59 | 41 | 36 | 20 | 7 | 2 | 1 |
| 1993 | 0 | 9 | 90 | 116 | 113 | 113 | 117 | 66 | 53 | 23 | 10 | 2 | 0 |
| 1996 | 0 | 2 | 45 | 146 | 123 | 100 | 107 | 92 | 34 | 17 | 3 | 1 | 1 |
| 1999 | 0 | 1 | 26 | 76 | 119 | 136 | 103 | 58 | 29 | 11 | 5 | 2 | 0 |
| 2001 | 0 | 9 | 87 | 120 | 106 | 81 | 84 | 64 | 34 | 15 | 8 | 3 | 1 |
| 2003 | 0 | 2 | 37 | 114 | 134 | 126 | 86 | 60 | 39 | 10 | 1 | 2 | 2 |
| 2005 | 0 | 0 | 64 | 87 | 83 | 78 | 84 | 39 | 21 | 6 | 4 | 1 | 1 |
| 2007 | 0 | 5 | 47 | 86 | 73 | 65 | 34 | 36 | 25 | 4 | 1 | 2 | 0 |
| 2009 | 0 | 1 | 60 | 120 | 105 | 86 | 47 | 19 | 16 | 5 | 4 | 0 | 0 |
| 2011 | 0 | 1 | 102 | 189 | 178 | 175 | 76 | 25 | 5 | 0 | 1 | 0 | 1 |

Table 2.11 Pacific cod abundance measured in biomass ( t ) and numbers of fish (1000s), as assessed by the GOA bottom trawl survey. Point estimates are shown along with coefficients of variation. The two right-hand sections show the total abundance divided into fish 27 cm or larger and fish smaller than 27 cm (totals are very slightly different in the first four years due to exclusion of tows with no length data from the strata extrapolations).

|  |  | All lengths |  | 27-plus |  |  | Sub-27cm |  |
| ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Year | Biomass(t) | CV | Abundance | CV | Abundance | CV | Abundance | CV |
| 1984 | 550,971 | 0.146 | 320,525 | 0.156 | 296,057 | 0.175 | 19,526 | 0.596 |
| 1987 | 394,987 | 0.130 | 247,020 | 0.185 | 238,165 | 0.234 | 6,772 | 0.374 |
| 1990 | 416,788 | 0.153 | 212,132 | 0.208 | 193,577 | 0.243 | 14,739 | 0.412 |
| 1993 | 409,848 | 0.179 | 231,963 | 0.190 | 214,244 | 0.210 | 17,021 | 0.372 |
| 1996 | 538,154 | 0.200 | 319,068 | 0.215 | 234,528 | 0.172 | 84,540 | 0.615 |
| 1999 | 306,413 | 0.126 | 166,584 | 0.112 | 157,019 | 0.118 | 9,565 | 0.272 |
| 2001 | 257,614 | 0.204 | 158,424 | 0.180 | 137,041 | 0.203 | 21,384 | 0.270 |
| 2003 | 297,402 | 0.150 | 159,749 | 0.129 | 153,895 | 0.134 | 5,854 | 0.231 |
| 2005 | 308,091 | 0.262 | 139,852 | 0.208 | 127,282 | 0.221 | 12,570 | 0.388 |
| 2007 | 233,310 | 0.139 | 192,025 | 0.175 | 134,261 | 0.163 | 57,764 | 0.425 |
| 2009 | 752,651 | 0.303 | 573,509 | 0.286 | 422,370 | 0.239 | 151,139 | 0.867 |
| 2011 | 500,975 | 0.136 | 348,060 | 0.177 | 339,410 | 0.178 | 8,650 | 0.347 |

Table 2.12. Number of parameters, negative log-likelihoods, and "effective q" for all model configurations (smaller indicates better fit to data; shaded areas indicate comparable categories).

|  | 2011 model | 2011 model no tc | Model 1 | Model 2 | Model 3 | Model 4 | Model 5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Number of parameters | 253 | 253 | 245 | 239 | 245 | 239 | 245 |
| Likelihood components (-ln) |  |  |  |  |  |  |  |
| Survey indices | 3 | 4 | 15 | 5 | 4 | -7 | 15 |
| Length compositions | 3,075 | 3,187 | 3,147 | 2,895 | 3,089 | 2,862 | 3,124 |
| Age compositions | 118 | 162 | 151 | 82 | 129 | 42 | 157 |
| Size-at-age | 623 | 637 | 642 | 501 | 0 | 0 | 557 |
| Recruitment | -22 | -22 | -23 | -24 | -22 | -21 | -23 |
| 27-plus survey indices | 13 | 14 | 12 | 5 | 2 | -7 | 12 |
| Sub-27 survey indices | -10 | -10 | 4 | - | 2 | 2 | 3 |
| Total | 3,797 | 3,967 | 3,931 | 3,459 | 3,199 | 2,875 | 3,831 |
| Length composition likelihoods (-ln) |  |  |  |  |  |  |  |
| Jan-Apr Trawl | 713 | 710 | 712 | 698 | 709 | 671 | 707 |
| Jan-Apr LL | 184 | 187 | 185 | 180 | 183 | 176 | 187 |
| Jan-Apr Pot | 285 | 288 | 286 | 281 | 287 | 277 | 288 |
| May-Aug Trawl | 598 | 609 | 612 | 571 | 593 | 586 | 616 |
| May-Aug LL | 112 | 117 | 117 | 107 | 114 | 111 | 117 |
| May-Aug Pot | 279 | 284 | 286 | 260 | 271 | 260 | 283 |
| Sep-Dec Trawl | 465 | 470 | 470 | 461 | 476 | 457 | 469 |
| Sep-Dec LL | 72 | 73 | 72 | 70 | 71 | 69 | 71 |
| Sep-Dec Pot | 180 | 184 | 184 | 179 | 184 | 179 | 184 |
| 27-plus survey | 109 | 130 | 127 | 89 | 100 | 78 | 116 |
| Sub-27 survey | 77 | 137 | 96 | - | 99 | - | 86 |
| Age compositions likelihoods (-ln) |  |  |  |  |  |  |  |
| Age 27-plus | 101 | 110 | 112 | 82 | 88 | 42 | 113 |
| Age sub-27 | 17 | 52 | 39 | - | 41 | - | 44 |
| Size-at-age likelihoods (-ln) |  |  |  |  |  |  |  |
| Size-at-age 27-plus | 552 | 568 | 572 | 501 | 0 | 0 | 557 |
| Size-at-age sub-27 | 71 | 68 | 69 | - | 0 | - | 0 |
| "Effective q" | 0.893 | 0.890 | 0.860 | 0.877 | 0.829 | 0.794 | 0.865 |

Table 2.13 - Fixed and estimated parameters and their asymptotic standard deviations for Model 2

| Label | Value | SD |
| :---: | :---: | :---: |
| NatM_p_1_Fem_GP_1 | 0.38 |  |
| L_at_Amin_Fem_GP_1 | 24.8646 | 0.293326 |
| L_at_Amax_Fem_GP_1 | 100.506 | 0.700256 |
| VonBert_K_Fem_GP_1 | 0.175614 | 0.002925 |
| CV_young_Fem_GP_1 | 3.13 |  |
| CV_old_Fem_GP_1 | 6.55 |  |
| Wtlen_1_Fem | 8.84E-06 |  |
| Wtlen_2_Fem | 3.07181 |  |
| Mat50\%_Fem | 4.35 |  |
| Mat_slope_Fem | -1.9632 |  |
| Eggs/kg_inter_Fem | 1 |  |
| Eggs/kg_slope_wt_Fem | 0 |  |
| RecrDist_GP_1 | 0 |  |
| RecrDist_A ${ }^{\text {rea_1 }}$ | 0 |  |
| RecrDist_Seas_1 | 0 |  |
| RecrDist_Seas_2 | 0 |  |
| RecrDist_Seas_3 | 0 |  |
| RecrDist_Seas_4 | 0 |  |
| RecrDist_Seas_5 | 0 |  |
| CohortGrowDev | 1 |  |
| AgeKeyParm1 | 1 |  |
| AgeKeyParm2 | 0.540713 | 0.010072 |
| AgeKeyParm3 | $9.50 \mathrm{E}-09$ |  |
| AgeKeyParm4 | 0 |  |
| AgeKeyParm5 | 0.096 |  |
| AgeKeyParm6 | 1.471 |  |
| AgeKeyParm7 | 0 |  |
| F-WL1_seas_1 | -0.00427 |  |
| F-WL1_seas_2 | -0.09781 |  |
| F-WL1_seas_3 | 0.260771 |  |
| F-WL1_seas_4 | 0.706346 |  |
| F-WL1_seas_5 | -0.20535 |  |
| F-WL2_seas_1 | 0.004059 |  |
| F-WL2_seas_2 | 0.005264 |  |
| F-WL2_seas_3 | -0.02684 |  |
| F-WL2_seas_4 | -0.05997 |  |
| F-WL2_seas_5 | 0.015615 |  |
| SR_LN(R0) | 12.406 | $0.03954 \overline{7}$ |
| SR_BH_steep | 1 | - |
| SR_sigmaR | 0.41 |  |
| SR_envlink | 0 |  |
| SR_R1_offset | -0.37431 | 0.126796 |
| SR_autocorr | 0 |  |
| Early_InitAge_13 | -0.24235 | 0.368934 |
| Early_InitAge_12 | -0.30545 | 0.360363 |
| Early_InitAge_11 | -0.36531 | 0.352674 |
| Early_InitAge_10 | -0.4119 | 0.346565 |
| Early_InitAge_9 | -0.41976 | 0.343688 |


| Label | Value | SD |
| :---: | :---: | :---: |
| Early_InitAge_8 | -0.35184 | 0.34594 |
| Early_InitAge_7 | -0.16307 | 0.354432 |
| Early_InitAge_6 | 0.122132 | 0.357053 |
| Early_InitAge_5 | 0.284351 | 0.342494 |
| Early_InitAge_4 | 1.11813 | 0.194055 |
| Early_InitAge_3 | 0.099255 | 0.21617 |
| Early_InitAge_2 | 0.093799 | 0.17006 |
| Early_InitAge_1 | -0.08392 | 0.177035 |
| Main_RecrDev_1977 | 1.1857 | 0.066999 |
| Main_RecrDev_1978 | -0.14115 | 0.107732 |
| Main_RecrDev_1979 | -0.25471 | 0.07766 |
| Main_RecrDev_1980 | -0.06752 | 0.061145 |
| Main_RecrDev_1981 | -0.06344 | 0.062944 |
| Main_RecrDev_1982 | 0.244591 | 0.065403 |
| Main_RecrDev_1983 | -0.05146 | 0.080939 |
| Main_RecrDev_1984 | -0.38361 | 0.093914 |
| Main_RecrDev_1985 | 0.401006 | 0.061908 |
| Main_RecrDev_1986 | -0.32373 | 0.077128 |
| Main_RecrDev_1987 | 0.099437 | 0.055324 |
| Main_RecrDev_1988 | 0.004479 | 0.056561 |
| Main_RecrDev_1989 | 0.141577 | 0.056315 |
| Main_RecrDev_1990 | 0.230005 | 0.052957 |
| Main_RecrDev_1991 | 0.089331 | 0.053531 |
| Main_RecrDev_1992 | -0.0721 | 0.059286 |
| Main_RecrDev_1993 | 0.035635 | 0.053221 |
| Main_RecrDev_1994 | 0.03759 | 0.050538 |
| Main_RecrDev_1995 | 0.037504 | 0.047647 |
| Main_RecrDev_1996 | -0.06485 | 0.048132 |
| Main_RecrDev_1997 | -0.35573 | 0.051844 |
| Main_RecrDev_1998 | -0.47488 | 0.054299 |
| Main_RecrDev_1999 | -0.19136 | 0.047833 |
| Main_RecrDev_2000 | -0.02899 | 0.044986 |
| Main_RecrDev_2001 | -0.35634 | 0.050699 |
| Main_RecrDev_2002 | -0.52407 | 0.054972 |
| Main_RecrDev_2003 | -0.31544 | 0.051633 |
| Main_RecrDev_2004 | -0.35593 | 0.058906 |
| Main_RecrDev_2005 | 0.099029 | 0.067964 |
| Main_RecrDev_2006 | 0.353826 | 0.082012 |
| Main_RecrDev_2007 | 0.300238 | 0.108483 |
| Main_RecrDev_2008 | 0.46615 | 0.122938 |
| Main_RecrDev_2009 | 0.392219 | 0.199382 |
| Main_RecrDev_2010 | -0.04233 | 0.392489 |
| Main_RecrDev_2011 | -0.05067 | 0.40328 |
| InitF_1 Jan-Apr_Trawl_Fishery | 0.075524 | 0.010688 |
| InitF_2May-Aug_Trawl_Fishery | 0 | - |
| InitF_3Sep-Dec_Trawl_Fishery | 0 | - |
| InitF_4Jan-Apr_Longline_Fishery | 0 |  |
| InitF_5May-Aug_Longline_Fishery | 0 | _ |
| InitF_6Sep-Dec_Longline_Fishery | 0 | - |
| InitF_7Jan-Apr_Pot_Fishery | 0 | - |


| Label | Value | SD |
| :---: | :---: | :---: |
| InitF_8May-Aug_Pot_Fishery | 0 | - |
| InitF_9Sep-Dec_Pot_Fishery | 0 |  |
| Q_envlink_10_27plus_Trawl_Survey | 0.545013 | 0.113976 |
| Q_base_10_27plus_Trawl_Survey | 0 |  |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery | 0 |  |
| SizeSel_1P_2_Jan-Apr_Trawl_Fishery | 0 |  |
| SizeSel_1P_3_Jan-Apr_Trawl_Fishery | 5.86635 | 0.028721 |
| SizeSel_1P_4_Jan-Apr_Trawl_Fishery | 0 | - |
| SizeSel_1P_5_Jan-Apr_Trawl_Fishery | -10 |  |
| SizeSel_1P_6_Jan-Apr_Trawl_Fishery | 10 |  |
| SizeSel_2P_1_May-Aug_Trawl_Fishery | 0 |  |
| SizeSel_2P_2_May-Aug_Trawl_Fishery | -9.67559 | 8.93329 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery | 0 |  |
| SizeSel_2P_4_May-Aug_Trawl_Fishery | 4.42432 | 0.445711 |
| SizeSel_2P_5_May-Aug_Trawl_Fishery | -10 |  |
| SizeSel_2P_6_May-Aug_Trawl_Fishery | 0 |  |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery | 0 |  |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery | 0 |  |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery | 0 |  |
| SizeSel_3P_4_Sep-Dec_Trawl_Fishery | 3.82519 | 0.580275 |
| SizeSel_3P_5_Sep-Dec_Trawl_Fishery | -10 |  |
| SizeSel_3P_6_Sep-Dec_Trawl_Fishery | -1.01729 | 0.266201 |
| SizeSel_4 $\mathrm{P}_{\text {_ }} 1$-Jan-Apr_Longline_Fishery | 0 |  |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery | 0 |  |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery | 0 |  |
| SizeSel_4P_4_Jan-Apr_Longline_Fishery | 3.98158 | 0.354505 |
| SizeSel_4P_5_Jan-Apr_Longline_Fishery | -10 | - |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery | 0 |  |
| SizeSel_5P_1_May-Aug_Longline_Fishery | 0 |  |
| SizeSel_5P_2_May-Aug_Longline_Fishery | -9.15336 | 19.8785 |
| SizeSel_5P_3_May-Aug_Longline_Fishery | 0 |  |
| SizeSel_5P_4_May-Aug_Longline_Fishery | 4.50634 | $0.49938 \overline{2}$ |
| SizeSel_5P_5_May-Aug_Longline_Fishery | -10 |  |
| SizeSel_5P_6_May-Aug_Longline_Fishery | 0 |  |
| SizeSel_6P_1_Sep-Dec_Longline_Fishery | 0 |  |
| SizeSel_6P_2_Sep-Dec_Longline_Fishery | -13.6793 |  |
| SizeSel_6P_3_Sep-Dec_Longline_Fishery | 0 |  |
| SizeSel_6P_4_Sep-Dec_Longline_Fishery | 0 |  |
| SizeSel_6P_5_Sep-Dec_Longline_Fishery | -10 |  |
| SizeSel_6P_6_Sep-Dec_Longline_Fishery | 0 |  |
| SizeSel_7P_1_Jan-Apr_Pot_Fishery | 0 |  |
| SizeSel_7P_2_Jan-Apr_Pot_Fishery | -14.3135 |  |
| SizeSel_7P_3_Jan-Apr_Pot_Fishery | 0 |  |
| SizeSel_7P_4_Jan-Apr_Pot_Fishery | 0 |  |
| SizeSel_7P_5_Jan-Apr_Pot_Fishery | -10 |  |
| SizeSel_7P_6_Jan-Apr_Pot_Fishery | 0 |  |
| SizeSel_8P_1_May-Aug_Pot_Fishery | 0 |  |
| SizeSel_8P_2_May-Aug_Pot_Fishery | $-9.05637$ | $21.644 \overline{5}$ |
| SizeSel_8P_3_May-Aug_Pot_Fishery | 0 |  |
| SizeSel_8P_4_May-Aug_Pot_Fishery | 5.08762 | $0.49577 \overline{3}$ |


| Label | Value | SD |
| :---: | :---: | :---: |
| SizeSel_8P_5_May-Aug_Pot_Fishery | -10 |  |
| SizeSel_8P_6_May-Aug_Pot_Fishery | -1.15075 | $0.59884 \overline{2}$ |
| SizeSē]_9 $\overline{\mathrm{P}}_{1} 1$ _Sep-Dec_Pot_Fishery | 0 |  |
| SizeSel_9P_2_Sep-Dec_Pot_Fishery | -5.45394 | 9.91802 |
| SizeSel_9P_3_Sep-Dec_Pot_Fishery | 0 |  |
| SizeSel_9P_4_Sep-Dec_Pot_Fishery | 4.62388 | $0.52381 \overline{6}$ |
| SizeSel_9P_5_Sep-Dec_Pot_Fishery | -10 |  |
| SizeSel_9P_6_Sep-Dec_Pot_Fishery | 0 |  |
| AgeSel_10P_1_27plus_Trawl_Survey | 0 |  |
| AgeSel_10P_2_27plus_Trawl_Survey | 0 |  |
| AgeSel_10P_3_27plus_Trawl_Survey | 0 |  |
| AgeSel_10P_4_27plus_Trawl_Survey | 0 |  |
| AgeSel_10P_5_27plus_Trawl_Survey | -10 |  |
| AgeSel_10P_6_27plus_Trawl_Survey | 0 |  |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery_BLK1repl_1977 | 62.6855 | $1.300 \overline{2}$ |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery_BLK1repl_1990 | 74.319 | 0.616703 |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery_BLK1repl_1995 | 75.8209 | 0.60921 |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery_BLK1repl_2000 | 70.2658 | 0.661814 |
| SizeSel_1P_1_Jan-Apr_Trawl_Fishery_BLK1repl_2005 | 71.2341 | 0.801381 |
| SizeSel_2P_1_May-Aug_Trawl_Fishery_BLK2repl_1977 | 55.3117 | 1.49943 |
| SizeSel_2P_1_May-Aug_Trawl_Fishery_BLK2repl_1985 | 62.6872 | 1.28172 |
| SizeSel_2P_1_May-Aug_Trawl_Fishery_BLK2repl_1990 | 67.23 | 1.14915 |
| SizeSel_2P_1_May-Aug_Trawl_Fishery_BLK2repl_2000 | 68.7701 | 1.66131 |
| SizeSel_2P_1_May-Aug_Trawl_Fishery_BLK2repl_2005 | 72.773 | 1.82712 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery_BLK2repl_1977 | 4.54026 | 0.220047 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery_BLK2repl_1985 | 5.15545 | 0.157182 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery_BLK2repl_1990 | 5.17156 | 0.120258 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery_BLK2repl_2000 | 5.7978 | 0.132544 |
| SizeSel_2P_3_May-Aug_Trawl_Fishery_BLK2repl_2005 | 5.93216 | 0.113571 |
| SizeSel_2P_6_May-Aug_Trawl_Fishery_BLK2repl_1977 | 0.135011 | 0.40069 |
| SizeSel_2P_6_May-Aug_Trawl_Fishery_BLK2repl_1985 | -0.98921 | 0.375539 |
| SizeSel_2P_6_May-Aug_Trawl_Fishery_BLK2repl_1990 | -2.48035 | 0.736248 |
| SizeSel_2P_6_May-Aug_Trawl_Fishery_BLK2repl_2000 | -1.10311 | 0.760783 |
| SizeSel_2P_6_May-Aug_Trawl_Fishery_BLK2repl_2005 | -1.38551 | 1.02363 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_1977 | 45.5521 | 4.3318 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_1980 | 56.972 | 3.6502 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_1985 | 61.9336 | 1.77117 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_1990 | 63.6057 | 3.05231 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_1995 | 75.3712 | 1.70177 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_2000 | 71.3116 | 2.4504 |
| SizeSel_3P_1_Sep-Dec_Trawl_Fishery_BLK3repl_2005 | 74.5874 | 1.52915 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_1977 | -2.36684 | 1.66575 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_1980 | -3.90494 | 5.18596 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_1985 | -9.16494 | 19.6762 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_1990 | -0.58724 | 0.273523 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_1995 | -9.27226 | 17.6423 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_2000 | -4.79988 | 9.87583 |
| SizeSel_3P_2_Sep-Dec_Trawl_Fishery_BLK3repl_2005 | -9.01982 | 22.2955 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_1977 | 3.74845 | 0.840923 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_1980 | 5.05386 | 0.360485 |


| Label | Value | SD |
| :---: | :---: | :---: |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_1985 | 5.70268 | 0.199345 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_1990 | 5.49662 | 0.246313 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_1995 | 6.26033 | 0.104742 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_2000 | 5.93847 | 0.141185 |
| SizeSel_3P_3_Sep-Dec_Trawl_Fishery_BLK3repl_2005 | 6.00674 | 0.089111 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_1977 | 57.6245 | 0.872187 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_1985 | 73.4914 | 1.31381 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_1990 | 70.6922 | 1.11366 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_1995 | 74.2486 | 0.804566 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_2000 | 69.3501 | 0.794999 |
| SizeSel_4P_1_Jan-Apr_Longline_Fishery_BLK4repl_2005 | 69.1428 | 0.500477 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_1977 | -0.45882 | 0.104061 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_1985 | -4.53878 | 4.74788 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_1990 | -5.10085 | 23.4596 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_1995 | -9.08594 | 21.1225 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_2000 | -4.50675 | 4.69456 |
| SizeSel_4P_2_Jan-Apr_Longline_Fishery_BLK4repl_2005 | -9.76238 | 6.76407 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_1977 | 4.67492 | 0.102688 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_1985 | 5.77666 | 0.076999 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_1990 | 5.30472 | 0.084134 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_1995 | 5.45472 | 0.061573 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_2000 | 5.12967 | 0.068966 |
| SizeSel_4P_3_Jan-Apr_Longline_Fishery_BLK4repl_2005 | 5.05858 | 0.045808 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_1977 | -0.83026 | 0.412807 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_1985 | 0.808976 | 0.27459 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_1990 | 1.19642 | 0.418355 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_1995 | 1.17252 | 0.477003 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_2000 | -0.11576 | 0.225128 |
| SizeSel_4P_6_Jan-Apr_Longline_Fishery_BLK4repl_2005 | 0.068151 | 0.216705 |
| SizeSel_5P_1_May-Aug_Longline_Fishery_BLK5repl_1977 | 56.3137 | 2.33386 |
| SizeSel_5P_1_May-Aug_Longline_Fishery_BLK5repl_1980 | 56.6541 | 0.849127 |
| SizeSel_-5P_1_May-Aug_Longline_Fishery_BLK5repl_1990 | 71.0133 | 2.23473 |
| SizeSel_5P_1_May-Aug_Longline_Fishery_BLK5repl_2000 | 72.2856 | 1.49444 |
| SizeSel_5P_3_May-Aug_Longline_Fishery_BLK5repl_1977 | 4.45164 | 0.333884 |
| SizeSel_5P_3_May-Aug_Longline_Fishery_BLK5repl_1980 | 4.37157 | 0.132142 |
| SizeSel_5P_3_May-Aug_Longline_Fishery_BLK5repl_1990 | 5.07006 | 0.257167 |
| SizeSel_5P_3_May-Aug_Longline_Fishery_BLK5repl_2000 | 4.99889 | 0.145532 |
| SizeSel_5P_6_May-Aug_Longline_Fishery_BLK5repl_1977 | 0.423222 | 0.664 |
| SizeSel_-5P_6_May-Aug_Longline_Fishery_BLK5repl_1980 | -0.52565 | 0.208549 |
| SizeSel_5P_6_May-Aug_Longline_Fishery_BLK5repl_1990 | -1.57636 | 1.18917 |
| SizeSel_5P_6_May-Aug_Longline_Fishery_BLK5repl_2000 | 1.05756 | 0.988512 |
| SizeSel_6P_1_Sep-Dec_Longline_Fishery_BLK6repl_1977 | 58.9952 | 1.958 |
| SizeSel_6P_1_Sep-Dec_Longline_Fishery_BLK6repl_1980 | 56.9678 | 0.387529 |
| SizeSel_6P_1_Sep-Dec_Longline_Fishery_BLK6repl_1990 | 68.8194 | 0.522085 |
| SizeSel_6P_3_Sep-Dec_Longline_Fishery_BLK6repl_1977 | 4.63818 | 0.232946 |
| SizeSel_6P_3_Sep-Dec_Longline_Fishery_BLK6repl_1980 | 4.32809 | 0.061538 |
| SizeSel_6P_3_Sep-Dec_Longline_Fishery_BLK6repl_1990 | 4.96458 | 0.052051 |
| SizeSel_6P_4_-Sep-Dec_Longline_Fishery_BLK6repl_1 1977 | 7.56207 | 0.847303 |
| SizeSel_6P_4_Sep-Dec_Longline_Fishery_BLK6repl_1980 | 4.16839 | 0.149849 |
| SizeSel_6P_4_-Sep-Dec_Longline_Fishery_BLK6repl_1990 | 3.86187 | 0.411326 |


| Label | Value | SD |
| :---: | :---: | :---: |
| SizeSel_6P_6_Sep-Dec_Longline_Fishery_BLK6repl_1977 | -8.45229 | 31.434 |
| SizeSel_6P_6_Sep-Dec_Longline_Fishery_BLK6repl_1980 | -1.37793 | 0.103302 |
| SizeSel_6P_6_Sep-Dec_Longline_Fishery_BLK6repl_ 1990 | -0.19865 | 0.211466 |
| SizeSel_7P_1_Jan-Apr_Pot_Fishery_BLK7repl_1977 | 69.0369 | 0.384451 |
| SizeSel_7P_1_Jan-Apr_Pot_Fishery_BLK7repl_1995 | 71.7177 | 0.362686 |
| SizeSel_7P_-_Jan-Apr_Pot_Fishery_BLK7repl_2000 | 67.999 | 0.377593 |
| SizeSel_7P_1_Jan-Apr_Pot_Fishery_BLK7repl_2005 | 68.2901 | 0.387619 |
| SizeSel_7P_3_Jan-Apr_Pot_Fishery_BLK7repl_1977 | 4.78708 | 0.045146 |
| SizeSel_7P_3_Jan-Apr_Pot_Fishery_BLK7repl_1995 | 4.93938 | 0.036873 |
| SizeSel_7P_3_Jan-Apr_Pot_Fishery_BLK7repl_2000 | 4.87376 | 0.040327 |
| SizeSel_7P_3_Jan-Apr_Pot_Fishery_BLK7repl_2005 | 4.72843 | 0.040613 |
| SizeSel_7P_4_Jan-Apr_Pot_Fishery_BLK7repl_1977 | 4.57339 | 0.16544 |
| SizeSel_7P_4_Jan-Apr_Pot_Fishery_BLK7repl_1995 | 4.17395 | 0.25953 |
| SizeSel_7P_4_Jan-Apr_Pot_Fishery_BLK7repl_2000 | 4.28071 | 0.222842 |
| SizeSel_7P_4_Jan-Apr_Pot_Fishery_BLK7repl_2005 | 4.06327 | 0.28019 |
| SizeSel_7P_-6_Jan-Apr_Pot_Fishery_BLK7repl_1977 | -2.13677 | 0.254541 |
| SizeSel_7P_6_Jan-Apr_Pot_Fishery_BLK7repl_1995 | -0.58293 | 0.187561 |
| SizeSel_7P_6_Jan-Apr_Pot_Fishery_BLK7repl_2000 | -0.48199 | 0.154792 |
| SizeSel_7P_6_Jan-Apr_Pot_Fishery_BLK7repl_2005 | 0.32856 | 0.200347 |
| SizeSel_8P_1_May-Aug_Pot_Fishery_BLK8repl_1977 | 64.6714 | 1.59067 |
| SizeSel_8P_1_May-Aug_Pot_Fishery_BLK8repl_1995 | 70.9111 | 1.10659 |
| SizeSel_8P_1_May-Aug_Pot_Fishery_BLK8repl_2000 | 65.5352 | 0.909074 |
| SizeSel_8P_3_May-Aug_Pot_Fishery_BLK8repl_1977 | 4.41747 | 0.260896 |
| SizeSel_8P_3_May-Aug_Pot_Fishery_BLK8repl_1995 | 4.86677 | 0.131258 |
| SizeSel_8P_3_May-Aug_Pot_Fishery_BLK8repl_2000 | 4.29032 | 0.162799 |
| SizeSel_9P_1_Sep-Dec_Pot_Fishery_BLK9repl_1977 | 72.0713 | 1.10771 |
| SizeSel_9P_1_Sep-Dec_Pot_Fishery_BLK9repl_1995 | 70.8309 | 1.21201 |
| SizeSel_9P_1_Sep-Dec_Pot_Fishery_BLK9repl_2000 | 65.9386 | 0.896044 |
| SizeSel_9P_1_Sep-Dec_Pot_Fishery_BLK9repl_2005 | 66.2771 | 0.714483 |
| SizeSel_9P_3_Sep-Dec_Pot_Fishery_BLK9repl_1977 | 5.29516 | 0.104601 |
| SizeSel_9P_3_Sep-Dec_Pot_Fishery_BLK9repl_1995 | 4.96429 | 0.137396 |
| SizeSel_9P_3_Sep-Dec_Pot_Fishery_BLK9repl_2000 | 4.79775 | 0.104363 |
| SizeSel_9P_3_Sep-Dec_Pot_Fishery_BLK9repl_2005 | 4.71262 | 0.082984 |
| SizeSel_9P_6_Sep-Dec_Pot_Fishery_BLK9repl_1977 | -1.58771 | 0.658867 |
| SizeSel_9P_6_Sep-Dec_Pot_Fishery_BLK9repl_1995 | 0.241672 | 0.606992 |
| SizeSel_9P_6_Sep-Dec_Pot_Fishery_BLK9repl_2000 | -0.41798 | 0.320675 |
| SizeSel_9P_6_Sep-Dec_Pot_Fishery_BLK9repl_2005 | -0.14559 | 0.31358 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1977 | 4.01359 | 0.008048 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1987 | 2.03644 | 0.009261 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1990 | 3.9154 | 0.240387 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1993 | 4.22328 | 0.329188 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1996 | 5.17891 | 0.278601 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_1999 | 4.23629 | 0.615017 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_2001 | 3.70345 | 0.254191 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_2003 | 4.25218 | 0.252552 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_2005 | 4.85767 | 0.205615 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_2007 | 2.81342 | 0.223011 |
| AgeSel_10P_1_27plus_Trawl_Survey_BLK10repl_2009 | 2.02855 | 0.014569 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1977 | -13.6558 |  |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1987 | -2.85092 | 0.330704 |


| Label | Value | SD |
| :---: | :---: | :---: |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1990 | -1.84823 | 0.130366 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1993 | -2.21592 | 1.32797 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1996 | -2.72899 | 0.333879 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_1999 | -1.08488 | 0.172209 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_2001 | -7.9805 | 37.9686 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_2003 | -3.20447 | 1.74529 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_2005 | -9.14005 | 20.1009 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_2007 | 0.159078 | 31.0306 |
| AgeSel_10P_2_27plus_Trawl_Survey_BLK10repl_2009 | -2.01586 | 0.034035 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1977 | 0.764168 | 0.057775 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1987 | -10 |  |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1990 | 0.694854 | 0.225812 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1993 | 1.20944 | 0.254599 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1996 | 1.7215 | 0.173851 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_1999 | 1.29626 | 0.498248 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_2001 | 0.962207 | 0.229782 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_2003 | 1.02672 | 0.211819 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_2005 | 1.47998 | 0.148214 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_2007 | 0.165322 | 0.327994 |
| AgeSel_10P_3_27plus_Trawl_Survey_BLK10repl_2009 | -9.11184 | 11.53 |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1977 | -9.99501 | 0.160709 |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1987 | 1.56168 | 0.380914 |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1990 | -9.97652 |  |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1993 | -1.90333 | $9.9027 \overline{6}$ |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1996 | -9.96742 |  |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_1999 | -10 |  |
| AgeSel_10P_4_-27plus_Trawl_Survey_BLK10repl_2001 | 3.04285 | $0.45507 \overline{6}$ |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_2003 | 1.85703 | 0.735973 |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_2005 | -18.0675 |  |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_2007 | -9.88607 |  |
| AgeSel_10P_4_27plus_Trawl_Survey_BLK10repl_2009 | -2.6374 | $36.341 \overline{3}$ |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1977 | -2.08124 | 0.180157 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1987 | -2.82415 | 0.512803 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1990 | -1.41489 | 0.469559 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1993 | -1.61748 | 0.501258 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1996 | -1.59947 | 0.548923 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_1999 | -10 |  |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_2001 | -8.36912 | 32.689 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_2003 | -9.74544 | 7.19514 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_2005 | 0.415027 | 0.578541 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_2007 | 7.70421 | 41.8543 |
| AgeSel_10P_6_27plus_Trawl_Survey_BLK10repl_2009 | -0.41167 | 0.429973 |

Table 2.14 - Estimated female spawning biomass ( t ) from the 2011 assessment and Model 2

|  | Last year |  | This year |  |
| :---: | :---: | :---: | :---: | :---: |
| Year | Spawning Biomass | Standard Deviation | Spawning Biomass | Standard Deviation |
| 1977 | 167,351 | 19,894 | 169,031 | 19,911 |
| 1978 | 195,614 | 21,366 | 204,602 | 21,738 |
| 1979 | 202,005 | 21,089 | 213,924 | 21,714 |
| 1980 | 202,656 | 19,838 | 211,323 | 20,347 |
| 1981 | 221,969 | 19,786 | 225,610 | 19,888 |
| 1982 | 248,968 | 20,576 | 257,987 | 20,899 |
| 1983 | 245,648 | 19,490 | 257,838 | 20,006 |
| 1984 | 227,027 | 17,444 | 238,070 | 17,937 |
| 1985 | 216,277 | 15,268 | 225,556 | 15,822 |
| 1986 | 213,715 | 13,230 | 221,341 | 13,901 |
| 1987 | 214,332 | 11,480 | 220,298 | 12,205 |
| 1988 | 205,381 | 9,968 | 208,697 | 10,634 |
| 1989 | 198,367 | 8,835 | 198,776 | 9,396 |
| 1990 | 187,980 | 7,930 | 188,967 | 8,496 |
| 1991 | 169,558 | 7,152 | 169,701 | 7,650 |
| 1992 | 150,833 | 6,655 | 149,600 | 7,046 |
| 1993 | 143,790 | 6,413 | 142,250 | 6,736 |
| 1994 | 148,362 | 6,339 | 145,558 | 6,645 |
| 1995 | 156,445 | 6,121 | 154,545 | 6,561 |
| 1996 | 150,276 | 5,816 | 149,391 | 6,318 |
| 1997 | 141,684 | 5,508 | 141,330 | 6,042 |
| 1998 | 129,846 | 5,389 | 129,266 | 5,923 |
| 1999 | 125,501 | 5,464 | 124,857 | 6,027 |
| 2000 | 112,891 | 5,501 | 113,669 | 6,172 |
| 2001 | 111,225 | 5,410 | 113,039 | 6,211 |
| 2002 | 104,223 | 5,211 | 106,423 | 6,064 |
| 2003 | 94,330 | 5,128 | 95,779 | 5,968 |
| 2004 | 92,875 | 5,338 | 93,901 | 6,183 |
| 2005 | 92,640 | 5,657 | 94,115 | 6,540 |
| 2006 | 87,802 | 5,788 | 89,918 | 6,685 |
| 2007 | 84,669 | 5,987 | 86,436 | 6,830 |
| 2008 | 80,278 | 6,552 | 81,890 | 7,319 |
| 2009 | 83,448 | 8,031 | 83,523 | 8,598 |
| 2010 | 95,784 | 11,105 | 94,670 | 11,440 |
| 2011 | 108,953 | 15,548 | 108,491 | 15,806 |
| 2012 | 121,180 | 18,284 | 123,986 | 21,106 |

Table 2.15 - Estimated age-0 recruits ( 000 's) from the 2011 assessment and Model 2

|  |  | Last year |  | Model 2 |
| ---: | ---: | ---: | ---: | ---: |
| Year | Age-0 | SD | Age-0 | SD |
| 1977 | 671,545 | 51,305 | 735,007 | 51,726 |
| 1978 | 158,776 | 17,249 | 195,006 | 21,277 |
| 1979 | 194,972 | 14,217 | 174,073 | 13,739 |
| 1980 | 215,021 | 13,158 | 209,907 | 13,148 |
| 1981 | 220,450 | 14,577 | 210,763 | 13,353 |
| 1982 | 304,994 | 18,701 | 286,796 | 18,383 |
| 1983 | 211,702 | 15,936 | 213,303 | 16,679 |
| 1984 | 211,084 | 14,512 | 153,020 | 14,258 |
| 1985 | 319,579 | 16,570 | 335,354 | 18,680 |
| 1986 | 171,859 | 11,039 | 162,463 | 12,040 |
| 1987 | 288,696 | 12,634 | 248,047 | 12,296 |
| 1988 | 213,873 | 11,618 | 225,576 | 11,881 |
| 1989 | 311,488 | 14,659 | 258,723 | 13,611 |
| 1990 | 275,951 | 13,772 | 282,643 | 13,593 |
| 1991 | 258,427 | 12,177 | 245,553 | 11,902 |
| 1992 | 212,641 | 11,492 | 208,946 | 11,678 |
| 1993 | 252,984 | 11,668 | 232,715 | 11,611 |
| 1994 | 251,854 | 11,581 | 233,170 | 11,544 |
| 1995 | 236,085 | 10,438 | 233,150 | 11,048 |
| 1996 | 212,948 | 9,303 | 210,467 | 10,416 |
| 1997 | 152,037 | 7,726 | 157,347 | 8,344 |
| 1998 | 158,598 | 8,120 | 139,673 | 7,974 |
| 1999 | 200,927 | 9,842 | 185,457 | 9,988 |
| 2000 | 228,817 | 11,275 | 218,151 | 11,685 |
| 2001 | 145,779 | 8,515 | 157,251 | 9,270 |
| 2002 | 151,163 | 9,182 | 132,968 | 8,680 |
| 2003 | 163,610 | 10,797 | 163,815 | 11,010 |
| 2004 | 177,435 | 12,943 | 157,315 | 11,918 |
| 2005 | 291,254 | 25,963 | 247,945 | 22,268 |
| 2006 | 324,737 | 34,283 | 319,899 | 33,409 |
| 2007 | 332,116 | 44,918 | 303,208 | 39,836 |
| 2008 | 372,360 | 53,770 | 357,927 | 51,699 |
| 2009 | 255,623 | 61,794 | 332,420 | 72,508 |
| 2010 | 361,467 | 95,517 | 215,261 | 87,437 |
| 2011 |  |  | 213,472 | 88,978 |
| Average | 250,319 |  | 238,765 |  |
|  |  |  |  |  |

Table 2.16 - Estimated numbers-at-age (millions) at the time of spawning for Model 2

| Year | 0 | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 | 15 | 16 | 17 | 18 | $19 \quad 20$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1977 | 735.0 | 97.1 | 79.3 | 54.5 | 102.7 | 30.2 | 17.4 | 8.8 | 4.9 | 3.1 | 2.1 | 1.5 | 1.1 | 0.8 | 0.7 | 0.5 | 0.3 | 0.2 | 0.20 | 0.10 .2 |
| 1978 | 195.0 | 502.6 | 66.4 | 54.2 | 37.0 | 69.6 | 20.5 | 11.8 | 6.0 | 3.3 | 2.1 | 1.4 | 1.0 | 0.7 | 0.5 | 0.5 | 0.3 | 0.20 | 0.20 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1979 | 174.1 | 133.4 | 343.7 | 45.2 | 36.4 | 24.7 | 46.5 | 13.7 | 7.9 | 4.0 | 2.2 | 1.4 | 1.0 | 0.7 | 0.5 | 0.4 | 0.3 | 0.20 | 0.20 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1980 | 209.9 | 119.0 | 91.2 | 233.4 | 30.2 | 24.0 | 16.3 | 30.8 | 9.1 | 5.2 | 2.7 | 1.5 | 0.9 | 0.6 | 0.5 | 0.3 | 0.2 | 0.20 | 0.20 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1981 | 210.8 | 143.5 | 81.4 | 61.9 | 153.2 | 19.2 | 15.5 | 10.6 | 20.2 | 6.0 | 3.4 | 1.8 | 1.0 | 0.6 | 0.4 | 0.3 | 0.2 | 0.20 | 0.10 | $\begin{array}{lll}0.1 & 0.2\end{array}$ |
| 1982 | 286.8 | 144.1 | 98.1 | 55.2 | 40.5 | 97.5 | 12.4 | 10.1 | 7.0 | 13.3 | 3.9 | 2.3 | 1.2 | 0.7 | 0.4 | 0.3 | 0.2 | 0.10 | 0.10 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1983 | 213.3 | 196.1 | 98.5 | 66.6 | 36.2 | 25.8 | 62.9 | 8.1 | 6.6 | 4.6 | 8.8 | 2.6 | 1.5 | 0.8 | 0.4 | 0.3 | 0.2 | 0.10 | 0.10 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1984 | 153.0 | 145.9 | 134.1 | 66.7 | 42.9 | 22.5 | 16.3 | 40.5 | 5.3 | 4.3 | 3.0 | 5.8 | 1.7 | 1.0 | 0.5 | 0.3 | 0.2 | 0.10 | 0.10 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1985 | 335.4 | 104.6 | 99.7 | 91.2 | 44.3 | 27.9 | 14.6 | 10.7 | 26.7 | 3.5 | 2.9 | 2.0 | 3.8 | 1.1 | 0.7 | 0.3 | 0.2 | 0.10 | 0.10 | $\begin{array}{ll}0.1 & 0.2\end{array}$ |
| 1986 | 162.5 | 229.3 | 71.5 | 67.9 | 61.4 | 29.4 | 18.4 | 9.7 | 7.1 | 17.7 | 2.3 | 1.9 | 1.3 | 2.5 | 0.8 | 0.4 | 0.2 | 0.10 | 0.10 | $\begin{array}{lll}0.1 & 0.1\end{array}$ |
| 1987 | 248.0 | 111.1 | 156.7 | 48.6 | 45.4 | 40.3 | 19.2 | 12.0 | 6.3 | 4.7 | 11.7 | 1.5 | 1.3 | 0.9 | 1.7 | 0.5 | 0.3 | 0.20 | 0.10 | 0.10 .1 |
| 1988 | 225.6 | 169.6 | 75.8 | 105.7 | 31.7 | 28.6 | 25.2 | 12.2 | 7.8 | 4.1 | 3.0 | 7.6 | 1.0 | 0.8 | 0.6 | 1.1 | 0.3 | 0.20 | 0.10 | 0.10 .1 |
| 1989 | 258.7 | 154.3 | 115.7 | 51.1 | 69.2 | 20.1 | 17.9 | 16.0 | 7.8 | 5.0 | 2.7 | 2.0 | 4.9 | 0.6 | 0.5 | 0.4 | 0.7 | 0.20 | 0.10 | $\begin{array}{lll}0.1 & 0.1\end{array}$ |
| 1990 | 282.6 | 176.9 | 105.4 | 78.5 | 33.4 | 42.7 | 12.0 | 10.8 | 9.7 | 4.8 | 3.1 | 1.7 | 1.2 | 3.1 | 0.4 | 0.3 | 0.2 | 0.40 | 0.10 | $\begin{array}{lll}0.1 & 0.1\end{array}$ |
| 1991 | 245.6 | 193.3 | 120.9 | 71.5 | 51.5 | 20.4 | 24.4 | 6.7 | 6.1 | 5.6 | 2.8 | 1.8 | 1.0 | 0.7 | 1.9 | 0.2 | 0.2 | 0.10 | 0.30 | 0.10 .1 |
| 1992 | 208.9 | 167.9 | 132.1 | 82.0 | 46.9 | 31.2 | 11.3 | 12.8 | 3.5 | 3.2 | 3.0 | 1.5 | 1.0 | 0.5 | 0.4 | 1.0 | 0.1 | 0.10 | 0.10 | 0.10 .1 |
| 1993 | 232.7 | 142.9 | 114.8 | 89.8 | 54.2 | 28.8 | 17.5 | 6.0 | 6.7 | 1.9 | 1.7 | 1.6 | 0.8 | 0.5 | 0.3 | 0.2 | 0.6 | 0.10 | 0.10 | 0.00 .1 |
| 1994 | 233.2 | 159.1 | 97.7 | 78.1 | 59.5 | 33.6 | 16.6 | 9.7 | 3.3 | 3.8 | 1.1 | 1.0 | 0.9 | 0.5 | 0.3 | 0.2 | 0.1 | 0.30 | 0.00 | 0.00 .1 |
| 1995 | 233.2 | 159.5 | 108.8 | 66.6 | 52.2 | 37.8 | 20.1 | 9.6 | 5.6 | 1.9 | 2.2 | 0.6 | 0.6 | 0.6 | 0.3 | 0.2 | 0.1 | 0.10 | 0.20 | 0.0 |
| 1996 | 210.5 | 159.4 | 109.0 | 74.0 | 44.1 | 32.4 | 21.6 | 10.9 | 5.1 | 3.0 | 1.0 | 1.2 | 0.3 | 0.3 | 0.3 | 0.2 | 0.1 | 0.10 | 0.00 | 0.10 .1 |
| 1997 | 157.3 | 143.9 | 109.0 | 74.1 | 48.9 | 27.2 | 18.3 | 11.6 | 5.7 | 2.7 | 1.6 | 0.6 | 0.6 | 0.2 | 0.2 | 0.2 | 0.1 | 0.10 | 0.00 | 0.00 .1 |
| 1998 | 139.7 | 107.6 | 98.3 | 73.8 | 48.3 | 29.1 | 14.5 | 9.1 | 5.6 | 2.8 | 1.3 | 0.8 | 0.3 | 0.3 | 0.1 | 0.1 | 0.1 | 0.0 | 0.00 | 0.00 .1 |
| 1999 | 185.5 | 95.5 | 73.5 | 66.8 | 48.4 | 29.0 | 15.8 | 7.4 | 4.6 | 2.9 | 1.5 | 0.7 | 0.4 | 0.2 | 0.2 | 0.1 | 0.0 | 0.0 | 0.00 | $0.0 \begin{array}{lll}0.0\end{array}$ |
| 2000 | 218.2 | 126.8 | 65.2 | 49.6 | 42.6 | 27.4 | 14.6 | 7.5 | 3.5 | 2.3 | 1.5 | 0.8 | 0.4 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | 0.0 0.0 |
| 2001 | 157.3 | 149.2 | 86.7 | 44.2 | 32.4 | 25.7 | 15.5 | 8.1 | 4.2 | 2.0 | 1.3 | 0.9 | 0.5 | 0.2 | 0.1 | 0.0 | 0.1 | 0.0 | 0.00 | 0.0 0.0 |
| 2002 | 133.0 | 107.5 | 101.9 | 58.6 | 28.6 | 19.3 | 14.3 | 8.5 | 4.5 | 2.4 | 1.2 | 0.8 | 0.5 | 0.3 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | 0.0 0.0 |
| 2003 | 163.8 | 90.9 | 73.5 | 68.9 | 37.5 | 16.3 | 10.0 | 7.2 | 4.4 | 2.4 | 1.3 | 0.7 | 0.4 | 0.3 | 0.2 | 0.1 | 0.0 | 0.0 | 0.00 | 0.0 0.0 |
| 2004 | 157.3 | 112.0 | 62.1 | 49.6 | 44.0 | 21.5 | 8.6 | 5.1 | 3.8 | 2.4 | 1.4 | 0.8 | 0.4 | 0.3 | 0.2 | 0.1 | 0.0 | 0.0 | 0.00 | 0.0 |
| 2005 | 247.9 | 107.6 | 76.5 | 42.0 | 32.0 | 25.7 | 11.5 | 4.5 | 2.7 | 2.1 | 1.4 | 0.8 | 0.4 | 0.2 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 .0 |
| 2006 | 319.9 | 169.6 | 73.5 | 51.9 | 27.3 | 18.9 | 13.8 | 6.0 | 2.4 | 1.5 | 1.2 | 0.8 | 0.5 | 0.3 | 0.1 | 0.1 | 0.1 | 0.0 | 0.00 | 0.0 0.0 |
| 2007 | 303.2 | 218.8 | 115.9 | 49.9 | 33.9 | 16.3 | 10.3 | 7.4 | 3.2 | 1.3 | 0.8 | 0.7 | 0.4 | 0.3 | 0.1 | 0.1 | 0.1 | 0.0 | 0.00 | 0.0 0.0 |
| 2008 | 357.9 | 207.3 | 149.4 | 78.4 | 32.1 | 19.4 | 8.3 | 5.1 | 3.7 | 1.7 | 0.7 | 0.5 | 0.4 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | 0.00 .0 |
| 2009 | 332.4 | 244.8 | 141.7 | 101.1 | 50.3 | 18.2 | 9.7 | 4.0 | 2.5 | 1.9 | 0.9 | 0.4 | 0.2 | 0.2 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | 0.0 |
| 2010 | 215.3 | 227.3 | 167.2 | 95.8 | 64.6 | 28.1 | 9.0 | 4.6 | 1.9 | 1.3 | 1.0 | 0.5 | 0.2 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | $0.0 \begin{array}{lll}0.0\end{array}$ |
| 2011 | 213.5 | 147.2 | 155.3 | 113.2 | 61.3 | 36.0 | 13.8 | 4.2 | 2.2 | 1.0 | 0.7 | 0.5 | 0.2 | 0.1 | 0.1 | 0.1 | 0.0 | 0.0 | 0.00 | $0.0 \begin{array}{lll}0.0\end{array}$ |
| 2012 | 244.3 | 146.0 | 100.6 | 105.3 | 73.2 | 35.2 | 18.4 | 6.8 | 2.1 | 1.2 | 0.5 | 0.4 | 0.3 | 0.1 | 0.1 | 0.0 | 0.0 | 0.0 | 0.00 | $0.0 \quad 0.0$ |

Table 2.17 - Estimates of "effective" fishing mortality $\left(=-\ln \left(N a+1, y+1 / N_{a, y}\right)-M\right)$ at age and year for Model 2

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.001 |  |  |  |  |  |  |  |  | 0.003 | 0.003 |  | 0.003 | 3 |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 0.00 | 0.016 | 0.026 |  |  |  |  |  |  | 0.019 | 0.018 | . 017 |  | 0.016 |  |  |  |  |
|  | 0.000 | 0.00 | 0.039 | 0.072 | 0.06 | 0.05 | 0.04 | 0.045 | 0.04 | 0.0 | 0.03 | 0.036 | . 03 | . 01 | 0.033 | 0.032 |  |  |  |
|  | 0.000 | 0.007 | 0.04 | 0.070 | 0.06 | 0.04 | 0.039 | 0.036 | 0.035 | 0.03 | 0.03 | 0.030 | 0.029 | 0.02 | 0.028 | 0.027 | 0.027 | 0.027 |  |
|  | 0.000 | 0.006 | 0. | 0. | 0.056 | 0.042 | 0. | 0.033 | 0.03 | 0. | 0.029 | 0.027 | 0.026 | 0.02 | 0.025 | 0.025 | 0.02 | 0.024 |  |
|  | 0.000 | 0.008 | 0.054 | . 92 | 0.079 | 0.05 | 0.050 | 0.047 | 0.045 | 0.0 | 0.0 | 0.040 | 0.038 | 0.03 | . 03 | . 03 | 0.03 | 0.036 |  |
|  |  | 0.00 | 0.029 | 0.053 |  | 0.04 | 0.03 | 0.03 |  |  |  |  |  |  | 0.027 | 0.026 |  |  |  |
|  | 0.00 | 0.00 | 0.01 | 0.022 |  | 0.02 | 0.02 | 0.02 |  | 0.02 | 0.022 | 0.022 |  | 0.02 | 0.02 | 0.021 |  |  |  |
|  | 0.000 | 0.004 | 0. | 0 | 0.048 | 0.04 | 0.04 | 0.044 | 0.04 | 0.04 | 0.038 | 0.038 | 0.037 | 0.03 | 0.037 | 0.036 | . 03 | 0.036 |  |
|  | 0.001 | 0.01 | 0.04 | 0.0 | 0.079 | 0.06 | 0.0 | 0. | 0. | 0.0 | 0.0 | 0.0 | 0.038 | 0. | 0.0 | 0.037 | 0.037 | 0.037 |  |
|  | 0.001 | . 008 | . 033 | . 067 | . 080 | 0.07 | 0.06 | . 055 | 0.051 | . 0 | 0.048 | . 047 | . 046 | 0.0 | . 046 | . 046 | 0.046 | 0.046 |  |
|  | 0.000 | 0.00 | 0.039 | 0.087 | 0.106 | 0.09 | 0.08 | 0.07 | 0.06 | 0.06 | 0.065 | 0.06 |  | 0.06 | 0.063 | 0.063 | 0.06 | 0.06 |  |
|  | 0.000 | . | 0. | 0.0 | 0.17 | , | . 1 | . 1 | 0.1 | 0.1 | 0.14 | 0.13 |  | 0.1 | 0. | . 1 | 0.12 | 0.124 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0 | 0.00 | 0 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.000 | . 002 | 0.019 | 070 | 0.146 | 0 | 0.2 | 19 |  | 0.1 | 0.1 | 0 | 0.164 | 0.16 | 0.162 | 0.161 | . 161 | 0.160 |  |
|  | 0.000 | 0.002 | . 01 | 0.053 |  | 0.15 | 0.16 | 0.160 |  | 0.1 | 0.138 | 0.13 |  | 0. | 0.132 |  |  |  |  |
|  | 0 | 0.00 | 0.01 | 0.06 | 0.14 | 0.2 | 0. | 0.248 | 0.23 | 0. | 0.2 | 0.21 | 0.212 | 0. | 0.2 | 0.209 | 0.209 | 0.208 |  |
|  |  | . | 0.01 |  |  |  |  |  |  |  |  |  |  |  |  |  |  | . |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  |  | 0.003 |  |  |  |  |  |  |  |  |  |  |  | 0. |  |  | 224 | 0.223 |  |
|  |  | 0.00 | 0.02 | 0.09 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.000 | 0.00 | 0.02 | 0.1 |  |  | 0.2 | 0.239 |  |  |  | 0.179 |  |  | 0.1 | . 17 | 0.172 | 0.1 |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.00 | 0.00 |  |  |  |  |  |  |  |  |  | 0.16 |  |  | 0.1 | , |  | 0.1 |  |
|  |  | 0.00 |  |  |  |  |  |  |  |  | 0.16 | , |  | 0.1 |  |  | 0.149 | 0.1 |  |
|  | 0.00 | 0.00 | 0.03 | 0.121 | 0.24 | 0.2 | 0.28 | 0.240 | 0.20 | 0.18 | 0.172 |  |  | . | 0.159 | , |  | 0.157 |  |
|  | - 0.00 | 0.003 | . |  |  | 0.24 | . | , |  |  |  |  |  | 0.1 | 0.16 | . 16 | 0.168 | . 1 |  |
|  | 0.000 |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
|  | 0.000 | 0.00 | 0.02 |  |  | 0.29 | 0.29 |  |  | 0.22 |  | . 205 |  | 0.20 | 0.199 | 0.199 | 0.1 | 0.19 |  |
|  | 0 | . 00 | . 03 | 132 | 0.27 | , | 0.35 | , |  | .25 | 0.246 | 23 |  | 0.23 | 0.231 | 23 | 0.230 | 0.2 |  |
| 2009 | 00 | 0.004 | . 02 | 0.107 | , | 0.30 | . 301 | . 26 | . 23 | 0.2 | . 2 | . 20 | . | 0.1 | 0.19 | . 19 | 0.195 | . 1 |  |
|  | 0.00 | 0.00 | 0.03 | 0.135 | 0.29 | 0.386 | 0.38 | 0.34 | 0.303 | 0.27 | 0.26 | 0.25 | 0.254 | 0.25 | 0.2 | . 249 | 0.248 | 0.247 |  |
|  | 0.0 | 0.003 | 0.0 | 0.130 |  |  |  |  |  |  | 0 | 0.24 | 0.2 | 0 | 0 | 0 | 0.2 | 0.2 |  |

Table 2.18 - Results for the projection scenarios for Model 2

| Scenarios 1 and 2, Maximum Tier 3 ABC harvest permissible |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 80,830 | 0.489 | 111,042 | 449,342 |
| 2014 | 84,284 | 101,181 | 84,284 | 0.489 | 112,955 | 440,370 |
| 2015 | 78,932 | 94,605 | 78,932 | 0.489 | 104,180 | 417,649 |
| 2016 | 72,033 | 86,364 | 72,033 | 0.489 | 96,676 | 400,856 |
| 2017 | 68,497 | 81,490 | 68,497 | 0.482 | 94,575 | 395,340 |
| 2018 | 67,177 | 80,065 | 67,177 | 0.470 | 94,516 | 395,040 |
| 2019 | 67,445 | 80,442 | 67,445 | 0.467 | 95,099 | 396,375 |
| 2020 | 67,759 | 80,842 | 67,759 | 0.466 | 95,492 | 397,395 |
| 2021 | 67,856 | 80,976 | 67,856 | 0.466 | 95,494 | 397,987 |
| 2022 | 67,995 | 81,125 | 67,995 | 0.467 | 95,684 | 398,687 |
| 2023 | 68,163 | 81,333 | 68,163 | 0.468 | 95,799 | 399,181 |
| 2024 | 68,270 | 81,464 | 68,270 | 0.468 | 95,973 | 399,308 |
| 2025 | 68,438 | 81,676 | 68,438 | 0.468 | 96,180 | 399,472 |
| Scenario 3, $F_{A B C}$ at average $F$ over the past 5 years |  |  |  |  |  |  |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 56,149 | 0.323 | 114,485 | 449,342 |
| 2014 | 91,344 | 109,592 | 63,703 | 0.323 | 125,880 | 463,899 |
| 2015 | 91,096 | 109,078 | 63,712 | 0.323 | 123,522 | 457,316 |
| 2016 | 86,422 | 103,556 | 60,383 | 0.323 | 118,853 | 448,253 |
| 2017 | 84,117 | 100,909 | 58,674 | 0.323 | 117,473 | 445,176 |
| 2018 | 83,803 | 100,532 | 58,444 | 0.323 | 117,250 | 444,900 |
| 2019 | 83,827 | 100,469 | 58,552 | 0.323 | 117,378 | 445,097 |
| 2020 | 83,872 | 100,492 | 58,660 | 0.323 | 117,446 | 445,267 |
| 2021 | 83,741 | 100,330 | 58,610 | 0.323 | 117,251 | 445,339 |
| 2022 | 83,716 | 100,291 | 58,603 | 0.323 | 117,329 | 445,746 |
| 2023 | 83,815 | 100,405 | 58,665 | 0.323 | 117,421 | 446,150 |
| 2024 | 83,950 | 100,570 | 58,750 | 0.323 | 117,608 | 446,313 |
| 2025 | 84,117 | 100,774 | 58,878 | 0.323 | 117,847 | 446,536 |
| Scenario 4, $1 / 2$ Maximum ABC harvest possible |  |  |  |  |  |  |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 39,180 | 0.218 | 116,727 | 449,342 |
| 2014 | 96,245 | 115,431 | 46,994 | 0.218 | 135,094 | 480,181 |
| 2015 | 100,275 | 119,998 | 49,220 | 0.218 | 138,509 | 487,187 |
| 2016 | 98,122 | 117,480 | 48,110 | 0.218 | 137,361 | 486,807 |
| 2017 | 96,813 | 116,038 | 47,342 | 0.218 | 137,783 | 488,213 |
| 2018 | 96,946 | 116,225 | 47,381 | 0.218 | 138,563 | 490,341 |
| 2019 | 97,412 | 116,783 | 47,609 | 0.218 | 139,313 | 492,010 |
| 2020 | 97,783 | 117,219 | 47,798 | 0.218 | 139,785 | 493,122 |
| 2021 | 97,857 | 117,300 | 47,841 | 0.218 | 139,846 | 493,802 |
| 2022 | 97,938 | 117,402 | 47,874 | 0.218 | 140,068 | 494,555 |
| 2023 | 98,070 | 117,560 | 47,939 | 0.218 | 140,252 | 495,171 |
| 2024 | 98,232 | 117,755 | 48,017 | 0.218 | 140,504 | 495,484 |
| 2025 | 98,439 | 118,002 | 48,121 | 0.218 | 140,798 | 495,824 |

Table 2.18 - (continued) Results for the projection scenarios for Model 2

| Scenario 5, No fishing ( $F_{A B C}=0$ ) |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 0 | 0 | 121,561 | 449,342 |
| 2014 | 107,692 | 129,064 | 0 | 0 | 157,343 | 518,061 |
| 2015 | 124,144 | 148,390 | 0 | 0 | 178,855 | 564,717 |
| 2016 | 131,798 | 157,556 | 0 | 0 | 192,561 | 597,961 |
| 2017 | 136,586 | 163,446 | 0 | 0 | 203,707 | 623,920 |
| 2018 | 140,610 | 168,337 | 0 | 0 | 212,154 | 643,473 |
| 2019 | 143,941 | 172,351 | 0 | 0 | 218,488 | 657,779 |
| 2020 | 146,476 | 175,393 | 0 | 0 | 223,071 | 668,151 |
| 2021 | 148,165 | 177,419 | 0 | 0 | 226,132 | 675,607 |
| 2022 | 149,380 | 178,887 | 0 | 0 | 228,486 | 681,193 |
| 2023 | 150,304 | 180,000 | 0 | 0 | 230,194 | 685,246 |
| 2024 | 151,044 | 180,892 | 0 | 0 | 231,542 | 688,032 |
| 2025 | 151,682 | 181,658 | 0 | 0 | 232,648 | 690,185 |
| Scenario 6, Whether Pacific cod are overfished - S $_{35 \%}=82,100$ |  |  |  |  |  |  |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 97,251 | 0.610 | 108,617 | 449,342 |
| 2014 | 79,637 | 95,643 | 95,643 | 0.610 | 104,680 | 424,824 |
| 2015 | 71,585 | 85,097 | 85,097 | 0.603 | 92,941 | 393,625 |
| 2016 | 60,175 | 71,249 | 71,249 | 0.554 | 85,753 | 375,009 |
| 2017 | 59,088 | 70,131 | 70,131 | 0.552 | 85,529 | 374,773 |
| 2018 | 59,574 | 70,907 | 70,907 | 0.550 | 86,134 | 376,715 |
| 2019 | 60,007 | 71,488 | 71,488 | 0.549 | 86,613 | 377,877 |
| 2020 | 60,207 | 71,740 | 71,740 | 0.549 | 86,800 | 378,393 |
| 2021 | 60,146 | 71,634 | 71,634 | 0.548 | 86,665 | 378,602 |
| 2022 | 60,224 | 71,733 | 71,733 | 0.549 | 86,812 | 379,181 |
| 2023 | 60,328 | 71,855 | 71,855 | 0.549 | 86,906 | 379,616 |
| 2024 | 60,486 | 72,074 | 72,074 | 0.550 | 87,063 | 379,740 |
| 2025 | 60,619 | 72,240 | 72,240 | 0.550 | 87,225 | 379,797 |
| Scenario 7, Whether Pacific cod are approaching an overfished condition |  |  |  |  |  |  |
| Year | ABC | OFL | Catch | F | SSB | Total Bio |
| 2013 | 80,830 | 97,251 | 80,830 | 0.489 | 111,042 | 449,342 |
| 2014 | 84,284 | 101,181 | 84,284 | 0.489 | 112,955 | 440,370 |
| 2015 | 78,932 | 94,605 | 94,605 | 0.610 | 101,626 | 417,649 |
| 2016 | 65,919 | 77,847 | 77,847 | 0.580 | 89,529 | 386,097 |
| 2017 | 60,388 | 71,659 | 71,659 | 0.559 | 86,526 | 377,481 |
| 2018 | 59,750 | 71,121 | 71,121 | 0.551 | 86,357 | 377,245 |
| 2019 | 60,002 | 71,486 | 71,486 | 0.549 | 86,665 | 377,977 |
| 2020 | 60,195 | 71,727 | 71,727 | 0.549 | 86,815 | 378,421 |
| 2021 | 60,141 | 71,629 | 71,629 | 0.548 | 86,671 | 378,614 |
| 2022 | 60,222 | 71,731 | 71,731 | 0.549 | 86,814 | 379,186 |
| 2023 | 60,327 | 71,854 | 71,854 | 0.549 | 86,906 | 379,617 |
| 2024 | 60,485 | 72,074 | 72,074 | 0.550 | 87,063 | 379,740 |
| 2025 | 60,619 | 72,240 | 72,240 | 0.550 | 87,225 | 379,797 |

Table 2.19 - Biological reference points from GOA Pacific cod SAFE documents for years 2001-2011

| Year | $\mathbf{S B}_{\mathbf{1 0 0 \%}}$ | $\mathbf{S B}_{\mathbf{4 0} \%}$ | $\mathbf{F}_{\mathbf{4 0} \%}$ | $\mathbf{S B}_{\mathbf{y}+\mathbf{1}}$ | $\mathbf{A B C}_{\mathbf{y}+\mathbf{1}}$ |
| :--- | :--- | ---: | ---: | ---: | ---: |
| 2001 | 212,000 | 85,000 | 0.41 | 82,000 | 57,600 |
| 2002 | 226,000 | 90,300 | 0.35 | 88,300 | 52,800 |
| 2003 | 222,000 | 88,900 | 0.34 | 103,000 | 62,810 |
| 2004 | 211,000 | 84,400 | 0.31 | 91,700 | 58,100 |
| 2005 | 329,000 | 132,000 | 0.56 | 165,000 | 68,859 |
| 2006 | 259,000 | 103,000 | 0.46 | 136,000 | 68,859 |
| 2007 | 302,000 | 121,000 | 0.49 | 108,000 | 66,493 |
| 2008 | 255,500 | 102,200 | 0.52 | 88,000 | 55,300 |
| 2009 | 291,500 | 116,600 | 0.49 | 117,600 | 79,100 |
| 2010 | 256,300 | 102,500 | 0.42 | 124,100 | 86,800 |
| 2011 | 261,000 | 104,000 | 0.44 | 121,000 | 87,600 |
| 2012 | 234,800 | 93,900 | 0.49 | 111,000 | 80,800 |



Figure 2.1 - Total catch, by season and gear type


Figure 2.2 - Estimated spawning biomass for all model configurations


Figure 2.3 - Estimated age-0 recruits for all model configurations


Figure 2.4 - Estimated 27-plus survey indices for all model configurations (log scale)


Figure 2.5 - Estimated sub-27 survey indices for all model configurations (log scale)

Spawning biomass (mt) with ~95\% asymptotic intervals


Figure 2.6 - Estimated spawning biomass for Model 2

Age-0 recruits (1,000s) with ~95\% asymptotic intervals


Figure 2.7 - Estimated age-0 recruits for Model 2

## Log index 27plus_Trawl_Survey



Figure 2.8 - Estimated 27-plus survey indices for Model 2


Figure 2.9 - Spawning biomass-recruitment for Model 2


Figure 2.10 - Fishery selectivity-at-length, by seasons (rows) and gear type (columns), for Model 2

Time-varying selectivity for 27plus_Trawl_Survey


Figure 2.11 - 27-plus survey selectivity-at-age for Model 2
length comps, sexes combined, whole catch, Jan-Apr_Trawl_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2
length comps, sexes combined, whole catch, Jan-Apr_Trawl_Fishery


Length (cm)
Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, Sep-Dec_Trawl_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, Jan-Apr_Longline_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, May-Aug_Longline_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, Sep-Dec_Longline_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, Jan-Apr_Pot_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, May-Aug_Pot_Fishery


Length (cm)
Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)
length comps, sexes combined, whole catch, Sep-Dec_Pot_Fishery


Figure 2.12 - Fits to fishery catch-at-length for Model 2 (continued)

# length comps, sexes combined, whole catch, 27plus_Trawl_Survi aggregated across seasons within year 



Figure 2.13 - Fits to 27-plus survey length comp data for Model 2
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 2.14 - Fits to 27-plus survey age comp data for Model 2
mean length at age, sexes combined, whole catch, 27plus_Trawl_Survey


Age (yr)
Figure 2.15 - Fits to 27-plus mean size-at-age for Model 2


Figure 2.16 - Estimated length-at-age for Model 2


Figure 2.17-1 - SPR for Model 2

## 2010 numbers-at-age



Figure 2.18 - Numbers-at-age (in thousands) at the time of spawning for 2010 and 2011 from the 2011 GOA Pacific cod SAFE document and Model 2


Figure 2.19 - Estimates of spawning biomass for 2003-2012 for Model 2

Percent difference from 2012


Figure 2.20 - Percent difference between spawning biomass for 2003-2011 and 2012 for Model 2

# Chapter 2 Appendix: An exploration of GOA Pacific cod stock assessment models for 2012 

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## Introduction

This document represents an effort to respond to comments made the GOA Plan Team, the joint BSAI and GOA Plan Teams, and the SSC on the 2011 assessments of the Pacific cod (Gadus macrocephalus) stocks in the Gulf of Alaska (Thompson et al., 2011). In order to allow for exploration of a wide variety of modeling assumptions, this preliminary overview focuses on model development rather than application of the same model(s) to multiple data sets. Specifically, the Stock Synthesis model configurations presented here are applied to the data used in the 2011 GOA Pacific cod stock assessment.

## Comments from the Plan Teams and SSC

## J oint Plan Team Comments from the May 2012 Minutes

JPT1: "For the GOA, the Teams recommend that the preliminary assessment include the following two models, which are in addition to any models that the authors wish to propose: Model 1 is last year's final model, and Model 2 is last year's final model with re-tuned catchability."
Response: The initial results from these model configurations are included as Models 1 and 1 Q ("Q" indicates the model was iteratively tuned so that mean catchability for $60-81 \mathrm{~cm}$ was 0.916 ).

JPT2: "For both the EBS and GOA, the Teams recommend that the authors attempt to explore the divergent ageing bias trends in the two regions and the impacts thereof."
Response: This issue will be addressed in the November SAFE document.
JPT3: "For both the EBS and GOA, the Teams recommend that the authors attempt to evaluate the biological basis for estimated patterns of seasonal weight at length."
Response: This issue will be addressed in the November SAFE document.
JPT4: "For both the EBS and GOA, the Teams recommend that the authors attempt to estimate catchability internally. This can be addressed as an option under Model 1 without developing and presenting a full set of results for an additional model (full results for the base case of Model 1 are requested, however)."
Response: The initial results from this model configuration are included as Model E.
JPT5: "For the GOA only, the Teams recommend that the authors reduce the number of parameters. This can be addressed as an option under Model 1 without developing and presenting a full set of results for an additional model (full results for the base case of Model 1 are requested, however)."
Response: Models B, BQ, C, CQ, and D have fewer estimated parameters than Models 1 and 1Q.
JPT6: "Seven model evaluation criteria were proposed: 1) fitting the age composition data (unanimous CIE recommendation); 2) internal estimation of aging error bias (much more efficient); 3)
correspondence between the model-estimated mean size-at-age and the empirical survey mean size-at-age of the first three modes of the average survey size composition; 4) correspondence of the product of survey catchability and survey selectivity (for the 61 to 80 cm size range) from the model and the value of 0.916 estimated by Nichol et al. (2007); 5) accounting for full variability in the observed length-at-age among individuals and years; 6) low temporal variability in survey selectivity and catchability; and 7) reasonable retrospective behavior. The Plan Team endorsed, and the SSC concurs, with these selection criteria, which are a distillation of past preferences and recommendations from the Plan Teams, CIE reviewers, the public, and the SSC. These criteria will be used when evaluating the results of model configurations."
Responses correspond to each of the JPT proposed model evaluation criteria listed above:
Criteria 1, The fits to the survey age composition data are presented for all models;
Criteria 2, All model configurations estimate ageing bias;
Criteria 3, Comparison of survey mean size-at-age with model-estimated survey mean size-at-age will be addressed in the November SAFE document;
Criteria 4, Calculated values for mean 27-plus survey catchability for $60-81 \mathrm{~cm}$ for all model configurations are presented in Table 2;
Criteria 5, This issue will be addressed in the November SAFE document;
Criteria 6, Five model configurations were explored that have two time blocks defined for catchability and selectivity for the sub-27 survey; one model configuration has two time blocks defined for selectivity for the 27-plus survey as well, as the base model configurations have two time blocks defined for catchability of the 27-plus survey (1984-1993 and 1996 - present) due to the switch from 30-minute to 15-minute tows in the survey design; and
Criteria 7, This issue will be addressed in the November SAFE document and when the Plan Team working group on retrospective analyses is completed.

SSC Comments from the June 2012 Minutes
SSC1: "As for the EBS, the SSC agrees with the choice of last year's final model (formerly Model 3, new model 1) as the baseline model for the Gulf of Alaska and a second model (model 2) that re-tunes catchability to match the empirical estimates from Nichol et al. (2007)."
Response: The initial results from these model configurations are included as Models 1 and 1 Q .

## Summary of the base model configuration

The base model configuration for 2012 is the 2011 Model 3 configuration. The software used to run all models was SS v3.24f as compiled on 03 August 2012 with ADMB v. 11 (Last year's models used v3.22b, as compiled on $8 / 3 / 11$ ).

## Model evaluation

## Model configurations for 2012

The following details attributes of the requested models:
Model 1: 2011 Model 3 (the base model)
Model 1Q: Model 1 with mean catchability for the 27-plus survey tuned iteratively to 0.916
These models include:

- Time-varying fishery selectivity-at-length for all gears and seasons;
- Two blocks for catchability for the 27-plus survey, 1984 - 1993 and 1996 - 2011;
- Time-varying catchability for the Sub-27 survey;
- Time-varying survey selectivity-at-age for the 27-plus survey;
- Constant survey selectivity-at-age for the Sub-27 survey; and
- Median recruitment before 1977 restricted to be less than the post-1976 median recruitment, as the pre-1977 recruitment deviation is restricted to be less than 0.0

On evaluation of these models it was suggested by Ian Stewart and Ian Taylor (NWFSC, pers. comm.) to change the Stock Synthesis option "comp_tail_compression" from 0.000001 (used previously) to -1 , which turns this feature off and uses all of the age and length composition data. This feature, when implemented as last year, binned the values in the composition "tails" for both the observed and expected values. This change is applied in subsequent models beginning with:

Model A: Model 1 (the base model) with tail compression turned off
Model AQ: Model A with mean catchability for the 27-plus survey tuned iteratively to 0.916
The four base models estimated very similar values for spawning biomass (Fig. 1), although the estimates of recruitment differ in recent years (Fig. 2), which is also seen in the fits to the 27 -plus and sub-27 survey indices (Figs. 3 and 4). The fits to the survey length composition data for Models 1 (Figs. 5 and 7) and A (Figs. 6 and 8) also differ substantially, which shows the impact of tail compression and is likely the reason for differences in recent recruitment estimates.

The following considerations for proceeding with alternative models should consider (and for which feedback would be appreciated) include:

- The importance using the ADF\&G nearshore trawl survey data as an additional index;
- The use of time-varying catchability coefficients for the sub-27 survey;
- Fitting the age composition for the sub-27 survey; and
- The use of time-varying selectivity for the 27-plus survey.

Alternative model configurations evaluated included:
Model B: $\quad$ Model A with sub-27 survey changed from time-varying Q and constant selectivity to two blocks for both Q and selectivity (split at 1996), and the initial value for the pre1996 Q deviation for both the 27-plus and sub-27 surveys set to 0.0

Model BQ: Model B with mean catchability for the 27-plus survey tuned iteratively to 0.916
Model C: $\quad$ Model B with the initial value for pre-1977 recruitment deviation changed to 0.0 and the upper bound increased to allow positive values

Model CQ: Model C with mean catchability for the 27-plus survey tuned iteratively to 0.916
Model D: Model C with 27-plus survey changed from 11 to 2 blocks for selectivity (split at 1996)

Model E: Model A with Q for the 27-plus survey estimated, and the initial value for the pre1996 Q deviation for both the 27-plus and sub-27 surveys set to 0.0 (See JPT4)

Model 1B: Model B with tail compression set to the value used in Model 1
Model 1C: Model C with tail compression set to the value used in Model 1

The changes for Q and selectivity for the sub- 27 survey characterizes the assumption that the data are representative of cohort strength, rather than survey variability. For comparison purposes only, the GOA Pacific cod length composition data from the ADF\&G crab and groundfish nearshore trawl survey are presented. Small fish, $5-15 \mathrm{~cm}$, which are assumed to be age- 0 , are sampled by the ADF\&G survey and are prominently featured in the data in some years. The NMFS GOA bottom trawl survey samples fish less than 27 cm , which are assumed to be primarily age- 1 fish. The results of a comparison of the proportions-at-length data from the ADF\&G survey and the NMFS survey suggest that the relative numbers of assumed age- 0 fish in the ADF\&G survey in a given year are similar to the relative numbers of assumed age- 1 fish in the NMFS survey the following year (Figs. 9 - 12). The change from timevarying to constant Q is featured in Models $\mathrm{B}, \mathrm{BQ}, \mathrm{C}, \mathrm{CQ}$, and D .

Similarly, the changes to selectivity for the 27 -plus survey characterize the assumption that the data are representative of the stock characteristics, rather than survey variability. These assumptions and model configurations will be investigated further. This feature is present in Model D.

To evaluate convergence, Models 1, 1Q, A, AQ, C, and CQ have been tested using the Stock Synthesis "jitter" functionality which generates random initial values based on parameter bounds and other properties. The jitter parameter was set to 0.1 and each model was run 50 times. If the jitter runs produced initial values which led to a lower value for the objection function, then these initial values were used as the starting point for another set of 50 jitter runs. To complement this work, the initial values of some deviation parameters were set to 0.0 to test convergence. Since deviation parameters are applied to other variables, setting the initial value to 0.0 represents the assumption that this deviation has no positive or negative impact on the variable. This was done for variables in Models B, BQ, C, CQ, D, and E.

Model configurations 1B and 1C were included as a comparison to Models 1 and 1 Q , as these model configurations have tail compression turned on.

## Results

The values for total likelihood for Models 1 and 1Q are lower than those for all new model configurations (Table 1), due to the change in the comp_tail_compression functionality. Model C has the lowest value of all new model configurations for both total likelihood and AIC; Model CQ has the lowest value for total likelihood and AIC of all new model configurations where the mean 27-plus survey catchability has been tuned iteratively to 0.916 .

The calculated values for the mean catchability for the 27-plus survey for $60-81 \mathrm{~cm}$ ranged from values in the neighborhood of 0.916 for tuned model configurations to above 0.94 for the other model configurations (Table 2); the values for the pre-1996 Q deviation for the 27-plus and sub-27 surveys were in the neighborhood of 0.5 and -0.5 , respectively, save for Model E , which estimated catchability for the 27 -plus survey. The (base) values for catchability for the sub-27 survey ranged from a low of 0.07 to a high of 0.16.

The estimates of spawning biomass are similar across all new model configurations, save for Model E, which has consistently lower estimates due to estimating the catchability for the 27 -plus survey to be significantly higher than all other model configurations (Figs. 13 and 14). The estimates of age-0 recruits are similar across the new model configurations, with Model D having higher estimates in recent years and Model E having lower estimates in recent years (Figs. 15 and 16).

The estimates of the (log) survey indices are similar for the 27-plus survey for Models $1,1 \mathrm{Q}, \mathrm{A}, \mathrm{AQ}, \mathrm{B}$, $\mathrm{BQ}, \mathrm{C}$, and CQ , as these model configurations model this survey catchability and selectivity in the same manner (Figs. 17 and 18); Models D and E have different patterns due to estimating fewer selectivity
curves and catchability, respectively. The base models, Models $1,1 \mathrm{Q}, \mathrm{A}$, and AQ, estimate catchability for each survey year for the sub-27 survey, so the estimated values are close to the actual values, as these model configurations essentially track the numbers of age-1 fish seen in the survey (Figs. 19 and 20). Models B, BQ, C, CQ, and D estimated constant catchability for the periods 1984 - 1993 and 1996 2011, which results in less variability in the estimates.

Models $1,1 \mathrm{Q}, 1 \mathrm{~B}$, and 1 C have tail compression turned on. The estimates of spawning biomass are similar for all models (Fig. 21), although the estimates of age-0 recruitment differ in recent years (Fig. 22). The estimates of the survey indices are similar for the 27 -plus survey (Fig. 23), as these model configurations model this survey catchability and selectivity in the same manner. The sub-27 survey estimates differ significantly (Fig. 24), as Models 1B and 1C estimate constant catchability for two periods rather than a separate catchability for each survey year.

The fits to the survey age composition data for all model configurations are shown in Figs. $25-48$.

## Tables

Table 20 - The likelihood components, number of parameters, and AIC values for the SS3 model configurations. Jitter convergence implies that the negative log likelihood total was lower (or as low) as any of the jittered runs.

| Model Total Survey | Length <br> comps | Age <br> comps <br> 27-plus | Age <br> comps <br> sub-27 | Size <br> at age | Recruit | Parameters | AIC | Jitter <br> convergence? |  |  |
| :--- | :--- | ---: | :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 3,842 | 4.34 | 3,116 | 80.67 | 16.61 | 645.73 | -21.16 | 252 | 8,189 | Yes |
| 1Q | 3,847 | 6.01 | 3,114 | 80.63 | 17.14 | 651.58 | -22.37 | 252 | 8,199 | Yes |
| A | 3,990 | 1.73 | 3,217 | 89.25 | 46.69 | 657.67 | -22.92 | 252 | 8,484 | Yes |
| AQ | 3,992 | 2.87 | 3,217 | 89.95 | 46.63 | 658.48 | -22.84 | 252 | 8,489 | Yes |
| B | 3,952 | 14.12 | 3,175 | 91.22 | 36.59 | 656.63 | -22.43 | 244 | 8,391 | Yes |
| BQ | 3,954 | 14.51 | 3,176 | 91.06 | 36.60 | 657.20 | -22.33 | 244 | 8,395 | Yes |
| C | 3,951 | 13.14 | 3,177 | 91.61 | 36.65 | 654.93 | -22.51 | 244 | 8,389 | Yes |
| CQ | 3,953 | 14.35 | 3,176 | 90.80 | 36.60 | 657.29 | -22.35 | 244 | 8,394 | Yes |
| D | 4,094 | -7.95 | 3,210 | 233.86 | 32.02 | 643.33 | -16.61 | 207 | 8,603 | Yes |
| E | 3,959 | -6.69 | 3,208 | 83.65 | 46.42 | 649.70 | -21.63 | 253 | 8,425 | Yes |
| 1B | 3,844 | 13.99 | 3,102 | 80.76 | 14.57 | 655.01 | -22.46 | 244 | 8,176 | Yes |
| 1C | 3,843 | 13.30 | 3,103 | 81.26 | 14.57 | 653.03 | -22.52 | 244 | 8,174 | Yes |

Table 21 - Fixed and estimated catchability parameters for the 27 -plus and sub-27 surveys (estimated values are in bold)

| Model | Q for 27-plus survey | Mean $Q$ for $60-81 \mathrm{~cm}$ for 27-plus survey (calc.) | Dev for $\mathbf{Q}$ for 27-plus survey <br> (1984-1993) | $\begin{aligned} & \text { Q for sub-27 } \\ & \text { survey } \end{aligned}$ | Min/max devs for $\mathbf{Q}$ for sub27 survey | Dev for $Q$ for sub-27 survey (1984-1993) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1.0400 | 0.9241 | 0.5106 | 0.0749 | -0.97 / 1.34 | - |
| 1Q | 1.0107 | 0.9159 | 0.5387 | 0.0760 | -1.01 / 1.31 | - |
| A | 1.0400 | 0.9419 | 0.4964 | 0.1318 | -0.86 / 1.37 | - |
| AQ | 1.0118 | 0.9168 | 0.5288 | 0.1304 | -0.85 / 1.35 | - |
| B | 1.0400 | 0.9425 | 0.5045 | 0.1164 | - | -0.51 |
| BQ | 1.0120 | 0.9154 | 0.5250 | 0.1145 | - | -0.49 |
| C | 1.0400 | 0.9265 | 0.4999 | 0.1144 | - | -0.49 |
| CQ | 1.0136 | 0.9165 | 0.5231 | 0.1145 | - | -0.49 |
| D | 1.0400 | 0.9630 | 0.5146 | 0.1077 | - | -0.39 |
| E | 1.8817 | 1.7314 | -0.0037 | 0.1592 | -1.00 / 1.83 | - |
| 1B | 1.0400 | 0.9432 | 0.5118 | 0.1156 | - | -0.93 |
| 1C | 1.0400 | 0.9303 | 0.5082 | 0.1140 | - | -0.92 |



Figure 21 - Estimates of spawning biomass for Models 1, 1Q, A, and AQ


Figure 22 - Estimates of age-0 recruits Models 1, 1Q, A, and AQ


Figure 23 - Log survey estimates and model fits for the 27-plus survey for Models 1, 1Q, A, and AQ


Figure 24 - Log survey estimates and model fits for the sub-27 survey for Models 1, 1Q, A, and AQ
length comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 25 - Proportions-at-length for the 27-plus survey for Model 1 (data in grey, model estimates in red)
length comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 26 - Proportions-at-length for the 27-plus survey for Model A (data in grey, model estimates in red)
length comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 27 - Proportions-at-length for the sub-27 survey for Model 1 (data in grey, model estimates in red)
length comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 28 - Proportions-at-length for the sub-27 survey for Model A (data in grey, model estimates in red)


Figure 29 - Length composition data from the ADF\&G survey for 1998 and 2000 (top row), length composition data from the NMFS GOA bottom trawl survey for 1999 and 2001 (bottom row)


Figure 30 - Length composition data from the ADF\&G survey for 2002 and 2004 (top row), length composition data from the NMFS GOA bottom trawl survey for 2003 and 2005 (bottom row)


Figure 31 - Length composition data from the ADF\&G survey for 2006 and 2008 (top row), length composition data from the NMFS GOA bottom trawl survey for 2007 and 2009 (bottom row)


Figure 32 - Length composition data from the ADF\&G survey for 2010 (top row), length composition data from the NMFS GOA bottom trawl survey for 2011 (bottom row)


Figure 33 - Estimates of spawning biomass for Models A, B, C, D, and E


Figure 34 - Estimates of spawning biomass for Models A, AQ, B, BQ, C, and CQ


Figure 35 - Estimates of age-0 recruitment for Models A, B, C, D, and E


Figure 36 - Estimates of age-0 recruitment for Models A, AQ, B, BQ, C, and CQ


Figure 37 - Log survey estimates and model fits for the 27-plus survey for Models A, B, C, D, and E


Figure 38 - Log survey estimates and model fits for the 27-plus survey for Models A, AQ, B, BQ, C, and CQ


Figure 39 - Log survey estimates and model fits for the sub-27 survey for Models A, B, C, D, and E


Figure 40 - Log survey estimates and model fits for the sub- 27 survey for Models A, AQ, B, BQ, C, and CQ


Figure 41 - Estimates of spawning biomass for Models $1,1 \mathrm{Q}, 1 \mathrm{~B}$, and 1 C


Figure 42 - Estimates of age-0 recruitment for Models 1, 1Q, 1B, and 1C


Figure 43 - Log survey estimates and model fits for the sub-27 survey for Models 1, 1Q, 1B, and 1C


Figure 44 - Log survey estimates and model fits for the sub-27 survey for Models 1, 1Q, 1B, and 1C
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 45 - Age composition data and model estimates for the 27-plus survey for Model 1
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 46 - Age composition data and model estimates for the sub-27 survey for Model 1
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 47 - Age composition data and model estimates for the 27-plus survey for Model 1Q
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 48 - Age composition data and model estimates for the sub-27 survey for Model 1Q
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 49 - Age composition data and model estimates for the 27-plus survey for Model A


Figure 50 - Age composition data and model estimates for the sub- 27 survey for Model A
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 51 - Age composition data and model estimates for the 27-plus survey for Model AQ


Figure 52 - Age composition data and model estimates for the sub-27 survey for Model AQ
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 53 - Age composition data and model estimates for the 27-plus survey for Model B


Figure 54 - Age composition data and model estimates for the sub-27 survey for Model B
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 55 - Age composition data and model estimates for the 27-plus survey for Model BQ
age comps, sexes combined, whole catch, Sub27_Trawl_Survey
aggregated across seasons within year


Figure 56 - Age composition data and model estimates for the sub-27 survey for Model BQ
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 57 - Age composition data and model estimates for the 27-plus survey for Model C
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 58 - Age composition data and model estimates for the sub-27 survey for Model C
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 59 - Age composition data and model estimates for the 27-plus survey for model CQ


Figure 60 - Age composition data and model estimates for the sub-27 survey for Model CQ
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 61 - Age composition data and model estimates for the 27-plus survey for Model D
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 62 - Age composition data and model estimates for the sub- 27 survey for Model D
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 63 - Age composition data and model estimates for the 27-plus survey for Model E


Figure 64 - Age composition data and model estimates for the sub-27 survey for Model E
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 65 - Age composition data and model estimates for the 27-plus survey for Model 1B
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 66 - Age composition data and model estimates for the sub-27 survey for Model 1B
age comps, sexes combined, whole catch, 27plus_Trawl_Survey aggregated across seasons within year


Figure 67 - Age composition data and model estimates for the 27-plus survey for Model 1C
age comps, sexes combined, whole catch, Sub27_Trawl_Survey aggregated across seasons within year


Figure 68 - Age composition data and model estimates for the sub-27 survey for Model 1C

